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Uplands Feasibility Study Report BNSF Wishram Railyard (Ecology Site Name BNSF Track Switching Facility) Wishram, Washington

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BNSF Railway Company 605 Puyallup Avenue Tacoma, Washington 98421

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Ecology Agreed Order No. DE 12897

Prepared by:

This report was prepared by the staff of Kennedy/Jenks Consultants, Inc. under the supervision of the engineer whose seal and signature appear below.

The findings, recommendations, specifications, or professional opinions presented in this report were prepared in accordance with the generally accepted professional engineering practice and within the scope of the project. No other warranty, either expressed or implied, is provided.

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Section 1: Introduction

On behalf of the BNSF Railway Company (BNSF), Kennedy/Jenks Consultants, Inc. (KJ) prepared this Draft Feasibility Study (FS) Report for the BNSF Wishram Railyard (Ecology Site Name: BNSF Track Switching Facility) in Wishram, Washington (Figure 1).

The FS has been prepared in accordance with the Washington State Department of Ecology (Ecology) Model Toxics Control Act (MTCA) regulations published in Washington Administrative Code (WAC) 173-340 (Ecology 2007) and pursuant to an Agreed Order (AO, No. DE 12897) between Ecology and BNSF dated 7 October 2015.

1.1 Objectives

This FS evaluates cleanup options for the uplands portion of the site, with the goal of identifying the most feasible cleanup strategy that is protective of human health and the environment and meets the requirements of Ecology's MTCA regulations [WAC 173-340].

1.2 Site History

The Wishram Railyard was originally developed by the Spokane, Portland, and Seattle (SP&S) Railway between 1910 and 1912. Existing and historical site features are shown on Figures 2 through 5. The primary use of the railyard was, and remains, railcar switching. Historically, industrial activities (locomotive fueling and watering, fuel storage, as well as engine and car repairs) occurred in the westernmost portion (approximately 1,100 feet) of the railyard (i.e., the site), covering an area of approximately 6 to 10 acres, as shown on Figures 2 and 3. The eastern portion of the site generally refers to the area east of the former Signal Office and former Oil House and near a former septic drainage field, as shown on Figures 4 and 5. A detailed history of the Wishram Railyard was included in the *Uplands Remedial Investigation Report BNSF Wishram Railyard (Ecology Site Name BNSF Track Switching Facility) Wishram, Washington* (Uplands RI Report) (KJ 2020) and is summarized below.

Historical Fueling Operations

Steam locomotive fueling using oil was conducted at the site from approximately 1912 through 1956. Fueling facilities included a 30,000-barrel oil aboveground storage tank (AST) located north of the mainline tracks, an Elevated Oil Service 28,000-gallon AST located south of the mainline tracks, an oil unloading trough east of the Power House, a concrete sump, as well as associated appurtenances and piping (Figures 2 and 3). The oil ASTs and appurtenances were removed circa-1957 after the transition to diesel-fueled locomotives in the 1950s.

Diesel locomotives were fueled at the site from the early 1950s to the late 1970s. Diesel fueling was performed along a fueling spur and at a concrete fueling island (installed in 1949), located north and south, respectively, of the mainline tracks. The diesel fuel was initially stored in one 15,000-gallon and one 20,000-gallon underground storage tanks (USTs). In the late 1950s, diesel was stored in two 100,000-gallon ASTs (constructed in or after 1955) located northwest of the Maintenance Shop. Diesel fuel was transferred to and from the tanks by underground

piping. The 100,000-gallon diesel ASTs were removed and fueling ceased at the site during the late 1970s. Based on available records in BNSF's internal UST database, the majority of the remaining tanks were removed in or prior to 1988.

Historical Maintenance Operations

Locomotives underwent maintenance and repairs in the former Engine House/Machine Shop. The former Engine House was constructed in 1911 as a rectangular run-through style building and underwent multiple additions until reaching its maximum footprint size in the 1940s (Figure 3). The former Engine House was demolished in the 1980s after it was no longer needed. Railcar repairs were performed in an area consisting of three tracks and the former (Car) Repair Shop, located to the southeast of the former Engine House/Machine Shop. The car repair shop building was removed in 1960.

Septic Drainage Field

A former septic drainage (leach) field and five septic tanks, located approximately 600 feet to the northeast of the former Engine House (Figure 5), was identified on a station layout map from 1959. Historical maps indicate that the septic system and drainage field treated and discharged wastewater generated on the railyard, as well as wastewater generated by the City of Wishram (single-family homes, a hotel, restaurant, etc.) starting sometime before 1962 and ceasing prior to 1996.

Lake Celilo

In 1957, the U.S. Army Corps of Engineers completed construction of The Dalles Dam, a hydroelectric dam just upstream of The Dalles, Oregon. The flood gates were closed in March 1957, and rising water created Lake Celilo, a 24-mile-long reservoir on the Columbia River. In the vicinity of the railyard, the surface water elevation increased by approximately 40 feet in just a few days, inundating formerly dry land south of the railyard (inundated lands), and significantly altering groundwater elevations and flow conditions beneath the railyard (see Figures 2 and 3). Operation of the hydroelectric dam results in daily fluctuations in the surface water elevation, impacting groundwater flow beneath the site. Adjacent to the railyard, daily surface water elevation changes have been recorded from a few inches up to approximately 4 feet.

1.3 Previous Investigations and Remedial Actions

A complete description of previous investigations and interim remedial actions is included in the Uplands RI Report (KJ 2020). Previous investigations and remedial actions are briefly summarized here. Areas where interim remedial actions were completed are shown on Figure 6.

Investigation activities were initiated onsite in 2002 to evaluate potential impacts to subsurface soils and groundwater from historical railyard activities. BNSF performed voluntary independent investigative and remedial actions through 2015. Site investigation and interim remedial activities included excavating and disposing soil containing petroleum hydrocarbons, removing and disposing former USTs, collecting soil and groundwater samples, conducting a laserinduced fluorescence (LIF) survey, and investigating light non-aqueous phase liquid (LNAPL)

mobility in the vicinity of the former Power House. An air sparge/soil vapor extraction system was installed in early 2012, and was converted to a bioventing system in June 2012, to address residual hydrocarbon impacts in soil north and west of the Maintenance Shop. Operation of the bioventing system continued through July 2019.

BNSF and Ecology entered into an AO (No. 12897) in 2015 to complete the RI and FS. Field activities performed under the AO were substantially conducted between August 2016 and August 2020 and included collection of soil and groundwater samples, installation of groundwater monitoring wells, installation of four oil head monitoring (OHM) wells, collection of LNAPL samples for mobility testing, monthly inspections of the nearshore Columbia River surface from the bank, collection of oil sheen/oil droplet samples from the surface of the Columbia River, and investigation of the presence and extent of LNAPL impacts in the nearshore inundated lands bordering the site. Additional field activities to aid in the feasibility evaluation of remediation alternatives were conducted between July and September 2019 and are summarized in Appendix A. Evaluations of groundwater flow conditions beneath and in the vicinity of the site, as well as the potential migration of dissolved hydrocarbons from the site to the Columbia River were also conducted as part of the RI activities and are summarized in the Uplands RI Report (KJ 2020).

The site is situated within the Columbia Hills Archaeological District as designated by the Washington State Department of Archaeology and Historic Preservation (DAHP), in a culturally significant area due to its proximity to Celilo Falls. The Geoarchaeological Monitoring of Additional Remedial Investigations report (Jacobs 2018) provides a detailed description of the historical background of the site and its potential cultural resources. A cultural resources management plan (CRMP) (AECOM 2016) was developed to establish protocols for managing cultural resources. The CRMP was submitted to the DAHP and the Confederated Tribes and Bands of the Yakama Nation. Subsurface investigative activities conducted in 2016 and 2018 were completed in accordance with the CRMP. Intrusive activities were performed under permit from DAHP.

1.4 Summary of Site Conditions

A complete description of current site conditions is included in the Uplands RI Report (KJ 2020). Site conditions are briefly summarized below.

Petroleum hydrocarbons are present in soil and groundwater in the vicinity of former USTs, former ASTs, and former infrastructure used to store and transfer fuel oils. LNAPL beneath the site is composed of both diesel and Bunker-C oil fuels; and is primarily located beneath the former underground oil pipelines and the former Power House (Figures 7 and 8). MTCA Method A and Method B Cancer and Noncancer Cleanup Levels (CULs) were used as screening values for chemicals reported in soil and groundwater samples during the RI.

A summary of conditions reported in the Uplands RI Report through 2019 and subsequent groundwater monitoring in 2020 follows:

• Petroleum hydrocarbons [reported as diesel-range organics (DRO) and oil-range organics (ORO)] are present in soil and groundwater at concentrations above MTCA

Method A CULs in areas associated with loading, unloading, and storage of fuel oil and diesel fuel.

- o Figures 9 and 10 show DRO/ORO results in unsaturated and saturated soil, respectively, in the main area of the site. DRO/ORO was not reported above MTCA Method A CULs in the eastern area of the site.
- o Petroleum hydrocarbon results in groundwater for DRO/ORO and total petroleum hydrocarbons – diesel range fraction (sum of DRO and ORO results) expressed as TPH-Dx are shown on Figures 11, 12, and 13 for the main area and Figures 14, 15, and 16 for the eastern area. DRO/ORO and TPH-Dx were reported in samples from monitoring wells and reconnaissance groundwater (RGW) sampling locations above groundwater MTCA Method A CULs, primarily in the historical fueling operations areas of the main area. TPH-Dx results were slightly above the CULs in samples from three RGW locations (no wells have been installed) near the former septic drainage field (eastern area, Figure 15).
- Suspected legacy sources of petroleum hydrocarbons have been decommissioned and removed from the site and impacted soil has been removed as part of interim remedial measure (IRM) activities. Where implemented, IRMs successfully removed petroleum hydrocarbons down to the water table or bedrock such that soil samples collected from all but 11 of 145 soil borings/excavation confirmation sampling locations in site areas in the unsaturated zone do not contain residual petroleum hydrocarbons above MTCA Method A CULs. A limited area of residual hydrocarbons is present near the berm (based on one sample) but does not significantly contribute to the overall presence of dissolved phase hydrocarbons in other areas of the site. Horizontal migration of vadose zone soil impacts is not expected because lateral gradients or geologic features that would result in horizontal movement are not present.
- The vertical and lateral extent of LNAPL beneath the site was delineated by conventional methods (advancing borings and installing monitoring wells) and the LIF survey. The LIF data were qualitatively evaluated with respect to field observations of the presence of LNAPL in soil borings, occurrence of measurable LNAPL thicknesses in monitoring wells, and interpretation of the LIF logs for fuel types(s) (e.g., diesel-like and Bunker C/oil-like). The data were quantitatively evaluated with respect to soil laboratory analytical results for petroleum hydrocarbons and estimates of residual LNAPL concentrations. These data were reviewed to assess the potential presence of LNAPL in the subsurface and to estimate LNAPL mobility. LNAPL mobility relates to the potential for LNAPL to flow from one location to another under an existing gradient. "Residual LNAPL" is present at or below LNAPL residual saturation and will not accumulate in a well or migrate across an area. "Mobile LNAPL" is present above the residual saturation and will accumulate in a well, but not migrate across an area. "Migrating LNAPL" is present above the residual saturation level and will migrate across an area under the appropriate hydraulic forces.

LNAPL with properties consistent with both diesel and Bunker-C oil is present south of the mainline tracks near the former underground oil pipelines and the former Power House (Figures 7 and 8). Diesel LNAPL is present at the groundwater surface and in

the periodically saturated vadose zone. Diesel and Bunker-C LNAPL is submerged at depths between the groundwater surface and up to 60 feet below the top of groundwater as a result of the formation of Lake Celilo. With the impoundment of Lake Celilo, the groundwater table rose rapidly, trapping LNAPL in the subsurface at depths similar to the pre-impoundment depth of groundwater. Submerged LNAPL is estimated to be greater than 60 years old based on known facility operations and the formation of Lake Celilo. LNAPL observed south of the mainline is classified as potentially recoverable, as evidenced by observations of measurable LNAPL in three of the four OHM wells and by laboratory measurements of LNAPL physical properties. LNAPL was not observed in the southernmost row of LIF borings bordering the berm separating the site from the Columbia River, nor in the monitoring wells installed along the berm during the RI (berm monitoring wells), and no evidence has been observed that the LNAPL body is migrating.

LNAPL with properties consistent with a weathered diesel fuel was historically present north of the mainline track in the vicinity of the Maintenance Shop. LNAPL in this area was located at the elevation of the pre-impoundment groundwater surface and in the periodically saturated vadose zone. Prior to January 2016, LNAPL was frequently observed in wells WMW-7 and WMW-8, located near the Maintenance Shop. Bioventing was implemented in this area as an IRM between 2012 and 2019. Except for a single event in November 2016 (0.10 foot measured in WMW-8), LNAPL has not been measured in either well since January 2016. A hydrocarbon sheen is inconsistently observed at the groundwater interface in wells WMW-7 and WMW-8. Soil gas measurements from 2019 feasibility field activities show oxygen at near-atmospheric concentrations and little to no carbon dioxide in this area (Appendix A), indicating that air exchange rates in the vadose zone are sufficient to sustain aerobic degradation of the petroleum hydrocarbons.

- Hydrocarbons in the gasoline range (GRO) were present above the MTCA Method A CUL in 12 of the 53 samples analyzed for GRO between 2004 and 2018. Elevated GRO concentrations in soil were located below the groundwater table in the vicinity of two former gasoline USTs (southwest of the Maintenance Shop) and the former Power House (Figure 17). GRO has not been reported at concentrations above its MTCA Method A CUL in groundwater samples since 2004, including those wells in close proximity to soil borings with GRO concentrations above the MTCA Method A CUL (Figure 18). Given their limited extent and concentrations, GRO does not pose risk to human health or the environment. No evidence of LNAPL associated with gasoline has been observed at the site.
- Volatile organic compounds (VOCs) typically associated with gasoline [benzene, toluene, ethylbenzene, and xylenes (BTEX compounds)], which typically pose the greatest potential risk to receptors, are present above MTCA Method A CULs in only two of 177 soil samples (benzene only); and do not exceed MTCA Method A CULs in groundwater samples collected since 2004. Fuel additives 1,2-dibromoethane [ethylene dibromide (EDB)] and 1,2-dichloroethane (EDC) were not reported above laboratory reporting limits in 128 soil samples collected and analyzed for these constituents. Chlorinated solvents and other VOCs were not reported at concentrations above MTCA Method A CULs in the 177 soil samples, nor in groundwater samples. The relative

absence of VOCs and the lack of buildings in or near impacted areas indicates vapor intrusion is an incomplete exposure pathway under current site conditions.

- Polycyclic aromatic hydrocarbons (PAHs), including carcinogenic PAHs (cPAHs) which were used to calculate Total cPAHs, were reported above applicable MTCA Method A or B CULs in less than 10 percent of soil and groundwater samples. PAHs above MTCA Method A CULs are associated with samples that contained DRO and ORO above MTCA Method A CULs.
- Metals reported in soil were below applicable MTCA Method A or B CULs in 126 of 127 samples. Metals reported in monitoring well groundwater samples above applicable MTCA Method A or B CULs were limited to dissolved and total arsenic, total barium (one sample only), dissolved iron, and dissolved manganese. Iron, manganese, and arsenic are present in groundwater in locations where petroleum hydrocarbons and residual organics affect groundwater geochemistry and liberate naturally occurring metals in soil into groundwater.

Section 2: Conceptual Site Model

A Conceptual Site Model (CSM) is included in the Uplands RI Report (KJ 2020). The CSM is briefly summarized below. Representative cross sections and a plan view CSM are presented on Figures 19 through 22.

2.1 Site-Related Constituents and Locations

Based on historical railroad operations and previous investigations, constituents of concern (COCs) identified for the site include Total Petroleum Hydrocarbons expressed as DRO, ORO, and to a limited extent, GRO. Dissolved and total arsenic, dissolved iron, and dissolved manganese are also present in site groundwater as a result of reducing conditions caused by the natural degradation of DRO and ORO in groundwater.

General areas within the present upland remediation investigation areas (i.e., not including inundated lands to the south of the railyard, beneath the Columbia River) where petroleum hydrocarbon-related constituents have been identified include the following (see Figure 22):

- Around the Mainline tracks (mainline track area) vicinity of the former Boiler House and its former UST, former Pump House (associated with former Diesel ASTs), and the current Maintenance Shop.
- South of Mainline tracks (eastern and western LNAPL areas and dissolved phase) vicinity of former diesel and oil fueling areas and underground piping, former Oil Unloading Track, former Oil Trough, and former Power House.
- Former Engine House/Machine Shop and vicinity.
- Berm Area south of the former Engine House/Machine Shop.
- Former Oil House east of the former Signal Office/former Store House.

Potential operational sources of petroleum hydrocarbon-related constituents in these areas (historical oil and diesel fueling operations and steam power production, storage of oil and diesel fuel in multiple ASTs and USTs onsite, transport of oil in associated underground piping systems) are no longer present. Current site conditions are the result of historical impacts.

2.2 Fate and Transport

Transport of COCs includes leaching from impacted saturated soils and submerged LNAPL bodies, downward migration from historically impacted unsaturated soils to the saturated zone during precipitation events, and migration in shallow-zone groundwater through advection and dispersion. Low concentrations of carbon dioxide and high concentrations of oxygen in soil gas (Appendix A, Figure A3) indicate that petroleum hydrocarbons adsorbed to unsaturated soils are naturally degrading through aerobic biodegradation. COCs adsorbed to saturated soils will migrate through dissolution and advective/dispersive forces.

As presented in the Uplands RI Report (KJ 2020), the Columbia River is a losing stream more than 80 percent (10 months) of the year. Appendix A includes graphs of the daily average hydraulic gradient (in feet per foot) calculated between select monitoring wells and the river during between December 2016 and April 2018 and between March 2019 and August 2021. A negative hydraulic gradient value results when the river elevation is greater than the groundwater elevation and implies water flowing away from the river (losing stream condition), while a positive value implies water flowing to the river (gaining stream condition). Over the two monitoring periods, losing stream conditions were observed more than 80 percent of the time in shallow berm wells (WMW-14, WMW-16, and WMW-18). Site groundwater discharges to surface water in the Columbia River during a very limited portion of the year. Therefore, transport of COCs offsite is limited.

2.3 Potential Exposure Pathways

Potentially complete exposure pathways for human and ecological receptors at the site include direct contact and/or incidental ingestion by construction workers and railyard workers of affected media (soil and groundwater). The vapor intrusion pathway has previously been identified as an incomplete exposure pathway due to lack of VOCs reported in soil and groundwater and the limited number of buildings (e.g., Maintenance Shop) onsite. A petroleum vapor intrusion (PVI) initial assessment presented in the Uplands RI Report (KJ 2020) concluded that PVI is not a risk in the Maintenance Shop. Although shallow groundwater at the site is not a current source of drinking water nor is it identified as a future drinking water source since potable water is supplied to the site by the City of Wishram, in accordance with WAC 173- 340-720, groundwater at the site is considered potable for current and future uses. Therefore, human consumption of shallow site groundwater is a potential exposure pathway. The former water supply wells (Well #2 and Well #3) were decommissioned between 22 March and 19 April 2022 in accordance with WAC 173-160-381.

The following exposure pathways are considered to be potentially complete for human receptors based on the existing site conditions and uses (Figure 23):

- Surface and subsurface soil direct contact and/or incidental ingestion by site, construction, and utility workers.
- Groundwater direct contact and/or incidental ingestion by site, construction, and utility workers [saturated conditions exist within approximately 10 to 15 feet below ground surface (bgs)].
- Consumption of groundwater by site, construction, and utility workers.
- Surface water direct contact and/or incidental ingestion by site, construction, and utility workers, and recreational users.
- Consumption of aquatic organisms by recreational users.

Direct contact and/or incidental ingestion exposure pathways for soil and groundwater can be controlled by institutional controls (ICs) and soil management/construction plans. The ICs would be put in place to protect onsite receptors.

Ecological exposures to site-related constituents in upland areas of the site are negligible, as gravel and asphalt cover render ecological exposure routes incomplete across much of the site. Ecological receptors may occupy the sparsely vegetated areas along the berm separating the site from the Columbia River. Potential exposure risks along the berm are negligible given the existing analytical data characterizing berm surface soil (i.e., no known impacts) and delineated depth of subsurface impacts beyond anticipated receptor exposure depth. Further, the primary COCs present in soil underlying the berm (i.e., petroleum hydrocarbon-related compounds) are not expected to bioaccumulate in the food chain.

Ecological exposures in the adjacent reach of the Columbia River are possible, specifically the potential for consumption of organisms by recreational users and other aquatic organisms. Investigation of environmental conditions in the area of the inundated lands in the adjacent reach of the Columbia River is ongoing.

Section 3: Proposed Cleanup Standards

The process of setting cleanup standards includes establishing Remedial Action Objectives (RAOs), identifying the CULs and Remediation Levels (RELs) that will be used to identify areas of the site requiring remediation based on current and future use and designated beneficial use, and identifying the point of compliance locations where remediation progress will be evaluated.

RAOs are narrative statements describing the goals of the remedial action. They identify the conditions targeted by the remedy in each environmental media to protect human health and the environment. RAOs are based on consideration of current and future land use, and groundwater and surface water beneficial use designations.

BNSF Wishram Railyard Property

The Klickitat County, Washington zoning map [https://www.klickitatcounty.org/284/Zoning-Map; (Klickitat County Zoning Map)], indicates the railyard is zoned as an "Industrial Park." The site is currently used as a railyard and meets the definition of an industrial property in WAC 173- 340-200 and the applicability criteria of soil cleanup standards for industrial properties under WAC 173-340-745(1)(a) parts (i)(A) through (F). The railyard and areas north of the mainline track area (e.g., the maintenance shop and bullpen) have controlled access, including fencing between the existing depot and maintenance shop, a fenced-in bullpen north of the maintenance shop, and are patrolled by railroad police. The primary potential exposures to potentially impacted surface and subsurface soil, groundwater and adjacent surface water are to adult railroad, construction, and utility workers.

The site is fully developed as an industrial railyard where buildings, pavement, rail lines, and surfaces (comprising approximately 94 percent of the land area) are designed and managed per federal regulations to remove and control vegetation, limit the potential for vegetation with deep root zones and use by wildlife. The railyard surface areas, covered by gravel, asphalt, or other impervious structures (e.g., buildings) minimize potential exposure to the soil. Along the Columbia River, engineered embankments (forming the berm area) composed of large riprap protect the banks from erosion and restrict potential deeper soil contact by occupants and wildlife. Foreseeable future use of the site is anticipated to remain the same, with railyard operations including railcar switching on tracks located just south of the Depot (Figure 2).

Off-Railyard Properties

The area to the north of the railyard property boundary (off-railyard properties) (see Figure 2), is zoned as "Rural Center". Potential petroleum hydrocarbon impacts related to the railyard are limited to the vicinity of the former boiler house and former heating oil UST within this area. The UST and approximately 750 tons of petroleum hydrocarbon-impacted soil were removed in April 2002 (see Figure 6 for approximate lateral extent of the excavation area). Petroleum hydrocarbon-impacted soils were excavated to the top of the bedrock surface (to the extent practicable) at a depth of approximately 16 feet below ground surface (bgs); groundwater was not encountered in the excavation. Confirmation soil samples from the north, east, and south sidewalls of the excavation indicated diesel- and oil-range petroleum hydrocarbon impacts remained in-place from approximately 14.5 to 15.5 feet bgs (see Figures 9 and 10).

Fire District #11 Wishram is listed as the property owner of the two property parcels located in this area (Figure 2). Current features include two warehouse storage-type buildings used by the fire department in the eastern parcel and a U.S. Post Office in the western parcel. Surface areas covered by gravel or impervious structures (buildings) comprise approximately 65 percent of the land area between the two parcels and minimize potential exposure to the subsurface soil. The primary potential exposures to potentially impacted subsurface soil and groundwater are to adult construction and utility workers.

In accordance with WAC 173-340-720, designated groundwater uses include potential drinking water source, although shallow groundwater at the site is not a current source of drinking water nor is it identified as a future drinking water source since potable water is supplied to the site by the City of Wishram. Beneficial use designations for the Columbia River near the site include water supply, spawning and rearing aquatic life, wildlife, and miscellaneous such as recreation, aesthetics, hydroelectric power generation, and commercial navigation and transportation.

3.1 Remedial Action Objectives

The RAOs for the Upland Area FS are based on the CSM, which identified the potential receptors and exposure pathways present at the site. The RAOs include the following:

- Protect site, construction, and utility workers from direct contact with and incidental ingestion of soil containing COCs at concentrations above CULs during typical operations and soil excavations.
- Protect site, construction, and utility workers from direct contact with and ingestion of groundwater containing COCs at concentrations above CULs.
- Due to the depth of groundwater greater than 6 feet bgs, terrestrial ecological receptors are not expected to encounter groundwater in the upland area of the site.
- Protect aquatic and recreational receptors immediately adjacent to the site from direct contact with and incidental ingestion of surface water containing COCs at concentrations above applicable surface water criteria discharging to the Columbia River.
- Protect recreational users and aquatic biota from ingestion of aquatic organisms containing COCs at concentrations above applicable standards.

3.2 Points of Compliance

Points of compliance are locations where site conditions are compared to numeric criteria to evaluate whether CULs have been met and are used during remediation to evaluate the performance of the remedy and confirm protection of identified exposure pathways and receptors. Generally, the standard point of compliance under MTCA is sitewide. As provided for in WAC 173-340-720(8)(c), where it can be demonstrated under WAC 173-340-350 through 173-340-390 that it is not practicable to meet the cleanup level throughout the site within a reasonable restoration time frame, Ecology may approve a conditional point of compliance that shall be as close as practicable to the source of hazardous substances, and except as provided under WAC 173-340-720(8)(d), not to exceed the property boundary. MTCA defines CPOCs,

including soil depths, for several potential receptor exposure pathways. Proposed points of compliance for soil and groundwater are discussed below.

3.2.1 Unsaturated Subsurface Soil

The standard point of compliance for soil based on human exposure via direct contact is sitewide to a depth of 15 feet bgs as the typical maximum depth of soil disturbing activities [WAC 173-340-740(6)(d)]. The unsaturated vadose zone within much of the railyard extends from the ground surface to 10 feet bgs and along the berm from the ground surface to approximately 15 feet bgs. A conditional point of compliance for soil, with adoption of ICs, is the water table depth of 10 feet bgs in the railyard.

BNSF Wishram Railyard Property. The site is fully developed as an industrial railyard where buildings, pavement, rail lines, and surfaces are designed and managed per federal regulations to remove and control vegetation, limit the potential for vegetation with deep root zones, and limit use by wildlife. Railroad operations control subsurface disturbance and can be formally established by ICs. Along the Columbia River, engineered embankments composed of large riprap protect the banks from erosion and restrict potential deeper soil contact by occupants and wildlife. As a result, soil between the ground surface and the groundwater table will be evaluated using applicable industrial MTCA Methods A and C cleanup levels, and soil concentrations protective of groundwater and protective of groundwater to surface water pathways.

Off-Railyard Property. The current use of the properties north of the maintenance shop area includes small business and commercial services uses (e.g., the U.S. Post Office and fire department storage buildings), although the zoning designation as "Rural Center" allows for other non-commercial uses including homes and eating/drinking establishments. Therefore, unsaturated zone soil in these areas will be evaluated using applicable MTCA Methods A and B unrestricted cleanup levels, along with consideration of soil concentrations protective of groundwater. As this area is not adjacent to surface water, soil concentrations protective of groundwater discharging to surface water will not be considered.

3.2.2 Saturated Subsurface Soil

The standard point of compliance for soil based on human exposure via direct contact is sitewide to a depth of 15 feet bgs as the typical maximum depth of soil disturbing activities [WAC 173-340-740(6)(d)]. The depth to groundwater in much of the railyard is 10 feet bgs. Groundwater was not encountered in the 2002 excavation adjacent to the former boiler house. However, based on 2016 and 2018 soil borings near the former 30,000-barrel Oil AST and the Maintenance Shop area, the depth to groundwater in the off-railyard properties to the north of the railyard was estimated as 14 feet bgs in the Uplands RI Report (KJ 2020).

BNSF Wishram Railyard Property. Soil between the groundwater table (approximately 10 feet bgs) and 15 feet bgs will be evaluated using applicable industrial MTCA Method A and C cleanup levels and, for those constituents likely to leach to groundwater, soil concentrations for the saturated zone that are protective of groundwater and protective of groundwater to surface water pathways. Impact to groundwater from subsurface soil in the saturated zone will be

evaluated based on direct measurement of groundwater conditions and remediation of soil in the saturated zone will be evaluated based on groundwater cleanup performance data.

Off-Railyard Property. Soil between the groundwater table (approximately 14 feet bgs) and 15 feet bgs will be evaluated as saturated zone soil using applicable unrestricted Method A and B cleanup levels for soil off-railyard property to the north of the maintenance shop (rural center zoned) and, for those constituents likely to leach to groundwater, soil concentrations for the saturated zone that are protective of groundwater. Potential impacts to groundwater from subsurface soil in the saturated zone will be evaluated based on direct measurement of groundwater conditions and remediation of soil in the saturated zone will be evaluated based on groundwater cleanup performance data.

3.2.3 Groundwater and Surface Water

Groundwater. The standard point of compliance for groundwater is sitewide. In accordance with WAC 173-340-720, designated groundwater uses include potential drinking water source, although shallow groundwater at the site is not a current source of drinking water nor is it identified as a future drinking water source since potable water is supplied to the site by the City of Wishram. It is unlikely that human or ecological receptors would contact groundwater during normal site use. Contact with groundwater will be managed through ICs, and RELs will be selected to protect human and ecological receptors.

Groundwater/Surface Water Interface. The proposed CPOC for monitoring groundwater at the interface with surface water is the line of existing shallow and deep monitoring wells installed on the berm bordering the Columbia River. The potential exposure pathway by which human or ecological receptors could encounter site groundwater is by flux of site groundwater into the Columbia River. Protection of surface water will be assessed at the berm, even though analysis performed during the RI demonstrated that flux of site groundwater toward the Columbia River occurs during only approximately 2 months of the year. The CPOC will be further defined in the draft Cleanup Action Plan (DCAP).

3.3 Proposed CULs and RELs

A summary of the development of proposed CULs and identification of COCs for soil, groundwater, and surface water is provided in Appendix B. Identification of COCs is based on a comparison of soil and groundwater sampling results presented in the Uplands RI Report and 2020 groundwater sampling results to the proposed CULs.

MTCA requires sites to be cleaned up to protect human health and the environment. WAC 173-340-704(1) through (3), provides summaries as to the applicability of MTCA Methods A, B, and C cleanup levels. Where site conditions meet MTCA Method A CULs for a site-related constituent, risks to human health and the environment are considered acceptable. MTCA Methods B and C provide alternative approaches to establishing CULs based on site-specific evaluation of potential risk to human and ecological receptors. CULs developed using MTCA Methods B and C must not be set at levels below the practical quantitation limit (PQL) or background.

Risk evaluations take into account site conditions that limit exposure or migration of residual site constituents. CULs may be based on the absence of exposure pathways. Once practical methods of treatment have been employed, residual constituents may remain. Proposed remedies for this site include a combination of active and passive measures, and ICs applied in series or independently. In areas where it may become infeasible to meet CULs through active remediation, passive methods and ICs will be applied once practicable remedial limits are met. RELs may be defined according to MTCA by quantitative or qualitative means.

The following provides a summary of the proposed CULs for the site based on MTCA (for applicable media), other applicable, relevant, and appropriate requirements (ARARs) for groundwater and surface water, and other relevant information pertinent to establishing sitespecific remedial goals. Available and applicable MTCA Methods A, B, and C cleanup levels and screening levels were obtained from Ecology's CLARC master data table (updated in July 2021) (Ecology 2021a). RELs are established to indicate what site conditions will be addressed with different cleanup methods and may be narrative or qualitative in nature. Table 1 summarizes the proposed COCs and CULs based on these ARARs. Development of the proposed CULs and constituents evaluated by media is presented in Appendix B. CULs and RELs will be established in the DCAP.

3.3.1 Soil CULs and RELs

Appendix B presents proposed CULs [in milligrams per kilogram (mg/kg)] for constituents reported in one or more soil samples collected from the vadose zone (Table B-3A) and the shallow (between 10 and 15 feet bgs) saturated zone (Table B-3B). Impact to groundwater from the deep (below 15 feet bgs) saturated zone soil will be evaluated based on direct measurement of groundwater conditions, and remediation of soil in the saturated zone will be evaluated based on groundwater cleanup performance data.

Proposed CULs were established by comparing applicable industrial MTCA Method A and C CULs for railyard property and MTCA Method A and B CULs for off-railyard property, background concentrations, PQLs, and for those constituents proven likely to leach to groundwater, soil concentrations (calculated using the fixed parameter 3-phase partitioning model described in WAC 173-340-747(4) and MTCA Equation 747-1) for the vadose or saturated (as applicable) zone that are protective of groundwater and the groundwater to surface water pathway. Development of proposed soil CULs is summarized in Section B3-4 of Appendix B. Identified COCs based on reported results above the proposed CULs are listed below and summarized in Table 1. The proposed CULs for the identified COCs are based on MTCA Method A CULs, which are the same concentrations for unrestricted and restricted (industrial) CULs for the COCs.

- GRO 30 mg/kg (Method A), based on the presence of benzene, although at low concentrations, in saturated soil samples collected from the site.
- DRO $-2,000$ mg/kg (Method A).
- ORO $-$ 2,000 mg/kg (Method A).
- Total petroleum hydrocarbon (TPH-Dx) 2,000 mg/kg (Method A).

The proposed soil RELs for petroleum hydrocarbon related constituents are equivalent to the above CULs. The soil CULs and RELs will be established in the DCAP.

3.3.2 Groundwater CULs and RELs

Beneficial use designations for the groundwater beneath the site include use as a potential drinking water source. Proposed groundwater CULs protective of drinking water were established from consideration of MTCA Method B cancer and noncancer CULs, ARARs including state and federal Maximum Contaminant Levels (MCLs), MTCA Method A CULs for select constituents (DRO, ORO, and TPH-Dx), background concentrations for arsenic, and PQLs. Development of proposed groundwater CULs is summarized in Section B3-1 and Table B-1 of Appendix B.

Remedies to address potential human exposure to DRO and ORO within the railyard area, including ICs and environmental covenants (ECs), are evaluated where groundwater contains DRO or ORO above the MTCA Method A CULs. Where potential exposure to groundwater is restricted, CULs will be selected based on the surface water pathway; proposed CULs for surface water are presented in Section 3.3.4.

Based on the CSM and the capacity for ICs to control direct contact and incidental ingestion pathways for groundwater within the rail yard, contact within the Columbia River is the only potentially complete unrestricted exposure pathway for site groundwater; therefore, the surface water screening levels will be applied to site groundwater with the potential to enter the Columbia River.

Results from groundwater monitoring conducted between November 2016 and August 2020 were selected as representative data for current dissolved phase groundwater conditions. Constituents reported in one or more groundwater monitoring well samples collected since November 2016 above the proposed CUL for groundwater are identified as COCs in Table 1. Identified COCs and proposed groundwater CULs [in micrograms per liter ($\mu q/L$)] are as follows:

- DRO 500 µg/L as an unrestricted use goal (Method A).
- ORO 500 µg/L as an unrestricted use goal (Method A).
- Total petroleum hydrocarbon (TPH-Dx) 500 µg/L as an unrestricted use goal (Method A).
- \bullet 1-methylnaphthalene 1.5 μ g/L (Method B Cancer).
- Total arsenic $-5 \mu g/L$ (Background¹).
- Total barium 2,000 µg/L (State and Federal MCL). Total barium has been reported above the proposed groundwater CUL in one groundwater sample.

¹ Background concentration for arsenic from MTCA Table 720-1 Method A CULs, subject to change based on potential policy updates resulting from Ecology's draft *Natural Background Groundwater Arsenic Concentrations in Washington State* (July 2021), Publication 14-09-044.

- Dissolved Iron 11,000 µg/L (Method B Noncancer).
- Dissolved manganese 750 ug/L (Method B Noncancer).

Groundwater RELs are established to indicate the nature and concentrations of COCs addressed with different cleanup methods. The proposed groundwater RELs for petroleum hydrocarbon-related constituents identify criteria for transitioning from active removal or destruction to monitored natural attenuation (MNA), and include the following:

- Statistically significant stable or decreasing trends in groundwater concentrations will be established using Mann-Kendall, Mann-Whitney, or similar analysis for a period of 3 years (or 12 sampling events). Statistical methods to evaluate groundwater concentration trends will be provided in the groundwater compliance monitoring plan in the DCAP.
- Groundwater geochemistry indicates conditions suitable for natural attenuation.

The groundwater CULs and groundwater RELs for transitioning from active to passive remedial action will be established in the DCAP.

3.3.3 Surface Water CULs

Surface water will be protected by remediating groundwater in the vicinity of the berm to applicable CULs. Surface water COCs are based on constituents identified in site groundwater, since groundwater is in direct communication with the river. The proposed CPOC for demonstrating that site groundwater meets surface water beneficial uses is the line of groundwater monitoring wells along the berm.

Beneficial use designations for the Columbia River near the site include use as potential drinking water supply, along with spawning and rearing aquatic life, wildlife, and miscellaneous use such as recreation, aesthetics, hydroelectric power generation, and commercial navigation and transportation. Proposed surface water CULs protective of drinking water and aquatic life were established based on surface water Method B cancer and noncancer cleanup levels, ARARs including fresh surface water concentrations that are protective of human health and protective of aquatic life under acute and chronic exposure conditions as established under state (WAC 173-201A-240) and federal laws [Section 304 of the Clean Water Act (CWA)], Environmental Effects-Based Concentrations (EEBCs) from Implementation Memo No. 23 (Ecology 2021b) for petroleum hydrocarbons (e.g., weathered diesel), Risk Assessment Information System (RAIS) concentrations, Method A for select constituents for beneficial use as potable water, background concentrations, and PQLs.

Development of proposed surface water CULs is summarized in Section B3-3 of Appendix B. Constituents reported above the proposed CUL for surface water in one or more groundwater samples collected since November 2016 from monitoring wells located along the berm are identified as COCs in Table 1. The berm wells include 10 shallow monitoring wells (WMW-14 through WMW-23) and six deep monitoring wells (RMD-1 through RMD-6).

The proposed surface water COCs and respective CULs (and their sources) for monitoring wells located along the berm are as follows:

- DRO 500 µg/L as an unrestricted use goal (Method A CUL) for human health exposure considerations and 3,000 µg/L (EEBC weathered DRO) for ecological receptor exposure considerations.
- ORO 500 µg/L as an unrestricted use goal (Method A CUL) for human health exposure considerations and 3,000 µg/L (EEBC weathered DRO) for ecological receptor exposure considerations.
- Total petroleum hydrocarbon (TPH-Dx) 500 µg/L as an unrestricted use goal (Method A CUL) for human health exposure considerations and 3.000μ g/L (EEBC) weathered DRO) for ecological receptor exposure considerations.
- 1-methylnaphthalene $-2.1 \mu g/L$ (RAIS).
- Total arsenic $-5 \mu g/L$ (Method A, Background).
- Total lead 2.5 µg/L (Surface Water Aquatic Life Fresh/Chronic WAC 173-201A and CWA §304). Total lead has been reported above the proposed surface water CUL in three groundwater samples.
- Dissolved iron 1,000 µg/L (Surface Water Aquatic Life Fresh/Chronic CWA §304).
- Dissolved manganese 50 µg/L (Surface Water Human Health Fresh Water CWA §304).

The surface water CULs will be established in the DCAP.

3.3.4 LNAPL RELs

The proposed petroleum hydrocarbon LNAPL RELs, to indicate when to transition from active removal or destruction to natural source zone depletion (NSZD), is based on the mobility of the LNAPL and potential for the LNAPL to migrate. LNAPL will be removed until:

- Ambient-temperature transmissivity is below 0.8 feet squared per day (ITRC 2018), or
- Removal rates approach asymptotic conditions.

The LNAPL RELs for transitioning from active to passive remedial action will be established in the DCAP.

Section 4: Estimates of Impacted Media Above Proposed Cleanup Levels

This section provides estimates of the extents of impacted site media (summarized in Table 2) based on the findings of the RI.

4.1 Shallow Soil

The estimated volume of shallow (above the groundwater table) petroleum hydrocarbonimpacted soil outside of the LNAPL body area, based on the proposed soil cleanup standards presented in Section 3, is approximately 170 cubic yards. Soil with DRO and ORO at concentrations above the proposed CULs occurs between 6 and 9 feet bgs on the northern side of the berm in the vicinity of boring B-16-01, as shown on Figure 9. The estimated spatial extent of impacts is also based on low-level LIF responses within the same depth interval in LIF borings CR4, CR4.5, and CR5 (Figure 7). Soil above 6 feet bgs in this area is not impacted.

Impact to groundwater from subsurface soil in the saturated zone will be evaluated based on direct measurement of groundwater conditions and remediation of soil in the saturated zone will be evaluated based on groundwater cleanup performance data.

4.2 LNAPL

LNAPL has been identified south of the mainline tracks near the former underground oil pipelines and the former Power House. As shown in cross section depictions of the LNAPL and residual TPH impacted areas (Figures 19, 20, and 21), the estimated vertical and horizontal LNAPL extents vary by depth and apparent thickness in the formation at each location. As described below and shown on Figure 22, eastern and western LNAPL areas have been defined based on the variations of depth and apparent LNAPL thickness, as well as depths to bedrock.

Area and volume estimates of LNAPL in the subsurface are presented in Table 2. LNAPL volume calculations assume a uniform soil porosity of 40 percent (upper bound measurement), LNAPL saturation values of 33 percent in the eastern LNAPL area and 17 percent in the western LNAPL area, and uniform distribution throughout the impacted area, including the outer limits where residual saturation concentrations are lowest. Based on these assumptions, LNAPL volumes represent a conservative estimate.

In the eastern LNAPL area, spatial and vertical LNAPL extents are relatively continuous. The eastern LNAPL mass, based on LIF data and soil borings, is approximately 17,000 square feet in area, with apparent LNAPL impacts ranging from approximately 5 feet in thickness (in soil boring B-12-8) to approximately 60 feet in thickness (in soil boring for OHM-1) and averaging 14 feet. Apparent LNAPL thicknesses of approximately 12 feet or more have been measured in wells OHM-1, OHM-2, and OHM-3, located within the eastern LNAPL area (KJ 2020). Depths to LNAPL impacts range from approximately 10 feet bgs to approximately 85 feet bgs (LIF location D06). The estimated total volume of LNAPL is approximately 203,000 gallons and the

estimated recoverable volume of LNAPL is approximately 56,900 gallons (approximately 28 percent of the pore volume).

In the smaller, western LNAPL area, LNAPL occurs in thin lenses approximately 0.5 to 4 feet in thickness at shallow (between approximately 11 and 13 feet bgs) and deep (between approximately 20 and 25 feet bgs) depth intervals. The western LNAPL mass is approximately 5,200 square feet in total area (2,600 square feet in each of shallow and deep zones). The estimated total volume of LNAPL in the shallow and deep intervals is approximately 4,100 gallons. The estimated recoverable volume of LNAPL for the two depth intervals is approximately 30 gallons (approximately 0.6 percent of the pore volume). LNAPL has not been measured in well OHM-4, located in the western LNAPL area (KJ 2020).

4.3 Dissolved-Phase Impacts in Groundwater

Dissolved phase DRO and ORO concentrations above the proposed CUL of 500 µg/L were identified in wells north and south of the mainline tracks (mainline track area), and at wells near the former Engine House area (Figures 11 to 16). The approximate spatial extent of dissolvedphase impacts in the mainline track area is 88,000 square feet, and the approximate spatial extent of dissolved-phase impacts near the former Engine House area is 24,000 square feet.

Section 5: Technology Screening and Alternative Development

This section presents the rationale for identifying remedial alternatives to address soil and groundwater containing COCs at concentrations exceeding site-specific CULs. Section 5.2 presents the initial screening of remedial technologies to identify potentially applicable process options. Remedial methods passing the initial screening process are combined to create potentially feasible remedial alternatives, which are described in Section 5.3.

5.1 Identification and Evaluation of Potential Remedial Methods

General response actions are broad categories of remedial methods that can address the cleanup of a specific matrix (e.g., soil, groundwater). Remedial technologies are various techniques within the general response actions. Process options are specific processes within each remedial technology category. General response actions, remedial technologies, and process options that may be appropriate for addressing site conditions and petroleum hydrocarbon site-related constituents in soil and groundwater are presented in Table 3.

Process options were initially screened using three criteria: effectiveness, implementability, and relative cost, as summarized below:

- Effectiveness involves consideration of a process option's ability to address the anticipated volume of soil and groundwater, meet cleanup standards, and protect human health and the environmental during construction and implementation.
- Implementability includes technical and administrative considerations. This criterion focuses on the ability to technically address COCs in a selected matrix at concentrations reported during the RI process. It also evaluates the permits necessary for onsite and offsite activities and discharges, and the availability of offsite facilities, services, and materials.
- Cost is based on engineering judgments rather than detailed estimates. Process options that are judged to be similar in effectiveness and ability to be implemented yet estimated to cost several times more than other process options in the same technology category, were eliminated from further consideration.

Process options that are not appropriate for site conditions, planned future site uses, or constituents and concentrations contained in soil and groundwater were eliminated from further consideration. In addition, process options that are innovative but unproven were also eliminated. If more than one process option in a remedial technology group was identified as potentially appropriate for the site, further screening was performed, and one process option was selected to represent that technology group.

5.2 Remedial Technologies Screening

General response actions that may be applicable for remediating impacted soil are summarized below and further evaluated in Table 4:

- Natural Source Zone Depletion (NSZD): Depletion of petroleum hydrocarbon related compounds through natural biological and abiotic processes. Data obtained from carbon trap samples indicate NSZD is occurring at rates between 350 and 6,000 gallons per acre per year near the submerged LNAPL body. These results demonstrate that NSZD is an active and effective process for consideration in each alternative (see Appendix A).
- Containment: Installation of a physical barrier that will encapsulate and isolate the impacted soils. The impacted soils will be left in place and the petroleum hydrocarbons will naturally degrade over time.
- Removal: Petroleum hydrocarbons between the ground surface and the top of the water table will be removed through physical processes and disposed offsite at a licensed facility. Removal processes could include:
	- \circ Excavation Using an excavator or large diameter auger to physically remove the soils from the subsurface. The excavated area will be backfilled with clean fill.
	- \circ Thermal Extraction Using heat to increase subsurface temperatures to 50 to 70 degrees Celsius (°C) to reduce the viscosity of the petroleum hydrocarbons combined with extraction technologies (e.g., physical removal or soil vapor extraction) to remove them from the subsurface. Residual petroleum hydrocarbons remaining in the subsurface will naturally degrade over time.
- *In Situ* Treatment: The use of physical or chemical processes to stabilize or destroy COCs in the subsurface. *In Situ* technologies include:
	- \circ Solidification/Stabilization Mixing chemicals (e.g., cement, fly ash, bentonite) with the soil at depth that bind with the petroleum hydrocarbons and soil to solidify the impacted area and control mobility/leachability of the COCs.
	- o Chemical Destruction Injecting or mixing oxidants (e.g., peroxide, ozone) into the subsurface soil at depth resulting in degradation or destruction of the petroleum hydrocarbons adsorbed to the soil.
	- \circ Thermal Destruction Applying conductive/resistive heating technologies to heat the subsurface above 100 °C to physically destroy in place the petroleum hydrocarbons or vaporize them for removal. Soil gasses produced in the process are extracted, treated if necessary, and vented to the atmosphere.

General response actions that may be applicable for remediating the LNAPL are summarized below and further evaluated in Table 4:

• NSZD: Depletion of LNAPL through natural biological and abiotic processes. Data obtained from carbon trap samples indicate NSZD is occurring at rates between 350 and

6,000 gallons/acre/year near the submerged LNAPL body. These results demonstrate that NSZD is an active and effective process for consideration in each alternative (see Appendix A).

- Containment: Installation of a physical or hydraulic barrier that will encapsulate and isolate the LNAPL. The LNAPL will be left in place and the petroleum hydrocarbons will naturally degrade over time.
- Removal: LNAPL at and below the water table (submerged LNAPL) will be removed through physical processes and disposed offsite at a licensed facility. Removal processes could include:
	- o Excavation Using an excavator or large diameter auger to physically remove the soils from the subsurface. The excavated area will be backfilled with clean fill.
	- \circ Extraction Applying physical processes such as skimming, vacuum enhanced skimming, vacuum extraction, and multi-phase extraction to remove the LNAPL from the subsurface. Residual petroleum hydrocarbons remaining in the subsurface will naturally degrade over time.
	- \circ Thermal Extraction Using electrical resistive heating to heat the subsurface to between 50°C to 70°C to reduce the viscosity of the petroleum hydrocarbons then remove them from the subsurface using extraction technologies. Residual petroleum hydrocarbons remaining in the subsurface will naturally degrade over time.
- *In Situ* Treatment: The use of physical or chemical processes to stabilize or destroy the LNAPL in the subsurface. *In Situ* technologies include:
	- \circ Chemical Destruction Injecting oxidants (e.g., peroxide, ozone) into with the subsurface that react with the LNAPL resulting in degradation or destruction of the petroleum hydrocarbons.
	- o Geochemical Stabilization Injection of chemicals to create a geochemical reaction between oxidant and soils resulting in mineralization around LNAPL globules to treat and encapsulate in place
	- o Thermal Destruction Applying sufficient energy (heat) and oxygen to the LNAPL body to initiate combustion resulting in the physical destruction of the petroleum hydrocarbons. Soil gasses produced in the process are extracted, treated if necessary, and vented to the atmosphere.
	- \circ Biological Destruction Biological destruction of LNAPL in the subsurface occurs through increasing oxygen concentrations and/or temperature in the impacted zone to increase biological activity. Biological degradation takes place at the interface between the LNAPL body and areas where sufficient oxygen supply is present and progresses inward into the interior of the LNAPL body.

- *Ex Situ* Treatment: LNAPL, water, and soil vapors removed from the subsurface could be treated using the following treatment options:
	- \circ Soil Vapor Treatment using granular activated carbon (GAC) or thermal/catalytic oxidizers to remove COCs from effluent to below permitted discharge limits. Vapors can also be condensed and combined with the liquid waste streams for treatment and disposal.
	- \circ Oil/Water Separation LNAPL and groundwater would be removed from the subsurface simultaneously. LNAPL and water would be separated and the LNAPL containerized, then shipped off-site for disposal or recycling. Water would be treated on-site using GAC then discharged to the local publicly owned treatment works (POTW) under an industrial wastewater permit.

General response actions that may be applicable for remediating dissolved phase petroleum hydrocarbons beneath the site are summarized below and further evaluated in Table 4:

- MNA: Natural attenuation occurs through dispersion, dilution, sorption, volatilization, and biological transformation or destruction of the petroleum hydrocarbons. Petroleum hydrocarbon biodegradation is more energetically favorable and complete under aerobic (oxidizing) than anaerobic (reducing) conditions. When groundwater is monitored for available electron acceptors and other parameters indicative of the natural attenuation process, it is referred to as monitored natural attenuation or MNA.
- Containment: Installation of a physical or hydraulic barrier that will contain and control the dissolved petroleum hydrocarbons. Removal of the dissolved phase constituents may or may not be combined with containment as part of the process.
- Removal: Dissolved phase petroleum hydrocarbons can be removed from the subsurface through groundwater extraction. Under certain conditions, extraction can be enhanced using surfactants.
- *In Situ* Treatment: The use of physical or chemical processes to stabilize or destroy the dissolved phase petroleum hydrocarbons. *In Situ* technologies include:
	- o Physical/Chemical Destruction Injecting oxidants into the subsurface that react with the petroleum hydrocarbons resulting in degradation or destruction of the petroleum hydrocarbons.
	- o Biological Destruction –. Injecting oxygen into groundwater to promote aerobic biological degradation processes is a proven effective means of degrading heavy end, long chain petroleum hydrocarbons in place.
- *Ex Situ* Treatment: Fluids extracted from wells are treated using GAC, ozone or other technology applicable to petroleum hydrocarbons. Treated effluent would be discharged to the local wastewater treatment plant under an industrial wastewater permit

Performance and/or confirmation monitoring are required components of all response actions. Performance monitoring includes sampling during removal or treatment to assess progress

and/or achievement of CULs. Groundwater confirmation monitoring is required to assess longterm effectiveness and compliance with CULs.

MTCA requires that the process options used minimize the amount of untreated COCs remaining at the site to the extent practicable, and that preference be given to a permanent solution and hierarchy of preferred remedial methods. In general, technologies that reuse, recycle, destroy, or detoxify hazardous substances will result in permanent solutions.

5.3 Development of Alternatives

This section identifies alternatives appropriate for addressing the COCs in soil and groundwater. These alternatives are identified using the requirements and expectations described in MTCA (WAC 173-340-360), which include:

- Meeting threshold requirements for remedial alternatives (refer to Section 7.1).
- Using permanent solutions to the maximum extent practicable.
- Providing for a reasonable restoration timeframe.

Ecology has the following expectations for cleanup action alternatives (WAC 173-340-370):

- Use treatment technologies whenever practicable.
- Minimize the need for long-term management of impacted materials by destroying, detoxifying, or removing hazardous substances that are above CULs.
- Recognize the need to use engineering controls, such as containment for sites with large volumes of relatively low levels of hazardous substances.
- Implement measures to control precipitation and runoff from contacting affected soils and waste materials.
- Consolidate hazardous substances to the maximum extent practicable if the hazardous substances remain onsite.
- Control/minimize releases to surface water via runoff and groundwater discharges exceeding CULs.
- Consider the use of natural attenuation of hazardous substances, which may be appropriate under some circumstances.
- Do not undertake cleanup actions that will result in a greater overall threat to human health and the environment than will other alternatives.

MTCA requires treatment, wherever practicable, for sites containing liquid wastes, areas containing high concentrations of hazardous substances, highly mobile materials, or discrete areas of hazardous substances that lend themselves to treatment. MTCA also recognizes that

engineering controls (such as covenants, containment, caps, and covers) are appropriate for sites or portions of sites that contain large volumes of materials with relatively low levels of hazardous substances where treatment is impracticable [WAC 173-340-370(3)].

Based on the regulatory considerations, site-specific conditions, and the assessment of the remedial technologies summarized in Section 5.2 and in Table 4, the following five remedial alternatives were developed for further evaluation (and summarized in Table 5). For Alternatives 1, 3, 4, and 5, the timeframe to achieve RELs is estimated. Based on the available data, the timeframe to achieve restoration is estimated to be greater than 30 years for each alternative. Alternative 1 is presented as the permanent solution (baseline alternative) against which the other alternatives are evaluated for the purpose of determining whether the cleanup action selected is permanent to the maximum extent practicable (173-340-350(8)(c)(ii)(A)). The baseline alternative is the most practicable, technically possible solution for the site. Remedial actions that are not technically possible to implement (e.g., excavation of deep soils) are identified in Tables 3 and 4.

Alternatives 1 through 5 include ICs and ECs as part of the remedy. Accordingly, Ecology will perform a periodic review once every 5 years after initiation of the cleanup action as required under WAC 173-340-440 and as described in WAC 173-340-420.

Alternative 1: Physical LNAPL Removal, Focused Biosparge, and MNA

- LNAPL will be removed periodically using a mobile vacuum unit in the eastern LNAPL area. Relatively thinner areas of LNAPL in the western LNAPL area will be treated through biosparging. After RELs have been achieved, LNAPL thickness, soil gas, and groundwater concentrations will be monitored to confirm and quantify NSZD.
- Dissolved phase petroleum in the southern end of the eastern LNAPL area, in the mainline track area and beneath the Engine House will be assessed for natural attenuation and followed by implementation of compliance groundwater monitoring. After RELs have been achieved, petroleum hydrocarbons will be monitored to confirm concentrations and extents are decreasing.
- Dissolved phase metals in groundwater will be addressed by remediating petroleum hydrocarbons and the corresponding return of measured geochemical parameters to their ambient conditions.
- Impacted soils in the vadose zone beneath the berm will be excavated and disposed of offsite.
- A Groundwater Covenant will be enacted to restrict future use of site groundwater.

Periodic LNAPL removal is expected to be implemented for approximately 10 years. Focused biosparging is expected to be implemented for less than 5 years. Once biosparging and physical LNAPL removal ceases, MNA and NSZD assessments will be conducted for a 3-year period, followed by compliance monitoring until CULs are achieved.

Alternative 2: Institutional Controls and Environmental Covenants with Compliance Groundwater Monitoring

- No active remediation is performed. Constituent concentrations will continue to decrease through NSZD and natural attenuation. ICs and ECs will be managed in perpetuity or until CULs are met.
- ICs are implemented to control site uses that could potentially expose receptors to impacted media including soil and groundwater.

Compliance groundwater monitoring will be implemented.

Alternative 3: LNAPL Containment, Biosparge, MNA, and Targeted Excavation

- The eastern and western LNAPL areas will be contained using a sheet pile wall on three sides, LNAPL thickness and soil gas concentrations will be monitored within the contained LNAPL areas to confirm and quantify NSZD.
- Dissolved phase petroleum hydrocarbons around the mainline tracks and beneath the Engine House will be assessed for natural attenuation. and treated through biosparging as a contingency. Biosparging will be implemented as a contingency to MNA to enhance aerobic biodegradation in areas where groundwater is not naturally aerobic. After RELs have been achieved, petroleum hydrocarbons will be monitored to confirm concentrations and extent are decreasing.
- Dissolved phase metals in groundwater will be addressed by remediating petroleum hydrocarbons and the corresponding return of measured geochemical parameters to their ambient conditions.
- Impacted soils in the vadose zone beneath the berm will be excavated and disposed offsite.
- A Groundwater Covenant will be enacted to restrict future use of site groundwater.

The timeframe for active remediation (biosparging) if implemented, is estimated to be less than 5 years.

Alternative 4: Physical LNAPL Removal, Biosparge, MNA, and Targeted Excavation

• LNAPL will be removed periodically using a mobile vacuum unit in the eastern LNAPL area. After physical removal has achieved practicable limits, biosparging may be implemented as necessary to enhance biological degradation and address dissolved phase DRO and ORO if present from residual LNAPL. Relatively thinner areas of LNAPL in the western LNAPL area and dissolved phase in the southern end of the eastern LNAPL area will be treated through biosparging. After RELs have been achieved, LNAPL thickness and soil gas will be monitored to confirm and quantify NSZD.

- Dissolved phase petroleum in the mainline track area and beneath the Engine House will be assessed for natural attenuation and treated through biosparging as a contingency. Biosparging will be implemented as a contingency to MNA to enhance aerobic biodegradation in areas where groundwater is not naturally aerobic. After RELs have been achieved, petroleum hydrocarbons will be monitored to confirm concentrations and extent are decreasing.
- Dissolved phase metals in groundwater will be addressed by remediating petroleum hydrocarbons and the corresponding return of measured geochemical parameters to their ambient conditions.
- Impacted soils in the vadose zone beneath the berm will be excavated and disposed of offsite.
- A Groundwater Covenant will be enacted to restrict future use of site groundwater.

Periodic LNAPL removal is expected to be implemented for approximately 10 years. If applied, biosparging to enhance residual LNAPL biodegradation is expected to be implemented for less than 5 years. Biosparging in dissolved phase impacted areas is expected to be implemented for less than 5 years. Once biosparging and physical LNAPL removal ceases, MNA and NSZD assessments will be conducted for a 3-year period, followed by compliance monitoring until CULs are achieved.

Alternative 5: Low-Temperature Thermal LNAPL Removal (LTTR), Biosparge, MNA, and Targeted Excavation

- The eastern LNAPL area will be heated to approximately 50 $^{\circ}$ C to 70 $^{\circ}$ C to reduce LNAPL viscosity and improve mobility. Removal and heating approaches will be adapted to target specific areas based on observations of improved LNAPL recovery, changes in thickness, and the effect of temperature. After physical removal has achieved practicable limits, remediation wells can be converted to implement biosparging as needed to enhance biological degradation in the areas where elevated temperatures are present and oxygen becomes limited due to increased biological activity. Relatively thinner areas of LNAPL in the western LNAPL area will be treated using biosparging. After active LNAPL removal and sparging have achieved RELs, LNAPL thickness and soil gas concentrations will be monitored within LNAPL areas to confirm and quantify NSZD.
- Dissolved phase petroleum hydrocarbons in the mainline track area and beneath the Engine House will be assessed for natural attenuation and treated using biosparging as a contingency. Biosparging will be implemented as a contingency to MNA to enhance aerobic biodegradation in areas where groundwater is not naturally aerobic. After RELs have been achieved, petroleum hydrocarbons will be monitored to confirm concentrations and extent are decreasing.
- Dissolved phase metals in groundwater will be addressed by remediating petroleum hydrocarbons and the corresponding return of measured geochemical parameters to their ambient conditions.
- Impacted soils in the vadose zone beneath the berm will be excavated and disposed of offsite.
- A Groundwater Covenant will be enacted to restrict future use of site groundwater.

The timeframe for active remediation (subsurface heating, LNAPL removal and biosparging) is expected to be less than 5 years. Once biosparging and physical LNAPL removal ceases, MNA and NSZD assessments will be conducted for a 3-year period, followed by compliance monitoring until CULs are achieved.

5.3.1 Alternative 1 – Physical LNAPL Removal, Focused Biosparge, MNA, and Targeted Excavation

Alternative 1 involves physical removal of LNAPL in the eastern LNAPL area, focused biosparging in the western LNAPL area, excavation of shallow impacted soils, and MNA for dissolved phase petroleum impacts. Monitored NSZD in the LNAPL areas and MNA for dissolved impacts will be implemented once RELs for active remediation have been met. During and following physical removal of LNAPL to achieve RELs, natural attenuation is expected to eliminate the transport of dissolved phase petroleum hydrocarbons to the river during high groundwater conditions (approximately 2 months out of the year). This alternative uses removal and destruction technologies, along with ICs and ECs, to control exposures to potential receptors (onsite workers, visitors, nearby residents, trespassers, and surface water).

The key components of Alternative 1 are depicted on Figure 24. This alternative consists of the following general actions:

- Design and Permitting: LNAPL removal tests will be conducted, and removal will be modeled using the American Petroleum Institute (API) LNAPL Distribution and Recovery Model (LDRM; API 2007a and 2007b) to evaluate extraction well spacing and removal timeframe. An LDRM estimate of LNAPL recoverability is provided in Appendix D and summarized in Section 5.3.6. A biosparging pilot study will be conducted to evaluate the radius of influence and sparge rate, identify the design parameters and design a fullscale system for the western LNAPL area. A Cleanup Action Plan (CAP) and a Joint Aquatic Resources Permit Application (JARPA) will be prepared for construction and excavation near the Columbia River. Permits will be obtained, including well permits for the LNAPL removal wells, biosparge wells, and associated monitoring wells; and building permits for the biosparge system(s). Subsurface intrusive activities will include cultural resources monitoring completed in accordance with the CRMP under permit from DAHP. Because the biosparge system will not result in discharges to the atmosphere, an air permit is not needed. Samples of the LNAPL and groundwater in the extraction wells will be collected and analyzed for waste profiling and to identify appropriate disposal methods. Discharge of effluent water to the local POTW will require completion (and approval by the Klickitat County PUD) of a feasibility study to establish acceptable discharge volume, concentration limits, and monitoring requirements.
- Site preparation. Activities will include grading for equipment access, delineation of work zones and laydown areas, and identification of underground utilities within the footprint

of the drilling and excavation areas. Utilities impacted by the remedial activities will be protected in-place, relocated, removed, or abandoned in-place as appropriate.

• Installation. In the eastern LNAPL area, existing OHM wells and new recovery wells installed within the LNAPL extent will be used for LNAPL removal. Current estimates assume up to 15 new LNAPL recovery wells will be installed within the LNAPL body, on approximately 40-foot centers. New recovery wells will be constructed of 4- to 6-inchdiameter, stainless steel wire-wrapped screen connected to stainless steel or polyvinyl chloride (PVC) risers.

In the western LNAPL area, five shallow biosparge wells will be installed to depths of approximately 25 feet bgs and three paired shallow (25 feet bgs) and deep (60 feet bgs) wells will be installed to address dissolved-phase and LNAPL impacts at varying depths throughout the area. Biosparge system installation includes constructing treatment pads/sheds, and furnishing air compressors, programmable logic controllers (PLCs), distribution manifolds and piping, and biosparge wells.

• Operation and Monitoring. LNAPL will be removed from wells by periodic high-vacuum extraction or other methods depending on the results of the LDRM and field implementation testing. Adequate time will be allowed for LNAPL to recharge prior to subsequent extraction events. Between extraction events, the presence and apparent thickness of the LNAPL will be monitored. Physical removal will allow for adaptive management strategies to increase or decrease the frequency of removal and to expand the extraction well network based on observed site conditions. Once LNAPL removal and biosparging operations have met their design objectives (i.e., RELs), active remediation will be discontinued and monitored NSZD will be implemented in the LNAPL areas (eastern and western). NSZD in the LNAPL areas is expected to eliminate the transport of petroleum hydrocarbons to the river during high groundwater conditions.

Concurrent with the implementation of the active remedy in the LNAPL areas, MNA will be implemented in the berm area south of the former Power House, the mainline track area, and beneath the former Engine House. MNA will be assessed in general accordance with Ecology's *Guidance on Remediation of Petroleum-Contaminated Ground Water by Natural Attenuation* (NA Guidance) (Ecology 2005) and will include collecting groundwater samples for DRO, ORO, and natural attenuation parameters semi-annually for 3 years. Compliance groundwater monitoring will be implemented following completion of the MNA assessment and will include annual to biennial monitoring for geochemical parameters and analysis of DRO and ORO.

• Groundwater Covenant. A restrictive covenant will be enacted to prohibit the installation of wells for the purpose of water supply within the site boundary, restrict the extraction of groundwater for purposes other than construction and hydraulic control dewatering, monitoring/investigation, or remediation, and require groundwater extracted within the site boundary be evaluated via data review and testing and the discharge managed in accordance with state and federal regulations.

Excavation of impacted soils along the berm, south of the Power House, and MNA of groundwater are common components of Alternatives 1, 3, 4, and 5. As such, these components are described in Sections 5.3.7 and 5.3.8, respectively.

Timeframe. The time to achieve RELs is estimated to be 10 years, and the time to reach CULs is estimated to be more than 30 years (depending on NSZD rates) following cessation of active remediation. LNAPL recovery is typically most productive during early phases of removal and becomes asymptotic as the rebound of LNAPL thickness decreases with time. Active LNAPL removal is estimated to be implemented over a 10-year timeframe. Biosparge operations are expected to be implemented over a 2- to 5-year period (concurrent with physical removal of the LNAPL) in the western LNAPL area. NSZD monitoring in the eastern LNAPL area will be conducted once physical removal actions cease and in the western LNAPL area once biosparging ceases. An MNA assessment will be implemented for approximately 3 years in the berm area south of the former Power House, the mainline track area, and beneath the former Engine House, concurrent with the implementation of the active remedies. Compliance groundwater monitoring will continue until CULs are met. ECs and ICs established for the site are expected to remain in place until a no further action determination is provided by Ecology.

Cost: The estimated cost for this alternative is summarized in Appendix D, Table D1.

5.3.2 Alternative 2 – Institutional Controls (ICs) and Environmental Covenants (ECs) with Compliance Groundwater Monitoring

This alternative involves developing ICs and ECs to limit exposures to potential receptors (onsite workers, visitors, nearby residents, and trespassers). Note that while the following describes what will be implemented as part of Alternative 2, the nature of the ICs and ECs describe herein is part of each of the five alternatives. Compliance groundwater monitoring will be implemented in select wells to monitor constituent concentrations. This alternative does not include active remediation; however, constituent concentrations will continue to decrease by natural processes (NSZD and natural attenuation). ICs will be implemented to restrict exposure to impacted soil and groundwater. To limit exposures during construction activities, ICs will be implemented and maintained using an EC developed in accordance with Ecology procedures. Specifically, the EC will:

- Control activities that may result in the release of residual COCs, create a new exposure to residual COCs, or disturb the subsurface environment through plans and procedures approved by Ecology.
- Provide notification that residual COCs may be present and include deed restrictions that limit future development and use of the site.
- Prohibit the installation of wells for the purpose of water supply within the site boundary.
- Restrict the extraction of groundwater for purposes other than construction and hydraulic control dewatering, monitoring/investigation, or remediation.

- Require that groundwater extracted within the site boundary be evaluated via data review and testing and the discharge managed in accordance with state and federal regulations.
- Require that utilities or other subsurface infrastructure within the site boundary where the depth is greater than the highest measured groundwater be designed and constructed based on groundwater conditions.

Timeframe: The timeframe to implement this alternative is less than 1 year. However, ICs and ECs will remain in place in perpetuity or until CULs are met.

Cost: The estimated cost for this alternative is summarized in Appendix D, Table D2.

5.3.3 Alternative 3 – LNAPL Containment, Biosparge, MNA, and Targeted Excavation

Alternative 3 includes installing a sheet pile wall to the east, south, and west of the LNAPL body to control downgradient transport of dissolved phase petroleum hydrocarbons during high groundwater conditions and allow the LNAPL body to naturally degrade (monitored NSZD). Biosparging will be implemented to remediate dissolved phase petroleum hydrocarbons in the mainline track area and beneath the former Engine House (contingent on an MNA assessment see Section 5.3.8). Excavation of shallow impacted soils will occur in a portion of the berm area. MNA will be implemented for dissolved phase petroleum hydrocarbons once RELs for active remediation have been met. This alternative uses physical containment, ICs, and ECs to control exposures to potential receptors (onsite workers, visitors, nearby residents, trespassers, and surface water).

The key components of Alternative 3 are depicted on Figure 25. This alternative consists of the following general actions:

- Design and Permitting. Borings will be advanced along the alignment of the sheet pile wall to collect geotechnical data and design the sheet pile wall. A biosparging pilot study will be conducted to evaluate the radius of influence and sparge rate, identify the design parameters, and design a full-scale system. A CAP and a JARPA will be prepared for construction near the Columbia River, as well as excavation activities. Permits will be obtained, including well permits for the geotechnical borings, biosparge wells, and associated monitoring wells; and building permits for the sheet pile wall and the biosparge system. Subsurface intrusive activities will include cultural resources monitoring completed in accordance with the CRMP under permit from DAHP. Because the biosparge system will not result in discharges to the atmosphere, an air permit is not needed.
- Site preparation. Activities will include grading for equipment access, delineation of work zones and laydown areas, and identification of underground utilities within the footprint of the wall and the excavation. Utilities impacted by the remedial activities will be protected in-place, relocated, removed, or abandoned in-place as appropriate.

- Installation. Sheet pile wall installation includes driving individual interlocked steel sheets to the desired depth along three sides of the LNAPL body and grading the surface for drainage to minimize infiltration of stormwater into the subsurface. Biosparge system installation includes constructing treatment pads/sheds, and furnishing air compressors, PLCs, distribution manifolds and piping, and biosparge wells. Forty-five biosparge wells will be installed to depths of approximately 25 feet bgs to address dissolved-phase impacts at varying depths throughout the remediation area (Figure 25). The number, location, and depth of biosparge wells in each area will be evaluated based on the MNA assessment.
- Operation and Monitoring. The sheet pile wall requires no operational activities. Groundwater pumping from within the sheet pile wall will not be performed as the groundwater gradient across the site is relatively flat throughout most of the year, and groundwater typically (approximately 10 months per year) flows away from the Columbia River. Groundwater inside the wall will be monitored, using existing or new monitoring wells (as needed) to evaluate NSZD, as well as document groundwater elevations inside the wall area. Groundwater outside the wall will be monitored to confirm groundwater elevations and that CULs are achieved.

The biosparging system will deliver ambient air to groundwater under either intermittent or continuous operation based on system monitoring results. Regular operations and maintenance (O&M) will be performed to assess the performance of the biosparging system. O&M activities will include measurement of system parameters (flow rate, pressure response, etc.), flow rate/time adjustments, and measurements of dissolved oxygen concentrations in groundwater. Monitoring wells will also be installed, as needed, to evaluate operations of the biosparging system. Once biosparging operations have met the design objectives, active remediation will be discontinued, and MNA followed by compliance groundwater monitoring will be implemented.

• Groundwater Covenant. A restrictive covenant will be enacted to prohibit the installation of wells for the purpose of water supply within the site boundary; restrict the extraction of groundwater for purposes other than construction and hydraulic control dewatering, monitoring/investigation, or remediation; and require groundwater extracted within the site boundary be evaluated via data review and testing and the discharge managed in accordance with state and federal regulations.

Excavation of impacted soils along the berm, and MNA of groundwater are further described in Sections 5.3.7 and 5.3.8, respectively.

Timeframe: As this alternative does not include active removal or destruction of the LNAPL, the time to achieve RELs is estimated to be greater than 30 years, and the time to achieve CULs is more than 50 years. Installation of the containment barrier will take less than 1 year. Biosparge operations are expected to be implemented over a 2- to 5-year period. MNA for groundwater outside the containment wall will be assessed for approximately 3 years following cessation of the biosparging. NSZD monitoring inside the containment wall and compliance groundwater monitoring of selected wells outside the containment wall are expected to continue in perpetuity

or until CULs are met. ECs and ICs established for the site are expected to remain in place in perpetuity.

Cost: The estimated cost for this alternative is summarized in Appendix D, Table D3.

5.3.4 Alternative 4 – Physical LNAPL Removal, Biosparge, MNA, and Targeted Excavation

Alternative 4 involves physical removal of LNAPL in the eastern LNAPL area, biosparging in the western LNAPL area and southern end of the eastern LNAPL area, biosparging contingent on an MNA assessment in the mainline track area and beneath the former Engine House, and excavation of shallow impacted soils in a portion of the berm area. Monitored NSZD in the LNAPL areas and MNA for dissolved impacts in the biosparge areas will be implemented once RELs for active remediation have been met. Based on soil gas monitoring for NSZD parameters in the eastern LNAPL area, LNAPL removal wells may be converted to biosparge wells after RELs for physical LNAPL recovery have been achieved. Biosparging is expected to enhance the attenuation of and to eliminate the transport of dissolved phase petroleum hydrocarbons to the river during high groundwater conditions. This alternative uses removal and destruction technologies, along with ICs and ECs, to control exposures to potential receptors (onsite workers, visitors, nearby residents, trespassers, and surface water).

The key components of Alternative 4 are depicted on Figure 26. This alternative consists of the following general actions:

- Design and Permitting: LNAPL removal tests will be conducted, and removal will be modelled using the API LDRM to evaluate extraction well spacing and removal timeframe. An LDRM estimate of LNAPL recoverability is provided in Appendix D and summarized in Section 5.3.6. A biosparging pilot study will be conducted to evaluate the radius of influence and sparge rate, identify the design parameters and design a fullscale system. A CAP and a JARPA will be prepared for construction and excavation near the Columbia River. Permits will be obtained, including well permits for the LNAPL removal wells, biosparge wells, and associated monitoring wells; and building permits for the biosparge systems. Subsurface intrusive activities will include cultural resources monitoring completed in accordance with the CRMP under permit from DAHP. Because the biosparge systems will not result in discharges to the atmosphere, an air permit is not needed. Samples of the LNAPL and groundwater in the extraction wells will be collected and analyzed for waste profiling and to identify appropriate disposal methods. Discharge of effluent water to the local POTW will require completion (and approval by the Klickitat County PUD) of a feasibility study to establish acceptable discharge volume, concentration limits, and monitoring requirements.
- Site preparation. Activities will include grading for equipment access, delineation of work zones and laydown areas, and identification of underground utilities within the footprint of the drilling and excavation areas. Utilities impacted by the remedial activities will be relocated, removed, or abandoned in-place as appropriate.
- Installation. In the eastern LNAPL area, existing OHM wells and new recovery wells installed within the LNAPL extent will be used for LNAPL removal. Current estimates

assume up to 15 new LNAPL recovery wells will be installed within the LNAPL body, on approximately 40-foot centers. New recovery wells will be constructed of 4- to 6-inchdiameter, stainless steel wire-wrapped screen connected to stainless steel or PVC risers.

• As with Alternatives 1 and 3, biosparge system installation includes constructing treatment pads/sheds, and furnishing air compressors, PLCs, distribution manifolds and piping, and biosparge wells. In the western LNAPL area, mainline track area, and beneath the former Engine House, 50 biosparge wells will be installed to depths between approximately 20 to 40 feet bgs to address dissolved-phase and LNAPL impacts at varying depths; and three paired shallow (25 feet bgs) and deep (60 feet bgs) wells will be installed. At the southern end of the LNAPL removal area, 10 paired (20 total) biosparge wells will be installed to depths of approximately 25 feet bgs (shallow) and 60 feet bgs (deep) to address dissolved-phase impacts (Figure 26). The number, location, and depth of biosparge wells in each area will be evaluated based on the MNA assessment.

Operation and Monitoring. LNAPL will be removed from wells by periodic high-vacuum extraction or other methods depending on the results of the LDRM and field implementation testing. Adequate time will be allowed for LNAPL to recharge prior to subsequent extraction events. Between extraction events, the presence and apparent thickness of the LNAPL will be monitored. Physical removal will allow for adaptive management strategies to increase or decrease the frequency of removal and to expand the extraction well network based on observed site conditions. Once LNAPL removal (eastern area) and biosparging (western area) operations have met their design objectives (i.e., RELs), active remediation will be discontinued and monitored NSZD will be assessed. Based on soil gas monitoring for NSZD parameters in the eastern LNAPL area, LNAPL removal wells may be converted to biosparge wells to further enhance biodegradation. MNA for dissolved impacts in the biosparge areas will be assessed once RELs for active remediation have been met. MNA will be assessed in general accordance with Ecology's NA Guidance (Ecology 2005) and will include collecting groundwater samples for DRO, ORO, and natural attenuation parameters semi-annually for 3 years. Compliance groundwater monitoring will be implemented following completion of the MNA assessment and will include annual to biennial monitoring for geochemical parameters and laboratory analysis for DRO and ORO.

• Groundwater Covenant. A restrictive covenant will be enacted to prohibit the installation of wells for the purpose of water supply within the site boundary, restrict the extraction of groundwater for purposes other than construction and hydraulic control dewatering, monitoring/investigation, or remediation, and require groundwater extracted within the site boundary be evaluated via data review and testing and the discharge managed in accordance with state and federal regulations.

Excavation of impacted soils along the berm, and MNA of groundwater are further described in Sections 5.3.7 and 5.3.8, respectively.

Timeframe. The time to achieve RELs is estimated to be 10 years, and the time to reach CULs is estimated to be more than 30 years (depending on NSZD rates) following cessation of active

remediation. LNAPL recovery is typically most productive during early phases of removal and becomes asymptotic as the rebound of LNAPL thickness decreases with time. Active LNAPL removal is estimated to be implemented over a 10-year timeframe. Biosparge operations are expected to be implemented over a 2- to 5-year period (concurrent with physical removal of the LNAPL) in the western LNAPL, the mainline track area, and beneath the former Engine House. NSZD monitoring in the eastern LNAPL area will be conducted while LNAPL is being removed. If implemented in the eastern area after physical LNAPL removal, biosparge operations are expected to be implemented over a 3- to 5-year period. MNA will be assessed for approximately 3 years following cessation of biosparge operations in the western LNAPL, mainline track, and Engine House areas, and for 3 years following cessation of physical LNAPL removal activities in the eastern LNAPL area (or following cessation of biosparging, if implemented). Compliance groundwater monitoring will continue until CULs are met. ECs and ICs established for the site are expected to remain in place until a no further action determination is provided by Ecology.

Cost: The estimated cost for this alternative is summarized in Appendix D, Table D4.

5.3.5 Alternative 5 – Low-Temperature Thermal Removal (LTTR), Biosparge, MNA, and Targeted Excavation

Alternative 5 involves heating the subsurface to enhance physical removal of the LNAPL in the eastern LNAPL area, biosparging in the western LNAPL area and southern end of the eastern LNAPL area, biosparging contingent on MNA assessment in the mainline track area and beneath the former Engine House, and excavation of shallow impacted soils in a portion of the berm area. The saturated subsurface in the eastern LNAPL area will be heated to an average temperature between 50 and 70°C to reduce the LNAPL viscosity and interfacial tension and facilitate its removal through multi-phase extraction wells. The effluent will be treated onsite, with the groundwater being discharged to the local POTW and LNAPL being disposed offsite at a permitted facility. Temperatures will be limited to less than 90°C at the heating points to avoid steam production/vaporization of groundwater. Active extraction during heating will result in an inward hydraulic gradient controlling LNAPL migration within the remediation area and biosparging along the southern end of the eastern LNAPL area is expected to eliminate the transport of dissolved phase petroleum hydrocarbons to the river, therefore physical containment is not needed. Alternative 5 assumes electrical resistive heating (ERH) will be used. ERH is the best way to control temperatures between wells and more uniformly distribute heat throughout the subsurface to alter LNAPL properties for efficient recovery. Monitored NSZD in the LNAPL areas and MNA for dissolved impacts in the biosparge areas will be implemented once RELs for active remediation have been met. Based on soil gas monitoring for NSZD parameters in the eastern LNAPL area, multi-phase extraction wells may be converted to biosparge wells after RELs for physical LNAPL recovery have been achieved. Increased temperatures will stimulate greater biological activity, which may benefit from additional oxygen delivery to the subsurface. This alternative uses removal and destruction technologies, along with ICs and ECs, to control exposures to potential receptors (onsite workers, visitors, nearby residents, trespassers, and surface water).

The key components of Alternative 5 are depicted on Figure 27. This alternative consists of the following general actions:

- Design and Permitting: A pilot test will be conducted to evaluate and confirm the low temperature heating mechanism and model the removal using the LDRM to evaluate extraction well spacing and removal timeframe. An LDRM estimate of LNAPL recoverability, including effects of increasing temperature, is provided in Appendix D and summarized in Section 5.3.6. A biosparging pilot study will be conducted to evaluate the radius of influence and sparge rate, identify the design parameters and design a fullscale system. A CAP and a JARPA will be prepared for construction and excavation near the Columbia River. Permits will be obtained including well permits for the LNAPL removal wells, biosparge wells, and associated monitoring wells; building permits for the heating and biosparge systems, and discharge permits (industrial wastewater, National Pollutant Discharge Elimination System, and/or air permit). Subsurface intrusive activities will include cultural resources monitoring completed in accordance with the CRMP under permit from DAHP. Because the LNAPL removal system is not expected to produce significant off-gas, and the biosparge system will not result in discharges to the atmosphere, an air permit is not needed. Samples of the LNAPL and groundwater in the extraction wells will be collected and analyzed for waste profiling and identification of appropriate disposal methods. Discharge of effluent water to the local POTW will require completion (and approval by the Klickitat County PUD) of a feasibility study to establish acceptable discharge volume, concentration limits, and monitoring requirements.
- Site preparation. Activities will include grading for equipment access, delineation of work zones and laydown areas, and identification of underground utilities within the footprint of the drilling and excavation areas. Utilities impacted by the remedial activities will be protected in-place, relocated, removed, or abandoned in-place as appropriate.
- Installation. Existing wells OHM-1, OHM-2, OHM-3, and WMW-1 are constructed of PVC and, therefore, will be decommissioned. New heating and recovery wells will be installed within the eastern LNAPL body. Preliminary evaluation of the LNAPL data indicates up to 54 electrode locations and 18 fluid recovery wells will be installed within the LNAPL body. Recovery wells will be constructed of 4-inch-diameter, stainless steel wire-wrapped screens and risers. The type and spacing of the thermal elements will be based on subcontractor recommendations. Thermally enhanced recovery is an adaptable approach following initial phases of operation. Both the temperature and heating areas can be adjusted through changes in the application of energy to the system. Depending on observed influence of the extraction wells, additional removal points can be installed if needed, and the frequency of LNAPL extraction can be selected adaptively to align extraction with the observed effects of heating. Following LNAPL recovery, extraction wells can be converted for use as biosparge points to add oxygen to the subsurface. Temperatures in the range of 50°C are optimal for biodegradation of petroleum hydrocarbons and will result in an oxygen-limited system following completion of the low-temperature LNAPL removal. The addition of oxygen while the subsurface temperature is elevated and during cool down will help maintain the active aerobic biodegradation of the residual hydrocarbons.
- As with Alternatives 1, 3, 4, and 5, biosparge system installation includes constructing treatment pads/sheds, and furnishing air compressors, PLCs, distribution manifold and piping, and biosparge wells. In the western LNAPL area, the mainline track area, and

beneath the former Engine House, 50 biosparge wells will be installed to depths between approximately 20 to 40 feet bgs, and three pairs of shallow (25 feet bgs) and deep (60 feet bgs) sparge wells, to address dissolved-phase and LNAPL impacts at varying depths. At the southern end of the LNAPL removal area, 10 paired (20 total) biosparge wells will be installed to depths of approximately 25 feet bgs (shallow) and 60 feet bgs (deep) to address dissolved-phase impacts (Figure 27). The number, location, and depth of biosparge wells in each area will be evaluated based on the MNA assessment.

- Operation and Monitoring: Fluids (LNAPL) will be removed until RELs are achieved. Biosparging operations will continue until groundwater results indicate it has met its design objectives. Following cessation of active remediation, monitored NSZD and MNA will be implemented for 3 years in the eastern and western LNAPL areas, in the mainline track area, beneath the former Engine House, and in the berm area south of the former Power House. MNA will be assessed in general accordance with Ecology's NA Guidance (Ecology 2005) and will include collecting groundwater samples for DRO, ORO, and natural attenuation parameters semi-annually for 3 years. Compliance groundwater monitoring will be implemented following completion of the MNA assessment and will include annual to biennial monitoring for DRO and ORO.
- Groundwater Covenant. A restrictive covenant will be enacted to prohibit the installation of wells for the purpose of water supply within the site boundary; restrict the extraction of groundwater for purposes other than construction and hydraulic control dewatering, monitoring/investigation, or remediation; and require groundwater extracted within the site boundary be evaluated via data review and testing and the discharge managed in accordance with state and federal regulations.

Excavation of impacted soils along the berm, and MNA of groundwater are described in Sections 5.3.6 and 5.3.7, respectively.

Timeframe. The time to achieve RELs is expected to be 2 to 5 years, and the time to reach CULs is estimated to be more than 30 years (depending on NSZD rates) following cessation of active remediation. Active thermally enhanced LNAPL removal is estimated to be implemented over a 1- to 2-year period. Biosparge operations are expected to be implemented over a 2- to 5-year period (concurrent with LNAPL removal) in the western LNAPL, the mainline track area, and beneath the former Engine House. Monitored NSZD and MNA will be assessed for approximately 3 years following cessation of active remediation. Compliance groundwater monitoring of selected wells will continue until CULs are met. ECs and ICs established for the site are expected to remain in place until a no further action determination is provided by Ecology.

Cost: The estimated cost for this alternative is summarized in Appendix D, Table D5.

5.3.6 LNAPL Recovery Estimate

The API LDRM was used to estimate the percentages of (a) recoverable LNAPL and (b) total LNAPL that could potentially be removed from the subsurface in the eastern LNAPL area under Alternatives 1, 4, and 5. The LDRM was populated with laboratory-measured physical

properties of LNAPL and water samples collected from wells OHM-1, OHM-2, and OHM-3; as well as soil properties for soil cores collected from the three OHM well borings. LNAPL removal rates for a representative well for three different temperature scenarios [roughly representing the physical (21 °C) and two (38 and 55 °C) low temperature thermal removal options] were estimated by varying the LNAPL viscosity (based on laboratory data) and elapsed time. The individual well results were then applied to a representative number of wells (18 wells, six of each well type representing the three OHM wells in the eastern LNAPL area) to estimate LNAPL removal from the submerged LNAPL body.

Relative to Alternative 1 and Alternative 4, which include physical LNAPL removal, LNAPL viscosity and density data collected at 21 °C was run for 10 years of LNAPL recovery well operation. This scenario was used to represent LNAPL recovery without subsurface heating. Relative to Alternative 5, which includes heating the subsurface to temperatures between 50 and 70 °C to enhance physical removal of the LNAPL, a LNAPL viscosity and density data collected at 55 °C was run for 1 year of LNAPL recovery well operation. A second temperature heating scenario, using LNAPL viscosity and density data collected at 38 °C, was also modeled for 1 year to assess the impact from heating the subsurface to a very low temperature.

The model estimates the initial total LNAPL and total recoverable LNAPL based on input LNAPL and soil properties, and the LNAPL recovered based on well characteristics and elapsed time. By individual well, the estimated percentage of recoverable LNAPL was higher for a well with a lithology of gravel with silt and sand (OHM-1) than wells with fine sand (OHM-2 and OHM-3).

Model results showed that the estimated percentages of recoverable submerged LNAPL removed were similar for the physical removal (21 °C) scenario after 10 years (20 percent removed) and the low temperature (55 °C) scenario after 1 year (21 percent). The percent recovery for the very low temperature scenario (38 °C) was roughly half (10 percent) of the higher temperature scenario after 1 year. Results indicate that the percent of LNAPL recovered under the physical removal (21 $^{\circ}$ C) scenario over a 10-year period will be roughly the same as the percent recovered over a 1-year period when the low-temperature thermal option is applied to the subsurface. However, the time difference between the two options is negligible when compared to the estimated time to achieve the overall cleanup objectives. The LDRM evaluation is further described in Appendix C.

5.3.7 Shallow Berm Excavation

This remedial process involves removing and disposing of impacted soils offsite at a permitted facility. Based on existing data, a small area with shallow petroleum-hydrocarbon impacted soil is present on the northern side of the berm south of the former Power House, in the vicinity of soil boring B-16-01 and LIF borings TG-CR4, TG-CR4.5, and TF-CR5 (Figures 19 through 22). Petroleum hydrocarbon-impacted soil is anticipated to be encountered at depths between approximately 6 and 9 feet bgs. The estimated volume of in-place petroleum hydrocarbonimpacted soil is 170 cubic yards (cy). The vertical and lateral extents of soils exceeding the CULs will be confirmed in advance of conducting the physical removal. This remedial process is a component of Alternatives 1, 3, 4, and 5 described above.

The excavation process will include removing and stockpiling soil from the upper 6 feet for testing for potential reuse as backfill material. The impacted soils estimated to be between

6 and 9 feet bgs will then be excavated, stockpiled separately, and characterized for disposal at an offsite licensed Subtitle D landfill facility as non-hazardous waste. Subsurface intrusive activities will include cultural resources monitoring, which will be completed in accordance with the CRMP under permit from DAHP.

Excavation sidewalls will be sloped appropriately for safety; however, workers will not be allowed to enter the excavation. *In situ* confirmation samples will be collected in advance of performing the removal activities, to minimize the time the excavation near the riverbank remains open. The excavation will be backfilled with the overburden material, as well as clean fill material (imported from a local quarry) and compacted to existing grade.

Utilities within the excavation area will be temporarily rerouted, as necessary. Dewatering is not expected to be needed based on the known depth of the impacted soils (9 feet bgs) and the average depth to groundwater (10 to 15 feet bgs beneath the berm).

Timeframe: the time required to complete this activity is expected to be less than 3 months.

Cost: The estimated cost is included in the cost estimate spreadsheets in Appendix B.

5.3.8 Groundwater Monitoring

MNA of impacted groundwater is a component of Alternatives 1, 3, 4, and 5 described above, with a contingency of biosparging to enhance biodegradation in Alternatives 3, 4, and 5. Since shallow source material in the unsaturated zone has been removed during previous interim remedial actions (Section 1.3), MNA is an appropriate remedial option for dissolved phase groundwater impacts. Natural attenuation will be evaluated prior to implementing biosparging as a contingency, and performed once RELs are achieved, in the mainline tracks area, the former Engine House area, and in the LNAPL areas and south of the former Power House.

MNA primarily relies on biological transformation or destruction of the petroleum hydrocarbons. Groundwater monitoring will be performed at selected locations to demonstrate conditions needed for biological degradation are present, natural attenuation is occurring, and the groundwater impacts are stable or decreasing. MNA includes collecting groundwater samples for field water quality parameters [e.g., dissolved oxygen (DO) and oxidation-reduction potential (ORP)] and conducting specific laboratory analyses to confirm the natural processes are occurring. Compliance monitoring will be implemented once sufficient MNA data is collected to substantiate MNA is occurring and evaluate the timeframe for concentrations to reach their CULs.

Natural attenuation of residual and dissolved petroleum hydrocarbons through aerobic biodegradation mechanisms relies on sufficient dissolved oxygen [typically greater than 2 milligrams per liter (mg/L)] in groundwater. Therefore, in areas where groundwater is naturally aerobic, biosparging will not be implemented. If dissolved oxygen is depressed (less than 2 mg/L, i.e., anaerobic conditions) and dissolved phase petroleum hydrocarbon concentrations are not stable or decreasing, then biosparging may be implemented to increase the dissolved oxygen concentrations in the aquifer. Following cessation of active remediation (i.e., LNAPL removal and biosparging), groundwater will be monitored to demonstrate remediation goals can be achieved through MNA and NSZD.

MNA primarily relies on biological transformation or destruction of the petroleum hydrocarbons. Groundwater monitoring will be performed at selected locations to demonstrate conditions needed for biological degradation are present, natural attenuation is occurring, and the groundwater impacts are stable or decreasing. MNA includes collecting groundwater samples for field water quality parameters (e.g., DO and ORP) and conducting specific laboratory analyses to confirm the natural processes are occurring. Compliance monitoring will be implemented once sufficient MNA data is collected to substantiate MNA is occurring and evaluate the timeframe for concentrations to reach their CULs.

Timeframe: An MNA assessment will be conducted for a 3-year period. The duration may differ in different areas of the site, depending on the results. Compliance monitoring will be implemented following completion of the MNA assessment and will continue until CULs are achieved.

Cost: Groundwater monitoring costs are included in the cost estimate spreadsheets in Appendix B. The cost estimate spreadsheet for Alternative 2 includes 30 years of compliance monitoring (without an MNA evaluation). The cost estimate spreadsheets for Alternatives 1, 3, 4, and 5 include groundwater monitoring during active remediation (e.g., biosparging, ambient temperature physical LNAPL removal, and/or thermally enhanced LNAPL removal activities) followed by 3 years of MNA and 27 years of compliance monitoring on an annual to biennial basis.

Section 6: Evaluation of Remedial Alternatives

This section presents a preliminary analysis of the remedial alternatives against the MTCA threshold criteria (Section 6.1), followed by detailed analyses (Section 6.2).

6.1 MTCA Threshold Criteria

A remedial action must meet certain threshold criteria to be considered under the MTCA [WAC 173-340-360 (2)(a)]. An alternative cannot be selected if it cannot meet the following threshold requirements:

- Protect human health and the environment
- Comply with cleanup standards
- Comply with applicable state and federal laws
- Provide for compliance monitoring.

A cleanup is presumed to be protective of human health and the environment if it achieves the CULs and/or mitigates exposure through controls. Compliance with cleanup standards involves achieving CULs at an appropriate point of compliance, implementing ICs in those area where remediation to CULs is not appropriate or feasible but RELs have been met. An alternative can comply with applicable federal and state laws by protecting human health and the environment through a combination of active and passive remedial measures and/or the implementation of ICs.

Compliance monitoring assesses the protection of human health and the environment during construction and the O&M period of a cleanup action. Compliance monitoring assesses whether a remedial action has met RELs and/or CULs and verifies its long-term effectiveness. Compliance with the threshold requirements does not imply untreated hazardous substances cannot remain onsite. MTCA recognizes non-treatment alternatives can comply with cleanup standards, provided compliance monitoring is included to confirm protection of human health and the environment.

Table 6 summarizes the evaluation of the alternatives in relation to MTCA's threshold criteria. Based on this evaluation, five proposed alternatives meet the threshold criteria - they can achieve CULs; have an acceptable point of compliance; and provide for compliance monitoring. No Action is included as required for comparison. While No Action will achieve CULs over a long timeframe, it is unlikely to be accepted as compliant with state and federal laws and does not provide for compliance monitoring.

6.2 Alternatives Analysis

MTCA specifies that when selecting a remedial action, preference shall be given to actions that are "permanent to the maximum extent practicable." To determine whether a remedial action uses permanent solutions to the maximum extent possible, a disproportionate cost analysis

(DCA) shall be used [WAC 173-340-360(3)(b)]. Costs are disproportionate to benefits if the incremental cost of the alternative over that of a lower cost alternative exceed the incremental degree of benefits achieved by the alternative over that of the lower cost alternative.

The most practical permanent alternative evaluated in the feasibility study shall be the baseline cleanup alternative against which other alternatives are compared. The permanency of alternatives is largely qualitative and is based on best professional judgment.

This section evaluates each remedial alternative against seven criteria set in WAC 173-340- 360(3)(f) in order to establish whether a cleanup is permanent to the maximum extent practical. The seven criteria are:

- 1. Protectiveness
- 2. Permanence
- 3. Cost
- 4. Effectiveness over the long-term
- 5. Management of short-term risks
- 6. Technical and administrative implementability
- 7. Consideration of public concerns.

These criteria, as well as a restoration timeframe and compliance with federal and state ARARs, are evaluated below. An evaluation of the five alternatives relative to these seven criteria is provided in Table 6. The following summarizes the analysis of the alternatives under each criterion ranked on a scale of 1 to 10. A score of 10 indicates the alternative meets the criterion significantly well and 1 indicates the alterative does not meet the criterion well or at all. Remediation elements common to Alternatives 1, 2, 3, 4, and 5 do not contribute to differential rankings as these common elements carry the same benefits and risks when applied similarly or identically. The most significant differences are observed in evaluating the LNAPL remedy for each alternative; therefore, the following evaluation is focused on the various proposed remedial alternatives for LNAPL.

6.2.1 Protectiveness

Protectiveness is the degree existing risks are reduced, the time required to reduce risk and attain cleanup standards, onsite and offsite risks resulting from implementing the alternative, and improvement of overall environmental quality.

Alternatives 2 and 3 are protective, but do not actively remove LNAPL, leaving higher potential for future exposure. Alternatives 1 and 4 will remove a fraction of the LNAPL mass, leaving residual mass for NSZD. Alternative 5 will remove LNAPL faster than Alternatives 1 and 4 but will still leave residual mass for NSZD. Active remediation via biosparging is included in Alternatives 1, 4, and 5 for the western LNAPL area and in Alternatives 3, 4, and 5 for dissolved phase impacts. Tradeoffs between timeframe and likelihood of effectiveness result in the differential scores among these alternatives. Alternatives 1, 4, and 5 are similar in intent to actively address LNAPL source and differ in timeframe and approach and are the most protective.

6.2.2 Permanence

A permanent cleanup achieves cleanup standards without requiring further action such as longterm monitoring or ICs. The remedial action alternatives were compared based on their adequacy in destroying hazardous substances, reducing or eliminating hazardous substance releases and sources, the irreversibility of waste treatment process, and the characteristics and quantity of treatment residuals generated.

Alternatives 2 and 3 do not provide permanent measures to address site conditions without long-term monitoring or ICs, though Alternatives 2 and 3 provide long-term protection with ongoing maintenance. Alternatives 1, 4, and 5 include LNAPL removal to the extent practicable using different methods. Each option includes long-term monitoring.

6.2.3 Cost

The costs to implement the alternatives, including the construction, operation, and the long-term monitoring, were estimated to assess practicability (see Section 7.2). Long-term costs include O&M costs, monitoring costs (including 30 years of compliance monitoring after cessation of active remediation), equipment replacement costs, and the costs of maintaining ICs.

Alternative 2 is the lowest cost alternative that includes measures to protect human health and the environment, which consist primarily of long-term monitoring and reporting once controls are in place. Alternatives 1, 3, 4, and 5 have similar or identical costs for some project elements, including planning, biosparging, shallow excavation, groundwater monitoring, and reporting elements that scale based on the number of locations included or the duration of activities. Alternatives 3 and 4 were ranked the same as their respective costs were within 5 percent. Costs for Alternatives 4 and 5 include biosparge treatment in five areas [western LNAPL area, eastern LNAPL area (after LNAPL removal), and dissolved phase in the southern end of eastern LNAPL area, mainline tracks area, and former Engine House] compared to one area for Alternative 1 (western LNAPL area) and two areas for Alternative 3 (mainline tracks area and former Engine House). The most significant differences in conceptual scope and cost among these alternatives are for the LNAPL remedy. The cost to implement physical removal (Alternatives 1 and 4) is significantly less than the cost to implement LTTR (Alternative 5). Estimated costs for each alternative are described in Section 7.

6.2.4 Long-Term Effectiveness

Long-term effectiveness is the degree of certainty that the alternative will be successful, the reliability of the alternative during the period of time hazardous substances are expected to remain onsite at concentrations that exceed CULs, the magnitude of residual risk with the alternative in place, and the effectiveness of controls required to manage treatment residues or remaining wastes.

Long-Term Effectiveness

Alternatives 2 and 3 provide effective long-term control but rely on controls to protect human health and the environment for a long period of time while hazardous substances remain on site, increasing uncertainty of future protection. Alternatives 1, 4, and 5 each remove LNAPL,

significantly reducing the likelihood of exposure in the future. Alternatives 1 and 4 remove LNAPL at ambient temperatures using simple and measurable methods; changes in LNAPL thickness and transmissivity can be measured over time and risks from unidentified isolated residuals or migration are limited. Alternative 4 uses biosparging as a more active remedy to address dissolved phase impacts. Alternative 5 uses heat to modify the physical properties of the LNAPL to shorten the extraction time; but is not expected to significantly increase the amount of LNAPL removed over the project lifecycle compared to Alternatives 1 and 4.

6.2.5 Short-Term Risks

Short-term risk is the risk to human health and the environment associated with each alternative during construction and implementation, and the effectiveness of measures that will need to be taken to manage such risks.

Short-Term Risks

Alternative 2 has little to no risk associated with implementation related to ICs and other planned activities are relatively limited in scale. Alternative 3 includes the most invasive and significant construction elements. Installation of sealed interlocking sheet piles to the proposed depths is challenging work that involves the management of significant health and safety risks to workers and the most significant potential environmental impact of the activities in each of the alternatives. Alternatives 1 and 4 include conventional drilling and LNAPL recovery methods that have relatively lower short-term risk ranking among the active remedial approaches. Alternative 5 includes construction risks similar to Alternative 4 with the addition of construction of treatment systems and high voltage electrical connections, and the addition of heat to the subsurface system. Operational risks for Alternative 5 include heating of the subsurface to moderate temperatures that pose limited thermal risks. Controlling the temperature to remain between 50 and 70°C limits LNAPL migration and vaporization risks. The use of ERH introduces electrical risk from current passing through the subsurface to produce heat.

6.2.6 Ability to Implement

The ability to implement includes technical feasibility; availability of necessary offsite facilities, services, and materials; administrative and regulatory requirements; scheduling; access

constraints; and integration with existing facility operations and other current or potential remedial actions.

Alternative 2 is easily implemented. Alternative 3 includes significant construction elements and is therefore, challenging to implement. Installation of sealed interlocking sheet piles to the proposed depths is challenging work. Alternatives 1 and 4 include conventional drilling and LNAPL recovery methods that are relatively easy to implement. Alternative 5 involves construction implementation methods similar to Alternatives 1 and 4 with the addition of treatment system construction and high voltage electrical connections. For Alternative 5, controlling the temperature to remain between 50 and 70°C is implementable using thermocouples throughout the wellfield.

6.2.7 Consideration of Public Concerns

Ecology may assist with considering public concerns during selection of the remedial action. A Public Notice and Participation period is required (WAC 173-340-600) before implementation of the action.

6.2.8 Restoration Timeframe

MTCA requires that remedial alternatives provide a reasonable restoration timeframe, which means:

- Assessing the remedial alternatives relative to:
	- o The practicability of achieving a shorter restoration timeframe
	- \circ The ability to achieve cleanup levels at the point of compliance with a greater degree of long-term effectiveness.
- Considering the impacts to current and future use of the site, surrounding areas, and associated resources that are, or may be, affected by releases from the site.
- Considering the availability of alternative water supplies.
- Evaluating the likely effectiveness and reliability of institutional controls.

The time required to attain RELs for each remedial alternative was estimated and is summarized below. Based on the available data, the timeframe to achieve restoration is estimated to be more than 30 years for each alternative.

Restoration Timeframe

The excavation element common to Alternatives 1, 3, 4, and 5 can occur quickly after a period of design and permitting and will be accomplished in less than 1 year, and concurrent with construction of other remedial systems. Biosparging elements common to these alternatives could likewise be designed and constructed in a short timeframe (less than 1 year) and will likely operate on the order of 2 to 5 years, if needed. Alternatives 2 and 3 do not actively remove LNAPL, leaving higher potential for future exposure. Alternatives 1 and 4 will remove a fraction of the LNAPL mass, leaving residual mass for NSZD. Physical removal will recover the mobile LNAPL within a period of less than 10 years with curtailed recovery events following the initial more intensive period. Alternative 5 will remove mobile LNAPL faster (within 1 to 2 years following implementation) than Alternatives 1 and 4 but will still leave residual mass for NSZD.

Assessment of monitored NSZD and MNA in applicable alternatives, will extend 3 years beyond cessation of the active remedy. Compliance monitoring will be implemented following completion of active remediation and the MNA assessment. For cost estimating purposes, compliance monitoring is estimated to extend for 27 years beyond completion of the 3-year monitored NSZD and MNA assessment period.

6.2.9 Summary of Alternative Ranking

The alternative rankings described above, including the calculated total score, are summarized in Table 6. Alternative 4 has the highest total score while Alternative 3 has the lowest. Because of tradeoffs between criteria, total scores for Alternatives 1, 4, and 5 are within a similar range of values.

6.2.10 Compliance with ARARs

Action-specific ARARs regulate technologies or activities associated with the implementation of the remedial action. Action-specific ARARs are typically technology- or activity-based requirements or limitations. Potential ARARs for the site include the soil and groundwater CULs provided by MTCA, as well as the non-numerical requirement LNAPL be removed to the extent necessary to mitigate the potential for migration. Alternative 2 does not satisfy these ARARs as they will not include removal or containment of site-related constituents. Alternative 3 will satisfy ARARs by improving groundwater conditions at the selected point of compliance to meet

groundwater CULs; however, Alternative 3 does not include removal of LNAPL as it is based on physical containment of the LNAPL body. Alternatives 1, 4, and 5 satisfy ARARs as these alternatives are intended to remediate site soil and groundwater in accordance with MTCA and include removal and biological degradation of the LNAPL mass.

Section 7: Comparative Analysis of Alternatives

7.1 Comparative Analysis

Based on the analysis of alternatives above, the following comparative analysis and disproportionate cost analysis (Section 7.2) contribute to the identification of the recommended alternative. The five alternatives are compared as most favorable to least favorable as:

- Alternative 4 Physical LNAPL Removal, Biosparge, MNA, and Targeted Excavation
- Alternative 5 LTTR, Biosparge, MNA, and Targeted Excavation
- Alternative 1 Physical LNAPL Removal, Focused Biosparge, MNA, and Targeted **Excavation**
- Alternative 3 LNAPL Containment, Biosparge, MNA, and Targeted Excavation
- Alternative 2 Institutional Controls and Environmental Covenants with Compliance Groundwater Monitoring

Alternative 4 – Physical LNAPL Removal is a proven and mature approach that is readily implementable with commonly available equipment. It also requires little specialized labor besides the installation of LNAPL removal wells, and its impact on site operations will be minimal. The frequency of LNAPL removal and other operational parameters will be adjusted to maintain efficiency over the duration of the remediation process (e.g., performing more frequent LNAPL removal in the summer months, adjusting removal frequency as LNAPL mass decreases). Using seasonal temperature change to maximize LNAPL recovery adds a sustainability element to the project as compared to electrical heating. While the treatment timeframe is longer than Alternative 5, Alternative 4 avoids the operational and implementability challenges and significant energy consumption and resulting carbon emissions that come with ERH.

Alternative 5 – LTTR using ERH is a technically advanced, proven, and mature approach that requires specialized equipment and the application of significant electrical power. Operation of ERH systems is challenging, but routinely achieves target temperatures and cleanup goals in conditions similar to those at the site. Operation is adaptable to target LNAPL removal in specific areas or enhance operation for variable timeframes based on performance. The equipment installation and operation are expensive and therefore, are only efficient when applied continuously without extended shutdowns that result in cooling. System efficiency declines as the timeframe needed for cleanup increases as high energy inputs continue while mass removal efficiencies decrease. High energy use is associated with higher carbon emissions and relatively lower sustainability.

Alternative 1 – Physical LNAPL Removal is a proven and mature approach that is readily implementable with commonly available equipment. It also requires little specialized labor besides the installation of LNAPL removal wells, and its impact on site operations will be minimal. The frequency of LNAPL removal and other operational parameters will be adjusted to

maintain efficiency over the duration of the remediation process (e.g., performing more frequent LNAPL removal in the summer months, adjusting removal frequency as LNAPL mass decreases). Using seasonal temperature change to maximize LNAPL recovery adds a sustainability element to the project as compared to electrical heating. The treatment timeframe is longer than Alternative 4 due to relying on MNA and NSZD processes to remediate dissolved phase impacts in several areas rather than additional biosparging systems. The treatment timeframe is also longer than Alternative 5; however, Alternative 1 avoids the operational and implementability challenges and significant energy consumption and resulting carbon emissions that come with ERH.

Alternative 3 – Three-Sided Sheet Pile Wall provides adequate protection of both human health and the environment and contains the dissolved phase petroleum hydrocarbons eliminating the migration pathway to the river but does not accelerate the timeframe for cleanup over natural processes and leaves LNAPL in place behind the sheet pile wall. It is also the most challenging from a construction and implementability perspective due to the methods involved and the quality control needed to achieve certainty of remedy performance. Equipment used to install sheet pile walls have relatively higher environmental impact and safety concerns. Therefore, while protective, it is not preferred over other alternatives.

Alternative 2 – Institutional Controls and Environmental Covenants with Compliance Groundwater Monitoring could feasibly protect human health over the timeframe needed for natural processes to address site-related compounds, but do not actively protect applicable potential environmental receptors via applicable potential pathways over the long-term. Alternative 2 is the most sustainable protective alternative with the lowest emissions and overall impact to the environment.

7.2 Disproportionate Cost Analyses

Costs for the five alternatives are summarized below and presented in detail in Appendix D, Tables D1 through D5 with a summary in Table D6. Several components of the cost estimates, including planning, permitting, administration and reporting; biosparging; shallow berm excavation; and monitoring have similar cost estimates and expected timeframes. The LNAPL remedies have the most significant differences in cost to achieve similar outcomes.

Alternatives 2 and 3 rely, in part, on the mitigation of the direct contact exposure pathway through institutional controls and reliance on groundwater monitoring to assess the natural

processes that may affect some of the contaminant concentrations. Alternatives 1, 4, and 5 involve a variety of cleanup actions with LNAPL removal and biosparging or MNA as their central component. The primary difference between these three alternatives is how LNAPL will be remove from the subsurface. The three alternatives each propose to remove LNAPL to the maximum extent practicable, consistent with the requirements under MTCA. Alternatives 1 and 4 propose to remove LNAPL to the extent practicable under ambient conditions while Alternative 5 involves applying heat to reduce LNAPL viscosity, thereby enhancing recovery at the extraction points. Under Alternative 5, RELs are expected to be achieved in less than 5 years, whereas under Alternatives 1 and 4, RELs are expected to be achieved over a 10-year to 15-year (Alternative 4 if eastern LNAPL biosparging is implemented) period. At the end of the active remedy phase, work will transition to passive remediation through MNA and NSZD. Restoration timeframes under each of these three alternatives is estimated to be more than 30 years once active remediation has ceased.

Regarding the DCA, as stated in WAC 173-340-360(3)(e)(ii)(B), the most practicable permanent solution evaluated in the feasibility study shall be the baseline cleanup action alternative against which cleanup action alternatives will be compared. Alternative 1 fits this criterion. The permanency of alternatives is largely qualitative and is based on best professional judgment. To document the qualitative analysis, weighting factors are assigned for each of the six non-cost benefits criteria listed in WAC 173-340-360(3)(f) to represent the importance of each benefit criterion and are expressed as a percent. Weighting factors, as provided by Ecology in its 16 November 2021 memorandum to BNSF, for each non-cost criteria are summarized below.

- **Protectiveness.** A weighting factor of 30% is assigned based on its overarching importance relative to the ultimate goal of environmental cleanup and protection of human health and the environment.
- **Permanence.** A weighting factor of 20% is assigned in association with the need or lack thereof for further action in the future.
- **Long Term Effectiveness.** A weighting factor of 20% is assigned in association with a measure of certainty related to the robustness of the action, as well as confidence in the technology used for the protection of human health and the environment.
- **Short Term Risk.** A weighting factor of 10% is assigned because the majority of shortterm risks can be managed through the use of best practices during process design and construction.
- **Ability to Implement.** A weighting factor of 10% is assigned because, although an important consideration, implementability is less associated with environmental concerns than with the above criteria.
- **Consideration of Public Concerns.** A weighting factor of 10% is assigned because the majority of public concern issues are incorporated in the protectiveness, permanence, and long-term effectiveness criteria.

The overall ranking (i.e., the benefit) of each alternative was compared to its cost to provide a benefit/cost ratio. Based on the benefit/cost ratio evaluation, presented in Table 7, Alternative 2

is the best alternative for the site. However, as Alternative 2 relies on the mitigation of the direct contact exposure pathway through institutional controls and does not actively remove the LNAPL to the maximum extent practicable, this alternative is eliminated from consideration. Alternatives 1 and 4 have similar benefit/cost ratios. As Alternative 4 ranks higher in its total weighted benefit, Alternative 4 is a better alternative for the site.

Section 8: Summary of Recommended Alternative

Alternative 4, which involves periodic physical LNAPL removal within the LNAPL body and targeted remediation for other site areas, is the recommended alternative. Below is a summary of the recommended remedial alternative and its key advantages.

Alternative 4 includes the following components:

- Eastern LNAPL area: Install approximately 15 LNAPL recovery wells within the LNAPL body. LNAPL will be removed periodically from these 15 wells plus three existing OHM wells by high-vacuum extraction or other methods depending on the results of the LDRM and field implementation testing. Adequate time will be allowed for LNAPL to recharge prior to subsequent extraction events. Once LNAPL removal has met the RELs, monitored NSZD will be assessed for a 3-year period to confirm residual hydrocarbons will continue to degrade. Biosparging may be implemented as necessary to enhance biological degradation and address dissolved phase DRO and ORO if present from residual LNAPL.
- Western LNAPL area: Install 11 biosparge wells in the western LNAPL area to increase *in situ* degradation rates. Implement biosparging until objectives for active remediation have been met. Biosparge wells will be installed to depths ranging from approximately 20 to 40 feet bgs (shallow) and 60 feet bgs (deep) to address dissolved-phase and LNAPL impacts at varying depths. Once LNAPL RELs have been achieved in the western LNAPL area, monitored NSZD will be implemented for a 3-year period to confirm residual hydrocarbons will continue to degrade.
- Groundwater impacts (multiple areas): Implement biosparging if needed until objectives for active remediation (sustained aerobic groundwater conditions) have been met. Install 65 biosparge wells, as needed, around the mainline tracks area, beneath the former Engine House, and at the southern end of the eastern LNAPL removal area, to depths ranging from approximately 20 to 65 feet bgs to address dissolved-phase impacts at varying depths; in some areas, paired shallow and deep wells will be installed. Once biosparging operations have met their RELs, active remediation will be discontinued. MNA will then be implemented for a 3-year period to confirm remaining dissolved phase hydrocarbon concentrations will continue to decrease. Groundwater samples will be analyzed for DRO, ORO, and natural attenuation parameters.
- Compliance groundwater monitoring: Select wells will be monitored for DRO, ORO, and other water quality parameters following completion of the MNA and monitored NSZD programs. Compliance groundwater monitoring will continue until site-related constituents in groundwater are at or below their respective CULs, or until otherwise directed by Ecology.
- Shallow soil impacts under berm area: A small area with shallow petroleum hydrocarbon-impacted soil is present below the northern side of the berm south of the former Power House. Petroleum hydrocarbon-impacted soil from this area will be excavated and disposed of offsite. The excavation process will include removing and

stockpiling soil from the upper 6 feet for testing for potential reuse as backfill material. The impacted soils between 6 and 9 feet bgs will then be excavated, stockpiled separately, and characterized for disposal at an offsite licensed Subtitle D landfill facility as a non-hazardous waste.

The selected alternative is protective of potential receptors and considers overall environmental impact and sustainability, while avoiding implementation challenges, performance uncertainty, and short-term impacts posed by Alternative 5. While LNAPL removal at ambient temperature results in a longer remediation timeframe, the estimated restoration timeframe is similar to Alternative 5 and recovers neither more nor less of the recoverable LNAPL. Because the LNAPL is not migrating towards the river, the additional time anticipated for physical removal does not represent an increased risk to the environment. As the site's use as an active railyard is not expected to change for the foreseeable future, the remedy will not result in impacts to current or future use of the site, and ICs and ECs will be easily maintained.

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Tables

SUMMARY OF CONSTITUENTS OF CONCERN (COCs) AND PROPOSED CLEANUP LEVELS (CULs) BNSF Wishram Railyard, Wishram, Washington

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Notes:

COC = Constituents of concern

CUL = Cleanup level. Refer to Appendix B to the Draft Feasibility Study Report for more information.

µg/L = micrograms per liter

mg/kg = milligrams per liter

Total TPH-Dx = Total TPH-Dx concentrations calculated by summing diesel-range organics (DRO) and oil-range organics (ORO) concentrations.

Cleanup Levels:

- Groundwater: Cleanup level values based on Model Toxics Control Act (MTCA) Method B values (B Cancer or B Non Cancer) and MTCA Method A values for groundwater (Table 720-1) based on Washington State Administrative Code (WAC) 173-340-720 from Cleanup Levels and Risk Calculation (CLARC) tables (Accessed February 2021), and Washington State (246-290 WAC) and Federal (40 CFR 141) Maximum Contaminant Levels (MCLs).
- Surface Water: Cleanup level values based on Ecology MTCA Method B values (B Cancer or B Non Cancer), MTCA Method A values and other applicable, relevant, and appropriate requirements (ARARs) under applicable state (173-201A-240 WAC) and federal laws [Section 304 of the Clean Water Act (CWA); 40 CFR Subpart D 131.45] for surface water based on WAC 173-340-730 from CLARC tables (Accessed February 2021).
- Environmental Effects-Based Concentrations, from Concentrations of Gasoline and Diesel Range Organics Predicted to be Protective of Aquatic Receptors in Surface Waters, Implementation Memorandum No. 23 (Ecology, 25 August 2021).
- Soil: Cleanup level values based on Ecology MTCA Method C values (Cancer and Noncancer) and MTCA Method A (Table 745-1) values for soil based on WAC 173-340-745 and 3-Phase Model Soil Protective of Groundwater Vadose (Eq. 747-1) and 3-Phase Model Soil Protective of Groundwater to Surface Water Vadose Fresh Water (Eq. 747-1). from CLARC tables (Accessed February 2021).

TABLE 2

SUMMARY OF REMEDIAL AREAS AND VOLUMESBNSF Wishram Railyard

Notes:

 $sf = square feet$ cy = cubic yards $bgs - below ground surface$ gal = gallons $cf = cubic$ feet LNAPL = light non-aqueous phase liquid

Volume of total fluids based on total porosity estimate of 40%.

Volume of LNAPL based on estimate of initial LNAPL saturation in soil cores of approximately 33% in eastern LNAPL area (geomean of 4 results) and 17% in western LNAPL areas (OHM-4 core).

Volume of Recoverable LNAPL by physical means from geomean of residual core plug saturations for the eastern LNAPL area (28% of pore volume) and the OHM-4 core result (0.6% of pore volume) in the western LNAPL area.

GENERAL RESPONSE ACTIONS AND REMEDIAL TECHNOLOGIES CONSIDERED BNSF Wishram Railyard

GENERAL RESPONSE ACTIONS AND REMEDIAL TECHNOLOGIES CONSIDERED BNSF Wishram Railyard

TABLE 3

GENERAL RESPONSE ACTIONS AND REMEDIAL TECHNOLOGIES CONSIDERED BNSF Wishram Railyard

TABLE 4

 \blacksquare

DETAILED SCREENING OF REMEDIAL TECHNOLOGIES BNSF Wishram Railyard

 \blacksquare

DETAILED SCREENING OF REMEDIAL TECHNOLOGIES BNSF Wishram Railyard

TABLE 4

DETAILED SCREENING OF REMEDIAL TECHNOLOGIES BNSF Wishram Railyard

TABLE 4

DETAILED SCREENING OF REMEDIAL TECHNOLOGIES BNSF Wishram Railyard

TABLE 4

 \blacksquare

DETAILED SCREENING OF REMEDIAL TECHNOLOGIES BNSF Wishram Railyard

SUMMARY OF REMEDIAL ALTERNATIVES BNSF Wishram Railyard

cubic yards. Laboratory testing to evaluate reuse of non-impacted

e time for recharge of LNAPL into extraction points between events. **ENAPL** is removed.

Sparge air to stimulate biological degradation of LNAPL impacts in area after cessation of biosparging.

ea (just north of the river berm). Monitor groundwater for natural ctive treatment (biosparging and LNAPL removal). Compliance

dents). Compliance groundwater monitoring will be implemented in

cubic yards. Laboratory testing to evaluate reuse of non-impacted

No active treatment of source area. A Groundwater Covenant will be

e air to stimulate biological degradation of dissolved-phase

olishing step following cessation of active treatment (biosparging).

cubic yards. Laboratory testing to evaluate reuse of non-impacted

e time for recharge of LNAPL into extraction points between events. . LNAPL is removed . Biosparging may be needed following

e and in the western LNAPL area, and install paired shallow (25 feet of LNAPL impacts (in the western LNAPL area) and dissolved-NAPL area after cessation of biosparging.

hing step following cessation of active treatment (biosparging and

cubic yards. Laboratory testing to evaluate reuse of non-impacted

Aboveground treatment of extracted fluids, off-site disposal of npacts in source area after drainable LNAPL is removed and system ns in saturated interval and further stimulate degradation of residual

e and in the western LNAPL area, and install paired shallow (25 feet of LNAPL impacts (in the western LNAPL area) and dissolved-NAPL area after cessation of biosparging.

hing step following cessation of active treatment (biosparging and

TABLE 6

ALTERNATIVES ANALYSIS SUMMARYBNSF Wishram Railyard

Abbreviations

ICs and ECs with Compliance GWM = Institutional Controls and Environmental Covenants with Compliance Groundwater Monitoring

LNAPL = light non-aqueous phase liquid

MNA = Monitored Natural Attenuation

LTTR = Low-Temperature Thermal Removal

TABLE 7

BENEFIT/COST ANALYSIS

BNSF Wishram Railyard

February 2022 2196120*06

Figures

Note: 1. Locations are approximate.

Kennedy/Jenks Consultants

BNSF Wishram Railyard Wishram, Washington

Site Location Map

2196120*06 December 2021

Figure 1

- $-$ Former Sewer Line (Potential)
-
- Existing Site Feature
	-
-
-
-

Note: 1. Locations are approximate.

Legend

- **W** Former Water Supply Well (Approximate)
- ---- Approximate BNSF Property Line
-
- T⁻⁻⁻] Former Site Feature
- **Former Bunker Fuel / Oil Pipeline**
- --- Former Oil Drain
- Former Oil Trough
-
- Stormwater Underdrain (A portion $\overline{}$ removed from service circa 1960)
- Stormwater Underdrain (Rerouted portion circa 1960)
- ----- Former Diesel Line
- $==$ Former Steam Line
- Off-Railyard Parcel Boundaries

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

2196120*06 May 2022 **Figure 2**

Current and Historical Site Features

-
- Stormwater Underdrain (A portion $\overline{}$ removed from service circa 1960) Stormwater Underdrain (Rerouted
- portion circa 1960)
- ----- Former Diesel Line
- $==$ Former Steam Line

Notes: 1. Locations are approximate. 2. Background image from U.S. Army Corps of Engineers, 1951.

Legend

- **W** Former Water Supply Well (Approximate) $-$ Former Sewer Line (Potential)
- ---- Approximate BNSF Property Line
- **----- Former Railroad Tracks**
- **Former Bunker Fuel / Oil Pipeline**
- ---- Former Oil Drain
- Former Oil Trough

Existing Site Feature

Former Site Feature

BNSF Wishram Railyard Wishram, Washington

Kennedy/Jenks Consultants

2196120*06 December 2021 **Figure 3**

Historical Site Features

- Former Bunker Fuel / Oil Pipeline Former Site Feature
- ---- Former Oil Drain
- Former Oil Trough
- $-$ Former Sewer Line (Potential)
- Stormwater Underdrain
- Stormwater Underdrain

$\frac{N}{2}$ N. 100 Scale: Feet

2196120*06 December 2021 **Figure 4**

Legend

---- Approximate BNSF Property Line

Existing Site Feature

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

> **Current and Historical Site Features - East Area**

Notes: 1. Locations are approximate. 2. Background image from U.S. Army Corps of Engineers, 1962.

- ---- Approximate BNSF Property Line
- Existing Site Feature
- [⁻⁻⁻] Former Site Feature

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

2196120*06 December 2021 **Figure 5**

Suspected Diesel Fueling Area Suspected Oil Fueling Area Approximate Previous Excavation Area Approximate Previous Excavation Area Approximate Previous Excavation Area

- **Existing Site Feature**
- Former Site Feature

 $\frac{N}{2}$ 0 50 100 Scale: Feet

Legend

- **A** Bioventing Injection Well
- \triangle Air Sparge (AS) Well
- --- Approximate BNSF Property Line
- **Former Bunker Fuel / Oil Pipeline**
- ---- Former Oil Drain
- Former Oil Trough
- $-$ Former Sewer Line (Potential) Stormwater Underdrain (A portion removed from service circa 1960)
- Stormwater Underdrain (Rerouted portion circa 1960) **?**
- Former Diesel Line
- $==$ Former Steam Line

 $\overline{}$

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

2196120*06 December 2021

Previous Interim Remedial Actions

Legend

- **.** LNAPL Observed in Boring
- **LIF Response (Diesel, Oil and/or Mixed)**
- e inferred Diesel/Oil-Like LNAPL (> 60% RE)
Inferred Diesel/Oil-Like LNAPL (> 60% RE) International Extent of Shallow (at a late of the United Shallow (at
- Residual LNAPL (20 to 60 % RE)
- No LNAPL or Residual LNAPL (< 20% RE)
- Inferred Lateral Extent of Shallow (at water table) Diesel-Like LNAPL Impacts
-
- % RE) $\frac{1}{1}$ $\frac{1}{1}$ water table) Oil-Like LNAPL Impacts
	- **----** Approximate BNSF Property Line

<u>Notes:</u>

BNSF Wishram Railyard Wishram, Washington

Kennedy/Jenks Consultants

Scale: Feet

2196120*06 December 2021 **Figure 7**

Combined Inferred Shallow LNAPL Extent Map

- 1. Locations are approximate.
- 2. LNAPL = light non-aqueous phase liquid
- 3. Inferred lateral extent of potentially mobile Diesel- or Oil-Like LNAPL based on interpretation of LIF waveforms (July 2013) and soil boring logs.

Legend

- LNAPL Observed in Boring
- **LIF Response (Diesel, Oil and/or Mixed)**
- Inferred Diesel/Oil-Like LNAPL (> 60 %RE)
■ Inferred Diesel/Oil-Like LNAPL (> 60 %RE)
■ Inferred Diesel/Oil-Like LNAPL (> 60 %RE)
- Residual LNAPL (20 to 60 % RE)
- No LNAPL or Residual LNAPL (< 20 %RE)
- Inferred Lateral Extent of Submerged (below
- water table) Diesel-Like LNAPL Impacts
-
- water table) Oil-Like LNAPL Impacts
- **----** Approximate BNSF Property Line

<u>Notes:</u>

Path: N:\BNSF Washington\Wishram\GIS\Events\2021Events\FS Figures Updated\Fig08_LIF_Dx_SUB.mxd © 2021 Kennedy/Jenks Consultants

BNSF Wishram Railyard Wishram, Washington

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2196120*06 December 2021 **Figure 8**

Combined Inferred Submerged LNAPL Extent Map

- 1. Locations are approximate.
- 2. LNAPL = light non-aqueous phase liquid
- 3. Inferred lateral extent of potentially mobile Diesel- or Oil-Like LNAPL based on interpretation of LIF waveforms (July 2013) and soil boring logs.

BNSF Wishram Railyard **Kennedy/Jenks Consultants**

Wishram, Washington

Legend

Notes:

- 1. Locations are approximate.
- 2. DRO = Diesel-Range Organics.
- ORO = Oil-Range Organics.
- 3. Results present for soil samples collected between 2002 and 2018.
- 4. MTCA Method A Cleanup level for diesel- and oil-range organics (DRO and ORO) in soil is 2,000 milligrams per kilogram (mg/kg).
- 5. Unsaturated depths are from approximately 0 to 10 feet below
- ground surface (bgs).

 $\frac{N}{2}$ N **DRO Result**

2. DRO = Diesel-Range Organics.

2. DRO = Diesel-Range Organics.

2. DRO = Oil-Range Organics.

3. Results present for soil samples collected between 2002 and 2018.
 DRO/ORO in Subsurface Soil
 DRO/ORO in 0 50 100 Scale: Feet

2196120*06 Decem b er 2021 **Figure 9**

-
- Result is above MTCA Method A CUL
- Result is below MTCA Method A CUL
- Result is below laboratory reporting limit

³

- Result is below laboratory reporting limit
- LNAPL Observed in Boring \bullet

Notes:

Legend

Wishram, Washington

BNSF Wishram Railyard **Kennedy/Jenks Consultants**

Scale: Feet

2196120*06 December 2021 **Figure 10**

DRO/ORO in Subsurface Soil (Saturated) - Main Area

- 1. Locations are approximate.
- 2. DRO = Diesel-Range Organics.
- ORO = Oil-Range Organics.
- 3. Results present for soil samples collected between 2002 and 2018.
- 4. MTCA Method A Cleanup level for diesel- and oil-range organics (DRO and ORO) in soil is 2,000 milligrams per kilogram (mg/kg).
- 5. Saturated depths are from approximately 10 feet below ground surface (bgs) to bedrock.
- 6. Not all points are labeled. See Figures 6B, 7, and 25 in Uplands Remedial Investigation Report.

Legend

 \bullet

DRO Result

Result is above MTCA Method A CUL

BNSF Wishram Railyard Wishram, Washington

Kennedy/Jenks Consultants

2196120*06 December 2021 **Figure 11**

Groundwater Sampling Results DRO and ORO (2012 - 2018) - Main Area

Note:

1. Locations are approximate.

- 2. "Diesel" = Diesel-Range Organics.
- "Oil" = Oil-Range Organics.
- 3. MTCA Method A Cleanup levels (CULs) for diesel- and oil-range organics (DRO and ORO) in groundwater is 500 micrograms per liter (µg/l).
- 4. Reconnaissance groundwater samples from 2012, 2014, 2016, and 2018 are shown. Samples from 2012 and 2014, and some samples from 2016, were analyzed with silica gel cleanup. Monitoring well sample resutls from August 2018 are shown.
- 5. Bold font indicates result reported above the laboratory reporting limit, blue font indicates result reported above the MTCA Method A CUL.

LNAPL observed, well not sampled

Result is below MTCA Method A CUL Result is below laboratory reporting limit

r - **Approximate Lateral Extent of Dissolved - -** Phase Diesel and/or Oil

Note: **F** - **Approximate Lateral Extent of Dissolved**

Phase Diesel and/or Oil

Legend

BNSF Wishram Railyard Wishram, Washington

2196120*06 December 2021

Figure 12

Groundwater Sampling Results Total TPH-Dx (2012 - 2018) - Main Area

- 1. Locations are approximate.
- 2. MTCA Method A Cleanup level (CUL) for total petroleum hydrocarbons
- (TPH) in groundwater is 500 micrograms per liter (µg/l).
- 3. Total TPH-Dx = Sum of diesel-range and oil-range organics results. 4. Reconnaissance groundwater samples from 2012, 2014, 2016, and 2018 are shown. Samples from 2012 and 2014, and some samples from 2016, were analyzed with silica gel cleanup. Monitoring well sample resutls from August 2018 are shown.
- 5. Bold font indicates result reported above the laboratory reporting limit, blue font indicates result reported above the MTCA Method A CUL.
- Total TPH-Dx Concentration Above MTCA \bullet Method A CUL
- Total TPH-Dx Concentration Below MTCA \bullet Method A CUL
- Total TPH-Dx Concentration Below
- $\mathbf O$ Laboratory Reporting Limit
- LNAPL Observed, Well Not Sampled

Legend

- <u>Note:</u>
- 1. Locations are approximate.
- 2. MTCA Method A Cleanup level (CUL) for total petroleum hydrocarbons
- (TPH) in groundwater is 500 micrograms per liter (µg/l).
- 3. Total TPH-Dx = Sum of diesel-range and oil-range organics results.
- 4. Monitoring well sample results from August 2019 are shown.
- 5. Bold font indicates result reported above the laboratory reporting limit, blue font indicates result reported above the MTCA Method A CUL.
- Total TPH-Dx Concentration Above MTCA \bullet Method A CUL
- Total TPH-Dx Concentration Below MTCA \bullet Method A CUL
- Total TPH-Dx Concentration Below \bullet
- Laboratory Reporting Limit
- **O** LNAPL Observed, Well Not Sampled

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Groundwater Sampling Results Total TPH-Dx (2019) - Main Area

> 1696120.02 July 2020 **Figure # Figure 13**

5. Bold font indicates result reported above the laboratory reporting limit, blue font indicates result reported above the MTCA Method A CUL.

 \bullet

100 Scale: Feet

1696120.02 October 2021

- $\mathbf O$ Laboratory Reporting Limit
- **O** LNAPL Observed, Well Not Sampled

analyzed with silica gel cleanup. Monitoring well sample resutls from August

2018 are shown. 5. Bold font indicates result reported above the laboratory reporting limit, blue font indicates result reported above the MTCA Method A CUL.

Figure 15

 $\frac{N}{N}$ N 0 50 100 Scale: Feet

Legend

BNSF Wishram Railyard Wishram, Washington

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- <u>Note:</u> 1. Locations are approximate.
- 2. MTCA Method A Cleanup level (CUL) for total petroleum hydrocarbons
- (TPH) in groundwater is 500 micrograms per liter (µg/l).
- 3. Total TPH-Dx = Sum of diesel-range and oil-range organics results.
- 4. Monitoring well sample results from August 2019 are shown.
- 5. Bold font indicates result reported above the laboratory reporting limit, blue font indicates result reported above the MTCA Method A CUL.
- Total TPH-Dx Concentration Above MTCA \bullet Method A CUL
- Total TPH-Dx Concentration Below MTCA \bullet Method A CUL
- Total TPH-Dx Concentration Below \bullet
- Laboratory Reporting Limit
- **O** LNAPL Observed, Well Not Sampled

Groundwater Sampling Results Total TPH-Dx (2019) - East Area

> 1696120.02 August 2020 **Figure # Figure 16**

-
- <u>Note:</u>
- 1. Locations are approximate.
- 2. GRO = Gasoline-Range Organics. 3. Soil sample results from 2004 - 2018 are shown.
- 4. Only locations analyzed for GRO are shown.
- 5. Samples from areas later excavated are not shown.

Legend

- GRO Concentration Above MTCA Method A
- **O** GRO Concentration Below MTCA Method A
- \mathbf{o} GRO Concentration Below Laboratory Reporting
- Approximate Previous Excavation Area
- **Approximate Previous Excavation Area**

BNSF Wishram Railyard Wishram, Washington

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2196120*06 December 2021 **Figure 17**

Soil Sampling Results GRO (2004 - 2018) - Main Area

-
- <u>Note:</u>
- 1. Locations are approximate.
- 2. GRO = Gasoline-Range Organics. 3. Most recent monitoring well and reconnaissance groundwater
- sample results from 2004 through 2019 are shown.
- 4. Only locations analyzed for GRO are shown.
- 5. Results from up to eight previous sampling events are shown.

Legend

- GRO Concentration Above MTCA Method A
- GRO Concentration Below MTCA Method A CUL
- **O** GRO Concentration Below Laboratory Reporting Limit

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Scale: Feet

2196120*06 December 2021 **Figure 18**

Groundwater Sampling Results GRO (2004 - 2019) - Main Area

Wishram, Washington

Hydrogeologic Cross Section Transect A to A' and Transect B to B'

-
-

BNSF Wishram Railyard Wishram, Washington

Hydrogeologic Cross Section Transect F to F'

Kennedy/Jenks Consultants

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Level Pre-1956 - Approximate Water

Notes:

- 1. All locations and depths are approximate. Lithologic interpretation inferred from boring logs, historical documents, maps, and/or field observations. 2. AMSL = above mean sea level.
-
- 3. MTCA = Washington Model Toxics Control Act 4. LNAPL = light non-aqueous phase liquid
-
- 5. TPH = total petroleum hydrocarbon
- 6. LIF = laser-induced fluorescence

7. Inferred LNAPL/residual TPH extent from interpretation of LIF waveforms and LNAPL field obervations in soil borings.

Figure 20

Columbia River

Conceptual Site Model

Kennedy/Jenks Consultants

Wishram, Washington

Figure 21

- imate Lateral Extent of Smear Zone Diesel **Oil Impacts** imate Lateral Extent of Dissolved Phase Diesel $-$ and/or Oil Impacts Above MTCA Method A CUL
- **P-PI** Approximate Previous Excavation $I - -I$ Areas (2002-2010)
- **O** Soil Boring
- **O** LIF Location
- an a **Shallow Berm Excavation** CSM Cross Section Line

<u>Note:</u> 1. Locations are approximate.

Legend

 $==$ Former Steam Line

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

Conceptual Site Model (CSM) Plan View Map

> 2196120*06 December 2021 **Figure 2 Figure 22**

BNSF Wishram Railyard Wishram, Washington

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Conceptual Site Exposure Model Diagram

Complete exposure pathway.

Incomplete exposure pathway.

Potentially complete exposure pathway or insufficient information.

NA

Exposure pathway considered not applicable for the listed receptors.

5. Precipitation and infiltration through vadose soil to shallow groundwater.

6. Leaching of contaminants in soil to groundwater and downgradient dissolved-phase transport in shallow groundwater.

7. Seepage from groundwater to surface water along bank area adjacent to the Columbia River. This is expected to be minimal as the river is a losing stream approximately 10 months a year.

1. Potential exposure pathways may differ after remediation and/or

redevelopment of the site.

2. Onsite employees performing routine tasks.

3. Onsite construction and/or utility workers performing invasive activities; workers performing environmental investigation or sampling activities. 4. Based on the Terrestrial Ecological Evaluation performed for the site.

2196120.00December 2021

Figure 23

Legend

- \bigoplus Shallow Monitoring Well $---$ Former Bunker Fuel / Oil Pipeline Existing Site Feature ---- Former Oil Drain ⁻⁻⁻1 Former Site Feature \bigoplus Deep Monitoring Well Former Oil Trough \bigoplus Oil Head Monitoring Well Approximate Lateral Extent of Dissolved ET. Phase Diesel and/or Oil Impacts Above $-$ – Former Sewer Line (Potential) **G** Bioventing Injection Well MTCA Method A CUL Stormwater Underdrain \bigoplus Air Sparge (AS) Well Inferred Lateral Extent of Submerged Diesel and/or Oil Impacts
- **Shallow Berm Excavation**
- \bigcirc Proposed LNAPL Removal Well
- Ω Shallow Biosparge Well: 15 foot ROI Paired Shallow and Deep Biosparge \overline{O} Well: 15 foot ROI
- **C** LNAPL Removal Well: 20 foot ROI

Notes:

and Targeted Excavation **Alternative 1: Physical LNAPL Removal, Focused Biosparge, MNA,**

Inferred Lateral Extent of Smear Zone

Diesel and/or Oil Impacts

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

Figure 24 Figure 24

1. Locations are approximate. 2. MNA = monitored natural attenuation; LNAPL = light non-aqueous phase liquid.

-
- Stormwater Underdrain
-

 \bigoplus Shallow Monitoring Well **Former Bunker Fuel / Oil Pipeline** Existing Site Feature Sheet Pile Wall Notes: ---- Former Oil Drain F⁻⁻1 Former Site Feature **Shallow Berm Excavation** 1. Locations are approximate. **C** Deep Monitoring Well 2. MNA = monitored natural attenuation Former Oil Trough Shallow Biosparge Well: 15 foot ROI \bullet \bigoplus Oil Head Monitoring Well Approximate Lateral Extent of Dissolved CT. Phase Diesel and/or Oil Impacts Above $-$ – Former Sewer Line (Potential) **G** Bioventing Injection Well MTCA Level A CUL Stormwater Underdrain Inferred Lateral Extent of Submerged \bigoplus Air Sparge (AS) Well Stormwater Underdrain Diesel and/or Oil Impacts

Legend

 $\frac{N}{2}$ N 0 50 100 Scale: Feet

MNA, and Targeted Excavation Alternative 3: LNAPL Containment, Biosparge,

Inferred Lateral Extent of Smear Zone

Diesel and/or Oil Impacts

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

Figure 23 Figure 25

Legend

 $\frac{N}{2}$ N 0 50 100 Scale: Feet

MNA, and Targeted Excavation **Alternative 4: Physical LNAPL Removal, Biosparge,**

 \bigoplus Shallow Monitoring Well \bigoplus Deep Monitoring Well \bigoplus Oil Head Monitoring Well Bioventing Injection Well Air Sparge (AS) Well $---$ Former Bunker Fuel / Oil Pipeline ---- Former Oil Drain Former Oil Trough $-$ – Former Sewer Line (Potential) Stormwater Underdrain Stormwater Underdrain **Existing Site Feature Former Site Feature** Approximate Lateral Extent of Dissolved Phase Diesel and/or Oil Impacts Above MTCA Method A CUL Inferred Lateral Extent of Submerged Diesel and/or Oil Impacts Inferred Lateral Extent of Smear Zone **Shallow Berm Excavation** \circledR Proposed LNAPL Removal Well Shallow Biosparge Well: 15 foot ROI Paired Shallow and Deep Biosparge Well: 15 foot ROI **C** LNAPL Removal Well: 20 foot ROI Notes: 1. Locations are approximate. 2. MNA = monitored natural attenuation; LNAPL = light non-aqueous phase liquid.

Diesel and/or Oil Impacts

BNSF Wishram Railyard Wishram, Washington

Figure 24 Figure 26

- \bigoplus Shallow Monitoring Well $---$ Former Bunker Fuel / Oil Pipeline Existing Site Feature ---- Former Oil Drain **Former Site Feature** \bigoplus Deep Monitoring Well Former Oil Trough Approximate Lateral Extent of Dissolved \bigoplus Oil Head Monitoring Well CT. Phase Diesel and/or Oil Impacts Above $-$ – Former Sewer Line (Potential) **G** Bioventing Injection Well MTCA Method A CUL Stormwater Underdrain \bigoplus Air Sparge (AS) Well Inferred Lateral Extent of Submerged Stormwater Underdrain
	- Diesel and/or Oil Impacts Inferred Lateral Extent of Smear Zone
	- Diesel and/or Oil Impacts
- **Shallow Berm Excavation**
- \bigcirc Proposed LNAPL Removal Well
- Shallow Biosparge Well: 15 foot ROI \bullet Paired Shallow and Deep Biosparge \overline{O} Well: 15 foot ROI
- LNAPL Removal Well: 20 foot ROI

Notes:

Legend

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

1. Locations are approximate. 2. MNA = monitored natural attenuation; LNAPL = light non-aqueous phase liquid.

$\frac{N}{2}$ N 0 50 100 **Scale: Feet December 27**

Alternative 5: Low-Temperature Thermal Removal (LTTR), Biosparge, MNA and Targeted Excavation

Appendix A

Summary of Bioventing and NSZD Testing Field Work and Hydraulic Gradient Monitoring

Summary of Bioventing and NSZD Testing Field Work

Appendix A

Page 1

Summary of Bioventing and NSZD Testing Field Work BNSF Wishram Railyard

Kennedy/Jenks Consultants, Inc. (KJ) has prepared this summary to document the results from bioventing and natural source zone depletion (NSZD) testing field work conducted in 2019 at the BNSF Wishram Railyard (site) located in Wishram, Washington. The railyard and historical site features are shown on Figure A1.

A1. Background

On 18 July 2019, Washington State Department of Ecology (Ecology) granted conditional approval of the *LNAPL Transmissivity, Bioventing Respirometry and NSZD Testing Work Plan* (Work Plan) submitted by BNSF. The final Work Plan was submitted to Ecology on 30 September 2019. Field work associated with the Work Plan was substantially conducted between July 2019 and September 2019.

The objectives of the field activities in the Work Plan were to 1) evaluate transmissivity of light non-aqueous phase liquid (LNAPL), 2) assess the performance of the existing bioventing system operating in the vicinity of the Maintenance Shop, 3) assess potential for bioventing in the vicinity of the submerged LNAPL south of the mainline, and 4) evaluate occurrence of NSZD. The procedures and results from the LNAPL transmissivity baildown tests were summarized in the *Uplands Remedial Investigation Report* (Uplands RI Report, KJ 2020). The results from the remaining activities are presented below.

A2. Bioventing Respirometry and NSZD Evaluation

Bioventing and NSZD evaluations were performed using several approaches to demonstrate and quantify biological degradation of hydrocarbons in the vadose and smear zones using multiple lines of evidence.

- A respirometry test in the vicinity of the existing bioventing system near the Maintenance Shop (north of the mainline tracks) was performed to compare conditions during and following operation and to evaluate performance.
- Soil gas measurements were collected from test and monitoring wells at multiple locations across the site, as shown on Figure A2, to evaluate current conditions.
- A bioventing injection test was performed south of the mainline tracks, near the submerged LNAPL (Submerged LNAPL area) (Figure A2). Test activities included 1) baseline soil gas measurements from select monitoring wells, 2) an air injection test to approximate system ROI, and 3) a respirometry test (following the injection test).
- Carbon traps were deployed at locations shown on Figure A2 and analyzed to evaluate NSZD using $CO₂$ flux related to hydrocarbon degradation.

Page 2

Testing methods and results from each evaluation element are presented below followed by overall conclusions based on the combined lines of evidence.

A3. Soil Gas Measurement Methods

A common component of the bioventing and NSZD evaluations was the collection of soil gas measurements from test and monitoring wells. Measurements included oxygen $(O₂)$, carbon dioxide $(CO₂)$, hydrogen sulfide, methane $(CH₄)$, and volatile organic compounds (VOCs). An RKI Eagle 2 multi-gas meter was used for soil gas measurements. Soil gas VOC concentrations were measured using a photoionization detector (PID). Monitoring wells were fitted with vapor monitoring well plugs or modified well caps with barbed fittings at least 1 day prior to soil gas measurements. Measurements were collected using the low purge volume well head method (Sweeney and Ririe 2017).

A4. Existing Bioventing System – Respirometry Test

A bioventing system operated between June 2012 and July 2019 to address residual petroleum hydrocarbons in soil north of the mainline tracks near the Maintenance Shop. The bioventing system operated by injecting ambient air into the unsaturated zone through four wells (SVE-12-1 through SVE-12-4) (Figure A2). The bioventing system operated in continuous mode (24 hours a day, 7 days a week) between June 2012 and April 2017, when the system blower failed. The blower was replaced on 28 November 2017, and continuous operation of the bioventing system was restarted. System operational data collected between November 2017 and May 2019 were summarized in the Work Plan, including an estimated radius of influence (ROI) of the system of 90 feet, based on measurements of induced pressures in wells WMW-7 and WMW-8, near injection well SVE-12-1 (see Figure A3).

The bioventing system was shut down on 24 July 2019 at 9:00 AM in preparation for the respirometry test. Immediately prior to system shutdown, soil gas (oxygen, carbon dioxide, hydrogen sulfide, and methane) and VOCs measurements were collected from injection wells SVE-12-1 through SVE-12-4 and monitoring wells WMW-7, WMW-8, and WMW-12. Wells SVE-12-1 through SVE-12-4 and monitoring wells WMW-7 and WMW-8 are located within the system's estimated ROI of 90 feet. Monitoring well WMW-12 is not located within the ROI but was included to provide background data for comparison. Soil gases were measured using the meters and methods described above. Soil gas measurements were collected at increasing time intervals following system shutdown through 29 July 2019 [approximately 122 hours (5 days) after shutdown] with an additional round of data collected on 8 August 2019 [approximately 369 hours (15 days) after shutdown]. Soil gas measurements are included in Table A1.

Respirometry test results were evaluated to estimate biodegradation rates based on oxygen utilization rates as described in Leeson and Hinchee (1996). Soil gas measurements from injection wells SVE-12-1 through SVE-12-4 and monitoring wells WMW-7, WMW-8, and WMW-12 indicated that oxygen levels remained high and carbon dioxide levels remained low for over 2 weeks after the system was shut off. Table A1 includes soil gas measurements from, well WMW-3 (located south of the mainline tracks outside of the pressure ROI), as a decrease in oxygen levels was observed between 26 July and 8 August.

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Biodegradation rates (K_b) were calculated based on oxygen utilization rates (K_o) provided in Table A1. K_b values were calculated using the following equation (Leeson and Hinchee1996):

 $K_b = (-K_0 / 100)^* (θ_a * ρO_2 * C / ρ_k)$

Wherein:

 θ_a = air-filled porosity = 0.32 for sand, assuming an effective porosity of 0.37 and residual water content of 0.05.

 $\rho O₂= O₂$ density = 1,365 milligrams per liter (mg/L), assuming standard temperature and pressure.

 $C =$ hydrocarbon to O_2 ratio = 0.2888, assuming a hydrocarbon molecular weight (MW) based on the hexadecane equivalent (MW = 226) (Leeson and Hinchee 1996). ρ_k = soil dry bulk density = 1.6 grams per cubic centimeter (g/cm³), sand.

The calculated average biodegradation rate was 0.065 mg hexadecane-equivalent/kg/day, and the calculated average oxygen utilization rate was 0.082 percent per day (%/day) (Table A1). According to the U.S. Environmental Protection Agency (EPA) (1995) bioventing design guidance documents, oxygen utilization rates greater than 1.0 %/day indicate bioventing may be feasible at a given site. Based on the respirometry test results, with an average calculated oxygen utilization rate more than an order of magnitude less than 1.0 %/day, the subsurface environment is sufficiently oxygenated for aerobic biodegradation to occur, and bioventing is no longer necessary in this area. The bioventing system remained off following the respirometry testing in July 2019.

A5. Sitewide Soil Gas Results

Soil gas conditions were measured in the Maintenance Shop, Submerged LNAPL, and former Engine House/Machine Shop areas without the influence of the existing bioventing system (shutdown 15 days earlier) on 8 August 2019 and are presented in Table A2 and on Figure A4. Measurements included oxygen, carbon dioxide, hydrogen sulfide, methane and VOCs using the methods described above.

Oxygen concentrations were generally between 19 and 20.9% oxygen, except for well WMW-3, which contained 18.1% oxygen. Carbon dioxide concentrations were generally less than 1%, other than wells WMW-3 and WMW-9, which contained 2.9% and 1.5% carbon dioxide, respectively. Methane and hydrogen sulfide concentrations were 0% of the lower explosive limit (LEL) and 0 parts per million (ppm), respectively, in the wells measured. VOC concentrations ranged from 0.10 ppm (well WMW-1) to 1.90 ppm (well WMW-29). Soil gas data indicate that the vadose zone contains oxygen at near-atmospheric concentrations, and limited or no carbon dioxide. These results further indicate that bioventing is no longer necessary in the Maintenance Shop area, as oxygen concentrations remained high (approximately 20%) 15 days after system shutdown, and that bioventing is not necessary to increase vadose zone oxygen concentrations in other site areas.

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A6. Submerged LNAPL Area Bioventing Injection and Respirometry **Tests**

A bioventing injection test was performed south of the mainline tracks, near submerged LNAPL areas (Figure A2) in August 2019. Test activities included 1) an initial soil gas survey of select monitoring wells, 2) an air injection test to approximate system ROI, and 3) a respirometry test (following the injection test).

Monitoring well WMW-11 was used as the bioventing injection test well due to the availability of open screen interval above the groundwater table (approximately 4 feet), proximity to inferred LNAPL and residual total petroleum hydrocarbon (TPH) impacts, presence of dissolved phase TPH concentrations above MTCA Method A CULs, proximity to other monitoring wells, and its location in a limited traffic and easy access area.

An air injection test was performed at well WMW-11 to assess air injection into the formation and the optimum injection flow rate for the multi-day injection test. A regenerative blower was used to apply four different injection flow rates for approximately 1 hour each: approximately 16 standard cubic feet per minute (scfm), 20 scfm, 31 scfm, and 35 scfm. Induced pressure responses were measured in wells located in the vicinity of WMW-11: WMW-1, WMW-3, WMW-10, WMW-14, WMW-15, WMW-16, and WMW-17. Pressure response was measured in the wells identified above at approximately 30-minute intervals during the step tests with a differential pressure gage.

After the completion of step testing, a 30-hour continuous injection test was conducted at well WMW-11 using an injection rate of approximately 35 scfm as identified during the step testing. A portable generator was used to supply power to the injection blower during the test. The injection test was shut down on 28 August 2019 and soil gas measurements were collected at increasing time intervals through 31 August 2019 [approximately 64 hours (2.6 days) after shutdown].

A.6.1. Injection Test Results

Pressure responses greater than 0.1 inches of water were measured in wells WMW-11 (injection well) and well WMW-15 during the injection step testing. Pressure responses greater than 0.001 inches of water, but less than 0.1 inches of water were observed in wells WMW-10, WMW-14, WMW-16, and WMW-17 during the injection tests. The calculated ROI from well WMW-11 ranged from 38 feet during the 20 cfm step test to 48 feet during the 31 cfm step test. During the 30-hour step test at 35 cfm, a ROI of 47 feet was calculated (Figure A4). Results from the injection step tests are presented in Table A3, and a summary of the injection test results is presented in Table A5.

A.6.2. Respirometry Test Results

In the submerged LNAPL area, baseline soil gas oxygen concentrations (measured prior to the air injection test) were high (19 to 20% oxygen) and carbon dioxide concentrations were low (less than 1%). Respirometry test results did not show a significant change in soil gas concentrations in the wells monitored (see Table A3). Wells WMW-10, WMW-11, WMW-14, and

Page 5

WMW-16 displayed a slight increase in oxygen concentrations and a slight decrease in carbon dioxide concentrations following the bioventing injection test.

The calculated average biodegradation rate was 0.05 mg hexadecane-equivalent/kg/day, and the average calculated oxygen utilization rate was 0.06 %/day as shown in Table A4. The results from the baseline soil gas measurement and bioventing and respirometry testing, with a calculated average oxygen utilization rate significantly lower than the 1.0 %/day criteria, indicate bioventing does not enhance biological degradation in the tested area.

A.6.3. Bioventing Tests Summary

Results from the respirometry test performed near the existing bioventing system in the Maintenance Shop area and the injection and respirometry test performed in the vicinity of the submerged LNAPL area are summarized in Table A5. The calculated injection ROIs in both site areas (approximately 90 feet in the Maintenance Shop area and approximately 47 feet in the submerged LNAPL area) indicate the vadose zone lithology is suitable for bioventing. However, baseline soil gas measurements of oxygen at near-atmospheric concentrations and calculated oxygen utilization rates in both areas less than 1.0 %/day criteria (EPA 1995), indicate bioventing is not necessary to increase vadose zone oxygen and is not an appropriate remedial technology for vadose zone conditions or submerged and residual LNAPL at the site.

A7. Carbon Traps NSZD Test

At sites impacted with petroleum hydrocarbons, LNAPL losses can occur through natural biodegradation processes such as methanogenesis (Amos et al. 2005), in which $CO₂$ and CH₄ are generated by an anaerobic process during natural degradation of organic materials such as petroleum hydrocarbons. As both these gases are transported from the LNAPL source toward the atmosphere, CH_4 encounters atmospheric O_2 and, through aerobic processes, generates $CO₂$ in the shallow surface soil layer and the vadose zone. Field screening tools have been developed to quantify biodegradation rates and estimate NSZD over time.

 $CO₂$ flux was measured by trapping $CO₂$ and storing it for laboratory analysis, using a technology developed at Colorado State University Center for Contaminant Hydrology (CCH), and now commercialized by E-Flux, LLC (E-Flux). The E-Flux method measures carbon (to estimate total $CO₂$ flux) and carbon isotopes (i.e., ¹⁴C) to estimate the contribution of petroleum hydrocarbon degradation to the total carbon flux from the soil to the ground surface (Zimbron et al. 2011). This process was used to assess and estimate natural LNAPL losses from biodegradation (i.e., NSZD). Research performed by CCH shows LNAPL losses on the order of thousands of gallons per acre per year at petroleum-impacted sites can be identified by measuring CO₂ flux. See the attached Standard Operating Procedure (SOP) for additional information about the $CO₂$ trap technology (Attachment A).

A.7.1. Carbon Trap Results

Carbon traps were deployed in 11 locations in the Maintenance Shop, Submerged LNAPL, and former Engine House/ Machine Shop areas on 12 and 13 August 2019 and retrieved on

Page 6

21 August 2019 (Figure A2). Each trap consisted of two sorbent elements: the bottom captured $CO₂$ from surface soil and the top intercepted atmospheric $CO₂$. CO₂ passing through the trap reacts to form carbonates. The sorbent elements were analyzed by the laboratory for total carbonate mass, and $CO₂$ flux was estimated by dividing the carbonate mass by the crosssectional area on the trap, taking into account the time the trap was deployed in the field. The estimated $CO₂$ flux was then converted to an estimated LNAPL biodegradation rate using the stoichiometric ratio between $CO₂$ and LNAPL. (Note: LNAPL biodegradation rates were corrected to subtract out background rates of $CO₂$ quantified by background sampling in a travel blank (TB) sample and/or 14C analysis).

The carbon trap results are summarized on Figure A5 and in further detail in the E-Flux summary report, included in Attachment A. Detections of fossil fuel $CO₂$ in the carbon traps from locations 1 to 10 ranged from 0.02 grams (g) at location 8 (on the berm near well WMW-16) to 3.51 g at location 4 (located in the western LNAPL area). Fossil fuel $CO₂$ flux ranged from 0.07 micromole per square meter per second (μ mol m⁻² s⁻¹) at location 8 to 12.47 μ mol m⁻² s⁻¹ at location 4. Calculated equivalent NSZD rates, based on a flux equivalence of 492.7 gallons/acre/year assuming an LNAPL density of 0.966 grams/milliliter (g/mL) and a 4-inch receiver pipe, ranged from 35 gallons/acre/year at location 8 (above dissolved phase petroleum impacts) to 6,146 gallons/acre/year at location 4 (above smear zone and submerged LNAPL).

Fossil fuel CO2 was not detected (ND) at location 11 (located adjacent to well WMW-26 in the former Engine House). The results in the E-Flux report show while the total $CO₂$ content at location 11 (16.85%) and total $CO₂$ flux (27.82 µmol m⁻² s⁻¹) were high compared to other samples, the $CO₂$ content was 98.3 % modern carbon (i.e., biological degradation of new sources of carbon, possibly from higher organic content in the backfill used in this area), and 1.7% old carbon (i.e., from biodegradation of low concentrations of dissolved petroleum hydrocarbons). The fossil fuel $CO₂$ results for location 11 were less than the estimated background (travel blank) sample, therefore the result was reported as ND.

As shown on Figure A5, the lower calculated equivalent NSZD rates (between ND and 147 gallons/acre/year) were measured in areas with dissolved phase petroleum impacts only (locations 2, 8, 9, and 11) or no dissolved phase or LNAPL impacts (locations 1 and 3). The higher calculated equivalent NSZD rates (between 364 and 6,146 gallons/acre/year) were measured in areas near or above the inferred extents of smear zone and/or submerged LNAPL (locations 4, 5, 6, 7, and 10). The carbon trap results provide evidence of biological activity (production of $CO₂$) from both petroleum hydrocarbon and natural sources; and show that biodegradation of petroleum hydrocarbons (i.e., NSZD) is occurring at measurable rates in areas where significant petroleum hydrocarbon mass exists in the subsurface.

A8. Conclusions

Soil gas measurements and bioventing testing results performed in the Maintenance Shop, Submerged LNAPL, and former Engine House/Machine Shop areas indicate the vadose zone across the site contains elevated oxygen (near-atmospheric concentrations) and low concentrations of carbon dioxide (generally less than 1%). Therefore, delivering additional oxygen using bioventing or soil vapor extraction is not expected to increase or enhance aerobic biological degradation of residual petroleum hydrocarbons in the vadose and smear zones. As a

Page 7

result, interim remedial measure bioventing operations were discontinued at the site. The results of the carbon trap evaluation indicated NSZD is occurring, with the highest responses in test locations near/above the inferred extents of smear zone and submerged LNAPL.

Multiple lines of evidence indicate aerobic biological degradation is currently effectively remediating the remaining residual petroleum hydrocarbons in vadose and smear zone soils without the active addition of oxygen via bioventing.

Enclosures:

- Table A1: Maintenance Shop Existing Bioventing System Respirometry Results
- Table A2: Baseline Soil Gas Data
- Table A3: Injection Step Test Results
- Table A4: Injection Test Respirometry Results
- Table A5: Summary of Bioventing System Data
- Figure A1: Current and Historical Site Features
- Figure A2: Bioventing and NSZD Test Locations
- Figure A3: Bioventing System Data
- Figure A4: Sitewide Soil Gas Concentrations (8 August 2019)
- Figure A5: Carbon Trap NSZD Estimates

Attachment A: E-Flux CO2 Flux and NSZD Rate Results Attachment B: Field Forms and Supplemental Data

References

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Tables

TABLE A1

MAINTENANCE SHOP EXISTING BIOVENTING SYSTEM RESPIROMETRY RESULTS BNSF Wishram Railyard

SYSTEM SHUT DOWN

> **Duration** System Off

Depth to Water

Ko - Oxygen Utilization Rate Kb - Biodegradation Rate (mg Hexadecane Equivalent/kg/day)

SVE-12-1

Injection Well #N/A

Well

April 2021 2096120.00

TABLE A1

MAINTENANCE SHOP EXISTING BIOVENTING SYSTEM RESPIROMETRY RESULTS BNSF Wishram Railyard

April 2021 2096120.00

MAINTENANCE SHOP EXISTING BIOVENTING SYSTEM RESPIROMETRY RESULTS BNSF Wishram Railyard

CO2 % and the contract of the

 $-$ = not measured % $CO₂$ = percent carbon dioxide mg/L = milligrams per liter g/cc = grams per cubic centimeter $\%$ /day = percent per day g/cc = grams per cubic centimeter %/day = percent per day ft = feet mg = milligram hrs = hours kg/day = kilograms per day % O_2 = percent oxygen $NA = not applicable$ STP = standard temperature and pressure

WMW-3: 168 feet from injection well SVE-12-1. Outside of Pressure ROI - Response in O2 measurements Screen Interval: 10 - 20 feet bgs

Notes:

(a) Literature values assumed for porosity and dry bulk density.

Tan shading indicates background data collected prior to study step testing.

Blue shading indicates data recorded prior to shut down of blower. Green shading indicates data used for estimating biodegradation.

TABLE A2 BASELINE SOIL GAS DATA BNSF Wishram Railyard

Notes:

 $O₂$ = oxygen

 $CO₂$ = carbon dioxide

 CH_4 = methane

 $% =$ percent

LEL = lower explosive limit

 H_2S = hydrogen sulfide

VOC = volatile organic compound

ppm = parts per million

TABLE A3

INJECTION STEP TEST RESULTS BNSF Wishram Railyard

INJECTION TEST RESPIROMETRY RESULTS BNSF Wishram Railyard

SYSTEM SHUT DOWN

8/28/19 16:13

Oxygen Utilization Rate < 1 %/day Baseline O2 > 20% in most wells

April 2021 2096120.00

TABLE A4

INJECTION TEST RESPIROMETRY RESULTS BNSF Wishram Railyard

INJECTION TEST RESPIROMETRY RESULTS BNSF Wishram Railyard

Notes:

(a) Literature values assumed for porosity and dry bulk density.

Tan shading indicates background data collected prior to study step testing.

Blue shading indicates data recorded prior to shut down of blower.

Green shading indicates data used for estimating biodegradation.

Grey shading indicates data not used in respirometry estimates.

TABLE A5

SUMMARY OF BIOVENTING SYSTEM DATA BNSF Wishram Railyard

Page 1 of 1

Existing Bioventing System Summary

Bioventing Injection Test Summary

Wells WMW-7 and WMW-8 frequently contain sheen in purge water, but no measurable LNAPL > 0.01 feet.

WMW-3 elevated dissolved phase DRO and ORO, decline in % O_2 observed.

Figures

Legend

- $\frac{1}{2}$ Approximate BNSF Property Line $\frac{1}{2}$ -
- Existing Site Feature
- **Former Site Feature**
- \blacksquare Former Bunker Fuel / Oil Pipeline \blacksquare --- Former Oil Drain
-
- Former Oil Trough
- Former Sewer Line (Potential) Stormwater Underdrain (A portion \mathcal{L}
- removed from service circa 1960) Stormwater Underdrain (Rerouted
- portion circa 1960) Former Diesel Line
- $==$ Former Steam Line

BNSF Wishram Railyard Wishram, Washington

Kennedy/Jenks Consultants

1896120.07 June 2019 **Figure A1**

Current and Historical Site Features

Legend

Kennedy/Jenks Consultants

EXTREMENTED BINSF Wishram, Washington

- Inferred Lateral Extent of Smear Zone Diesel and/or Oil Impacts
- Inferred Lateral Extent of Submerged Diesel and/or Oil Impacts
- Approximate Lateral Extent of Dissolved Phase Diesel and/or Oil
- \bigoplus Shallow Monitoring Well
- **Deep Monitoring Well**
- \bigoplus Oil Head Monitoring Well
- \bigoplus Air Sparge Well
- Soil Vapor Extraction Well
-
- \diamond WMW-16 Well included in bioventing/NSZD testing
- Q WMW-7 Well included in respirometry study

1996120*07 April 2021 **Figure A2**

Bioventing and NSZD Test Locations

1. Locations are approximate.

Approximate Bioventing Radius of Influence Inferred Lateral Extent of Smear Zone Diesel

---- Approximate BNSF Property Line

<u>Notes:</u>

Legend

Shallow Monitoring Well

- \bigoplus Deep Monitoring Well
- \bigoplus Oil Head Monitoring Well
- \bigoplus Air Sparge Well
- Soil Vapor Extraction Well
- **Inferred Lateral Extent of Submerged Diesel** $\frac{1}{1}$ - - - and/or Oil Impacts **P** - Approximate Lateral Extent of

and/or Oil Impacts

Dissolved Phase Diesel and/or Oil

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

1996120.00 April 2021 **Figure A3**

Bioventing Systems Data

- 1. Locations are approximate.
- 2. LNAPL = light non-aqueous phase liquid
- 3. TPH = total petroleum hydrocarbons
- 4. Ko = oxygen utilization rate
- 5. Inferred lateral extent of Diesel- or Oil-Like LNAPL based on interpretation of LIF waveforms (July 2013) and soil boring logs.
- 5. Existing bioventing system measurements collected on 24 July 2019 prior to system shutdown.

Bioventing test measurements collected on 28 August 2019 after approximately 24 hours of system operation.

Legend

<u>Notes:</u>

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

1996120.00 December 2020

Sitewide Soil Gas Concentrations (8 August 2019)

- 1. Locations are approximate.
- 2. LNAPL = light non-aqueous phase liquid
- 3. TPH = total petroleum hydrocarbons
- 4. Inferred lateral extent of Diesel- or Oil-Like LNAPL based on interpretation of LIF waveforms (July 2013) and soil boring logs.
- 5. All soil gas measurements collected on August 8, 2019. Bioventing system shutdown on July 24, 2019.

Legend

<u>Notes:</u>

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

1996120.00 April 2021 **Figure A5**

Carbon Trap NSZD Estimates

- 1. Locations are approximate.
- 2. LNAPL = light non-aqueous phase liquid
- 3. TPH = total petroleum hydrocarbons
- 4. NSZD = natural source-zone depletion
- 5. ND = not detected
- 6. Inferred lateral extent of Diesel- or Oil-Like LNAPL based on interpretation of LIF waveforms (July 2013) and soil boring logs.
- 7. NSZD estimates based on carbon trap site-specific flux results.

Attachment A

E-Flux CO2 Flux and NSZD Rate Results

Confidential Report CO² Flux and NSZD Rate Results

> RYAN HULTGREN KENNEDY JENKS PROJECT: WISHRAM, WA SAMPLING DATES: 8/12/2019 - 8/21/2019

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> Report Date: 9/23/2019 © 2019 All Rights Reserved.

The purpose of this document is to provide sample calculations for the reported results and to explain the method for differentiating petroleum hydrocarbon-derived CO² from that produced from natural soil respiration processes. The value of the ¹⁴C analysis, site-specific study results and applicable notes, calculation explanations, and references are included.

The Value of the ¹⁴C Analysis

How to differentiate between petroleum hydrocarbon-derived CO² and natural process-derived CO² using CO² flux traps:

Unimpacted soils naturally produce $CO₂$ due to microbial root zone activity and/or the degradation of natural organic matter. Thus, the total measured $CO₂$ flux at an impacted location is a function of the rates of both natural soil respiration and LNAPL degradation (Sihota and Mayer, 2012). The latter, which is caused by Natural Source Zone Depletion (NSZD), can be estimated by subtracting measured $CO₂$ fluxes at unimpacted locations from the total measured $CO₂$ fluxes at LNAPL-impacted locations (Sihota and Mayer, 2012). This spatial "background correction" assumes that bio-based $CO₂$ fluxes are similar at both impacted and unimpacted locations. This approach is complicated to implement, given that at many industrial facilities it is difficult to find unimpacted areas and vegetation cover can vary across a site. Alternatively, carbon isotope analysis can be used to carry out a location-specific correction for total measured $CO₂$ fluxes, and this approach effectively overcomes the limitations of the background correction.

Theory of Carbon Isotope Analysis:

Our method for NSZD rate estimation relies on the analysis of ¹⁴C, an unstable carbon isotope with an absolute half-life of 5,730 years. ¹⁴C is generated by cosmic rays in the atmosphere and is quickly oxidized to ¹⁴CO₂; thus, bio-based living carbon is ¹⁴C-rich, while ancient fossil fuel carbon is completely ¹⁴C-depleted. Additionally, bio-based organic carbon and the atmosphere have the same characteristic amount of ¹⁴C. The short half-life of ¹⁴C only allows for dating of samples younger than 60,000 years using accelerator mass spectrometry (Stuiver and Polach, 1977). ¹⁴C analysis can therefore be used to differentiate between anthropogenic (i.e., fossil fuel) and natural sources of atmospheric carbon (see Klouda and Connolly, 1995; Levin et al., 1995; Avery et al., 2006), and this analysis is the basis for ASTM D6866-18.

For samples that contain both bio-based and fossil fuel-derived carbon, such as E-Flux's fossil fuel traps, measurement of ¹⁴C enables quantitation of *both* source contributions. The fossil fuel-derived percentage of the sample (ff*sample*) and the bio-based percentage (*1-ffsample,* or *bbsample*) are related by the following two-component mass balance (modified from Avery, Jr. et al., 2006):

$$
Fm_{sample} = (ff_{sample})(Fm_{ff}) + (1 - ff_{sample})(Fm_{atm})
$$

Here, *Fm_x* represents the fraction modern, a measure of how close the present ¹⁴C/¹²C ratio of the sample is to the ratio from 1950, which is derived from a pre-industrial era standard. *Fmsample* is the total measured fraction modern of the sample. *Fm*^{*f*} is the fraction modern of only the fossil fuel portion of the sample. This number is 0, as there is no ¹⁴C in fossil fuelderived CO2. *Fmatm* is the fraction modern of the part of the sample derived from natural soil respiration processes. This value, currently equal to **1.02** (Cerling et al., 2016, Larsen et al., 2018), has been experimentally determined and is a fixed value at each point in time. By convention, the results of carbon isotope analysis are reported based on a 1950 NBS oxalic acid standard, and so *Fmsample* is reported as if the analysis took place in 1950. Due to nuclear testing, current ¹⁴C atmospheric levels are now higher than they were in 1950. This means that *Fmatm* is counter-intuitively larger than 1, as the ¹⁴C/¹²C sample ratio is higher now than it would have been in 1950.

¹⁴C Calculations:

Conversion of Fraction Modern Carbon to Fossil Fuel Carbon:

The equation for calculating the percentage of fossil fuel carbon (*ffsample*) is derived from the following mass balance:

$$
Fm_{sample} = (ff_{sample})(Fm_{ff}) + (1 - ff_{sample})(Fm_{atm})
$$

Solving for *ffsample* yields:

$$
ff_{sample} = 1 - \frac{F m_{sample}}{F m_{atm}}
$$

Fraction modern (*Fmsample*, from ¹⁴C analysis) is reported by convention based on ¹⁴C levels from 1950. Because of atomic testing, current environmental ¹⁴C levels are approximately 2% higher than they were in 1950 (Cerling et al., 2016, Larsen et al., 2018) and *Fmatm* is equal to 1.02. This equation then becomes:

$$
ff_{sample} = 1 - \frac{F m_{sample}}{1.02}
$$

As percentages must add to 1, the percentage of bio-based carbon (*bbsample.)* can then be calculated using the following equivalence:

$$
bb_{sample} = 1 - ff_{sample} = 1 - \left(1 - \frac{Fm_{sample}}{1.02}\right) = \frac{Fm_{sample}}{1.02}
$$

Converting Carbon Flux to Equivalent LNAPL Loss Rate:

The intermediate reactions for LNAPL mineralization include methanogenesis, leading to production of methane and $CO₂$, and the subsequent aerobic oxidation of methane into $CO₂$:

$$
C_8H_{18} + 3.5 H_2O \rightarrow 6.25 CH_4 + 1.75 CO_2
$$
 (methanogenesis)

6.25
$$
CH_4 + 12.5 O_2 \rightarrow 6.25 CO_2 + 12.5 H_2O
$$
 (methane oxidation)

Note: Example conversion factor for gasoline LNAPL

 C_8H_{18} +12.5 $O_2 \rightarrow 9H_2O + 8$ CO_2 (overall reaction)

Assuming a conservative LNAPL density of 0.77 g mL⁻¹ (upper range of gasoline) and using the molecular *y*veight of C $_8$ H $_{18}$ (octane, 114.23 g mol $^{\text{-1}}$), µmol m $^{\text{-2}}$ s $^{\text{-1}}$ of CO $_2$ can then be converted into gal. acre $^{\text{-1}}$ yr $^{\text{-1}}$ of LNAPL:

$$
1 \frac{\mu \text{mol } CO_{2}}{\text{m}^{2} \text{ s}} \cdot \left(\frac{1 \mu \text{mol } C_{8}H_{18}}{8 \mu \text{mol } CO_{2}}\right) \left(\frac{1 \text{ mol } C_{8}H_{18}}{1 \times 10^{6} \mu \text{mol } C_{8}H_{18}}\right) \left(\frac{114 \text{ g } C_{8}H_{18}}{1 \text{ mol } C_{8}H_{18}}\right) \left(\frac{1 \text{ mL } C_{8}H_{18}}{0.77 \text{ g } C_{8}H_{18}}\right)
$$

$$
\left(\frac{1 \text{ L}}{1000 \text{mL}}\right) \left(\frac{1 \text{ gal.}}{3.785 \text{ L}}\right) \left(\frac{4.046 \text{ m}^{2}}{1 \text{ acre}}\right) \left(\frac{3600 \text{ s}}{1 \text{ h}}\right) \left(\frac{24 \text{ h}}{1 \text{ d}}\right) \left(\frac{365 \text{ d}}{1 \text{ yr}}\right).
$$

$$
= 625.2 \frac{\text{gal. } C_{8}H_{18}}{\text{acre} \cdot \text{yr}}
$$

Note that both the LNAPL formula and its density are assumed, and so this conversion is subject to uncertainty. However, site-specific data can be used if available. Using alternative representative hydrocarbon formulas and densities generally results in conversion factors that are within 10-15% of 625.2 gal. acre⁻¹ yr^{-1.}. Therefore, the uncertainty associated with these values does not preclude an acceptable estimate.

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Expected Results and Recommendations:

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¹⁴C-based techniques offer a built-in, location-specific correction as an alternative to the standard background location correction. Early work on a limited number of samples suggested that ¹⁴C-corrected results are equivalent to background-corrected results (Sihota and Mayer, 2012; McCoy et al., 2015). However, a more recent comparison spanning 4 different sites suggests that measured carbon fluxes can differ by up to five times among different locations within the same site (Zimbron and Kasyon, 2015). Depending on the location, the resulting difference between background-corrected and ¹⁴C-corrected NSZD rate estimates can be up to one order of magnitude. In contrast, the background correction assumes that the non-fossil fuel CO₂ flux is constant across an entire site; large errors in final estimated NSZD rates might therefore be introduced if the background correction is used. Because the ¹⁴C measurement is co-located with the $CO₂$ flux measurement, it is unbiased by spatial uncertainties related to the background location(s) (e.g., vegetation, lithology, unknown impacts, different gas transport regimes, soil moisture).

The fossil fuel $CO₂$ content of unexposed sorbent as used in the traps is typically around 30% (as of today) and likely results from material processing and handling (e.g., exposure to fossil fuel fumes). This small mass of fossil fuel $CO₂$ is removed from samples by carrying out a ¹⁴C travel blank correction. ¹⁴C analysis is performed on CO₂ sorbent subsamples after homogenization of the entire bottom sorbent layer (see McCoy et al., 2015). The mass of fossil fuel $CO₂$ in the unexposed travel blank trap (TB) is then subtracted from the mass of fossil fuel $CO₂$ in each field-deployed trap.

The results in this report are based on proprietary technology used to measure soil gas efflux. All information contained herein is strictly confidential to the customer.

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Project: Wishram, WA

Customer: Kennedy Jenks

Customer Contact: Ryan Hultgren

Report Date: 23-Sep-2019

• The flux equivalence is 1 µmol m⁻² s⁻¹ = 492.7 gallons acre⁻¹ yr⁻¹, assuming a representative hydrocarbon density of 0.966 g mL⁻¹ with the formula $C_{22}H_{4e}$.Trap cross-sectional area is 8.11 \times 10⁻³ m² (based on a 4-inch receiver pipe).

- Carbonate analysis of each trap/sample is based on method ASTM 4373-14, which does do not provide acceptable variability (CV) standards. Similar methods (e.g., ASTM D513-16) allow typical errors of ≤ 20%. Analysis is therefore conducted in duplicate if the coefficient of variation (CV) of the duplicates is < 5%. If CV ≥ 5%, duplicate analyses are repeated until CV < 5%.
- NA = Not Applicable; ND = Not Detectable.
- **a** Raw and ¹⁴C Results are not TB-corrected.
- **b.** Final CO₂ and NSZD Results are TB-corrected.
- **c.** Refers to the measured weight percentage of CO₂ with respect to the total dry sorbent mass.
- **d.** Refers to the coefficient of variation of CO₂ measurements for each sample: CV = [standard deviation of %CO₂ measurements] / [average %CO₂ measurement]
- **e.** Refers to the reported fraction modern (*Fm_{sample}*). As is standard in radiocarbon reporting, this value has not been corrected to account for present-day ¹⁴C atmospheric levels. This number is originally reported as pMC (percent modern carbon) and is converted into *Fm* for our calculations using the relation 100.0 pMC = 1.0 *Fm* = 100% *Fm*.
- **I** "As of Today" means that the value has been adjusted to account for the difference between atmospheric 14C levels from the 1950s and today (Stenström et al., 2011). *bb_{sample}* is the percentage of the total CO₂ that is derived from bio-based (non-fossil fuel) sources. *ff_{sample}* refers to the percentage of CO₂ that is derived from fossil fuels. The values reported in the 14C Results section are not TB-corrected, but those in the NSZD Results section are.

Results Snapshot:

- **o** The Travel Blank (TB) concentration is **0.95%**; typically, this number is < 2%.
- **o** Trap tops are not saturated with CO2 (sorbent saturation is 30%). The maximum measured (raw) top concentration is **6.06***%* (sample **WIWA-R1-CO2-11 top**).
- **o** Bio-based carbon fluxes represent the CO₂ contributions from natural soil respiration processes to the total carbon flux; the ¹⁴C analysis corrects for this contribution. Average bio-based CO₂ flux is **4.39** µmol m⁻² s⁻¹, and the coefficient of variation is **185%**. The range of bio-based CO₂ fluxes is between **0.75** and **27.85** µmol m⁻² s⁻¹. If these interferences were not removed using the results of the radiocarbon analysis, the errors in the NSZD rate estimates would be between **370** and **13723** gallons acre-1 yr-1 .
- Sample **WIWA-R1-CO2-11** shows non-detectable (ND) fossil fuel CO₂ flux. The entire CO₂ flux for this sample is likely derived from non-fossil fuel sources.

Site-specific Sample Calculations:

Grams of Fossil Fuel CO2:

The mass of fossil fuel-derived $CO₂$ in each trap is calculated by subtracting the total fossil fuel $CO₂$ in the travel blank (TB) from the total fossil fuel CO₂ in the trap. Only data that are **not** TB-corrected (i.e., f_{sample} As of Today and raw % $CO₂$) are used in this calculation. Using Sample 1 as an example:

$$
(g \, \text{CO}_{2(\text{ff})})_{sample\ 1} = g_{\text{sorbent}} \cdot [((\% \, \text{CO}_2)_{\text{sample}}(f f_{\text{sample}})) - ((\% \, \text{CO}_2)_{\text{TB}}(f f_{\text{TE}}))]
$$
\n
$$
(g \, \text{CO}_{2(\text{ff})})_{sample\ 1} = 42.15 \, g \cdot [(6.42 \, \% \cdot 6.38 \, \%) - (0.95 \, \% \cdot 33.79 \, \%)]
$$
\n
$$
(g \, \text{CO}_{2(\text{ff})})_{sample\ 1} = 0.0369 \, g
$$

Here, gsorbent is the mass of sorbent used in the bottom layer of the trap, $%CO₂$)sample is the average weight percentage of CO₂ in the sample, *ff_{sample}* is the percentage of carbon in the sample derived from fossil fuels, (%CO₂)_{TB} is the average weight percentage of CO_2 in the travel blank, and ff_{TB} is the percentage of carbon in the travel blank that is derived from fossil fuels. In this example, Sample 1 contains **0.0369 g** of fossil-fuel derived CO2.

Fossil Fuel CO² Flux:

Converting grams of CO² to CO² flux requires the cross-sectional area of the receiver (**8.11 × 10-3** m² for a **4**-inch receiver), the number of days that the trap was deployed in the field, and the molecular weight of CO₂ (44 g mol⁻¹). Using Site 1 as an example:

Fossil Fuel CO₂ Flux =
$$
\frac{\text{g fossil fuel CO}_2 \cdot \frac{1 \text{ mol CO}_2}{44 \text{ g CO}_2} \cdot \frac{1,000,000 \text{ µmol CO}_2}{1 \text{ mol CO}_2}}{\text{days in the field} \cdot \frac{24 \text{ hr}}{\text{day}} \cdot \frac{3600 \text{ s}}{\text{hr}} \cdot (\text{receiver area})}
$$

$$
\text{Fossil}\ \text{Fuel}\ \text{CO}_2\ \text{Flux}\ = \frac{\mathbf{0.0369}\ \text{g}\ \text{fossil}\ \text{fuel}\ \text{CO}_2\ \cdot\ \frac{1\ \text{mol}\ \text{CO}_2}{44\ \text{g}\ \text{CO}_2}\ \cdot\ \frac{1,000,000\ \text{µmol}\ \text{CO}_2}{\text{mol}\ \text{CO}_2}}{\mathbf{8.32}\ \text{days}\ \cdot\ \frac{24\ \text{hr}}{\text{day}}\ \cdot\ \frac{3600\ \text{s}}{\text{hr}}\ \cdot\ (\mathbf{8.11}\times\mathbf{10^{-3}\ m^2})}
$$

Fossil Fuel CO₂ Flux = **0. 14** $\frac{\mu \text{mol CO}_2}{m^2 \cdot s}$ $m^2 \cdot s$

EFLUX

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CO2 TRAP SHIPMENT AND INSTALLATION LOG LNAPL NATURAL ATTENUATION STUDY

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Attachment B

Field Forms and Supplemental Data

System Parameter Field Form

Project: BNSF Wishram

... stem: Bioventing Date: 7/23/19

 mp

Date: 16711

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CALIBRATION FIELD FORM BNSF WISHRAM

Calibration Field Form
Project: VVISWVOIV Date:

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BIOVENTING WELL. CAD MEASUREMENTS
BNSF WISHRAM

Meter: <u>LXI Cagle 2</u> S/N E2 H491
Screening Equipment: <u>Fini Lae Lite</u>

Purging Equipment: Refer pupi p

Purging Flowrate:

* Tubing installed

BNSF WISHRAM

System Parameter Field Form Project: WISH System: **R.f**

Date:

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July 2019 1996120*07
BIOVENTING WELC EAD MEASUREMENTS **BNSF WISHRAM** 101^o \mathbf{I} \overline{a}

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BIOVENTING WELLHEAD MEASUREMENTS BNSF WISHRAM

Screening Equipment:

Purging Equipment: Purging Flowrate:

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BIOVENTING WELLHEAD MEASUREMENTS BNSF WISHRAM

System Parameter Field Form Project: USSHamm System: **BIDVENTIAL** Date: $\frac{2}{2}$

Meter: RX1 EAGLE 2 5/4 EZH491
Screening Equipment: MINI RAE UTE 5/4590-905052

Purging Equipment: MGTELS

Purging Flowrate:

BIOVENTING WELGEAD MEASUREMENTS

BNSF WISHRAM

7/25 @1010AMea around wMW-8 & WMW-12 V. active, starting exavation near wMW-12

BIOVENTING WELLHEAD MEASUREMENTS

BNSF WISHRAM RAGGL2/PID **Screening Equipment:**

Purging Equipment: NOTETS

Purging Flowrate:

BIOVENTING WELLHEAD MEASUREMENTS BNSF WISHRAM

Meter: RK1 Eagle 2/PID

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Purging Flowrate:

SOIL GAS WELLHEAD MEASUREMENTS

BNSF WISHRAM

Motor: RKI ECRAL 2 (PID **Screening Equipr**

Chaid Co

Purging Equipment:

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BNSF WISHRAM

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SOIL GAS WELLHEAD MEASUREMENTS **BNSF WISHRAM**

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SOIL GAS WELLHEAD MEASUREMENTS

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SOIL GAS WELL HEAD MEASUPEMENTS

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ATTACHMENT B1 SOIL GAS DATA BNSF Wishram Railyard

Notes:

Soil gas measurements collected following installation of vapor well plugs / well caps in areas outside existing bioventing system. August 8, 2019 measurements for baseline soil gas conditions in natural source zone depletion and respirometry evaluations.

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start measurement on UHM-3 1625 start OHM-1 reading 1305 10 33 OHM-1 DTW = 18.25 K+ 1315 OHM-3 $DTW = 14.50 F4$ 1325 $DTP = 9.43 +$ $DTP = 11.42R$ 10 yo Rouning 1520-1330 Rank Inspection, Probe to freeze 1045 no sheep, stopped raining. 1110 Probe in OHM-2 waiting 1535 adding silicone tubing for it to than to unev-10 and which 1140 OHM - 2 OFF 18.74 Pt $(40)(1)$ (1) (1) (2) (1) (2) $($ $\frac{1}{\left(\cos\theta\right)}$ $DTP =$ well lime Press. Of CD1 CH4 H2S Vocs Purge $(4H_{20})$ 13. well Time Pressure of con city 45 120 Pin 20.9 0 0 0 m WHW-11 1649 0 MWO .. -10 1417 0 20.6000 18.32 $WW-W-1111214$ $\overline{}$ 20.9 00 0 0 0.72 16200 20.60.10 0 $-M$ 1.3 $\boldsymbol{2}$ -15 1423 0 90.8 0.200 45 WM11-101216 20.90000 0.62 \circ \mathbf{c} 20.90100 052 20.90000.42 1219 the nelle $-_U$ \circ -151 1221 20.50100 209 0.2 0 0 1.42 -17 11290 \circ $1.3.2$ -161240 14320 20.8000 1.52 20403000.22 -1 20.40.2000.32 -17 226 -3 16350 2090007 1648 GIRG OPFSIte 1229 \overline{O} -1 10.9 0.0 0065 2 -3 1231 \boldsymbol{O} 20.9 0.0 00 0.9 2 1240 Frozen probe in 0HM-3 1247=1152 Bank Inspection from while to WMW-18, no sheen present. Small Daves, with precipiation scale: 1 square 11 10 flve Mer Scale: 1 square = Rete in the Rain

 $408/30/19$ 0710 GIRGI annives on ste den PPE $OHM-1$ 0812 $DTP = 9.37R$ 0715 Turn on, meters. $OPT = 11.69 - 4$ 0820 $OHH - 2$ 0725 RKI Eagle calibrated $UHM-3$ 0825 $DTP = 11.42$ Mini Rat calibrated 08 30-0835 NO sheen in river, (μ_{H20}) (i) (μ_{H20}) (i) (μ_{H20}) (pan) (μ_{H20}) DTW CP4) MMe well Well press Time 0, 02 city H2S vas purge $SVE-1 NN.W.$ 0842 $M/MwH - 0730$ 20.9 0 0 0.6 2 $SUE-2$ 8.83 0852 $-10 - 0133$ n.9 0.1 0 0 $0,9$ $\mathbf{2}$ $SVE-3$ $N.W.$ 0858 $-14 - 07362090.200$ 0.52 $SVE - 4$ N.W 0902 -01399090700 -15 1.1 \mathbf{r} $-16 - 074220.6040$ \mathcal{O} 0.02 Had to remove surface water Apm inside of SVE-2, could $-17 - 0742$ 20.70.200 0.42 -0746000000 0.62 $-3 - 01502000000$ 0.92 0800 SPOKE With RH yesterday (8/29) 1450 GIRGI back on site. Q6:00 and spoke about scope of work 1500 Move drums over OHM for the next two day. take Wells. readings 8 hrs about = on 8130 1515 Prepare meters for readings and one Final reading 8/3/ 1516 Readings table on the Also take DTP measurements back page and, check DTW in SVE wells. 0806 A lot of activity on site 0808 Start measurement of OHM Wells Scale: 1 square = Rete in the Rain Scale. 1 square $=$

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Hydraulic Gradient Monitoring

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Scale: Feet

1996120.00 November 2020 **Figure 1**

-
-
- semiannual collection of samples for cations, anions, arsenic (total and dissolved), and natural attenuation parameters.

BNSF Wishram Railyard Wishram, Washington

Groundwater and Surface Water Elevations - 3/14/2019 to 8/5/2021

Groundwater and Surface Water Elevations and Hydraulic Gradients - 3/14/2019 to 8/5/2021: WMW-16 compared to WMW-14

BNSF Wishram Railyard Wishram, Washington

Appendix B

Proposed CULs and Identification of COCs

Appendix B: Proposed CULs and Identification of COCs

B1. Introduction

This Appendix to the Uplands Feasibility Study Report (FS Report) summarizes the development of proposed cleanup levels (CULs) for groundwater, surface water, and soil and identifies constituents of concern (COCs) for the BNSF Wishram Railyard [Washington State Department of Ecology (Ecology) Site Name: BNSF Track Switching Facility] (site) in Wishram, Washington. Identification of COCs is based on comparison of soil and groundwater sampling results [presented in the *Uplands Remediation Investigation Report BNSF Wishram Railyard* (Uplands RI Report; Kennedy Jenks 2020)] and 2020 groundwater sampling results to the proposed CULs.

This summary has been prepared in response to Ecology's memorandum titled *Wishram Cleanup Site: Preliminary Cleanup Levels and Contaminants of Potential Concern for Groundwater, Surface Water and Upland Soils* (Ecology PCULs Memo) (Ecology 2021a), received by BNSF on 17 August 2021, Ecology's comments to the draft Uplands Feasibility Study Report, received by BNSF on 31 August 2021, and Ecology's comments to the revised draft Uplands Feasibility Study Report, received by BNSF on 15 and 16 March 2022.

B2. Current and Future Use

The Klickitat County, Washington zoning map [\[https://www.klickitatcounty.org/284/Zoning-Map;](https://www.klickitatcounty.org/284/Zoning-Map) (Klickitat County Zoning Map)], indicates the railyard is zoned as an "Industrial Park." The area to the north of the railyard property boundary (illustrated as a green line on Figures B-1 and B-2), is zoned as "Rural Center".

BNSF Wishram Railyard Property

The BNSF Wishram Railyard Property (site) is currently used as a railyard and meets the definition of an industrial property in Washington Administrative Code (WAC) 173-340-200 and the applicability criteria of soil cleanup standards for industrial properties under WAC 173-340- $745(1)(a)(i)(A)$ through (F). The railyard and areas north of the mainline track area (e.g., the maintenance shop and bullpen) have controlled access, including fencing between the existing depot and maintenance shop, a fenced-in bullpen north of the maintenance shop, and are patrolled by railroad police. The primary potential exposures to potentially impacted surface and subsurface soil, groundwater and adjacent surface water are to adult railroad, construction, and utility workers.

The site is fully developed as an industrial railyard where buildings, pavement, rail lines, and surfaces (comprising approximately 94 percent of the land area) are designed and managed per federal regulations to remove and control vegetation, limit the potential for vegetation with deep root zones and use by wildlife. The railyard surface areas, covered by gravel, asphalt, or other impervious structures (e.g., buildings) minimize potential exposure to the soil. Along the Columbia River, engineered embankments (forming the berm area) composed of large riprap protect the banks from erosion and restrict potential deeper soil contact by occupants and wildlife. Railroad operations control subsurface disturbance and can be formally established by institutional controls (ICs). The site is anticipated to remain a railyard for the foreseeable future,

with railyard operations comprising of railcar switching on track spurs located south of the Depot (see Figure 2 in the FS Report).

Off-Railyard Properties

Areas to the north of the railyard property boundary (off-railyard properties) are zoned as Rural Center. Potential petroleum hydrocarbon impacts related to the railyard are limited to the vicinity of the former boiler house and former heating oil UST within this area. The UST and approximately 750 tons of petroleum hydrocarbon-impacted soil were removed in April 2002 (see Figure B-2 for approximate lateral extent of the excavation area). Petroleum hydrocarbonimpacted soils were excavated to the top of the bedrock surface (to the extent practicable) at a depth of approximately 16 feet below ground surface (bgs); groundwater was not encountered in the excavation. Confirmation samples from the north, east, and south sidewalls of the excavation indicated diesel- and oil-range petroleum hydrocarbon impacts remained in-place from approximately 14.5 to 15.5 feet bgs (see Figures 6, 9, and 10 in the FS Report).

Fire District #11 Wishram is listed as the property owner of the two property parcels located in this area (Figure B-2). Current features include two warehouse storage-type buildings used by the fire department in the eastern parcel and a U.S. Post Office in the western parcel. Surface areas covered by gravel or impervious structures (buildings) comprise approximately 65 percent of the land area between the two parcels and minimize potential exposure to the subsurface soil. The primary potential exposures to potentially impacted subsurface soil and groundwater are to adult construction and utility workers.

B3. Proposed CULs

Proposed site CULs for groundwater, surface water, and soil are based on Ecology's Model Toxics Control Act (MTCA), other applicable, relevant, and appropriate requirements (ARARs) and other relevant information pertinent to establishing site-specific remedial goals. Ecology's *Supporting Material for Cleanup Levels and Risk Calculation (CLARC) - Soil Cleanup Levels to Protect Groundwater* (CLARC Guidance) guidance document was used in the development of proposed groundwater, surface water, and soil CULs along with applicable sections of WAC 173-340. Available and applicable CULs and screening levels were obtained from Ecology's CLARC master data table (updated in July 2021) (Ecology 2021b).

MTCA requires sites to be cleaned up to the extent that they no longer pose an unacceptable threat to human health and the environment. Establishing CULs to meet that objective is described generally under WAC 173-340-700. The applicability of MTCA Methods A, B, and C cleanup levels, is summarized under WAC 173-340-704, -705, and -706, respectively. According to WAC 173-340-700(5)(a)-(c), except where institutional controls (ICs) are required by WAC 173-340-440(4), where cleanup meet Method A and/or Method B cleanup levels, a site may be used without future restrictions on the property. Site cleanups with Method C cleanup levels may have restrictions (ICs) placed on the property to ensure future protection of human health and the environment. Cleanup levels developed using Methods B and C must not be set at levels below the practical quantitation limit (PQL) in accordance with WAC 173-340-707 or background concentrations.

Method A cleanup levels may be used at sites that have few hazardous substances and are either undergoing a routine cleanup action as defined in WAC 173-340-200 or where numerical standards are available for all indicator hazardous substances in the media for which the

Method A cleanups levels are to be used. Method A cleanup levels for unrestricted land use (listed in MTCA Table 740-1) and for restricted (industrial) land use (listed in MTCA Table 745-1) are based on the protection of groundwater for drinking water beneficial uses.

Method B is applicable to all sites and can be used to develop cleanup levels unless one or more conditions for using Method A or Method C is demonstrated to exist and the person conducting the cleanup action elects to use that method. Method B cleanup levels are based on unrestricted land use and the protection of groundwater.

Method C cleanup levels are protective of human health and the environment for specified uses and conditions. Method C cleanup levels use exposure assumptions and risk levels for restricted land uses, including industrial land uses. A site (or portion of a site) that qualifies for Method C cleanup levels for one medium (e.g., soil), may not necessarily qualify for a Method C cleanup level in other media. As stated by Ecology (2016a), a key difference between Method B and Method C soil cleanup levels is that under Method C, the direct soil contact exposure pathway is based on healthy workers being exposed to soil contamination instead of children being exposed in a residential setting. Per WAC 173-340-745(4), Method C is the standard method for establishing soil cleanup levels at industrial sites and its use is conditioned upon the continued use of the site for industrial purposes. Soil CULs developed under Method C must still be protective of groundwater and surface water.

Groundwater and Surface Water CULs

Beneficial use designations for the Columbia River near the site include water supply, spawning and rearing aquatic life, wildlife, and miscellaneous such as recreation, aesthetics, hydroelectric power generation, and commercial navigation and transportation. In accordance with WAC 173- 340-720, designated groundwater uses include potential drinking water source, although shallow groundwater at the site is not a current source of drinking water nor is it identified as a future drinking water source as potable water is supplied to the site by the City of Wishram. Proposed groundwater and surface water CULs protective of drinking water were based on Method B cleanup levels, Method A cleanup levels for select petroleum hydrocarbons [gasolinerange organics (GRO), diesel-range organics (DRO), oil-range organics (ORO), and total petroleum hydrocarbons – diesel range fraction (TPH-Dx, sum of DRO and ORO)], and ARARs. Surface water CULs were also selected to be protective of aquatic life.

Soil CULs - BNSF Wishram Railyard Property

Based on the current and future use of the site as a railyard, with activities continuing to meet the definition of an industrial property, proposed CULs for the unsaturated and saturated zone soil on railyard property are based on Methods A and C cleanup levels for industrial properties, along with consideration of soil concentrations protective of groundwater (e.g., leaching) including leaching with the potential to discharge to surface water.

Soil CULs - Off-Railyard Properties

The current use of the properties north of the maintenance shop area includes small business and commercial services uses (e.g., the U.S. Post Office and fire department storage buildings), although the zoning designation as "Rural Center" allows for other non-commercial uses including homes and eating/drinking establishments. Therefore, proposed CULs for the unsaturated and saturated zone soil in these areas are based on Methods A and B unrestricted cleanup levels, along with consideration of soil concentrations protective of groundwater (e.g.,

leaching). As this area is not adjacent to surface water, soil concentrations protective of groundwater discharging to surface water were not considered.

CUL Tables

Tables B-1, B-2, and B-3 present the proposed CULs for groundwater, surface water, and soil, respectively. The selection process of the CULs is discussed in the following sections. Because the soil CULs are based in part on protection of groundwater and surface water (via the soil leaching to groundwater pathway), groundwater and surface water CULs are presented first in the following sections. Constituents with one or more samples exceeding the applicable proposed CUL for that media are also identified as COCs in the tables. Table B-4 presents a summary of the CULs for identified COCs by matrix. The CULs for COCs will be further evaluated in the draft Cleanup Action Plan (CAP).

Preliminary PQLs included in Tables B-1, B-2, and B-3 were based on one of the following: (1) recommended PQLs as available from Table 7.3 of Ecology's *Guidance for the Remediation of Petroleum Contaminated Sites* (Ecology 2016a); (2) PQLs provided in the Ecology PCULs Memo; or (3) PQLs and method reporting limits (MRLs) provided by Pace Analytical for its laboratories in Mount Juliet, Tennessee, and Minneapolis, Minnesota. Identification of PQLs for constituents in the draft CAP will include evaluation of MRLs and method detection limits (MDLs) (which may change on at least an annual basis) provided by several Washington State-Accredited laboratories.

B3-1. Remedial Investigation Data

Soil and groundwater sampling results presented in the Uplands RI Report and groundwater sampling results from 2020 were used in the evaluation of COCs by media.

Groundwater Data. Results from groundwater sampling events of monitoring wells conducted between November 2016 and August 2020 were selected as representative data for current dissolved phase groundwater conditions. Depending on monitoring well installation date (pre-2016, 2016, or 2018), groundwater samples were collected from monitoring wells during five events (deep wells installed in 2018 and sampled semiannually) and 15 events (shallow wells installed in 2016 along the berm). Screening "grab" reconnaissance groundwater (RGW) samples were collected from temporary wells as a component of Uplands RI field investigation activities in 2016 and 2018; however, the evaluation of groundwater sampling results for identifying COCs considers monitoring well samples only, as they are reproducible and representative of current groundwater conditions.

Evaluation of the RGW groundwater data indicated that, other than four metals, constituents identified at concentrations above drinking water standards in the RGW samples were also identified in the samples collected from monitoring wells. Total metals results for barium, cadmium, chromium, and/or lead in fifteen (15) 2016 RGW samples were above respective CULs; however, dissolved metals results in the 2016 and 2018 RGW samples were below CULs. As discussed in Section 2.3.3.3.7 of the Uplands RI Report, it was determined that dissolved metals (and not total metals) concentrations in reconnaissance groundwater samples from temporary wells from 2016 and 2018 are representative of site conditions. The RGW samples with total and/or dissolved metals results above CULs are generally within approximately 50 to 100 feet of a cross-gradient or downgradient monitoring well. The

monitoring wells bound the lateral extent of the dissolved-phased constituents that exceeded applicable screening levels in RGW samples.

Additional considerations for select constituents are summarized below, related to evaluation of CULs for soil and groundwater.

Total Chromium. As reported in the Uplands RI Report, there is no source for hexavalent chromium at the site. Therefore, as stated in the CLARC chemical-specific considerations – July 2021, "If chromium VI is NOT present at the site, then the site assessor may assume that the measured concentration of total chromium is the concentration of chromium III." Accordingly, the CULs proposed for chromium in soil, groundwater and surface water are the CULs for chromium III or total chromium, as available.

Carcinogenic PAHs. Ecology policies and procedures for implementing WAC 173-340- 708(8)(e) in the MTCA rule requires that mixtures of cPAHs be considered a single hazardous substance (total cPAH) when establishing and determining compliance with cleanup levels. Results of the cPAH compounds [benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene, chrysene, dibenz(a,h)anthracene, and indeno(1,2,3-c,d)pyrene] were used to calculate total cPAH concentrations using the Toxicity Equivalency Factor (TEF) summation method. If an individual chemical was not detected, a value of one-half the method reporting limit (MRL) was used as the concentration in the calculation, except when all chemicals used in the calculation were not detected then one half the lowest MRL was used as the total concentration. Calculated total cPAH results are presented in tables in this Appendix for comparison to proposed CULs.

Total Naphthalenes. According to Ecology's *Guidance on Remediation of Petroleum-Contaminated-Sites* (Ecology 2016a) and MTCA Method A cleanup level Tables 720-1, 740-1, and 745-1, under the MTCA rule, "naphthalenes" for comparison to Method A cleanup levels are the total of naphthalene, 1-methylnaphthalene, and 2-methylnaphthalene. Results of the 1-methylnaphthalene, 2-methylnaphthalene, and naphthalene analyses (as available) were used to calculate total naphthalenes concentrations, which are presented in tables for soil and groundwater analytical results. Total naphthalenes were calculated by summing the individual naphthalenes concentrations. If an individual chemical was not detected, a value of one half the MRL was used as the concentration in the calculation, except when all chemicals used in the calculation were not detected, then one half the lowest MRL was used as the total concentration. Naphthalene, 1-methylnaphthalene, 2-methylnaphthalene and calculated total naphthalenes results are presented in Tables B-1, B-2, and B-3 in this Appendix for comparison to proposed CULs.

B3-2. Groundwater CULs and COCs

Beneficial use designations for the groundwater beneath the site include use as a potential drinking water source. Proposed groundwater CULs protective of drinking water were established from consideration of the following values:

• "Groundwater Target Cleanup Level for Soil to Ground Water Pathway" (groundwater target CUL), which is included in the CLARC tables and derived, as described in the CLARC Guidance, from evaluation of the groundwater Method B Cancer and Noncancer cleanup levels, ARARs including state and federal Maximum Contaminant Levels (MCLs), and/or adjustments to ARARs as applicable.

- Method A cleanup levels for groundwater for select constituents (e.g., GRO, DRO, ORO, and TPH-Dx).
- Background (for arsenic), and PQLs. The background concentration for arsenic is from MTCA Table 720-1 Method A CULs and is subject to change based on potential policy updates resulting from Ecology's January 2022 publication *Natural Background Groundwater Arsenic Concentrations in Washington State Study Results*.

Proposed groundwater CULs are summarized in Table B-1, along with the source (e.g., Method A, Method B Cancer or Noncancer, ARARs, etc.) of the value. Constituents included in the table were detected in one or more groundwater samples collected from monitoring wells between November 2016 and August 2020, representative of current groundwater conditions. Dibenz(a,h)anthracene was detected in one sample from well WMW-30 in the 2016 to 2020 time period; the other six cPAH constituents were not detected in groundwater samples during this time period. All seven cPAHs are shown in Table B-1 along with Total cPAH based on the TEF summation method; the proposed groundwater CUL is for benzo(a)pyrene and Total cPAH in accordance with WAC 173-340-708(8)(e).

For groundwater, the total chromium federal and Washington State MCL is 100 micrograms per liter (µg/L). As reported in the Uplands RI Report, there are no known or potential sources of hexavalent chromium at the site; therefore, it is assumed that all the chromium is trivalent and the Method A concentration for total chromium is 100 µg/L.

Constituents reported in one or more groundwater monitoring well samples collected since November 2016 above the proposed CUL for groundwater are identified as COCs in Table B-1.

Identified groundwater COCs and respective proposed CULs (and their sources) are as follows:

- DRO 500 µg/L as an unrestricted use goal (Method A CUL).
- ORO 500 µg/L as an unrestricted use goal (Method A CUL).
- Total petroleum hydrocarbon (TPH-Dx) 500 µg/L as an unrestricted use goal (Method A CUL).
- 1-methylnaphthalene -1.5μ g/L (Method B Cancer CUL).
- Total arsenic $-5 \mu g/L$ (Background).
- Total barium $-2,000 \mu g/L$ (State and Federal MCL).
- Dissolved iron 11,000 µg/L (Method B Noncancer CUL).
- Dissolved manganese 750 µg/L (Method B Noncancer CUL).

Discussion. Total naphthalenes results in groundwater samples collected between November 2016 and August 2020 were less than the proposed CUL of 160 µg/L (Method A); however, results for the individual constituent 1-methylnaphthalene were above its proposed CUL of 1.5 µg/L (Method B Cancer) and therefore, 1-methylnaphthalene was identified as a groundwater COC. Based on review of results from groundwater sampling events between November 2016 and August 2020, GRO and Chromium (total) were not identified as COCs in

Table B-1. These constituents were included as COCs in Ecology's Comment No. 28 to the Draft FS. Ecology's process for identifying COCs for groundwater included comparing the maximum reported concentration for each constituent in groundwater samples to the Ecology preliminary CUL (PCUL). It does not appear that consideration was given to the age of the sample result, concentration trends or sampling artifacts as discussed below:

- GRO was reported at a concentration (1,790 μ g/L) above its Method A CUL of 1,000 µg/L (GRO without presence of benzene in the sample) in the sample collected from monitoring well WMW-07 in July 2004. GRO results in samples collected from well WMW-07 since July 2007 have been below the CUL of 1,000 µg/L. GRO has not been reported above its CULs (800 µg/L with benzene or 1,000 µg/L without benzene present) in samples from other monitoring wells.
- Chromium (total) has not been reported above the proposed CUL of 100 µg/L in groundwater samples from monitoring wells. Total chromium (total) was reported at concentrations above the proposed CUL in four grab samples from temporary wells in 2016; dissolved chromium (total) was not reported above laboratory reporting limits in these four samples. As discussed in Section 2.3.3.3.7 of the Uplands RI Report, it was determined that dissolved metals (and not total metals) concentrations in reconnaissance groundwater samples from temporary wells from 2016 and 2018 are representative of site conditions.

B3-3. Surface Water CULs

Beneficial use designations for the Columbia River near the site include use as potential drinking water supply, along with spawning and rearing aquatic life, wildlife, and miscellaneous such as recreation, aesthetics, hydroelectric power generation, and commercial navigation and transportation. Proposed surface water CULs protective of drinking water and aquatic life were established from consideration of the following values and are summarized in Table B-2:

- "Surface Water Target Cleanup Level for Soil to Surface Water Pathway Fresh Water" (surface water target CUL), which is included in the CLARC tables and is derived from evaluation of surface water Method B Cancer and Noncancer cleanup levels and ARARs including fresh surface water concentrations that are protective of human health and protective of aquatic life under acute and chronic exposure conditions as established under state (WAC 173-201A-240) and federal laws [Section 304 of the Clean Water Act (CWA)].
- Environmental Effects-Based Concentrations (EEBCs) from Implementation Memo No. 23 (Ecology 2021c) for petroleum hydrocarbons (e.g., weathered diesel). As presented in the Uplands RI Report, NWTPH-Dx chromatograms from soil and groundwater samples collected from the site (areas north and south of the mainline tracks associated with historical fueling operations and along the berm) represent weathered diesel or a mixture of weathered diesel and/or weathered Bunker C.
- Method A CULs for select constituents (e.g., GRO, DRO, ORO, and TPH-Dx) for beneficial use as potable water.
- Surface water values for select constituents from the Risk Assessment Information System (RAIS) or from the Great Lakes Water Quality Initiative (GLWQI) Clearinghouse

[\(https://www.epa.gov/gliclearinghouse\)](https://www.epa.gov/gliclearinghouse) (if no surface water ecological ARAR is available).

• Background concentrations and PQLs.

Proposed surface water CULs are summarized in Table B-2. Constituents included in the table were detected in one or more groundwater samples collected from monitoring wells between November 2016 and August 2020. Samples collected in this timeframe are considered representative of current groundwater conditions and were used to identify COCs for surface water adjacent to the railyard.

Constituents reported above the proposed CUL for surface water in one or more groundwater samples collected since November 2016 from monitoring wells located along the berm separating the site from the Columbia River are identified as COCs in Table B-2. The berm wells include 10 shallow monitoring wells (WMW-14 through WMW-23) and six deep monitoring wells (RMD-1 through RMD-6).

The proposed surface water COCs and respective CULs (and their sources) for monitoring wells located along the berm are as follows:

- DRO 500 µg/L as an unrestricted use goal (Method A CUL) for human health exposure considerations and 3,000 µg/L (EEBC weathered DRO) for ecological receptor exposure considerations.
- ORO 500 µg/L as an unrestricted use goal (Method A CUL) for human health exposure considerations and 3,000 µg/L (EEBC weathered DRO) for ecological receptor exposure considerations.
- Total petroleum hydrocarbon (TPH-Dx) 500 µg/L as an unrestricted use goal (Method A CUL) for human health exposure considerations and $3,000 \mu g/L$ (EEBC) weathered DRO) for ecological receptor exposure considerations.
- 1-methylnaphthalene $-2.1 \mu g/L$ (RAIS).
- Total arsenic 5 µg/L (Background).
- Total lead 2.5 µg/L (Surface Water Aquatic Life Fresh/Chronic WAC 173-201A and CWA §304).
- Dissolved iron 1,000 µg/L (Surface Water Aquatic Life Fresh/Chronic CWA §304).
- Dissolved manganese 50 µg/L (Surface Water Human Health Fresh Water CWA §304).

Discussion. Between November 2016 and August 2020, total cPAHs were detected in one sample from one site well (WMW-30, not located on the shoreline berm), but not in the five subsequent samples from the well. Total cPAHs were not detected in the shoreline berm monitoring wells, and therefore, the groundwater to surface water pathway is not complete for total cPAHs. Similarly, total barium was reported above its proposed CUL in one of 74 groundwater samples from site wells (WMW-30, August 2019) at a concentration that appeared anomalous (16,500 µg/L) compared to other reported concentrations in samples from

the same well before and after August 2019 (ranged 34.2 to 45.5 µg/L) and other site wells. Total barium was not reported above the proposed groundwater and surface water CULs in the shoreline berm wells, and therefore, the groundwater to surface water pathway is not complete. Total cPAHs and total barium do not merit further evaluation as COCs in surface water.

B3-4. Soil CULs

Based on the property zoning designations, the proposed unsaturated (vadose) and saturated zone soil CULs were established from either unrestricted or restricted (industrial) cleanup levels as available for protection of human health (direct contact), and soil protective of groundwater (for highest beneficial use) and soil protective of groundwater with potential to discharge to water (for highest beneficial use).

For off-railyard properties (rural center-zoned), direct contact cleanup levels were established from unrestricted cleanup levels. As the railyard meets the criteria for an industrial property (WAC 173-340-200 and WAC 173-340-745), direct contact cleanup levels for soil within the railyard property boundary were established from industrial cleanup levels. Points of compliance are presented below for soil.

Unsaturated (Vadose) Subsurface Soil [0 to 10 feet below ground surface (bgs)]

The standard point of compliance for soil based on human exposure via direct contact is sitewide to a depth of 15 feet bgs as the typical maximum depth of soil disturbing activities [WAC 173-340-740(6)(d)]. The unsaturated vadose zone within much of the railyard extends from the ground surface to 10 feet bgs and along the berm from the ground surface to approximately 15 feet bgs. As such, a conditional point of compliance for vadose zone soil at the site, with adoption of ICs, is the water table depth of 10 feet bgs within the railyard.

Vadose zone soil will be evaluated using:

- Applicable unrestricted Method A and B CULs for soil off-railyard property to the north of the maintenance shop (rural center zoned); or
- Applicable industrial Method A and C CULs for soil within the railyard (industrial park zoned).

For those constituents proven likely to leach from soil to groundwater, cleanup levels will be established from soil concentrations (calculated using the fixed parameter 3-phase partitioning model described in WAC 173-340-747(4) and MTCA Equation 747-1) for the vadose zone that are protective of groundwater and the groundwater to surface water pathway.

Saturated Subsurface Soil (10 to 15 feet bgs)

The depth to groundwater in much of the railyard is 10 feet bgs. Groundwater was not encountered in the 2002 excavation adjacent to the former boiler house. However, based on 2016 and 2018 soil borings near the former 30,000-barrel Oil AST and the Maintenance Shop area, the depth to groundwater in the off-railyard properties to the north of the railyard was estimated as 14 feet bgs in the Uplands RI Report. Soil between the groundwater table (approximately 10 feet bgs on the railyard and 14 feet bgs off-railyard) and 15 feet bgs will be evaluated as saturated zone soil using:

- Applicable unrestricted Method A and B CULs for soil off-railyard property to the north of the maintenance shop (rural center zoned); or
- Applicable industrial Method A and C CULs for soil within the railyard (industrial park zoned).

For those constituents proven likely to leach to groundwater, cleanup levels will be established using soil concentrations (calculated using the fixed parameter 3-phase partitioning model described in WAC 173-340-747(4) and MTCA Equation 747-1) for the saturated zone that are protective of groundwater and the groundwater to surface water pathway. Impact to groundwater from subsurface soil in the saturated zone will be evaluated based on direct measurement of groundwater conditions and remediation of soil in the saturated zone will be evaluated based on groundwater cleanup performance data.

As reported in Section 2.3.4 of the Uplands RI Report, the site qualified for completion of a Simplified Terrestrial Ecological Evaluation (TEE) from which it was determined that the site does not pose a substantial threat of significant adverse effects to terrestrial ecological receptors. As such, a TEE is excluded from this evaluation.

Proposed soil cleanup levels are presented in Table B-3A for samples collected above the groundwater table (i.e., vadose zone) and in Table B-3B for samples collected below the groundwater table (i.e., saturated zone). The proposed CULs were based on the following:

- Method A direct contact cleanup levels for unrestricted land use from Table 740-1 for unrestricted land use (areas north of the railyard property boundary) or from Table 745-1 for industrial (restricted) land use for selected petroleum hydrocarbon-related constituents (GRO, DRO, ORO, and TPH-Dx). The numerical values for the unrestricted (Table 740-1) and restricted (Table 745-1) Method A CULs for these constituents are equivalent.
- Lowest of Method B Cancer and Noncancer unrestricted land use direct contact cleanup levels from the CLARC tables.
- Lowest of Method C Cancer and Noncancer industrial land use direct contact cleanup levels from the CLARC tables.
- Calculated soil concentrations protective of potable groundwater (i.e., leaching) and protective of groundwater that may discharge to surface water in accordance with the CLARC Guidance document.
- Background concentrations and PQLs.

Empirical Demonstration

In accordance with WAC 173-340-747(9), an empirical demonstration is presented below to document that the measured soil concentrations of several constituents in the vadose zone and saturated zone at the site have not and will not cause an exceedance of the applicable groundwater cleanup levels. The requirements for an empirical demonstration as presented in Ecology's Implementation Memorandum No. 15, Frequently Asked Questions (FAQs) Regarding Empirical Demonstrations and Related Issues (Ecology June 2016b) include that:

- The measured groundwater concentrations must be less than or equal to the applicable groundwater cleanup levels; and
- The measured soil concentrations will not cause an exceedance of the applicable groundwater cleanup levels at any time in the future.

"This requires applicants to demonstrate a) that enough time has elapsed for hazardous substances to migrate from soil to groundwater, and b) that the characteristics of the site (e.g., depth to groundwater and infiltration) are representative of future site conditions."

Railyard Former and Current Conditions. The operational history of the railyard was presented in the Uplands RI Report along with a summary of current site conditions based on extensive site investigations conducted since 2002. A summary is presented here, relative to the sources and estimated age of impacts.

Fueling of steam locomotives with Bunker-C oil was conducted at the site from approximately 1912 through 1956 and fueling of locomotives using diesel fuel was conducted from the early 1950s to the 1970s. Petroleum hydrocarbon impacts were identified in soil and groundwater in the vicinity of former USTs, former ASTs, and former infrastructure used to store and transfer fuel associated with former locomotive fueling operations. Where implemented, interim remedial measures (IRMs) conducted between 2002 and 2010 successfully removed petroleum hydrocarbons from unsaturated zone areas down to the water table or bedrock, as applicable, such that soil samples collected from 134 of 145 confirmation sampling locations in the unsaturated zone did not contain residual petroleum hydrocarbons. Based on the fueling and petroleum storage history of the site, these soil impacts were likely greater than 50 to 60 years old at the time of removal.

Groundwater monitoring has been conducted since 2003. Groundwater samples and/or depth to water/LNAPL measurements were collected from up to 41 groundwater monitoring wells during groundwater monitoring events. Groundwater monitoring data demonstrate that the average depth to groundwater is 10 feet bgs and varies by only 1 or 2 feet throughout the year. Submerged LNAPL is estimated to be greater than 60 years old based on known facility operations and the formation of Lake Celilo. The site is anticipated to remain a railyard for the foreseeable future. The railyard surface areas, covered by gravel, asphalt, or other impervious structures (e.g., buildings) are anticipated to remain the same for the future as well, minimizing changes to groundwater conditions (groundwater depth and infiltration) that may be affected by surface coverings.

Vertical infiltration rates through the unsaturated zone were calculated following guidance in Appendix 1 of Implementation Memorandum No. 15. Using a gravimetric water content of 5 percent (based on moisture content laboratory data in the Uplands RI Report), a bulk density of 1.49 grams per cubic centimeters (g/cc) (based on soil core data in Table 25 of the Uplands RI Report), and residual saturation and fully saturated values provided in Appendix 1 for sand, an effective saturation of 0.077 was calculated. Figure 4 of Appendix 1 was used to estimate a hydraulic conductivity of 1 centimeter per day (cm/day) (0.033 feet/day). The estimated time for impacts at the ground surface to reach the water table at a depth of 10 feet bgs (railyard property) is 0.8 years, and at a depth of 14 feet bgs (off-railyard properties) is 1.2 years.

Travel times for horizontal transport from saturated soil source material to a downgradient well were calculated without and with consideration for chemical retardation following guidance from Appendix 2 of Implementation Memorandum No. 15. Using an average estimated hydraulic

conductivity of 6.39 feet per day (Table 12 of Uplands RI Report), an estimated hydraulic gradient of 0.003 feet/foot [average positive hydraulic gradient (toward the river) [1](#page-189-0) from four wells (WMW-01, WMW-03, WMW-14, and WMW-18) between 14 March 2019 and 5 August 2021], and an effective porosity of 0.28 (average of soil core data in Table 25 of the Uplands RI Report), the calculated groundwater seepage velocity is 0.068 feet per day. Assuming no chemical retardation, travel times for dissolved phase impacts in groundwater would be 0.8 years to travel 20 feet or 4 years to travel 100 feet to a downgradient well.

Chemical properties of 1-methylnaphthalene were used to estimate the chemical retardation factor for petroleum hydrocarbon impacts to soil. Using an organic carbon/water partition coefficient of 2,530 milliliters per gram (mL/g) for 1-methylnaphthalene (CLARC tables), estimated fraction of organic carbon of 0.001 for sand aquifer material, bulk density of 1.49 grams per cubic centimeters (g/cc), and effective porosity of 0.28, the calculated retardation factor is 14.5. Dividing the groundwater seepage velocity by the retardation factor, the calculated chemical velocity is 0.0047 feet per day. At this velocity, travel times are approximately 12 years for 20 feet and 58 years for 100 feet.

Based on the estimated age of impacts to soil and groundwater (greater than 60 years), measured soil concentrations of potential COCs remaining in soil (both vadose zone and saturated zone) which have not already resulted in impacts to groundwater are not anticipated to cause an exceedance in the future.

The notation "Empirical" under the "3-Phase Model" headings in Tables B-3A and B-3B for select constituents indicates that the soil protective of groundwater and soil protective of groundwater to surface water pathways are not complete. For constituents with the empirical demonstration, proposed soil CULs were based on the human health direct contact exposure pathway. The empirical demonstration is based on one of the following (Additional notes have been added to the Empirical notation Tables 3A and 3B and below to clarify the basis for the empirical demonstration):

- Constituents without a cleanup level for indicated pathways. ("Empirical No CUL")
- Constituents detected in groundwater samples at concentrations below the respective proposed groundwater and surface water CULs. ("Empirical - Below GW&SW CUL")
- Constituents detected in one or more soil samples from the vadose and/or saturated zone but not detected in groundwater samples. ("Empirical – ND in wells")
- 1-Methylnaphthalene: The highest reported concentrations of 1-methylnaphthalene were in soil samples containing visible sheen and LNAPL collected at 12 feet bgs from borings advanced adjacent to LIF borings TG-D0 and TG-D1. Elevated 1-methylnaphthalene concentrations reported in wells south of these borings (RMD-1 and WMW-16) are attributed to dissolution from the LNAPL. 1-methylnaphthalene was also reported above the reporting limit in shallow samples from borings B-18-18 (1.5-2.0 feet bgs) and WMW-29 (2.0-2.5 feet bgs) but not in deeper soil samples from those two borings

 1 As presented in the Uplands RI Report (KJ 2020), the Columbia River is a losing stream more than 80 percent (10 months) of the year. The average hydraulic gradient was calculated from positive gradient values only when the hydraulic gradient is toward the river (gaining stream).

Uplands Feasibility Study Report, BNSF Wishram Railyard Page 12 of 14 m:\wp\2021\2196120.06_wishram_fs\fs_rpt_rev\appb_culs\appb1_bnsf wishram culs and cocs_202205.docx

(B-18-18 at 14.0-14.5 feet bgs and WMW-29 at 9.5-10.0 feet bgs). 1-methylnaphthalene was not reported in groundwater samples from monitoring well WMW-21, located approximately 40 feet south of soil boring B-18-18 nor in the well installed in boring WMW-29. Absence of 1-methylnaphthalene in the deeper soil samples from both borings and in groundwater samples from the two wells indicates that the leaching pathway from soil to groundwater is not complete. ("Empirical – Data Based")

• Barium: As discussed previously, total barium was reported above the proposed groundwater CUL in one of 74 groundwater samples from site wells (WMW-30, August 2019) at an anomalous concentration compared to other reported concentrations in samples from the same well and other site wells. Total barium was below the CULs in groundwater samples from berm wells. Barium concentrations in soil samples from WMW-30 were below the soil concentration protective of groundwater indicating the soil leaching to groundwater pathway is not complete. ("Empirical – Data Based")

Constituents reported in one or more soil samples above applicable (i.e., vadose or saturated zone) proposed soil CULs were identified as COCs. As shown in Table B-3A and Table B-3B, the proposed soil COCs were different for vadose and saturated zone conditions. The proposed soil COCs and respective CULs (and their sources) are as follows:

- GRO 30 milligrams per kilogram (mg/kg) as an unrestricted use goal (Method A CUL); the unrestricted and restricted (industrial) CULs are the same. Applicable to saturated soil only.
- DRO 2,000 mg/kg as an unrestricted use goal (Method A CUL); the unrestricted and restricted (industrial) CULs are the same. Vadose and saturated soil.
- ORO 2,000 mg/kg as an unrestricted use goal (Method A CUL); the unrestricted and restricted (industrial) CULs are the same. Vadose and saturated soil.
- TPH-Dx 2,000 mg/kg as an unrestricted use goal (Method A CUL); the unrestricted and restricted (industrial) CULs are the same. Vadose and saturated soil.

B3-5. Identification of COCs

Table B-4 summarizes COCs in groundwater, surface water, and soil (vadose and saturated zones for areas on- and off-railyard property), proposed CULs and their source, as identified in this appendix.

References

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Attachments

Table B-1: Proposed Groundwater Cleanup Levels

Table B-2: Proposed Surface Water Cleanup Levels

Table B-3A; Proposed Vadose Zone Soil Cleanup Levels

Table B-3B: Proposed Saturated Zone Soil Cleanup Levels

Table B-4: Summary of Constituents of Concern (COCs) and Proposed Cleanup Levels (CULs)

Figure B-1: Current and Historical Site Features and Previous Interim Remedial Actions

Figure B-2: Zoning Designations and Sample Locations (2002-2018) – Main Area

PROPOSED GROUNDWATER CLEANUP LEVELS

BNSF Wishram Railyard, Wishram, Washington

Cleanup level values based on Model Toxics Control Act (MTCA) Method B values (B Cancer or B Non Cancer) and MTCA Method A values for groundwater (Table 720-1) based on Washington State Administrative Code (WAC) 173-340-72 tables (Accessed July 2021), and Washington State (246-290 WAC) and Federal (40 CFR 141) Maximum Contaminant Levels (MCLs).

Preliminary Practical Quantitation Limit (PQL) based on values provided in one of the following: (1) Table 7-3 of Ecology's Guidance for Remediation of Petroleum Contaminated Sites. Toxics Cleanup Program. Publication No.

Table presents proposed cleanup levels (CULs) and CUL sources for constituents detected in one or more groundwater samples from monitoring wells between 2016 and 2020.

Diesel-Range Organics **Detected Constituents** above the proposed CUL are shaded blue. These constituents are identified as Constituents of Concern (COCs) for groundwater.

Abbreviations and Symbols

" - -" denotes not measured, not available, or not applicable. µg/L = micrograms per liter

There are no known or potential sources of hexavalent chromium at the site; therefore, it is assumed that all of the chromium is trivalent and the Method A number for total chromium is 100 µg/L in accordance with WAC 246-2

Total cPAHs = Possible total cPAHs are based on the relative toxicity of each cPAH to benzo(a)pyrene and were calculated by multiplying the individual cPAH concentrations by a toxicity equivalency factor (TEF) and summing

Total Naphthalenes = Total Naphthalenes concentrations were calculated by summing 1-Methylnaphthalene, 2-Methylnaphthalene, and Naphthalene concentrations. Non-detects were included as noted.

(HalfDL_WA) = If an individual chemical was not detected, a value of one half the method reporting limit was used as the concentration in the calculation, except when all chemicals used

in the calculation were not detected then one half the lowest method reporting limit was used as the total concentration.

Cleanup Levels

(2) Ecology Memorandum "Wishram Cleanup Site: Preliminary Cleanup Levels and Contaminants of Potential Concern for Groundwater, Surface Water and Upland Soils" dated 28 May 2021; or

(3) Method reporting limits (MRLs) provided by Pace Analytical for Pace Analytical National in Mount Juliet, TN (provided 4/27/2022) and Pace Analytical in Minneapolis, MN (provided 5/9/2022).

PROPOSED SURFACE WATER CLEANUP LEVELS

BNSF Wishram Railyard, Wishram, Washington

Table presents proposed cleanup levels (CULs) and CUL sources for constituents detected in one or more groundwater samples from monitoring wells located along the Columbia River berm between 2016 and 2020.
Diesel-Range Org

Detected constituents above the proposed CUL are shaded blue. These constituents are identified as Constituents of Concern (COCs) for surface water.

Cleanup level values based on Washington State Department of Ecology (Ecology) Model Toxics Control Act (MTCA) Method B values (B Cancer or B Non Cancer), MTCA Method A values and other applicable, relevant, and appropriat (WAC) 173-340-730 from Cleanup Levels and Risk Calculation (CLARC) tables (Accessed July 2021).

Abbreviations and Symbols

" - -" denotes not measured, not available, or not applicable.

yg/L = micrograms per liter

There are no known or potential sources of hexavalent chromium at the site; therefore, it is assumed that all of the chromium is trivalent and the Method A number for total chromium is 100 µg/L in accordance with WAC 246-2

Total cPAHs = Possible total cPAHs are based on the relative toxicity of each cPAH to benzo(a)pyrene and were calculated by multiplying the individual cPAH concentrations by a toxicity equivalency factor (TEF) and summing

Total Naphthalenes = Total Naphthalenes concentrations were calculated by summing 1-Methylnaphthalene, 2-Methylnaphthalene, and Naphthalene concentrations. Non-detects were included as noted.

Total Xylenes = Total Xylenes concentrations were calculated by summing Xylene, m,p- and Xylene, o- concentrations. Non-detects were included as noted.

(HalfDL_WA) = If an individual chemical was not detected, a value of one half the method reporting limit was used as the concentration in the calculation, except when all chemicals used

in the calculation were not detected then one half the lowest method reporting limit was used as the total concentration.

Cleanup Levels

"WQC_HH" = Surface water quality cleanup level for Human Health. "WQC_AL" = Surface water quality cleanup level for Aquatic Life.

Other Ecological Sources include: Great Lakes Water Quality Initiative (GLWQI) Clearinghouse (https://www.epa.gov/gliclearinghouse), The Risk Assessment Information System (RAIS) (https://rais.ornl.gov/tools/eco_search.php),

"EEBC" = Environmental Effects-Based Concentrations, from Concentrations of Gasoline and Diesel Range Organics Predicted to be Protective of Aquatic Receptors in Surface Waters, Implementation Memorandum No. 23 (Ecology, A

Preliminary Practical Quantitation Limit (PQL) based on values provided in one of the following: (1) Ecology's Guidance for Remediation of Petroleum Contaminated Sites. Toxics Cleanup Program. Publication No. 10-09-057. Re

(2) Ecology Memorandum "Wishram Cleanup Site: Preliminary Cleanup Levels and Contaminants of Potential Concern for Groundwater, Surface Water and Upland Soils" dated 28 May 2021; or

(3) Method reporting limits (MRLs) provided by Pace Analytical for Pace Analytical National in Mount Juliet, TN (provided 4/27/2022) and Pace Analytical in Minneapolis, MN (provided 5/9/2022).

PROPOSED VADOSE ZONE SOIL CLEANUP LEVELS

BNSF Wishram Railyard, Wishram, Washington

Table presents proposed cleanup levels (CULs) and CUL sources for constituents detected in one or soil samples collected between 2002 and 2018. Results include samples collected from unsaturated vadose zone soil.
Diesel-Ra

Preliminary Practical Quantitation Limit (PQL) based on values provided in one of the following: (1) Ecology's Guidance for Remediation of Petroleum Contaminated Sites. Toxics Cleanup Program. Publication No. 10-09-057. Re (3) Method reporting limits (MRLs) provided by Pace Analytical for Pace Analytical National in Mount Juliet, TN (provided 4/27/2022) and Pace Analytical in Minneapolis, MN (provided 5/9/2022).

Abbreviations and Symbols

 $-$ -" denotes not measured, not available, or not applicable. mg/kg = milligrams per kilogram per kilograms per kilogram

Where "Empirical" is listed in the row, it has been assumed that an empirical demonstration indicates that the contaminant is not reaching groundwater/surface water at values above what is considered protective; based on g

Empirical Below CUL = constituent not reported above the proposed groundwater (GW) and/or surface water (SW) CUL in groundwater samples. Above CUL = constituent reported above the proposed groundwater CUL in one or more gr There are no known or potential sources of hexavalent chromium at the site; therefore, it is assumed that all of the chromium is trivalent and the Method A number for total chromium is 100 ug/L in accordance with WAC 246-2

Total cPAHs = Possible total cPAHs are based on the relative toxicity of each cPAH to benzo(a)pyrene and were calculated by multiplying the individual cPAH concentrations by a toxicity equivalency factor (TEF) and summing

Total Naphthalenes = Total Naphthalenes concentrations were calculated by summing 1-Methylnaphthalene, 2-Methylnaphthalene, and Naphthalene concentrations. Non-detects were included as noted.

Total Xylenes = Total Xylenes concentrations were calculated by summing Xylene, m,p- and Xylene, o- concentrations. Non-detects were included as noted.

(HalfDL_WA) = If an individual chemical was not detected, a value of one half the method reporting limit was used as the concentration in the calculation, except when all chemicals used

in the calculation were not detected then one half the lowest method reporting limit was used as the total concentration.

Cleanup Levels

Cleanup level values based on Washington State Department of Ecology (Ecology) Model Toxics Control Act (MTCA) Method C values (Method C C = Cancer or Method C NC = Non Cancer) and MTCA Method A values for soil (Table 745-

and 3-Phase Model Soil Protective of Groundwater Vadose (Eq. 747-1) and 3-Phase Model Soil Protective of Groundwater to Surface Water Vadose Fresh Water (Eq. 747-1) from Cleanup Levels and Risk Calculation (CLARC) tables (

PROPOSED SATURATED ZONE SOIL CLEANUP LEVELS

BNSF Wishram Railyard, Wishram, Washington

Table presents proposed cleanup levels (CULs) and CUL sources for constituents detected in one or soil samples collected between 2002 and 2018. Results include samples collected from unsaturated vadose soil.

Diesel-Range Organics **Detected Constituents above the proposed CUL** are shaded blue. These constituents are identified as Constituents of Concern (COCs) for saturated soil between 10 and 15 feet below ground surface.

Abbreviations and Symbols

" - -" denotes not measured, not available, or not applicable. mg/kg = milligrams per kilogram

Where "Empirical" is listed in the row, it has been assumed that an empirical demonstration indicates that the contaminant is not reaching groundwater/surface water at values above what is considered protective; based on g

Empirical Below CUL = constituent not reported above the proposed groundwater (GW) and/or surface water (SW) CUL in groundwater samples. Above CUL = constituent reported above the proposed groundwater CUL in one or more gr There are no known or potential sources of hexavalent chromium at the site; therefore, it is assumed that all of the chromium is trivalent and the Method A number for total chromium is 100 µg/L in accordance with WAC 246-2

Total cPAHs = Possible total cPAHs are based on the relative toxicity of each cPAH to benzo(a)pyrene and were calculated by multiplying the individual cPAH concentrations by a toxicity equivalency factor (TEF) and summing

Total Naphthalenes = Total Naphthalenes concentrations were calculated by summing 1-Methylnaphthalene, 2-Methylnaphthalene, and Naphthalene concentrations. Non-detects were included as noted.

Total Xylenes = Total Xylenes concentrations were calculated by summing Xylene, m,p- and Xylene, o- concentrations. Non-detects were included as noted.

(HalfDL_WA) = If an individual chemical was not detected, a value of one half the method reporting limit was used as the concentration in the calculation, except when all chemicals used

in the calculation were not detected then one half the lowest method reporting limit was used as the total concentration.

Cleanup Levels

Cleanup level values based on Washington State Department of Ecology (Ecology) Model Toxics Control Act (MTCA) Method C values (Method C C = Cancer or Method C NC = Non Cancer) and MTCA Method A values for soil (Table 745-

and 3-Phase Model Soil Protective of Groundwater Saturated (Eq. 747-1) and 3-Phase Model Soil Protective of Groundwater to Surface Water Saturated Fresh Water (Eq. 747-1) from Cleanup Levels and Risk Calculation (CLARC) ta

Preliminary Practical Quantitation Limit (PQL) based on values provided in one of the following: (1) Ecology's Guidance for Remediation of Petroleum Contaminated Sites. Toxics Cleanup Program. Publication No. 10-09-057. Re (3) Method reporting limits (MRLs) provided by Pace Analytical for Pace Analytical National in Mount Juliet, TN (provided 4/27/2022) and Pace Analytical in Minneapolis, MN (provided 5/9/2022).

SUMMARY OF CONSTITUENTS OF CONCERN (COCs) AND PROPOSED CLEANUP LEVELS (CULs) BNSF Wishram Railyard, Wishram, Washington

Page 1 of 1

Notes:

COC = Constituents of concern

CUL = Cleanup level. Refer to Appendix B to the Draft Feasibility Study Report for more information.

"µg/L" = micrograms per liter

"mg/kg" = milligrams per liter

Total TPH-Dx = Total TPH-Dx concentrations calculated by summing diesel-range organics (DRO) and oil-range organics (ORO) concentrations.

Cleanup Levels:

Groundwater: Cleanup level values based on Model Toxics Control Act (MTCA) Method B values (B Cancer or B Non Cancer) and MTCA Method A values for groundwater (Table 720-1) based on Washington State Administrative Code (WAC) 173-340-720 from Cleanup Levels and Risk Calculation (CLARC) tables (Accessed July 2021), and Washington State (246-290 WAC) and Federal (40 CFR 141) Maximum Contaminant Levels (MCLs).

Surface Water: Cleanup level values based on Ecology MTCA Method B values (B Cancer or B Non Cancer), MTCA Method A values and other applicable, relevant, and appropriate requirements (ARARs) under applicable state (173-201A-240 WAC) and federal laws [Section 304 of the Clean Water Act (CWA); 40 CFR Subpart D 131.45] for surface water based on WAC 173-340-730 from CLARC tables (Accessed July 2021).

Risk Assessment Information System (RAIS) (https://rais.ornl.gov/tools/eco_search.php),

Environmental Effects-Based Concentrations (EEBC), from Concentrations of Gasoline and Diesel Range Organics Predicted to be Protective of Aquatic Receptors in Surface Waters, Implementation Memorandum No. 23 (Ecology, August 25, 2021).

Proposed Restricted / Industrial CULs for soil On BNSF Railyard Property

Soil: Cleanup level values based on Ecology MTCA Method C values (Cancer and Noncancer) and MTCA Method A (Table 745-1) values for soil based on WAC 173-340-745 and 3-Phase Model Soil Protective of Groundwater Vadose (Eq. 747-1) and 3-Phase Model Soil Protective of Groundwater to Surface Water Vadose Fresh Water (Eq. 747-1). from CLARC tables (Accessed July 2021).

Proposed Unrestricted CULs for soil in non-industrial use (zoned or otherwise) areas Off BNSF Railyard Property.

Soil: Cleanup level values based on Ecology MTCA Method B values (Cancer and Noncancer) and MTCA Method A (Table 740-1) values for soil based on WAC 173-340-740 and 3-Phase Model Soil Protective of Groundwater Vadose (Eq. 747-1) and 3-Phase Model Soil Protective of Groundwater to Surface Water Vadose Fresh Water (Eq. 747-1). from CLARC tables (Accessed July 2021).

 $\frac{N}{2}$ 0 50 100 Scale: Feet

Legend

- \bigoplus Bioventing Injection Well
- \bigoplus Air Sparge (AS) Well
- ---- Approximate BNSF Property Line
- **Fig. 2** Former Bunker Fuel / Oil Pipeline
- ---- Former Oil Drain
- Former Oil Trough
- $-$ Former Sewer Line (Potential) Stormwater Underdrain (A portion $\overline{}$ removed from service circa 1960)
- Stormwater Underdrain (Rerouted portion circa 1960) **?**
- Former Diesel Line
- $==$ Former Steam Line
- **Suspected Diesel Fueling Area**
- Suspected Oil Fueling Area
- Approximate Previous Excavation Area
- Approximate Previous Excavation Area
- Approximate Previous Excavation Area
- Existing Site Feature
- Former Site Feature

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants Current and Historical Site Features and Previous Interim Remedial Actions**

2196120*06 May 2022 **Figure 4 Figure B-1**

- ---- Approximate BNSF Property Line
- **O** LIF Location Within BNSF Property Boundary
- \odot LIF Location North of BNSF Property Boundary
- **Location North of BNSF Property Boundary**
- **•** Excavation Soil Sample
- **.** Soil Boring Location
- Off-Railyard Parcel Boundaries
- **Approximate 2002 Excavation** Approximate 2005/2010 Excavation
- **Location Within BNSF Property Boundary**
- Abandoned Monitoring Well
- \bigoplus Air Sparge
- \bigoplus Soil Vapor Extraction
-
- **Excavation Soil Sample** \bigoplus OHM-3R
- \bigoplus OHM Well
- \bigoplus Shallow Monitoring Well
- \bigoplus Deep Monitoring Well
- **O** Soil Boring Location
- Notes:
- 1. Selected parcel boundaries and property owners (shown) based on review of Klickitat County zoning maps through Interactive Next

Mapping Program, accessed 27 April 2022.

Selected parcel boundaries and property owners (shown) based

on review of Klickitat County zoning maps through Interactive

Mapping Program, accessed 27 April 2022.

Legend

BNSF Wishram Railyard Wishram, Washington **Kennedy/Jenks Consultants**

2296120*00 May 2022 **Figure # Figure B-2**

Zoning Designations and Sample Locations (2002-2018) - Main Area

Appendix C

LNAPL Recoverability Estimate

Appendix C: LNAPL Recoverability Estimate

Introduction

This Appendix to the Uplands Feasibility Study Report (FS Report) presents an estimate of the recoverability of submerged light non-aqueous phase liquids (LNAPL) from beneath the BNSF Wishram Railyard (Ecology Site Name: BNSF Track Switching Facility) (site) in Wishram, Washington.

LNAPL with properties consistent with both diesel and Bunker-C oil is present south of the mainline tracks near the former underground oil pipelines and the former Power House (see Figures 7, 8, and 20 in the FS Report). As shown on Figure 20 in the FS Report, eastern and western LNAPL areas have been defined based on the variations of depth and apparent LNAPL thickness, as well as depths to bedrock. The LNAPL recoverability estimate focuses on the eastern LNAPL area, which is characterized by primarily Bunker-C oil LNAPL submerged at depths from approximately 10 feet below ground surface (bgs) to 85 feet bgs. Submerged LNAPL observed south of the mainline is classified as potentially recoverable, as evidenced by apparent LNAPL thicknesses measurements of 12 feet or more in three of the four oil head monitoring (OHM) wells (OHM-1, OHM-2, and OHM-3) and by laboratory measurements of LNAPL physical properties.

LNAPL Recoverability Estimate Procedures

American Petroleum Institute's LNAPL Distribution and Recovery Model (LDRM) (API 2007) was used to estimate the percentages of recoverable LNAPL and total LNAPL that could potentially be removed from the subsurface in the eastern LNAPL area, and the amount of residual LNAPL that would remain. The LDRM was populated using laboratory-analyzed physical properties for LNAPL and water samples collected from wells OHM-1, OHM-2, and OHM-3 and soil properties for soil cores collected from the three OHM well borings. The LNAPL and soil property data used in the LDRM runs are summarized in Table C-1 [previously reported in Tables 24 and 25 of the *Uplands Remedial Investigation Report BNSF Wishram Railyard* (Kennedy Jenks 2020)].

The LDRM was used to estimate the removal of LNAPL from wells OHM-1, OHM-2, and OHM-3 under various conditions. The results were then applied to a representative number of wells with similar LNAPL conditions that would be needed to capture LNAPL from the entire submerged LNAPL body. The LDRM was run for three separate subsurface temperature scenarios and corresponding laboratory analyzed LNAPL density and viscosity measurements (Table C-1):

- (1) LNAPL removal under a temperature of 70 degrees Fahrenheit (°F) [21 degrees Celsius (°C)] and LNAPL viscosity ranging from 696 to 1989 centipoise (cp) for the three wells.
- (2) LNAPL removal at a heated subsurface temperature of 100 °F (38 °C) and reduced LNAPL viscosity ranging from 188 to 421 cp for the three wells.
- (3) LNAPL removal at a heated subsurface temperature of 130 \degree F (55 \degree C) and lowest measured LNAPL viscosities ranging from 70 to 133 cp for the three wells.

The scenarios were modeled for 1, 10, and/or 1,000 years of LNAPL skimmer recovery well operation as shown in Table C-2. Inputs for the skimmer recovery model included a well radius of 0.25 foot and a radius of capture of 20 feet. The only LDRM input parameters changed between the three scenarios were the LNAPL viscosity and LNAPL density, which decrease with increasing temperature. The estimated total recoverable LNAPL for each temperature scenario by well was based on a skimming operating time of 1,000 years (with one exception) to achieve a minimum of 99 percent recovery of the recoverable LNAPL. For well OHM-2 under scenario (1) at 70 °F, the model was run for the maximum model allowable time (9,999 years) to achieve greater than 99 percent recovery.

The lowest temperature (70 °F) scenario (1) was run for a duration of 10 years. Scenario (1) is representative of FS Alternatives 1 and 4, which include 10 years of LNAPL skimmer operation with no subsurface heating. [Note that 70 °F is slightly higher than ambient groundwater temperatures (between 56 and 66 °F) measured between March 2019 and August 2021 in five monitoring wells]. The LDRM estimate for 1 year of skimmer operation is shown in Table C-2 for comparison to scenarios (2) and (3).

The highest temperature (lowest viscosity) scenario (3) was run for a duration of 1 year. Scenario (3) provides an estimate for FS Alternative 5, which includes 1 year of LNAPL removal operation with subsurface heating at temperatures between 122 and 158 °F. The 100 °F temperature scenario (2) was also modeled for 1 year of LNAPL removal to evaluate the differences between low and moderate temperature heating scenarios.

LDRM Results

Based on the selected LNAPL and soil physical properties, estimates of the percent of recoverable LNAPL removed and the percent of total LNAPL removed were derived from the LDRM. The LDRM results also provided an estimate of the recoverable LNAPL as a percent of the total LNAPL mass modeled. LDRM results are summarized in Table C-2.

For the simulated conditions of each well, reductions in the LNAPL viscosity resulted in increases in LNAPL recoverability; LNAPL density values had less impacts on LNAPL removal estimates. As shown in Table C-2, the estimates of LNAPL removed compared to total recoverable LNAPL and total LNAPL mass modeled varied significantly between the three OHM wells. The estimated LNAPL removed as a percent of total LNAPL mass modeled was highest for well OHM-2 (42 to 55 percent); achievable after 1,000 years of operation or more. However, for the tested scenarios and durations applicable to FS Alternatives (e.g., 1 or 10 years), the percent of recoverable LNAPL removed and the percent of total LNAPL removed were consistently higher for well OHM-1 than wells OHM-2 and OHM-3.

Alternatives 1, 4, and 5 in the FS include operation of 18 LNAPL removal wells to address the submerged LNAPL impacts. The estimated percent of the recoverable LNAPL removed from the eastern LNAPL area of the site was based on skimmer operation of six of each OHM well (with applicable LNAPL and soil conditions), for a total of 18 LNAPL removal wells.

As shown in Table C-2, the estimated percentages of recoverable LNAPL removed were similar for the low temperature scenario (1) after 10 years (20 percent) and high temperature scenario (3) after 1 year (21 percent). The percent recovery for scenario (2) was roughly half (or less) of that of scenarios (1) and (3) (10 percent). The percent of total LNAPL removed was similar for scenarios (1) and (3), with results of 4.5 and 6.7 percent, respectively.

Discussion

The LDRM is limited in its ability to evaluate oil removal, based on the conceptual site model. The LNAPL at the site is submerged (trapped) below the water table, not floating on top of it. One of the assumptions made in the LDRM is that the water, oil, and air phases are in vertical equilibrium – i.e., the intermediate fluid (oil) sits on top of the denser fluid (water) and below the less dense air. The LDRM uses the observed LNAPL thickness in the well to define the capillary pressure as a function of elevation above the LNAPL/water interface. The capillary properties control the associated saturations of water, LNAPL, and air (API 2007). In cases where vertical equilibrium does not exist, the model can only generally represent the distribution of the various phases, as the LDRM formulation is not able to accurately account for residual LNAPL trapped beneath the water table.

The higher viscosity and molecular weight of the LNAPL, compared to groundwater, result in the LNAPL being trapped in the pore spaces below the water table. An example in the API LDRM guidance (API 2007) illustrated the inverse relationship between LNAPL viscosity on LNAPL recovery: increasing viscosity 10-fold from 0.5 cp to 5 cp resulted in initial LNAPL recovery rates decreasing by 90 percent.

Based on their review of documented LNAPL remediation examples, the authors of the API LDRM guidance (API 2007) indicated that for most environmental remediation cases, hydraulic recovery of more than 30 percent of the LNAPL in-place would be the exception rather than the rule. For finer grained materials, the authors state that recovery of more than 15 percent of the LNAPL volume in-place would be unusual. Factoring in the higher viscosity (up to 1,989 cp at 70°F) of the site Bunker-C oil LNAPL, which is approximately 400- to 4,000-times higher than the LDRM guidance example data range (0.5 to 5 cp, representative of liquids including gasoline, water, and jet fuel) and the fine sands lithology in OHM-2 and OHM-3 soil cores, the low LNAPL recovery percentages estimated by the LDRM for the site appear reasonable.

References

- American Petroleum Institute (API). 2007. LNAPL Distribution and Recovery Model (LDRM) Volume 2: User and Parameter Selection Guide. API Publication 4760. January 2007.
- Kennedy/Jenks Consultants. 2020. Uplands Remedial Investigation Report BNSF Wishram Railyard (Ecology Site Name BNSF Track Switching Facility) Wishram, Washington. Prepared by Kennedy/Jenks Consultants for BNSF Railway Company. 20 October 2020.

Attachments

Table C-1 Summary of LNAPL Distribution and Recovery Model (LDRM) Inputs
Table C-2 Summary of Recoverable LNAPL Estimates Summary of Recoverable LNAPL Estimates

TABLE C-1

SUMMARY OF LNAPL DISTRIBUTION AND RECOVERY MODEL (LDRM) INPUTS BNSF Wishram Railyard

Notes:

Lab = Analyses for fluid properties performed by PTS Laboratories, Inc. of Santa Fe Springs, California in 2017 (OHM-2 data), and PTS Laboratories, Inc. of Houston, Texas,

in 2019 (OHM-1 and OHM-3 data). LNAPL = light non-aqueous phase liquid

Maximum LNAPL thickness between December 2016 and November 2018 was measured on 21 August 2018 for all three wells.

 $ft^{-1} = 1/feet$

 $g/cm³$ = grams per cubic centimeter LIF = Laser Induced Fluorescence

 $dyne/cm = dyne$ per centimeter cp = centipoise

 ft^3/ft^2 = cubic feet per square foot

-
-

 $ft = feet$

TABLE C-2

SUMMARY OF RECOVERABLE LNAPL ESTIMATES BNSF Wishram Railyard

Notes:

Dynamic viscosities: 2016 data for OHM-2, 2019 data for OHM-1 and OHM-3. Light non-aqueous phase liquid (LNAPL) not collected in 2019 from OHM-2 for analysis.

Run LDRM recovery for Skimmer recovery system, variable recovery times (1, 10, 1000 years), Radius of well = 0.25 foot,

Radius of Capture = 20 feet (based on wells on 40-foot spacing)

1 year is estimated time for low-temperature thermal removal (LTTR) operation (50-70 °C) [estimated time LTTR system would run]; FS Alternative 5.

10 years is estimated time for physical LNAPL removal without heating (70 °F); FS Alternatives 1 and 4.

1,000 years is used as the end point for achieving extent practicable LNAPL Removal of the Recoverable LNAPL (targeting minimum of 99% recovery).

* = For OHM-2 at 70 °F, the model was run for 9,999 years to achieve 99% removal of the recoverable LNAPL.

YR - Year

% - Percentage

18 wells options assume 6 wells of each of OHM-1, OHM-2, and OHM-3 to calculate totals.

OHM-1, 70°F, 10 years OHM-1, 70°F, 1000 years

OHM-2, 70°F, 9999 years **OHM-1, 70°F**, 1 year

OHM-2, 70°F, 10 years OHM-2, 70°F, 1000 years

OHM-3, 70°F, 1000 years OHM-2, 70°F, 1 year

OHM-3, 70°F, 1 year OHM-3, 70°F, 10 years

OHM-1, 100°F, 1 year OHM-1, 100°F, 1000 years

OHM-2, 100°F, 1 year OHM-2, 100°F, 1000 years

OHM-3, 100°F, 1 years **OHM-3, 100°F, 1000** years

OHM-1, 130°F, 1 year OHM-1, 130°F, 1000 years

OHM-2, 130°F, 1 year OHM-2, 130°F, 1000 years

OHM-3, 130°F, 1 years **OHM-3, 130°F, 1000** years

Appendix D

Summary of Remedial Alternatives Cost Estimates

TABLE D1 ALTERNATIVE 1 - PHYSICAL LNAPL REMOVAL, FOCUSED BIOSPARGE, MNA, AND TARGETED EXCAVATION

BNSF Wishram Railyard

Notes:

1. Estimated cost was prepared at -30/+50% for relative comparison amongst alternatives. The prepared cost estimate is not intended for budgetary purposes.

2. An engineering cost estimate will be prepared in conjunction with CAP preparation and design (technical specifications and drawings).

Uplands Feasibility Study Report, BNSF Wishram Railyard (Ecology Site Name BNSF Track Switching Facility) M:\WP\2021\2196120.06_Wishram_FS\FS_Rpt_Rev\AppD_Costs\CostTables_Assumptions_DraftFS2021_2021Edits.xlsx

TABLE D2

ALTERNATIVE 2 - INSTITUTIONAL CONTROLS AND ENVIRONMENTAL CONVENANTS WITH COMPLIANCE GROUNDWATER MONITORING BNSF Wishram Railyard

Notes:

1. Estimated cost was prepared at -30/+50% for relative comparison amongst alternatives. The prepared cost estimate is not intended for budgetary purposes.

2. An engineering cost estimate will be prepared in conjunction with CAP preparation and design (technical specifications and drawings).

TABLE D3 ALTERNATIVE 3 - LNAPL CONTAINMENT, BIOSPARGE, MNA, AND TARGETED EXCAVATION BNSF Wishram Railyard

Notes:

1. Estimated cost was prepared at -30/+50% for relative comparison amongst alternatives. The prepared cost estimate is not intended for budgetary purposes.

2. An engineering cost estimate will be prepared in conjunction with CAP preparation and design (technical specifications and drawings).

TABLE D4 ALTERNATIVE 4 - PHYSICAL LNAPL REMOVAL, BIOSPARGE, MNA, AND TARGETED EXCAVATION BNSF Wishram Railyard

Notes: 1. Estimated cost was prepared at -30/+50% for relative comparison amongst alternatives. The prepared cost estimate is not intended for budgetary purposes.

2. An engineering cost estimate will be prepared in conjunction with CAP preparation and design (technical specifications and drawings).

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TABLE D5 ALTERNATIVE 5 - LOW-TEMPERATURE THERMAL REMOVAL (LTTR), BIOSPARGE, MNA, AND TARGETED EXCAVATION BNSF Wishram Railyard

Notes: 1. Estimated cost was prepared at -30/+50% for relative comparison amongst alternatives. The prepared cost estimate is not intended for budgetary purposes.

2. An engineering cost estimate will be prepared in conjunction with CAP preparation and design (technical specifications and drawings).

Uplands Feasibility Study Report, BNSF Wishram Railyard (Ecology Site Name BNSF Track Switching Facility) M:\WP\2021\2196120.06_Wishram_FS\FS_Rpt_Rev\AppD_Costs\CostTables_Assumptions_DraftFS2021_2021Edits.xlsx

TABLE D6 COST SUMMARY BNSF Wishram Railyard

