Final Report

BASELINE HUMAN HEALTH AND ECOLOGICAL RISK ASSESSMENT, UPLAND OPERABLE UNIT

Bradford Island Cascade Locks, Oregon

April 2016

Prepared for:



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ACRONYMS AND ABBREVIATIONS

| % | percent/percentile |
|------------|---|
| AAF | absorption adjustment factor |
| ACH | air changes per hour |
| ADAF | age-dependent adjustment factors |
| AF | attenuation factors |
| ALM | adult lead model |
| ACPC | |
| | area of potential concern |
| AST | aboveground storage tank |
| AT | average time |
| ATSDR | Agency for Toxic Substances and Disease Registry |
| AUF | area use factor |
| BAF | bioaccumulation factor |
| BaPeq | benzo(a)pyrene-equivalent carcinogenic potency |
| BHHRA | baseline human health risk assessment |
| BERA | baseline ecological risk assessment |
| bgs | below ground surface |
| BHC | benzene hexachloride |
| BISB | Bradford Island Service Building |
| BW | body weight |
| CDC | U.S. Center for Disease Control |
| CEC | contaminant of ecological concern |
| CEM | conceptual exposure model |
| CERCLA | Comprehensive Environmental Response, Compensation, and Liability Act |
| CF | conversion factor |
| COC | contaminant of concern (human health) |
| COI | contaminant of interest |
| COPC | contaminant of potential concern |
| cPAH | carcinogenic polycyclic aromatic hydrocarbon |
| CPEC | contaminant of potential ecological concern |
| C_{soil} | concentration in soil |
| CTE | central tendency exposure |
| CVOC | chlorinated volatile organic compound |
| DCE | dichloroethene |
| DDD | dichloro-diphenyl-dichloroethane |
| DDE | dichloro-diphenyl-dichloroethylene |
| DDT | dichloro-diphenyl-trichloroethane |
| DDx | DDT, DDD, and DDE |
| DEHP | bis(2-ethylhexyl)phthalate dL deciliter |
| EcoSSL | Ecological Soil Screening Level |
| ECSI | Environmental Cleanup Site Information |
| EF | exposure frequency |
| ELCR | excess lifetime cancer risk |
| EPC | exposure point concentration |
| FI | fraction ingested |
| foc | fraction of organic carbon |
| | |



ACRONYMS AND ABBREVIATIONS

| FS | feasibility study |
|----------------------|---|
| ft | foot/feet |
| GSDI | geometric standard deviation |
| HHRA | human health risk assessment |
| HI | hazard index |
| HMSA | hazardous material storage area |
| HPAH | high molecular weight polycyclic aromatic hydrocarbon |
| HQ | hazard quotient |
| IEUBK | Integrated Exposure Uptake Biokinetic |
| IR | ingestion rate |
| IRS | soil ingestion rate |
| IRAF | infant risk adjustment factors |
| IRIS | Integrated Risk Information System |
| IUR | inhalation unit risk |
| JEM | Johnson and Ettinger (Vapor Intrusion) Model |
| K _d | partitioning coefficient for inorganics |
| kg | kilogram(s) |
| K _{oc} | partitioning coefficient for organic chemicals |
| L | liter(s) |
| LD50 | lethal dose to 50% of test organisms |
| LOAEL | lowest observable adverse effect level |
| logK _{ow} | octanol-water partition coefficient |
| μg um | microgram(s) micrometer(s) |
| μm m ³ | cubic meter |
| MDL | method detection limit |
| mg | milligram(s) |
| MP | Management Plan |
| MW | monitoring well |
| NOAEL | no observable adverse effect level |
| ODEQ | Oregon Department of Environmental Quality |
| ORNL | Oak Ridge National Laboratory |
| OU | operable unit |
| PAH | polycyclic aromatic hydrocarbon |
| PbB | blood lead concentration |
| PbB0 | baseline blood lead concentration |
| PCB | polychlorinated biphenyl |
| PCE | tetrachloroethylene |
| PPRTV | provisional peer-reviewed toxicity value |
| PRG | preliminary remediation goals |
| Q/C | inverse of the mean concentration at the center of a 0.5 acre-square source |
| RA | Risk Assessment |
| RAGS | Risk Assessment Guidance for Superfund |
| RAIS | Risk Assessment Information System |
| RAL | remedial action levels |
| RAO | remedial action objective |
| | |



ACRONYMS AND ABBREVIATIONS

| RBA | relative bioavailability factor |
|------------------|---|
| RBC | Risk-Based Concentrations |
| RfD _o | Reference Dose |
| RI | remedial investigation |
| RI/FS | remedial investigation/feasibility study |
| RM | river mile |
| RME | reasonable maximum exposure |
| RSL | Regional Screening Levels |
| SFo | oral slope factors |
| SLERA | Screening Level Ecological Risk Assessment |
| SLV | screening level value |
| SVOC | semivolatile organic compound |
| TAG | Technical Advisory Group |
| TCE | trichloroethene |
| TEF | toxicity equivalence factors |
| TPH | total petroleum hydrocarbons |
| TRV | toxicity reference value |
| TRW | Technical Review Workgroup |
| UCL | upper confidence limit |
| UPL | upper prediction limit |
| URS | URS Corporation |
| USACE | United States Army Corps of Engineers |
| USACHPPM | U.S. Army Center for Health Promotion and Preventive Medicine |
| USEPA | U.S. Environmental Protection Agency |
| USFWS | United States Fish and Wildlife Service |
| VDEQ | Virginia Department of Environmental Quality |
| VF | volatilization factor |
| VOC | volatile organic compound |
| WP | work plan |
| | rom Paul |



EXECUTIVE SUMMARY

This report presents the baseline human health risk assessment (BHHRA) and the baseline ecological risk assessment (BERA) for the Upland Operable Unit (OU), Bradford Island, Bonneville Dam Complex. The Portland District of the United States Army Corps of Engineers (USACE) has characterized and evaluated the contamination arising from historical practices at Bradford Island in Oregon. Bradford Island is part of the Bonneville Dam complex, which is located on the Columbia River at river mile (RM) 146.1, approximately 40 miles east of Portland, Oregon. The site is a multipurpose facility that consists of the First and Second Powerhouses, the old and new navigation locks, and a spillway with a capacity of 1.6 million cubic feet per second.

USACE and their contractors have performed numerous investigations since 1997, focusing on two OUs: the Upland OU, and the River OU. The investigations have identified four areas of potential concern (AOPCs) in the Upland OU: the Landfill, Sandblast Area, Pistol Range, and the Bulb Slope. The primary contaminants of interest (COIs) that have been identified in soil and/or groundwater in the four AOPCs include certain metals; polychlorinated biphenyl (PCBs); semivolatile organic compounds (SVOCs), including polycyclic aromatic hydrocarbon (PAHs); butyltins; volatile organic compounds (VOCs); and a few pesticides and herbicides.

The Final Remedial Investigation (RI) report (URS Corporation [URS] 2012) documented the investigation, identified source areas at Bradford Island, defined the nature and extent of the environmental contamination, and identified the contaminants of potential concern (COPCs) for human health and contaminants of potential ecological concern (CPECs) in the media from the two OUs. On the basis of the screening level risk assessments, the Final RI recommended site-specific BHHRAs and BERAs for the Landfill and Sandblast Area AOPCs and site-specific BERAs for the Pistol Range and Bulb Slope AOPCs. After the Final RI was completed, USACE later elected to include a Fishing Platform scenario for human health at all four AOPCs as part of the BHHRA.

The Upland OU BHHRA and BERA build on the data and findings of the Final RI (URS 2012) and will be attached as an appendix to the Upland OU Feasibility Study (FS). Due to the anticipated outcome of the baseline risk assessments (RAs) that will document the need for risk management in an Upland OU FS, this RA document goes beyond the traditional assessment of the presence/absence and magnitude of baseline risk. To maximize use of these site datasets and develop an RA that is most beneficial to the FS, site-specific risk-based concentrations (RBCs) were calculated for the chemicals recommended for further evaluation in the Upland OU FS. Exceedances of these RBCs were illustrated for purposes of risk interpretation and to allow for general observations of the spatial distribution of potentially impacted areas.

Only upland exposure pathways were addressed in these baseline RAs. The Upland OU to River OU pathways (i.e., potential mass wasting and soil erosion) that were evaluated at a screening level in the Final RI were not addressed herein, as these possible pathways will be considered in the Upland OU FS or the River OU FS.

Baseline Human Health Risk Assessment

The screening RA performed in the Final RI (URS 2012) evaluated only occupational scenarios at the Upland OU AOPCs and recommended further evaluation of occupational exposures at the



Landfill and Sandblast Area AOPCs. Following the later decision by USACE to include a Fishing Platform scenario, it was necessary to perform a risk screening for all four of the Upland OU AOPCs for the Fishing Platform receptor, followed by a site-specific baseline risk assessment, as warranted. Therefore, this BHHRA includes the limited occupational scenario evaluation for only the Landfill and Sandblast Area AOPCs (as recommended in the Final RI) and evaluation of the Fishing Platform receptor at all of the four Upland AOPCs (as decided after completion of the Final RI).

Media and COPCs evaluated at the Landfill included metals (arsenic, chromium, lead); carcinogenic PAHs (cPAHs); and chlorinated volatile organic compounds (CVOCs), including tetrachloroethylene (PCE), trichloroethylene (TCE), and associated degradation products in soil and groundwater, as well as mercury, total PCBs as Aroclors, MCPP, and naphthalene in soil only. The Sandblast Area AOPC included similar COPCs in soil, groundwater, and soil gas. A larger number of COPCs were evaluated for the Hypothetical Fishing Platform User than for the occupational scenarios. For both the Pistol Range and Bulb Slope AOPCs, the only COPC was lead in soil for the Hypothetical Fishing Platform User.

In accordance with the Final RI findings, the human receptors evaluated included the following: outdoor maintenance workers exposed to shallow soils, defined as 0 to 3 feet (ft) below the ground surface (bgs), by direct contact at the Landfill and Sandblast Area AOPCs; construction and excavation workers exposed to deeper soils, defined as 0 to10 ft bgs; excavation and trench workers exposed by incidental contact with shallow groundwater at the Landfill and Sandblast Area AOPCs; and indoor office workers exposed by vapor intrusion of soil gas at the Sandblast Area AOPC. Vapor intrusion at the Landfill AOPC is not a concern since there are currently no buildings at the Landfill AOPC, and construction of future buildings is not considered feasible from an engineering standpoint.

Subsequent to the approval of the Update to RA Work Plan (WP), Upland OU (WP Update) (URS 2014), newly identified receptors with potential for direct contact with shallow soils in the Upland OU include adults and children of the four treaty tribes who, in the future, may exercise their treaty rights to engage in fishing from Bradford Island as a usual and accustomed fishing ground. Their activities may include the construction and use of fishing platforms along the upland shoreline of the Island as well as camping and other recreational activities in the interior of the Island. Fishing platform usage does not currently occur at Bradford Island, and the likelihood of such usage in the future is unknown. It is therefore considered a future hypothetical exposure scenario in its potential to result in direct contact with soils at the four Upland OU AOPCs.

Due to the lack of site-specific information to adjust the exposure durations to more realistic values, the Hypothetical Fishing Platform Users were conservatively treated as comparable to a residential scenario in the BHHRA. For this reason, risks to the Hypothetical Fishing Platform User should be understood to be overly conservative in nature. Additionally, in the future, soil management measures to meet acceptable risk levels for occupational uses may be implemented at the Landfill and Sandblast Area AOPCs. In this case, future concentrations of COPCs in soil may be different and lower than under current conditions.

Nursing infants who may be exposed through ingestion of maternal milk were also included as newly added receptors for selected bioaccumulative COPCs. This is consistent with Oregon Department of Environmental Quality (ODEQ) guidance (ODEQ 2010a) published after the



Remedial Investigation/Feasibility Study (RI/FS) Management Plan (MP) (URS 2007) was approved (URS 2007).

Both reasonable maximum exposure (RME) and central tendency exposure (CTE) were evaluated for the occupational receptors, to represent a range of upper-end and average exposures. Excess lifetime cancer risk (ELCR) and noncancer hazard quotient (HQ) and summed hazard index (HI) were reported for all chemicals. The estimated risks for each receptor-pathway combination were characterized with regard to whether they were:

- 1) less than the U.S. Environmental Protection Agency (USEPA) risk level of 1 x 10⁻⁶ (also expressed as one in a million) or HI of 1, whereby risk at or below these thresholds have an insignificant contribution to risk (i.e., *de minimis*);
- 2) within the USEPA acceptable risk range of 1×10^{-4} to 1×10^{-6} ; or
- 3) exceeding the USEPA acceptable risk range, i.e., greater than 1×10^{-4} .

In addition, ODEQ's acceptable risk thresholds of 1×10^{-6} for individual carcinogens and 1×10^{-5} for multiple carcinogens were also considered. Due to this, individual chemicals associated with risk levels greater than the ODEQ thresholds of 1×10^{-6} or HQ greater than 1 were identified as contaminants of concern (COCs), except arsenic, for which background levels were also considered. Since the individual compounds that make up the cPAHs all act by the same mode of action, risks for PAHs were calculated by individual compound and also summed to present risks for cPAHs as a group. It is important to note that, since ELCRs are only expressions of likelihood of cancer incidence, and HIs are estimated ratios to safe doses, exceedance of the ODEQ risk thresholds or USEPA acceptable risk range does not automatically mean that adverse effects may have occurred or will occur (e.g., does not automatically mean that the COC will be recommended for further evaluation in the FS).

For occupational receptors, estimated risks were either below the ODEQ threshold or within the USEPA acceptable risk range at both the Landfill and Sandblast Area AOPCs for all exposure media. At both AOPCs, RME exposure to surface and deeper soils resulted in risks that fell within the USEPA acceptable risk range for outdoor maintenance workers and construction workers. Risks for construction workers were lower than ODEQ's threshold of 1×10^{-5} for multiple carcinogens. Incidental exposure to groundwater for excavation/trench workers were *de minimis* at both AOPCs. Risks related to vapor intrusion and indoor inhalation were also *de minimis* for both current and future indoor workers at the Sandblast Area AOPC. Only outdoor maintenance workers at the Landfill AOPC had a CTE risk within the USEPA acceptable risk range and for all other receptors, CTE risks were insignificant contributors to risk (i.e., below 1 x 10^{-6}).

For the future Hypothetical Fishing Platform User, estimated risks exceeded the USEPA acceptable risk range under current conditions for the Landfill and Sandblast Area AOPCs. Risks and noncancer hazards for nursing infants were close to or less than the ODEQ threshold of 1 x 10^{-6} for selected bioaccumulative COPCs under current conditions at both of these AOPCs. Risks related to lead at the Pistol Range and Bulb Slope AOPCs were at acceptable levels.

Table ES-1 summarizes COCs in soil that are recommended for further evaluation in the Upland OU FS.



| АОРС | Outdoor Maintenance Worker | Construction Worker | Excavation/ Trench Worker | Indoor Office Worker | Hypothetical Fishing Platform User |
|----------------|----------------------------------|------------------------|---------------------------------|----------------------------|--|
| Landfill | cPAHs | Benzo(a)pyrene | None | None | Arsenic, PCBs, cPAHs |
| Sandblast Area | cPAHs | Benzo(a)pyrene | None | None | Arsenic, PCBs, cPAHs, DEHP |
| Pistol Range | NA | NA | NA | NA | None |
| Bulb Slope | NA | NA | NA | NA | None |

Table ES-1Summary of Upland OU Soil COCs for Further Evaluation in the Upland OU FS

DEHP = bis(2-ethylhexyl)phthalate

No further evaluation or action is warranted for groundwater at either the Landfill or Sandblast Area AOPCs or for soil gas at the Sandblast Area AOPC. No further evaluation of soils at the Pistol Range or Bulb Slope is warranted for the upland evaluation. Further evaluation in the Upland OU FS is recommended for soils at the Landfill and Sandblast Area AOPCs based on risks falling above the ODEQ risk thresholds, although they are within the USEPA acceptable risk range for the receptors listed in Table ES-1.

Site-specific RBCs, corresponding to a target risk of 1×10^{-6} for cPAHs (expressed as benzo(a)pyrene equivalents) for the outdoor maintenance worker and for benzo(a)pyrene as an individual chemical for the construction worker, were calculated for the COCs identified in Table ES-1. Figures 2-5 and 2-6 identify locations at the two AOPCs that exceed the RBCs for the occupational scenarios.

The smaller sub-set of locations that also exceed ODEQ's acceptable risk level of 1×10^{-5} for multiple carcinogens (cPAHs) are also indicated in these figures.

Exceedances of cPAHs above the occupational RBCs are limited to a few areas, as follows:

- Landfill AOPC northwestern section of the Landfill, primarily within and around the Gully Test Pit and Lead Hot Spot Test Pit #1, Mercury Vapor-Lamp Test Pit, far eastern portion of the Landfill, and one isolated spot northeast of the Pesticide/Herbicide Wash Area (Figure 2-5).
- Sandblast Area AOPC –the Erodible Unit, Current Hazardous Material Storage Area (HMSA), Equipment Laydown Area, and the Former HMSA (Figure 2-6).

Site-specific RBCs were also calculated for the COCs identified for the Hypothetical Fishing Platform User (arsenic, PCBs, DEHP, cPAHs), as shown in Table ES-1. Figures 2-7 and 2-8 illustrate the locations where RBCs for the Hypothetical Fishing Platform User were exceeded in shallow soils at the Landfill and Sandblast Area AOPCs under current site exposure conditions. These are more numerous than the exceedances of occupational RBCs. At most locations with exceedances, several COCs exceeded the USEPA acceptable risk range, indicating that measures to address multiple COCs may be possible. As noted earlier, risks to the Hypothetical Fishing Platform User should be understood to be overly conservative in nature since that receptor scenario was treated as comparable to a residential scenario. In addition, soil management measures to meet acceptable risk levels for occupational uses may be implemented in the future



at the Landfill and Sandblast Area AOPCs. In this case, future concentrations of COPCs in soil may be different and lower than under current conditions.

The primary purpose of adding the Hypothetical Fishing Platform User receptor scenario is to help guide site management decisions in the Upland OU FS. The RBC for cPAHs for the Hypothetical Fishing Platform User is lower than the Reference upper prediction limit (UPL) for cPAHs. The RBC is also conservative since it assumes an overly high degree of oral and dermal bioavailability of PAHs. Incorporation of literature-based or site-specific measures of bioavailability of PAHs may result in more realistic RBCs.

Baseline Ecological Risk Assessment

All four AOPCs in the Upland OU were retained for evaluation in the BERA. Only soil was identified as a medium of concern for terrestrial ecological receptors. The CPECs carried into the BERA include metals, total high-molecular-weight PAHs (HPAHs), tributyltin, organochlorine pesticides, VOCs, and SVOCs for the Landfill and Sandblast Area, lead for the Pistol Range AOPC, and lead and mercury for the Bulb Slope AOPC. These CPECs were included in the Upland OU-wide evaluation in which wide-ranging receptors were assumed to forage in all four AOPCs combined. Risk estimates were calculated for each CPEC for all receptors potentially present at a given AOPC.

The following list of receptors and exposure pathways were included in the Upland BERA:

- Terrestrial plants and soil invertebrates exposed through direct contact with surface and shallow soil.
- Canada goose (*Branta canadensis*) exposed through incidental ingestion of surface and shallow soil, prey (100% plants), and water.
- American robin (*Turdus migratorius*) exposed through incidental ingestion of surface and shallow soil, prey (100% soil invertebrates), and water.
- American kestrel (*Falco sparverius*) exposed through incidental ingestion of surface and shallow soil, prey (100% small mammals), and water.
- Vagrant shrew (*Sorex vagrans*) exposed through incidental ingestion of surface and shallow soil, prey (100% soil invertebrates), and water.
- American mink exposed through incidental ingestion of surface soil, upland prey (15% small mammals), and water.

Based on the available data, surface soil is defined as 0 to 1 ft bgs for all AOPCs except for the Pistol Range, for which surface soil is defined as 0 to 1.5 ft bgs. Shallow soil is defined as 0 to 3 ft bgs, and this depth interval only applies to the Landfill and Sandblast Area AOPCs, as deeper data are not available for the Bulb Slope and Pistol Range.

Both low screening level values (SLVs)/no observable adverse effect levels (NOAELs) and high SLVs/lowest observable adverse effect levels (LOAELs) were selected for each receptor to develop a range of HQs and HIs for consideration by risk managers. Table ES-2 summarizes the



AOPCs and contaminants of ecological concern (CECs) identified in surface through the BERA that are recommended for further evaluation in the Upland OU FS. All of the NOAEL- and LOAEL-based HQs and HIs are less than 1.0 for the wide-ranging receptors (kestrel and mink), and no additional evaluation is needed for these receptors on an individual AOPC basis or on a combined AOPC/Upland OU-wide basis.

Table ES-2Summary of Upland OU Soil CECs for Further Evaluation in the Upland OU FS

| AOPC | Terrestrial Plants | Soil Invertebrates | American Robin | Vagrant Shrew |
|-------------------|--------------------------------------|-----------------------|--|---|
| Landfill | Mercury, Nickel, Chlordane, HPAHs | Mercury and HPAHs | Chromium, Copper, Lead, Mercury, Nickel, Chlordane | Chromium, Copper, Lead, Mercury, Nickel, Chlordane, HPAHs |
| Sandblast Area | Nickel and HPAHs | None | Chromium, Lead, Nickel | Antimony, Chromium, Lead, Nickel, HPAHs |
| Pistol Range | None | None | Lead | None |
| Bulb Slope | None | None | None | None |

Chlordane = chlordane (technical mixture and metabolites)

No further upland evaluation of the Bulb Slope is recommended based on the low risk estimates. Further evaluation in the Upland OU FS is recommended for the three remaining AOPCs based on exceedances of site-specific RBCs for one or more of the ecological receptors listed in Table ES-2.

Site-specific ecological RBCs were used to identify the areas within each AOPC that warrant specific attention in the Upland OU FS, which are summarizes as follows:

- Landfill AOPC northwestern section of the Landfill, primarily within and around the Gully Test Pit and Lead Hot Spot Test Pit #1, Mercury Vapor-Lamp Test Pit, and two locations within and northeast of the Pesticide/Herbicide Wash Area.
- Sandblast Area AOPC Spent Sandblast Grit Disposal Area, northern boundary of the Former HMSA, Catch Basin #1, localized areas within the Erodible Unit Sampling Area, and Equipment Laydown Area.
- **Pistol Range AOPC** around the approximate location of the backstop, and north and east of the approximate location of the Former Firing Shed.

To protect ecological receptors, AOPC-wide average concentrations for each AOPC should be below the lowest site-specific RBCs. When the 95% UCL concentrations for the identified CECs are below the RBCs for each AOPC, acceptable levels of ecological risk will be achieved.



1.0 INTRODUCTION

This report presents the Upland Operable Unit (OU) baseline human health risk assessment (BHHRA) and a baseline ecological risk assessment (BERA) for the Bradford Island, Bonneville Dam Complex. The Portland District of the United States Army Corps of Engineers (USACE) has characterized and evaluated the contamination arising from historical practices at Bradford Island in Oregon. Bradford Island is part of the Bonneville Dam Complex, which is located on the Columbia River at river mile (RM) 146.1, approximately 40 miles east of Portland, Oregon (Figure 1-1). The site is a multipurpose facility that consists of the First and Second Powerhouses, the old and new navigation locks, and a spillway with a capacity of 1.6 million cubic feet per second (USACE 2000). Figure 1-2 shows features of the Bonneville Dam Complex.

Site investigations on Bradford Island began with the Landfill. The Landfill was used from the early 1940s until the early 1980s. The USACE informed the U.S. Environmental Protection Agency (USEPA) and the Oregon Department of Environmental Quality (ODEQ) of the presence of the Landfill in 1996. The Landfill was added to the ODEQ Environmental Cleanup Site Information (ECSI) database in April 1997, and the Bonneville Dam Project Manager signed an ODEQ Voluntary Cleanup Agreement letter for the Landfill in February 18, 1998, under the ODEQ Voluntary Cleanup Program. In 2005, USACE continued investigation of Bradford Island under the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA). The USACE is currently working with the ODEQ to address the state's concerns regarding this investigation and any associated cleanup.

USACE and their contractors have conducted numerous investigations since 1997, focusing on two OUs: the Upland OU and the River OU (Figure 1-3). The investigations identified four areas of potential concern (AOPCs) in the Upland OU: the Landfill, Sandblast Area, Pistol Range, and the Bulb Slope (Figure 1-4). The primary contaminants of interest (COIs) that have been identified in soil and/or groundwater in the four AOPCs include certain metals; polychlorinated biphenyl (PCBs); semivolatile organic compounds (SVOCs), including polycyclic aromatic hydrocarbon (PAHs); butyltins; volatile organic compounds (VOCs); and a few pesticides and herbicides.

The Final Remedial Investigation (RI) report (URS Corporation [URS] 2012) documented the investigation, identified source areas at Bradford Island, defined the nature and extent of the environmental contamination, and identified the contaminants of potential concern (COPCs) for human health and contaminants of potential ecological concern (CPECs) in the media from the two OUs. On the basis of the screening level risk assessments (RAs), the Final RI recommended site-specific BHHRAs and BERAs for the Landfill and Sandblast Area AOPCs, and site-specific BERAs for the Pistol Range and Bulb Slope AOPCs.

The USACE and the external stakeholders for the project are collectively referred to as the Technical Advisory Group (TAG) and include federal, state, and Tribal natural resource trustees. Section 2.6 of the Final RI (URS 2012) details these officials, responsibilities, and involvement on the project. The federal and state TAG trustees actively participating in this project include the USACE, United States Fish and Wildlife Service (USFWS), National Oceanic and Atmospheric Administration, ODEQ, Oregon Department of Fish and Wildlife, Oregon Department of Human Services, and Washington Department of Ecology. Several Indian Tribes



have interests in the Columbia River and the Bradford Island site, including the Yakama Nation, the Warm Springs Tribe, the Cowlitz Tribe, the Chinook Nation, the Nez Perce Tribe, and the Confederated Tribes of the Umatilla Reservation. At the request of the TAG, USACE elected to evaluate a Hypothetical Fishing Platform User scenario for human health at all four AOPCs in the BHHRA, subsequent to the completion of the Final RI.

Soil exposure at all four AOPCs was evaluated in the BHHRA and BERA, whereas exposure to lagoon sediment at the Pistol Range AOPC, soil gas at the Sandblast Area AOPC, and groundwater at the Landfill and Sandblast Areas AOPCs are only evaluated in the BHHRA. In addition, the BERA evaluated exposure of soil from all four AOPCs combined for wide-ranging mobile receptors. As described in the Final RI (URS 2012), some soil samples in the northern portion of the Sandblast Area AOPC were divided into two subgroups based on samples that were noted for having higher soil or higher sandblast grit composition (noted as "soil" and "grit" in the Sandblast Area AOPC Section 5 tables of the Final RI); however, both subgroups were utilized as soil in the screening-level RAs in the Final RI and in these baseline RAs.

Tables 1-1 through 1-4 summarize statistics for the COPCs/CPECs in soil for the Landfill, Sandblast Area, Pistol Range, and Bulb Slope AOPCs, per relevant depth intervals. Consistent with the Final RI (URS 2012), the depth intervals for Sandblast and Landfill AOPCs are characterized as surface (0-1 foot [ft] below ground surface[bgs]), shallow (0-3 ft bgs) and deeper (0-10 ft bgs), and the Bulb Slope and Pistol Range AOPCs only have surface soil depth intervals of 0-1 ft bgs and 0-1.5 ft bgs, respectively. Table 1-5 presents the statistical summary for the CPECs in soil for all four AOPCs combined (Upland OU-wide for mobile receptors). Table 1-6 lists the summary statistics for the COPCs in soil gas, and Tables 1-7 and 1-8 list the summary statistics for the COPCs in groundwater at the Landfill and Sandblast Area AOPCs. Table 1-9 provides summary statistics for chemicals in the Reference Area soil and groundwater at Bradford Island.

The Upland OU BHHRA and BERA presented herein build upon on the data and findings of the Final RI (URS 2012) and will be attached as an appendix to the Upland OU FS. Due to the anticipated outcome of the baseline RAs that will document the need for further evaluation in an Upland OU FS, this RA document goes beyond the traditional assessment of the presence/absence and magnitude of baseline risk. To maximize use of these site datasets and develop an RA that is most beneficial to the FS, site-specific risk-based concentrations (RBCs) were calculated for the chemicals recommended for further evaluation in the Upland OU FS. Exceedances of these RBCs were illustrated for purposes of risk interpretation and to allow for general observations of the spatial distribution of potentially impacted areas.

The following topics were presented in the RI and are not repeated herein unless an approach has been modified or updated:

- Site Description (RI Section 3)
- Conceptual Site Model (RI Section 4)
- Pre-RI and RI Investigations (RI Sections 5 and 6)
- RI Data Quality (RI Section 7)
- RI Screening Human Health Risk Assessment (HHRA) Discussion (RI Section 11)
- RI Screening HHRA Tables (RI Appendix M)



- RI Screening Level Ecological Risk Assessment (SLERA) Discussion (RI Section 12)
- RI SLERA Tables (RI Appendix N)
- RI Data Management (RI Appendix H)
- RI Data Sensitivity (RI Appendix K)
- RI Uncertainty Section (RI Appendix O)

The methodology for the BHHRA and BERA was originally presented in the Remedial Investigation/Feasibility Study (RI/FS) Management Plan (MP) (URS 2007) and recently refined in the Update to RA Work Plan (WP), Upland OU (WP Update) (URS 2014). The WP Update is included as Appendix C. As described in the WP Update (URS 2014), only upland exposure pathways were addressed in the BHHRA and BERA. The Upland OU to River OU pathways (i.e., potential mass wasting and soil erosion) that were evaluated at a screening level in the Final RI (URS 2012) were not addressed in the Upland OU baseline RAs. These possible pathways will be considered in the Upland Feasibility Study (FS) or River FS.

1.1 Objectives

The objectives of this report are as follows:

- Conduct a BHHRA and BERA to identify any unacceptable risks to occupational human and ecological receptors at the Landfill and Sandblast Area AOPCs.
- Conduct a BHHRA to identify any unacceptable risks to future Hypothetical Fishing Platform Users (under current conditions of site exposure point concentrations) at the Landfill, Sandblast Area, Pistol Range, and Bulb Slope AOPCs.
- Conduct a BERA to identify any unacceptable risks to ecological receptors at the Pistol Range and Bulb Slope AOPCs.
- Determine which contaminants of concern (COCs) for human health/contaminants of ecological concern (CECs) in which portions of the AOPCs need to be addressed in the subsequent FS.
- Determine which COPCs/CPECs require no additional risk evaluation and will not be carried forward to the FS.

This report will provide the basis for the Upland OU FS to be conducted later. The objectives of the Upland FS will include the following:

- Identify Remedial Action Objectives (RAOs).
- Identify and evaluate cleanup alternatives.
- Recommend proposed cleanup remedies.



1.2 Organization

This report is organized as follows:

Executive Summary

Section 1 – Introduction

Section 2 – BHHRA

Section 3 – BERA

Section 4 – Summary

Section 5 – References



2.0 BASELINE HUMAN HEALTH RISK ASSESSMENT

The BHHRA is the next step in the screening process, following the Upland OU screening-level RA presented in the Final RI (URS 2012).

At the time of completion of the Upland OU screening HHRA in the Final RI, no further evaluation of occupational exposures was recommended for the Bulb Slope and Pistol Range AOPCs. Only the Landfill and Sandblast Area AOPCs were recommended to be carried forward for the BHHRA for occupational exposures. For the Landfill and Sandblast Area AOPCs, Tables 11-1 and 11-2 of the Final RI identified the potentially exposed occupational receptors, exposure media (soil, groundwater, and soil gas), and the COPCs. However, all AOPCs, including the Bulb Slope and Pistol Range AOPCs, were considered in this BHHRA due to the addition of non-occupational new receptors as discussed below.

Updates and additions to the WP Update (URS 2014 and subsequent discussions) are as follows:

- Eliminated use of groundwater as a hypothetical potable water supply source based on ODEQ's acceptance that such use of on-site groundwater was not feasible under current or future conditions (Bob Schwarz [ODEQ], pers. comm. 2014).
- Updated exposure factors based on changes in the USEPA standard default exposure factors (USEPA 2014).
- Evaluated detected VOCs in groundwater for the trench exposure scenario.
- Added trivalent chromium as a COPC for the Landfill AOPC Outdoor Maintenance Worker.
- Added Hypothetical Fishing Platform User as a new receptor for all four AOPCs, as described in Section 2.1 and Section 2.2. This future receptor scenario is a conservative evaluation under current site exposure conditions.
- Added Nursing Infant pathway for all four AOPCs, as described in Section 2.1 and Section 2.2.

The BHHRA follows the methods described in the RI/FS MP (URS 2007) and the WP Update (URS 2014), primarily following USEPA guidance (1989, 1991, 1992a, 2002a, 2002b, 2013a, 2014, 2015a). ODEQ (2010a, 2010b) guidance documents were used during the RI HHRA screening process and are also considered in the BHHRA.

2.1 AOPCs, Exposure Media, and COPCs

Exposure media and associated COPCs for occupational receptors were identified in the Final RI (URS 2012) based on a rigorous screening exercise using occupational screening level values. For the newly added Hypothetical Fishing Platform User, a new COPC selection process using residential screening values was initiated, as described in Section 2.1.1.

As identified at the end of the Final RI, the contaminated media for occupational receptors are soil and groundwater at the Landfill AOPC and soil, groundwater, and soil gas at the Sandblast Area AOPC. The media-specific chemical groups of COPCs are as follows:



Media and COPCs for Landfill AOPC

- Soil Metals: arsenic, chromium, lead
 - Carcinogenic PAHs (cPAHs): benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, benzofluoranthenes, total, dibenz(a,h)anthracene, indeno(1,2,3-cd)pyrene
 - Chlorinated VOCs (CVOCs): tetrachloroethylene (PCE) and PCE degradation products
- **Groundwater** (incidental exposure in trench setting for Excavation/Trench Worker only) Metals: antimony, iron, mercury, thallium, zinc Phthalates: di-n-octyl phthalate VOCs: seven detected VOCs in groundwater (see Table 1-7)

Media and COPCs for Sandblast Area AOPC

- Soil Metals: arsenic, chromium, lead cPAHs: benzo(a)pyrene, benzofluoranthenes, total CVOCs: PCE and PCE degradation products
- **Groundwater** (incidental exposure in trench setting for Excavation/Trench Worker only) Metals: vanadium VOCs: 18 detected VOCs in groundwater (see Table 1-8)
- Soil Gas CVOCs: PCE and PCE degradation products, trichloroethylene (TCE)

TCE is a solvent used in degreasing and a degradation product of PCE. TCE was identified as a COPC when it was detected above its screening level in specific media (e.g., soil gas at Sandblast Area AOPC). TCE is also included as a COPC, even if it was not detected or detected at levels below TCE screening values, in a medium where PCE was detected (e.g., Landfill and Sandblast Area soils and groundwater). Other degradation products of PCE that are included as COPCs, even if they were not detected or detected below their screening levels, include cis-1,2-dichloroethene, 1,1-dichloroethene, and vinyl chloride. The rationale for including degradation products as COPCs is based on ODEQ guidance that degradation products may have the potential to occur at higher concentrations in the future as long as the parent compound is present under current conditions.

2.1.1 Hypothetical Fishing Platform User - COPC Selection

At the request of the TAG, an additional exposure scenario is included in this BHHRA for the Upland OU that was not included in the Final RI (URS 2012) or WP Update (URS 2014). A Hypothetical Fishing Platform User was considered in the BHHRA because adults and children of the four treaty tribes may, in the future, exercise their treaty rights to engage in fishing from Bradford Island as a usual and accustomed fishing ground. The tribes include the Confederated Tribes and Bands of the Yakama Nation, the Confederated Tribes of the Warm Springs Reservation, the Nez Perce Tribe, and the Confederated Tribes of the Umatilla Indian Reservation, and are referred to herein as the four treaty tribes. Their activities may include erecting fishing platforms along the shoreline, fishing with nets or lines from the platforms, and



camping at the Island. These uses may bring adults and children into direct contact with surface soils at the Island. Fish ingestion was considered in the River OU BHHRA.

At the current time, fishing platform activities do not occur at Bradford Island. If such activities were to be carried out in the future, they would likely be limited to the south shore of Bradford Island, which is the most accessible area for fishing. However, under treaty rights, such uses may be extended to the entire shoreline and the entire eastern portion of the Island. Therefore, the objective of this risk evaluation is to assess whether there are unacceptable risks to adults and children under worst-case assumptions, i.e., assuming unrestricted access to all four of the Upland OU AOPCs resulting in exposures generally comparable to residential levels to shallow soils at the AOPCs. Thus, exposure to a lifetime of unrestricted direct contact with shallow soils (i.e., incidental ingestion, dermal contact, and outdoor inhalation of dusts and vapors) was assumed. No exposure to groundwater or indoor exposure to soil gas is likely.

In summary, the Hypothetical Fishing Platform User scenario does not occur under current conditions and is hypothetical even under future conditions. However, it is included in the BHHRA as a potential, allowable use using current surface soil concentrations.

A new COPC selection process was carried out for this receptor since the earlier COPC selection process completed during the RI used only occupational screening values. For the Hypothetical Fishing Platform User, COPCs were selected by comparing COI concentrations in soil with modified residential screening values to account for an overall more conservative exposure scenario.

Consistent with ODEQ guidance, the COIs identified for shallow soils (0-3 ft bgs) in the Final RI (URS 2012) omitted chemicals detected below a 5% detection frequency and inorganics with concentrations lower than the Reference Area. The soil exposure point concentrations (EPCs) for each AOPC were the lesser of the 95th percentile (%) upper confidence limit (UCL) and the maximum detected concentration (see Tables 2-1 to 2-4).

When calculating the 95% UCL using the ProUCL software, the software may recommend a method and value for the 95% UCL that is a higher percentile (e.g., recommending a 99% UCL). The UCL recommended by the ProUCL software was automatically selected, following USEPA's statistical methodology guidance (2013b). For simplicity and clarity, the term used throughout the text is "95% UCL"; however, the specific UCLs recommended by ProUCL are shown in Tables 1-1 through 1-9.

At the request of the TAG, for the modified residential screening level values (SLVs), it was assumed that tribal members may spend 365 days a year for 26 years on the Island, which is an intentionally more conservative assumption than the RSLs. Since the residential RSLs (USEPA 2014, 2015a) assume an exposure duration of 350 days/year for 26 years, the residential soil RSLs (USEPA 2014, 2015a) were modified to account for 365 days exposure frequency (EF) for the Hypothetical Fishing Platform scenario.

(RSL Scaling Factor) = 350 days / 365 days =0.96

The generic residential soil RSL was multiplied by the scaling factor to yield a modified screening value that was lower (more stringent) than the generic RSL. COPCs for the Hypothetical Fishing Platform scenario were selected if the EPC exceeded the modified RSL for carcinogenic COIs. To account for multi-chemical exposure, non-carcinogenic COIs were



selected as COPCs if the ratio of exceedance of the modified RSL was greater than 0.1, consistent with ODEQ guidance and Final RI methodology.

A few COIs required special handling:

- Lead Lead was included as a COPC for all the AOPCs since a linear adjustment of the RSL may not be appropriate. Lead is evaluated independently using adult and child lead models.
- **Chromium** Due to the lack of speciation data, chromium was compared to the RSL for hexavalent chromium for the initial COPC selection. This is consistent with the process followed for other receptors in the Final RI. However, there is no reason to believe that hexavalent chromium may be present at the site and chromium is treated as trivalent chromium in the BHHRA.
- **PCBs** Three individual Aroclors were detected at the site: Aroclors 1248, 1254, and 1260. Total PCBs represent the sum of these three detected Aroclors, using their method detection limits (MDLs) when not detected in an individual sample total. Total PCBs and Aroclor 1248 exceeded the screening RA in the Final RI. To avoid double counting for Aroclor 1248, only Total PCBs were carried forward to the BHHRA.
- Degradation Products of PCE & TCE PCE and TCE were selected as COPCs for the BHHRA. The degradation products for these volatile compounds were never detected in landfill soils (see Table 9-1 of the Final RI; URS 2012). At the Sandblast Area, only two degradation products were detected (cis-1,2-dichloroethene [DCE] and trans 1,2-DCE), both at low detection frequencies (2-4%) and at concentrations that were orders of magnitude lower than the SLVs (see Table 2-2). Therefore, they were not included at COPCs for this receptor. The uncertainty regarding degradation products is discussed in Section 2.6.
- **Total Petroleum Hydrocarbons (TPH)** TPH compounds were excluded as COPCs due to petroleum hydrocarbon exclusion under the CERCLA process. This is discussed further in the Uncertainty Assessment (Section 2.6).

For each of the AOPCs, the screening of soil EPCs to modified USEPA soil RSLs are presented in Tables 2-1 through 2-4. The COPCs identified for the BHHRA for the Hypothetical Fishing Platform scenario are presented in Table 2-5 and are summarized below:

- Landfill AOPC Inorganics (arsenic, chromium, lead, mercury), PCBs, VOCs (naphthalene), pesticides (MCPP), and cPAHs
- **Sandblast Area AOPC** Inorganics (arsenic, chromium, lead, nickel), PCBs, VOCs (PCE, TCE), SVOCs (bis(2-ethylhexyl)phthalate [DEHP]), and cPAHs
- Pistol Range AOPC
 - \circ Soils Lead
 - o Lagoon Sediments -Lead
- Bulb Slope AOPC Lead



Based on these screening results, COPCs at the Landfill and Sandblast Area AOPCs were carried forward to the quantitative risk assessment. Lead was evaluated for all four AOPCs for adult and child.

2.1.2 COPC Selection for the Nursing Infant

The nursing infant was added as a new receptor to the BHHRA, following the publication of ODEQ's HHRA guidance (ODEQ 2010a). This receptor may be exposed to selected bioaccumulative COPCs in maternal milk during the first year of life and is recommended for inclusion as a receptor at any site where maternal exposure to selected bioaccumulative chemicals may occur, particularly by dietary and ingestion pathways.

ODEQ (2010a) published screening-level infant risk adjustment factors (IRAFs) for PCBs and for pesticides DDx (i.e., DDT, DDD, and DDE), which are presented and discussed later in Section 2.4. The nursing infant's cancer risks are typically estimated to be similar to, or lower than, the mother's risks. The noncancer hazard quotient (HQ) for the infant, however, may significantly exceed the mother's adult exposure-based HQ for some chemicals such as PCBs (ODEQ 2010a, Appendix D).

The IRAFs from ODEQ (2010a) were used to modify the residential and commercial RSLs to derive SLVs for PCBs and DDx, which were used to screen for potential Nursing Infant COPCs at each of the four AOPCs (see Table 2-6). Since PCB data for the Upland OUs are available as Aroclor data, the IRAF for PCBs as Aroclors was used. The modified residential RSLs were used to represent the Hypothetical Fishing Platform User, and the modified commercial RSLs were used to represent the occupational exposures.

DDx was eliminated as a potential Nursing Infant COPC for all receptors at all four AOPCs. PCBs were retained as a Nursing Infant COPC for the Hypothetical Fishing Platform User at both the Landfill and Sandblast Area AOPCs, but eliminated for Pistol Range and Bulb Slope AOPCs. PCBs were eliminated for occupational exposure at all four AOPCs.

2.2 Receptors and Exposure Pathways Evaluated in BHHRA

The eastern portion of Bradford Island is a restricted-access area managed by USACE. Due to the industrial nature of land use at the site, there are no residential receptors. On-site groundwater is not used at Bradford Island for any purpose. Potable use of groundwater is infeasible under both current and future conditions (Bob Schwarz [ODEQ], pers. comm. 2014).

Activities by employees on the eastern side of the Island include grounds maintenance, vegetation clearing, painting, building maintenance, and administrative duties. The grounds crew typically works up to 3 days per week during peak season. The building crews typically work up to 40 hours per week. The workers are supplied with drinking water. Construction activities (less than 1-year duration) and short-term excavation involving utility repair or other types of soil-disturbing activities in a trench setting may occur.

Although there are no such current uses, fishing platform activities by tribal members may result in exposures to shallow soils by adults and children at any of the AOPCs in the future. Exposure pathways may include incidental ingestion, dermal contact, and inhalation of dusts and vapors from shallow soils (0-3 ft bgs).



The human health conceptual exposure models (CEMs) depicted in Figures 2-1 through 2-4 present potential exposures of humans to various media (i.e., soil gas, soil, and groundwater). Those receptors and pathways that were recommended for further evaluation at the end of the RI are described below.

2.2.1 Landfill AOPC

The landfill is fully vegetated with no bare soils. Vegetation clearing occurs along a 3 to 4 ft wide strip along roadsides when needed. There are no built or occupied structures at the landfill or in its vicinity. There are no current plans to construct enclosed structures at this AOPC nor is it considered feasible from an engineering standpoint. Therefore, vapor intrusion pathways for indoor inhalation are incomplete and will remain incomplete.

Receptors who may be exposed to COPCs in soil and groundwater (incidental exposure in trench setting only) at the Landfill AOPC (Figure 2-1) and will be evaluated in the BHHRA include the following:

- Adult Outdoor Maintenance Worker engaged in site maintenance activities that do not involve a significant degree of soil disturbance (e.g., landscape workers). These receptors at the landfill may be exposed to COPCs in surface soil (0-3 ft bgs) by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil.
- **Construction Workers** may be exposed to COPCs in surface and subsurface soil (0-10 ft bgs) by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil. Occasionally, Construction Workers and short-term excavation workers also may be exposed to COPCs in shallow groundwater by incidental ingestion and dermal contact if they undertake activities at depths that may bring them in contact with shallow groundwater due to groundwater elevations as shallow as 6 ft bgs (as noted in the Final RI Appendix B logs and Appendix D tables [URS 2012]). These pathways were evaluated by assuming that a worker may be in a trench or excavation subject to pooling water seeping in from the shallow groundwater. The exposure assessment assumes employment of safe work practices will prevent anything more than incidental contact to groundwater. Based on the COPCs identified in the initial screening, at the Landfill AOPC, workers in a trench may come into dermal contact with metals and phthalate COPCs in groundwater and may experience some minimal incidental ingestion of groundwater, dermal contact, and inhalation of vapors volatilizing from groundwater.
- **Hypothetical Fishing Platform Users** were presumed to make temporary camps near their fishing locations, and their exposure would be limited to surface soil (0-3 ft bgs). For the Upland OU, these receptors at Bradford Island may be exposed to COPCs in surface soil by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil. Since fishing excursions may include all the members of the household, the exposure for all the age groups for this receptor (adult, child, and nursing infant) were evaluated.

2.2.2 Sandblast Area AOPC

The Sandblast Area AOPC is a mixture of vegetated and non-vegetated areas. The former Sandblast Building has been demolished, and the former Hazardous Material Storage Area



(HMSA) exists only as a lean-to area that is open to the outdoors. The two remaining buildings in the vicinity are the Service Building and the Equipment Building, and both are in active occupational use. The Service Building provides work space for approximately 27 people and may support additional full-time employees in the future. The Equipment Building includes office space for two people. The receptors and pathways for the Sandblast Area AOPC (Figure 2-2) include:

- Adult Outdoor Maintenance Worker engaged in site maintenance activities that do not involve a significant degree of soil disturbance (e.g., landscape workers). These receptors at the Sandblast Area AOPC may be exposed to COPCs in surface soil (0-3 ft bgs) by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil.
- Construction Workers and Short-term Excavation Workers may be exposed to COPCs in surface and subsurface soil (0-10 ft bgs) by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil. Occasionally, they also may be exposed to COPCs in shallow groundwater by incidental ingestion and dermal contact. Based on the results of the RI screening, incidental contact with vanadium in groundwater within a trench scenario at the Sandblast Area AOPC is evaluated in this BHHRA.
- Adult Indoor Workers may be exposed to VOCs emanating from soil gas entering the indoor environment by vapor intrusion in the existing Service Building and Equipment Building. Vapor intrusion pathways would also be complete if new buildings were to be constructed at this AOPC in the future.
- **Hypothetical Fishing Platform Users** were presumed to make temporary camps near their fishing locations; therefore, their exposure will be limited to surface soil. Exposure is assumed to be in an outdoor setting away from the buildings, and any incidental exposure to VOCs emanating from the ground would be intermixed with ambient air and insignificant. These receptors may be exposed to COPCs in surface soil (0-3 ft bgs) by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil. Since fishing excursions may include all the members of the household, the exposure for all the age groups for this receptor (adult, child, and nursing infant) are evaluated.

2.2.3 Pistol Range AOPC

In the Final RI (URS 2012), no potential COPCs were identified for occupational receptors at the Pistol Range AOPC. The Hypothetical Fishing Platform Users may access all of the Upland OU and, therefore, the Pistol Range AOPC was also considered (Figure 2-3). As noted earlier, this receptor is presumed to camp near fishing locations and may be exposed to contaminants in surface soil (incidental ingestion, dermal contact, and outdoor inhalation). Since fishing excursions may include all the members of the household, the exposure for all the age groups for this receptor (adult, child, and nursing infant) are evaluated.



2.2.4 Bulb Slope AOPC

In the Final RI, no potential COPCs were identified for occupational receptors at the Bulb Slope AOPC. The Hypothetical Fishing Platform Users may access all of the Upland OU and, therefore, the Bulb Slope AOPC was also considered (Figure 2-4). As noted earlier, this receptor is presumed to camp near fishing locations and may be exposed to contaminants in surface soil (incidental ingestion, dermal contact, and outdoor inhalation). Since fishing excursions may include all the members of the household, the exposure for all the age groups for this receptor (adult, child, and nursing infant) are evaluated.

2.3 Exposure Assessment

Quantifying exposure involves estimating chemical intake rates based on the evaluation of chemical releases from the site and estimation of EPCs for specific pathways.

The methods for calculating potential chemical intakes from soil, groundwater, and soil gas (vapor intrusion into indoor air) for the populations and exposure pathways selected for quantitative evaluation primarily followed USEPA guidance and also considered ODEQ guidance, as appropriate. Exposure factor values are drawn from the Risk Assessment Guidance for Superfund (RAGS) Part A (USEPA 1989) and all succeeding guidance documents including USEPA guidance (2014) and the most current updates to exposure factors in developing RSLs (USEPA 2015a). ODEQ's HHRA guidance, Vapor Intrusion guidance, and current tables for Calculating RBCs for Individual Chemicals were also consulted (ODEQ 2010a, 2010b, 2012).

There is generally close agreement between USEPA and ODEQ for the exposure factor values. Where exposure factor values are not available from USEPA, ODEQ's recommended values were used.

2.3.1 Exposure Point Concentrations

The EPC is a chemical-specific and media-specific value that represents a reasonable maximum exposure (RME) or central tendency exposure (CTE) estimate of the concentration to which a receptor is exposed. In accordance with USEPA guidance (USEPA 1989), and to address comments received on the Final RI (URS 2012), both RME and CTE estimates were used in the BHHRA in the following manner: RME risk estimates were developed for all receptors and pathways that were evaluated. CTE estimates were developed only for those receptors and pathways where the RME risk or hazard levels exceeded *de minimis* risk levels (cancer risk of 1 x 10^{-6} or hazard index [HI] of 1), as described in Section 2.5. If the RME-based risk levels were below *de minimis* risk levels, CTE estimates were not developed. The following sections discuss development of EPCs for specific media.

Data management rules and procedures are presented in detail in the Final RI (URS 2012) and summarized below.

2.3.1.1 Soil

As described in Appendix A of the RI/FS MP (URS 2007) and in the WP Update (URS 2014; Appendix C) and in accordance with the most recent USEPA guidance regarding statistical methodology to be used in EPC estimation (USEPA 2002a), the 95% UCL (soil and groundwater for trench setting) was used as the EPC representing the RME. In general, if duplicates exist, the



average of the duplicate results is used as a single data point. As described in Section 5.1 of the Final RI (URS 2012), detected values from field duplicates were averaged with corresponding primary detected sample values to create a single result; if the analyte was detected in only one of the pair, the detected result was used.

Where samples sizes are less than eight, the maximum or single location data may be used, as appropriate. As previously agreed, the 95% UCL is also acceptable to ODEQ and is more conservative than ODEQ's suggested 90% UCL. The EPC representing the CTE scenario also used the 95% UCL (USEPA 1992b). The EPCs were estimated using statistical methods and values recommended by USEPA's ProUCL software, as represented in the Final RI (URS 2012). EPCs for soil were based on 0 to 3 ft bgs depth interval for Outdoor Maintenance Workers, and 0 to 10 ft bgs depth interval for Construction Workers. The soil exposure depth for the Hypothetical Fishing Platform User is assumed to be 0 to 3 ft bgs. This is consistent with ODEQ (2010a) guidance for developing exposure point concentrations for residential and occupational uses (other than construction activities). Therefore, soil data from the 0 to 3 ft bgs interval were used to evaluate the Hypothetical Fishing Platform scenario. There were two exceptions to this usage:

- 1) At the Pistol Range and Bulb Slope AOPCs, the total soil depth extends only to 1.5 ft bgs and 1 ft bgs, respectively. These data were used to represent EPCs for all receptors at these two AOPCs since they represent the entire soil column at these AOPCs.
- 2) At the Sandblast Area AOPC, additional data were collected from the 0 to 1 ft bgs interval for lead alone, using methodologies based on USEPA's guidance for lead, as described in the WP (URS 2007). These lead data from sieved and unsieved soil samples were used for the blood-lead modeling.

2.3.1.2 Groundwater

Maximum detected concentrations in groundwater were used to represent incidental ingestion and dermal contact with groundwater by an excavation/trench worker.

The groundwater screening performed in the RI (URS 2012) utilized ODEQ RBCs that only consider dermal contact in their groundwater exposure in a trench. To be consistent with general USEPA practices, the evaluation of the Trench/Excavation Worker at the Landfill and Sandblast Area AOPCs includes an evaluation of inhalation of VOCs volatilizing from groundwater, even though VOCs were not identified as dermal COPCs in groundwater for the trench setting. The Virginia Department of Environmental Quality (VDEQ 2014) trench model was used to derive trench air concentrations for VOCs in groundwater. Details of the VDEQ model and model tables are presented in Appendix B-1. In summary, the model assumed a trench dimension of 3 ft wide and 8 ft deep, which yielded a width to depth ratio of 0.38. This ratio corresponded with the use of the most conservative exchange rate of 2 air changes per hour (ACH), which is similar to an indoor space.

2.3.1.3 Soil Gas

Vapor intrusion is a dynamic exposure pathway with multiple factors that influence the concentrations in indoor air. Ever-changing conditions such as weather and groundwater flow, variations in elevation, and building interior climate changes due to heating, ventilating, and air conditioning, and even individual habits such as opening a window or closing an office door could directly and dramatically change a receptor's exposure. Static features such as the design



and condition of the building foundation, utilities design, and the site-specific soil properties are all key factors that either mitigate or enhance the migration of vapors from the subsurface into the office buildings. As described in the WP Update (URS 2014), a lines-of-evidence approach that included vapor intrusion modeling was used to evaluate this pathway, consistent with USEPA and ODEQ guidance.

The following constitute the lines-of-evidence approach for soil gas exposures:

- Develop site-specific risk estimates based on the Johnson and Ettinger (Vapor Intrusion) Model (JEM).
- Consider VOC data from all the media focusing on spatial correlations.
- Consider other site data such as historical activity and identified sources.
- Consider the strengths and uncertainties associated with the results.

Modeled Risk Estimates

Risk estimates were quantified using USEPA's current versions of the JEM for soil gas data for deriving site-specific attenuation factors (AFs) (USEPA 2004a, 2004b, 2013c). Model tables are presented in Appendix B. The AF is the ratio of the concentration in indoor air to the concentration in soil gas. The JEM-derived AF incorporates the attenuating effects of the subsurface soil and building foundation/floor and the diluting effects of the indoor space, but does not account for biodegradation of the soil gas. The AFs were used to estimate the indoor air concentration or EPC for the risk calculations as follows:

Indoor Air Concentration = Soil Gas Concentration x AF

The two existing buildings at the Sandblast Area AOPC and a hypothetical future building were modeled using JEM. A combination of USEPA default and site-specific values (i.e., dimensions of existing inhabited buildings and available site-specific soil parameters) were used to represent building and soil properties.

The data from the soil gas sample locations that were closest to the existing buildings were used to represent the EPC for those buildings. This included location SB-13 to represent the Service Building and location SB-14 to represent the Equipment Building. The maximum detected concentrations from all five soil gas samples were used to represent future buildings. The maximum concentrations for PCE, TCE, and vinyl chloride were reported at location SB-12, near the former Sandblast Building, which has now been demolished. In keeping with USEPA methodology, only detected chemicals were included in the vapor intrusion risk estimation. However, inclusion of non-detected degradation compounds was also considered per ODEQ guidance and as described in Section 2.1. These results are presented in the Uncertainty Assessment (Section 2.6).

Spatial Comparison of VOC Detections in All Media

Within each medium (i.e., soil, soil gas, and groundwater), the detections of VOCs were evaluated spatially to identify areas or depths of elevated concentrations. From the Sandblast Area AOPC, there were five soil gas samples (Table 1-6 and also Figure 9-8 of the RI), 49 soil samples for depth range 0-10 ft bgs, of which 46 were surface samples of 0-3 ft bgs (Table 1-2 and also Figures 9-5j and 9-5k of the RI), and10 groundwater samples (Table 1-8 and also Figure



9-7b of the RI). Next, the VOC detection locations were compared between all three media at the Sandblast Area AOPC for patterns or correlations.

Other Site Data and Overall Vapor Intrusion Assessment

The historical activities and RI findings were reviewed to identify any correlations to VOC sample data. The nature and location of known VOC sources were reviewed in relation to the available data in all three media.

Strengths and Uncertainties

All of the various lines of evidence were evaluated comprehensively to characterize receptor risk to vapor intrusion concerns. The strengths and weaknesses associated with convergence or divergence of the findings drawn from each line of evidence were described and final conclusions were drawn.

2.3.2 Exposure Factors

Tables 2-7a and 2-7b list the exposure factors for occupational exposures and Hypothetical Fishing Platform User exposures, respectively, that were used in the BHHRA for the receptors noted above. USEPA (2014, 2015a) RME exposure factors were used as the primary source. When USEPA values are not available, such as for the CTE exposure factors, ODEQ values are used. USEPA updated its exposure factors in June of 2015 (USEPA 2015a), after consideration of recommendations developed in the Exposure Factors Handbook (USEPA 2011). Where different, the updated factors are used in place of those listed in the WP Update (URS 2014).

2.3.2.1 Occupational Exposures

The selection of RME and CTE values generally exceeds the likely exposures of on-island personnel with respect to exposure duration, EF, and other uptake parameters (Table 2-7a). For example, there are both part-time and full-time employees at Bradford Island. Also it is doubtful that workers will truly have an exposure duration of 25 years on the Island. Many of the maintenance staff are seasonal or have part-time work schedules, as described in Section 2.2.

The key differences between the exposure assumptions for the CTE Outdoor Maintenance Worker are 6 years of exposure compared to 25 years for the RME, half the soil ingestion rate (IRS), and a lower soil adherence factor (ODEQ 2010a). For the Construction Worker, the CTE scenario assumes a half year of exposure compared to 1 year for the RME, a third of the IRS, and a lower soil adherence factor (ODEQ 2010a).

For outdoor inhalation, a regionally site-specific value for Q/C (defined as the inverse of the mean concentration at the center of a 0.5 acre-square source) was used in place of the nation-wide default value in the calculation of the volatilization factor (VF). The Q/C measured in Salem, Oregon, was used as recommended by USEPA (2002b).

Another factor involved bioavailability of arsenic. As noted in Sections 9.2.1 and 9.3 of the Final RI (URS 2012), arsenic in Landfill and Sandblast Area soils is considered to be naturally occurring, although it is slightly higher than background levels. A relative bioavailability factor (RBA) of 60% was applied to arsenic in soil, following USEPA recommendations (USEPA 2012a, 2012b). In the absence of site-specific bioavailability data, this represents an upper-bound estimate by USEPA, in recognition of the fact that soil-bound arsenic is not 100% bioavailable.



Lead is a special case COPC with toxicity evaluated based on statistical probabilities of exceeding target blood lead levels instead of standard cancer risk and noncancer hazard estimates. Exposure to soil-borne lead is derived primarily from incidental ingestion of lead associated with fine soil particles that may adhere to the hands and face when a person comes in direct contact with soil. USEPA guidance (USEPA 2003b) recommends analysis of lead in sieved samples (<250 micrometers [μ m]) from shallow surficial soils to obtain the most likely exposure concentrations. Therefore, the EPC to evaluate lead exposure was developed from shallow surface samples (0-1 and 0-3 ft bgs).

2.3.2.2 Hypothetical Fishing Platform User Exposures

The exposure factors used for the Hypothetical Fishing Platform User scenario generally corresponded to the RME values used for residential receptors under current USEPA guidance (USEPA 2014, 2015a). They included 6 years of childhood and 20 years of adult exposure duration for 24 hours a day. One factor that is more conservative than USEPA usage is the assumption of 365 days/year EF. The CTE values differed from the RME mainly by less exposure duration (3 years for adult), less EF (152 days per year), less conservative soil adherence factors, and a lower IRS (50 milligrams (mg) per day for adult and 100 mg/day for child). The exposure factor values are provided in Table 2-7b. Similar to the other receptors in Section 2.0, each AOPC was evaluated independently, i.e., the receptors were conservatively assumed to be present at each AOPC for 100% of the exposure duration.

For the Pistol Range and Bulb Slope AOPCs, the only retained COPC was lead, which was evaluated differently from other chemicals (see Section 2.4.1).

2.3.3 Dose Calculations

The overall dose for each receptor and exposure pathway depends on receptor specific exposure factors (Table 2-7a and Table 2-7b) and the concentration of the chemical in the exposure medium. Once the dose is calculated, it can be applied to chemical specific toxicity data to estimate either cancer risk or noncancer health hazard. Example equations for a worker (w) are shown below.

The following equation is an example of the cancer risk and noncancer health hazard dose equations and risk equations used in the BHHRA (in this case, soil ingestion by a worker). The equations and variables vary for the different receptors, media, and exposure pathways and are presented and defined in their respective tables in Appendix A.

$$\begin{split} Risk &= SF_{o} \times \left\{ C_{soil} \times \left[\frac{IRS_{w} \times EF_{w} \times ED_{w} \times FI_{w} \times CF_{o}}{BW_{a} \times AT_{c} \times 365 day / year} \right] \right\} \\ Hazard &= \frac{1}{RfD_{o}} \times \left\{ C_{soil} \times \left[\frac{IRS_{w} \times EF_{w} \times ED_{w} \times FI_{w} \times CF_{o}}{BW_{a} \times AT_{nc,a} \times 365 day / year} \right] \right\} \end{split}$$



Where,

| SFo | = | Slope Factor (oral): (mg per kilogram [kg]-day) ⁻¹ |
|----------------------------|-----|---|
| C_{soil} | = | COPC Concentration in soil: mg/kg |
| IRS_{w} | = | Soil Ingestion Rate (worker): mg/day |
| EF_{w} | = | Exposure Frequency (worker): day/year |
| ED_{w} | = | Exposure Duration (worker): years |
| FI_{w} | = | Fraction Ingested (worker): unitless |
| CFo | = | Conversion Factor (oral): units dependent |
| BW_a | = | Body Weight (adult): kg |
| AT_{c} | = | Average Time (cancer): days |
| RfD_{o} | = | Reference Dose (oral): mg/kg-day |
| AT _{nc,a} | . = | Average Time noncancer (adult): days |

Nursing Infant

The Nursing Infant evaluation does not calculate dose for the infant; rather to calculate the infant's cancer risk and noncancer hazard, the mother's cancer risk and noncancer HI is modified by an IRAF that is both chemical- and pathway-specific. The Nursing Infant cancer risk and noncancer hazard are calculated using the following equation from ODEQ (2010a):

Infant Cancer Risk = Mother's Risk x $IRAF_c$

Infant Noncancer Hazard = Mother's HQ x $IRAF_{nc}$

For certain scenarios where only the child noncancer HQ is available, the child HQ is modified using the ODEQ Residential Soil (Direct Contact) IRAFs. See Section 2.4.7 for further discussion on IRAFs.

Relative Bioavailability

An RBA of 60% was applied to arsenic in soil, using current USEPA defaults (USEPA 2012a, 2012b). Although other COPCs in soil may also have less than 100% bioavailability, no other bioavailability factors were applied. This is discussed further in the Uncertainty Section (Section 2.6).

2.4 Toxicity Values

The selection of toxicity values followed the hierarchy of sources that is recommended by USEPA (2003a) and represented in the listing provided in USEPA (2015a), as follows:

- Integrated Risk Information System (IRIS) (USEPA)
- Provisional Peer-Reviewed Toxicity Values (PPRTVs) (USEPA)
- Minimal Risk Levels (Agency for Toxic Substances Disease Registry [ATSDR])
- Chronic Reference Exposure Levels (California Environmental Protection Agency)
- Appendices to PPRTVs (USEPA)
- Health Effects Assessment Summary Tables

This hierarchy follows ODEQ's hierarchy with some minor variations as discussed below in Sections 2.4.1 to 2.4.5. Toxicity values for most of the COPCs are available from IRIS or as



PPRTVs. A few selected COPCs are discussed in more detail below. The toxicity values used in the BHHRA are presented in Table 2-8 and within receptor- and media-specific tables of the risk calculations tables in Appendix A.

2.4.1 Lead

The evaluation of lead as a COPC differed for occupational exposures and Hypothetical Fishing Platform User exposures. Appendix B presents the lead models and their supporting tables and output figures.

Occupational Exposure

Exposure to lead in Sandblast Area AOPC soils was evaluated using the current version of the Adult Lead Model, as recommended by USEPA (2015a, undated). Risk from exposure to lead in surface soil by outdoor maintenance workers was evaluated using USEPA's adult lead model (ALM) spreadsheet (USEPA 2009a). Four exposure scenarios were evaluated, including Outdoor Maintenance Workers exposed to:

- Surface material (0-1 ft bgs) in the Landfill AOPC
- Surface and subsurface material (0-3 ft bgs) in the Landfill AOPC
- Surface material (0-1 ft bgs) in the Sandblast Area AOPC
- Surface and subsurface material (0-3 ft bgs) in the Sandblast Area AOPC

Surface soil from the Sandblast Area AOPC was sieved to $<\!\!250~\mu m$ (USEPA 2000) and then analyzed for lead.

USEPA (2003b) default parameter values for commercial/industrial workers were used. The arithmetic mean lead concentrations were used as the EPCs, as recommended by USEPA (2010). Updated ALM default values for baseline blood lead concentration (PbB0) (1.0 microgram [µg] per deciliter [dL]) and individual blood lead geometric standard deviation (GSDi) (1.8) were used in the ALM as recommended by USEPA (2009b). PbB0 and GSDi values have been updated by USEPA (2009b) based on an analysis of more recent blood lead concentration (PbB) data (USEPA 2002e) from the Third National Health and Nutrition Examination Survey, which showed decreases in PbBs among age and ethnic groups. USEPA's Technical Review Workgroup (TRW) has indicated that the updated PbB0 and GSDi values are appropriate for lead RAs for non-residential exposures both in assessing risk and in developing preliminary remediation goals (PRGs) for sites (USEPA 2010).

A target fetal PbB of 10 μ g/dL was recommended by USEPA (2003b) for use in the ALM, based on the assumption that the PbB of concern for fetuses is the same as that for children. The National Research Council (1993) has supported this PbB of concern for fetuses. More recent concerns that adverse effects may occur in children at levels below 10 μ g/dL may also pertain to the ALM. USEPA does not have a PbB of concern for teenagers or adults. USEPA's ALM (USEPA 2003b) was used to provide conservative estimates of PbBs in fetuses of women of child-bearing age who are exposed to lead in surface material. PbBs that are protective of fetuses and children are also considered protective for teenagers and adults (USEPA 2010).

However, the TRW has cautioned that lead RAs that include non-residential land use should report new toxicity information from USEPA (2006) indicating adverse health effects at PbBs



below 10 μ g/dL down to 5 μ g/dL and possibly lower (USEPA 2010). The new toxicity information for lead has not, as yet, been incorporated into any USEPA methodologies for evaluating the exposure of children or adults to lead. USEPA currently considers 10 μ g/dL to be the PbB of concern for fetuses and children. For non-residential settings, the U.S. Center for Disease Control (CDC) has replaced its former Level of Concern of 10 μ g/dL with a reference value of 5 μ g/dL (CDC 2012).

The ALM predicts the chance that fetuses of adults exposed to surface material at the site would have a PbB exceeding a level of concern. USEPA's target for women of child-bearing age is to limit the risk to a typical developing fetus of a woman exposed at a site to no more than a 5% chance of exceeding the 10 μ g/dL PbB of concern (USEPA 2003a).

To be conservative, the non-residential ALM model was run with both the USEPA target PbB and the CDC Level of Concern PbB.

Hypothetical Fishing Platform User Exposure

Lead was evaluated for adults using the ALM as noted above and for children using the Integrated Exposure Uptake Biokinetic (IEUBK) Model for Lead in Children. USEPA (2009a) default parameter values for were used for the adult resident, with the exception of the EF value. ALM guidance recommends using CTE values for EF (e.g., the CTE value for EF for residents is 234 days/year, USEPA 1993). However, to be protective and consistent with the EF used to evaluate non-lead chemicals at the site, an EF of 365 days/year was used in the ALM.

USEPA (2010) default parameter values in the IEUBK model were used for the child resident. The exception was that the EPCs for lead in soil and sediment in AOPCs were used as the outdoor soil concentration instead of the default value of 200 mg/kg. In addition, the concentration of lead in indoor household dust was estimated using the model's Multiple Source Analysis, which assumes that the concentration of lead in indoor household dust is 0.7 times the concentration of lead in outdoor soil plus the contribution from soil in outdoor air (10 mg/kg). For example, in the Sandblast Area AOPC, the EPC for lead in soil was 300 mg/kg. The 300 mg/kg value was entered into the IEUBK model as the outdoor soil concentration. The concentration of lead in indoor household dust was assumed to be 220 mg/kg, based on the following equation:

$$220 \text{ mg/kg} = (300 \text{ mg/kg x } 0.7) + 10 \text{ mg/kg}$$

To be protective, young children were also assumed to be exposed to background levels of lead in air at the residence, background lead in the diet, and background lead in drinking water. Risk from exposure to lead in soil or sediments by child receptor was evaluated using USEPA's IEUBK model (USEPA 2010).

Ten exposure scenarios were evaluated, including exposure of child and adult to:

- Soil (0-1 ft bgs) in the Landfill AOPC
- Soil (0-1 ft bgs) in the Sandblast Area AOPC
- Soil in the Pistol Range AOPC
- Sediment in the Pistol Range AOPC
- Soil in the Bulb Slope AOPC



2.4.2 Polycyclic Aromatic Hydrocarbons

cPAHs were evaluated by the use of Toxicity Equivalence Factors (TEFs) relative to benzo(a)pyrene, as listed below from USEPA (2015a) and ODEQ (2010a).

| Compound | TEF |
|-------------------------|-------|
| Benzo(a)pyrene | 1.0 |
| Benz(a)anthracene | 0.1 |
| Benzo(b)fluoranthene | 0.1 |
| Benzo(k)fluoranthene | 0.01 |
| Chrysene | 0.001 |
| Dibenz(a,h)anthracene | 1.0 |
| Indeno(1,2,3-c,d)pyrene | 0.1 |

PAHs are evaluated as individual chemicals and their cancer risk is presented individually and also summed separately from other COPCs to provide risk to cPAHs as a group.

2.4.3 Mutagenic Mode of Action

Certain cPAHs are considered to be mutagenic in activity and may be more potent during earlylife-stage exposures (USEPA 2005b). Mutagenic toxicity was assumed for the risk estimation process for cPAHs consistent with USEPA (2005b) and ODEQ recommendations (ODEQ 2010a) methods by applying age-dependent adjustment factors (ADAFs) to the dose equation to account for the greater vulnerability of younger age groups.

The following equation was used for mutagenic risk (See Section 2.3.3 for general definitions of terms and Tables 2-7a and 2-7b for receptor specific values):

$$\begin{aligned} Risk &= SF_{O} \times \left[C_{soil} \times EF_{r} \times FI \times CF_{O} \\ & \times \left(\frac{IRF_{c} \times ED_{0-2} \times ADAF_{0-2}}{BW_{c} \times AT_{c} \times 365 \ days/year} + \frac{IRFc \times ED_{2-6} \times ADAF_{2-6}}{BW_{c} \times AT_{c} \times 365 \ days/year} + \frac{IRF_{a} \times ED_{6-16} \times ADAF_{6-16}}{BW_{a} \times AT_{c} \times 365 \ days/year} + \frac{IRF_{a} \times ED_{16-26} \times ADAF_{16-26}}{BW_{a} \times AT_{c} \times 365 \ days/year} \right] \end{aligned}$$

The following ADAFs and EDs were used:

| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 | dimensionless |
|--|-----------------------|----|---------------|
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 | dimensionless |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 | dimensionless |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 | dimensionless |
| Exposure Duration, child 0-2 | ED ₀₋₂ | 2 | years |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 4 | years |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 | years |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 | years |



2.4.4 Trichloroethylene

Toxicity values for TCE are based on USEPA's current recommendations as listed in IRIS and USEPA (2015b). These include the following:

- SFo 4.6 x 10⁻⁰² (mg/kg-day)-1
- Inhalation Unit Risk -4.1×10^{-06} (µg per cubic meter [m³])-1
- Oral Reference Dose $-5.0 \times 10^{-04} \text{ mg/kg/day}$
- Inhalation Reference Concentration $-2.0 \times 10^{-03} \text{ mg/m}^3$

2.4.5 Benzofluoranthenes, Total

Total benzofluoranthenes were listed as a COPC in the WP Update (URS 2014) and represent an older sampling event that did not report the individual benzofluoranthenes. The newer data report individual benzofluoranthenes but do not report total benzofluoranthenes. Due to the lack of toxicity data and chemical properties for (total) benzofluoranthenes, benzo(b)fluoranthene values were used as a surrogate to evaluate total benzofluoranthenes.

The total benzofluoranthene data includes benzo(b), benzo(j), and benzo(k)fluoranthenes, with different toxicity values for each isomer. Benzo(b)fluoranthene has a SFo that is10 times more potent than benzo(k)fluoranthene.

In the initial risk screening, total benzofluoranthenes were assumed conservatively to consist entirely of the most potent benzo(b)fluoranthene. For this BHHRA, the total benzofluoranthene soil data was reviewed to derive a scaling factor to account for the percent of benzo(b)fluoranthene in the total benzofluoranthenes. Benzo(j)fluoranthene has not been detected at the site. Using individual benzo(b)fluoranthene and benzo(k)fluoranthene concentrations from the more recent data where individual benzofluoranthenes are reported, a percent benzo(b)fluoranthene of total benzofluoranthenes was calculated for each sample. Then the average percentage was derived for all samples (30 samples from both Landfill and Sandblast Area AOPCs). The percent benzo(b)fluoranthene was 60%. The scaling was applied as follows:

(Benzofluoranthene, Total EPC) x (60%) = New Benzofluoranthene, Total EPC

2.4.6 Chromium

Chromium was identified as a COPC in the Final RI (URS 2012) due to initial use of SLVs based on hexavalent chromium. Total chromium levels were similar in both Landfill (594 mg/kg) and Sandblast Area (579 mg/kg) soils. As noted in Section 9 of the Final RI, the source of the chromium is thought to be nickel-chromium coatings for equipment that was likely sand-blasted from the Sandblast Building. In groundwater at both AOPCs, total chromium was detected in a few samples at levels of 0.03 mg per liter (L) or lower. Speciation data for the total chromium are not available for either soil or groundwater. Trivalent chromium is the most stable form of chromium in the environment and has very low solubility (ATSDR 2012a). Dissolved chromium was not detected at all in groundwater at the Landfill AOPC and was detected at very low concentrations in groundwater at the Sandblast Area AOPC (maximum concentration of 4 μ g/L) (see Final RI [URS 2012], Tables I-1 and I-2). Based on these considerations, chromium in soil is most likely to be trivalent chromium and corresponding toxicity values for trivalent chromium



were used. The uncertainty related to the possibility of hexavalent chromium is discussed further in the Uncertainty Assessment.

2.4.7 Nursing Infant Risk

The Nursing Infant risk was evaluated using the ODEQ (2010a) approach of applying IRAF for PCBs to calculate maternal cancer risk and noncancer hazard. ODEQ (2010a) assumes that the concentration of a chemical in milk can be calculated from the long-term body burden in the mother and provides IRAFs for conversion of the mother's risk and hazard estimates to Nursing Infant risks and hazards.

The mother's cancer risks for the scenarios included in this BHHRA assume time-integrated exposure during childhood and adulthood. The IRAFs used in this risk assessment are as follows:

| Chemical | Residential IRAF Source | |
|-----------------------------|-------------------------|------------|
| Carcinogenic IRAF | | |
| Total PCB | 0.6 | ODEQ 2010a |
| Noncancer IRAF Total PCB | 4 | ODEO 2010a |

Nursing Infant Risk Adjustment Factors

Only PCBs were identified as COPCs. In general, total PCB cancer risk for the Nursing Infant would be lower than the mother. For total PCBs, the noncancer HQ for the Nursing Infant would be four times greater than the mother's hazard for total PCBs.

2.5 Risk Characterization

The BHHRA evaluated the Landfill, Sandblast Area, Pistol Range, and Bulb Slope AOPCs of the Upland OU. Both excess lifetime cancer risks and, if appropriate, noncancer hazards were estimated for the carcinogenic COPCs, for each combination of receptor, exposure medium and exposure scenario as presented in the WP Update (URS 2014). For non-carcinogenic chemicals, only noncancer hazards were estimated. The results were summed to provide quantitative estimates of multi-pathway and multi-media risks and hazards for each receptor for the RME and CTE scenarios. The CTE estimates are presented to provide a range and represent the average exposure whereas RME represents the more conservative, upper-bound reasonable maximum estimates. The estimated risks, HQs, and HIs are presented in the context of whether they were:

- 1) less than the USEPA risk level of 1×10^{-6} (also expressed as one in a million) or HI of 1, whereby risk at or below these thresholds have an insignificant contribution to risk (*i.e.*, *de minimis*);
- 2) within the USEPA's acceptable risk range of 1×10^{-6} to 1×10^{-4} for cancer and HQ or HI of 1 for noncancer COPCs (USEPA 1991); or
- 3) exceeding the USEPA acceptable risk range, i.e., greater than 1×10^{-4} .

In addition, ODEQ's acceptable risk thresholds of 1×10^{-6} for individual carcinogens and 1×10^{-5} for multiple carcinogens were also considered (ODEQ 2010a). Due to this, individual chemicals associated with risk levels greater than the ODEQ thresholds of 1×10^{-6} or noncancer



HQ greater than 1 were identified as COCs, except arsenic, for which background levels were also considered. Since the individual compounds that make up the cPAHs all act by the same mode of action, risks for PAHs were calculated by individual compound and also summed to present risks for cPAHs as a group. It is important to note that, since excess lifetime cancer risks (ELCRs) are only expressions of likelihood of cancer incidence, and HIs are estimated ratios to safe doses, exceedance of the ODEQ risk thresholds or USEPA acceptable risk range does not automatically mean that adverse effects may have occurred or will occur (e.g., does not automatically mean that the COC will be recommended for further evaluation in the FS).

Cancer risks are presented to one significant figure and noncancer hazards to two significant figures (ODEQ 2010a, USEPA 1989).Cancer risks are discussed using 1×10^{-6} , 1×10^{-5} , and 1×10^{-4} convention in the text and are represented using 1E-06, 1E-05, and 1E-04 scientific notation in the tables for clarity and legibility. The terms are equivalent to each other.

Tables 2-9 through Table 2-24 present summary cancer risk and noncancer hazards for each receptor. Table 2-25 presents a summary of the findings of the blood-lead evaluation. Table 2-26 summarizes the cancer risk and noncancer health hazard for each receptor by AOPC. Results are grouped by AOPC and presented by media and receptors for RME and then CTE. Where the cancer risk and/or the noncancer health hazard exceed the ODEQ cancer risk threshold of 1 x 10⁻⁶ risk or HI of 1, the primary contributors to risk are described further and may be considered as COCs recommended for further evaluation in the FS. Results of the lead evaluation are also presented for each AOPC with the modeling details presented in Appendix B.

2.5.1 Landfill AOPC

The BHHRA evaluated exposures to COPCs in soil and groundwater at the Landfill AOPCs. Three occupational receptors and the Hypothetical Fishing Platform User were evaluated, and Appendix A presents the equations and risk calculations for each receptor. The receptors and their summary cancer risk and noncancer health tables are listed below:

- Outdoor Maintenance Worker: RME Summary (Table 2-9)
- Outdoor Maintenance Worker: CTE Summary (Table 2-10)
- Construction Worker: RME Summary (Table 2-11)
- Construction Worker: CTE Summary (Table 2-12)
- Hypothetical Fishing Platform User: RME Summary (Table 2-13)
- Hypothetical Fishing Platform User: CTE Summary (Table 2-14)
- Excavation/Trench Worker, Exposure to Groundwater: RME Summary (Table 2-15)

2.5.1.1 Landfill AOPC Soil

Landfill soils were evaluated for risks to the Outdoor Maintenance Worker (0-3 ft bgs), Construction Worker (0-10 ft bgs), and Hypothetical Fishing Platform User (0-3 ft bgs).



2.5.1.1.1 Landfill – Outdoor Maintenance Worker

RME

As presented in Table 2-9, the Outdoor Maintenance Worker exposure to surface soil had RME cancer risk of 6 x 10^{-5} , which exceeds the ODEQ cancer threshold of 1 x 10^{-6} but falls within the USEPA's acceptable risk range of 1 x 10^{-6} to 1 x 10^{-4} . The noncancer HI was acceptable at 0.02. The cumulative risk level associated with cPAHs exceeds the ODEQ threshold for multiple carcinogens of 1 x 10^{-5} . The primary contributors to risk are arsenic and cPAHs.

Arsenic – Incidental ingestion was the only significant exposure route for arsenic, yielding 3 x 10^{-6} cancer risk, and was based on a surface soil EPC of 10.5 mg/kg. Following USEPA guidance (USEPA 2002c), the contributions from background arsenic and site-related arsenic were evaluated further. The Reference Area upper prediction limit (UPL) for arsenic is 5.4 mg/kg (Table 1-9). Therefore, the incremental risk from site-related arsenic is 1 x 10^{-6} , which is at the ODEQ threshold, as shown below:

| Data Grouping | Arsenic EPC (mg/kg) | Arsenic Cancer Risk | Arsenic HQ |
|------------------------|------------------------|------------------------|------------|
| RME Landfill | 10.5 | 3 x 10 ⁻⁶ | 0.02 |
| Reference UPL | 5.4 | 2 x 10 ⁻⁶ | 0.01 |
| Site-related Increment | 5.1 | 1 x 10 ⁻⁶ | 0.01 |

Summary of Total, Background, and Site-Related Risk for Arsenic Outdoor Maintenance Worker

Based on this comparison, arsenic is not considered a COC since the site-related contribution to risk is at acceptable levels and is barely distinguishable from background risk. This is consistent with USEPA guidance that further evaluation is not warranted for chemicals whose risks do not exceed background levels (USEPA 2002c, 2002d).

cPAHs – Ingestion and dermal contact with cPAHs were the significant exposure routes, with the ingestion route yielding approximately twice the risk as dermal contact. Benzo(a)pyrene was the primary contributor to risk among the cPAHs, and its contribution to risk from cPAHs is approximately an order of magnitude greater than any other cPAH or 60% of the risk from cPAHs (6×10^{-5}). Benzo(a)pyrene was detected consistently throughout the surface soil of the Landfill AOPC (31 detections out of 33 samples) with the EPC being 11.2 mg/kg as compared to the Reference Area 95% UPL of 0.037 mg/kg. The Reference Area UPL for all cPAHs was 0.051 mg/kg benzo(a)pyrene-equivalent carcinogenic potency (BaPeq) (Table 1-9).

Based on RME estimates, cPAHs are the only COC (i.e., primary contributors to risk) for surface soil in the Landfill AOPC. The individual PAH COCs are benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, dibenz(a,h)anthracene, indeno(1,2,3-cd)pyrene, and benzofluoranthenes, total.

CTE

As presented in Table 2-10, the Outdoor Maintenance Worker had CTE cancer risk of 6 x 10^{-6} , which exceeds the ODEQ threshold of 1 x 10^{-6} but falls within the USEPA's acceptable risk range of 1 x 10^{-6} to 1 x 10^{-4} . The noncancer HI was acceptable at 0.009. The primary contributor



to risk is one individual cPAH, benzo(a)pyrene, with ingestion risk yielding 3×10^{-6} , or 57% of the total cPAH risk (5×10^{-6}).

Exposure to Lead in Surface Soil

USEPA's target for women of child-bearing age is to limit the risk to a typical developing fetus of a woman exposed at a site to no more than a 5% chance of exceeding the 10 μ g/dL PbB of concern (USEPA 2003a). For both available surface soil depth ranges, the USEPA ALM estimated acceptable percent probability of exceeding the target blood lead concentration for the Outdoor Maintenance Worker due to exposure to surface soil (see Table 2-25 and the summary of results below). A complete description of the lead modeling is provided in Appendix B.

| Depth Range | Lead Concentration in Soil | Percent Chance of Exceeding Acceptable Threshold (10 µg/dL PbB) |
|-------------|-------------------------------|--|
| 0-1 ft bgs | 211 mg/kg | 0.013% |
| 0-3 ft bgs | 342 mg/kg | 0.032% |

Landfill Surface Soil Adult Lead Model (ALM) Summary

Notes: Appendix B presents the ALM inputs and outputs

In addition, it was shown that a typical fetus of an adult commercial/industrial worker exposed to lead in surface and subsurface material at the site would have a 0.65 - 1.3% chance of PbBs exceeding the CDC reference value of 5 µg/dL (CDC 2012). Therefore, lead does not pose a threat of unacceptable PbB levels to fetuses in adult commercial/industrial receptors exposed to surface and subsurface material in the Landfill AOPC (Appendix B).

2.5.1.1.2 Landfill – Construction Worker

RME

As presented in Table 2-11, the Construction Worker exposure to deep soil had a RME cancer risk of 5 x 10^{-6} , which exceeds the ODEQ threshold of 1 x 10^{-6} but falls within the USEPA's acceptable risk range of 1 x 10^{-6} to 1 x 10^{-4} . The cumulative cancer risk is below ODEQ's acceptable risk level of 1 x 10^{-5} . The noncancer HI of 1 was at the ODEQ and USACE threshold, primarily based on inhalation exposure for PCE. Due to the dilution and mixing that occur in outdoor situations and since the HI is at the risk threshold, this HI is acceptable and PCE was not selected as a COC.

The primary contributor to risk is benzo(a)pyrene through ingestion (3×10^{-6}) and dermal contact (1×10^{-6}) . No other cPAHs were COPCs for this receptor exposure to deep soil.

Based on RME estimates, the COC for subsurface soil in the Landfill AOPC is benzo(a)pyrene.

CTE

As presented in Table 2-12, CTE resulted in a low cancer risk of 7 x 10^{-7} , which is below the ODEQ cancer threshold and had an acceptable HI of 0.02.



2.5.1.1.3 Landfill – Future Hypothetical Fishing Platform User

RME

As presented in Table 2-13, the Hypothetical Fishing Platform User exposure to current surface soil concentrations resulted in a RME risk of 1×10^{-3} and HI of 1. Cancer risks exceeded the USEPA acceptable risk range of 1×10^{-4} to 1×10^{-6} . The primary contributors to risk were the cPAHs (primarily benzo(a)pyrene), with minor contributions from arsenic.

The noncancer hazard was at the noncancer threshold value of 1. No single chemical exceeded a HQ of 1. The greatest relative contributions to hazard came from total PCBs, arsenic, and mercury, although their individual HQs were less than 0.5. The target organs and effects associated with the non-cancer effects of these chemicals vary widely and are unlikely to result in cumulative effects. For example, arsenic may primarily affect the skin (hyperpigmentation, keratosis), mercury may affect the kidney (auto-immune effects), and PCBs may affect ocular glands (exudates) and digits (distorted fingers and toes) (USEPA 2015b). Based on the low potential for toxicity to individual target organs and an HI equal to the threshold, risk is considered insignificant.

Following USEPA guidance (USEPA 2002c), the contribution from background arsenic and siterelated arsenic were evaluated further. The Reference Area UPL for arsenic is 5.4 mg/kg, which is almost half of the arsenic EPCs for both the Landfill and Sandblast Area AOPCs. Therefore, the incremental risk from site-related arsenic is 1×10^{-5} , which exceeds the ODEQ threshold, but falls within the USEPA acceptable risk range.

The Nursing Infant receptor was evaluated for this receptor using the ODEQ (2010a) IRAF, as detailed in Section 2.4.7. As presented in Table 2-14, the RME Nursing Infant cancer risk was acceptable at 1×10^{-6} , but had an HI slightly greater than 1 (2) based on PCB exposure.

CTE

As presented in Table 2-14, the Hypothetical Fishing Platform User exposure to surface soil had a CTE risk of 2×10^{-4} and HI of 0.46. The cancer risks exceeded the USEPA acceptable risk range of 1×10^{-4} to 1×10^{-6} , but the noncancer HI was acceptable. The primary contributors to risk, similar to RME, were cPAHs. The Nursing Infant CTE cancer risk was acceptable at 1×10^{-7} and the noncancer HI was acceptable at 0.3.

Lead

Based on the lead evaluation, the probability of exceeding the target blood lead levels for the Hypothetical Fishing Platform User who may be exposed to Landfill soils was 0.034% for the fetus of an adult residential-like exposure and 0.36% for a child residential-like exposure for the evaluated data sets (Table 2-25). Therefore, lead concentrations are not a concern. See Appendix B for model details.

2.5.1.2 Landfill AOPC Groundwater

Excavation/Trench Worker

As presented in Table 2-15, the Excavation/Trench Worker RME exposure to contaminants in groundwater yielded cancer risk of 5×10^{-8} and noncancer HI of 0.04. Both are well below acceptable thresholds. This pathway considered incidental ingestion and dermal contact to



COPCs in groundwater and inhalation of VOCs that volatilized from the groundwater into trench air. CTE calculations were not performed due to low RME results and because this is a minor exposure pathway.

2.5.2 Sandblast Area AOPC

The BHHRA evaluated exposures to COPCs in soil, soil gas, and groundwater at the Sandblast Area AOPC. Four occupational receptors and the Hypothetical Fishing Platform User were evaluated and Appendix A presents the risk equations and calculations tables. The receptors and their summary cancer risk and noncancer health tables are listed below:

- Outdoor Maintenance Worker: RME Summary (Table 2-16)
- Outdoor Maintenance Worker: CTE Summary (Table 2-17)
- Construction Worker: RME Summary (Table 2-18)
- Construction Worker: CTE Summary (Table 2-19)
- Hypothetical Fishing Platform User: RME Summary (Table 2-20)
- Hypothetical Fishing Platform User: CTE Summary (Table 2-21)
- Excavation /Trench Worker, Exposure to Groundwater: RME Summary (Table 2-22)
- Indoor Worker, Exposure to Indoor Air: RME Summary (Table 2-23)
- Indoor Worker, Exposure to Indoor Air: CTE Summary (Table 2-24)

2.5.2.1 Sandblast Area AOPC Soil

2.5.2.1.1 Sandblast Area – Outdoor Maintenance Worker

RME

As presented in Table 2-16, the Outdoor Maintenance Worker exposure to surface soil had RME cancer risk of 1 x 10^{-5} , which exceeds the ODEQ threshold of 1 x 10^{-6} but falls within the USEPA's acceptable risk range of 1 x 10^{-6} to 1 x 10^{-4} . The noncancer HI was acceptable at 0.07. The cumulative cancer risk was greater than ODEQ's acceptable risk level of 1 x 10^{-5} (ODEQ 2010a). The primary contributors to risk are arsenic and cPAHs.

An evaluation of total, background, and site-related incremental risk was performed, similar to Landfill soils. The results are summarized below:

| Data Grouping | Arsenic EPC (mg/kg) | Arsenic Cancer Risk | Arsenic HQ |
|------------------------|------------------------|------------------------|------------|
| RME Sandblast Area | 9.7 | 3 x 10 ⁻⁶ | 0.02 |
| Reference UPL | 5.4 | 2 x 10 ⁻⁶ | 0.01 |
| Site-related Increment | 4.3 | 1 x 10 ⁻⁶ | 0.01 |

Summary of Total, Background, and Site-Related Risk for Arsenic Outdoor Maintenance Worker¹

¹Rounded values



Based on this comparison, arsenic is not considered a COC since the site-related contribution to risk is at acceptable levels (USEPA 2002c, 2002d).

Ingestion and dermal contact of cPAHs were the significant exposure routes with the ingestion route yielding approximately twice the risk as dermal contact. Risk from benzo(a)pyrene was about double that of benzofluoranthenes, total, and the overall cPAHs yielded a risk of 1×10^{-5} . Benzo(a)pyrene was detected consistently throughout the surface soil of the Sandblast Area AOPC (36 detections out of 40 samples) (Table 1-2) with an EPC of 2.4 mg/kg, as compared to the Reference Area UPL of 0.037 mg/kg.

Based on RME estimates, the only COCs for surface soil in the Sandblast Area AOPC are cPAHs, consisting of benzo(a)pyrene and benzofluoranthenes, total.

CTE

As presented in Table 2-17, CTE resulted in a low cancer risk of 1×10^{-6} , which is at the ODEQ cancer threshold, and an acceptable HI of 0.008.

Exposure to Lead in Surface Soil

For both available surface soil depth ranges, the USEPA ALM estimated acceptable percent chance of exceeding the target blood lead concentration for the Outdoor Maintenance Worker due to exposure to surface soil (see Table 2-25 and the summary of results below).

Sandblast Area Surface Soil Adult Lead Model (ALM) Summary

| Depth Range | Lead Concentration in Soil | Percent Chance of Exceeding Acceptable Threshold (10 µg/dL PbB) |
|-------------|-------------------------------|--|
| 0-1 ft bgs | 300 mg/kg | 0.025% |
| 0-3 ft bgs | 202 mg/kg | 0.013% |

Note: Appendix B presents the ALM inputs and outputs

Similar to the Landfill, it was shown that a typical fetus of an adult commercial/industrial worker exposed to lead in surface and subsurface material at the site would have a 0.65 to 1.1% chance of PbBs exceeding the CDC reference value of 5 μ g/dL (CDC 2012). Therefore, lead does not pose a threat of unacceptable PbB levels to fetuses in adult commercial/industrial receptors exposed to surface and subsurface material in the Sandblast Area AOPC (Appendix B).

2.5.2.1.2 Sandblast Area – Construction Worker

RME

As presented in Table 2-18, the Construction Worker exposure to sub-surface soil had a RME cancer risk of 2×10^{-6} , which exceeds the ODEQ threshold of 1×10^{-6} but falls within the USEPA's acceptable risk range of 1×10^{-6} to 1×10^{-4} . The noncancer HI was acceptable at 1. The cumulative cancer risk is less than ODEQ's threshold of 1×10^{-5} .

The primary contributor to risk is benzo(a)pyrene, with a cumulative multi-pathway cancer risk of 2×10^{-6} . No other cPAHs were COPCs for this receptor exposure to sub-surface soil.

Based on RME estimates, the only COC for subsurface soil in the Sandblast Area AOPC is benzo(a)pyrene.

CTE



As presented in Table 2-19, the CTE evaluation resulted in a low cancer risk of 3×10^{-7} , which is below the ODEQ cancer threshold, and an acceptable HI of 0.01.

2.5.2.1.3 Sandblast Area – Future Hypothetical Fishing Platform User

RME

As presented in Table 2-20, the Hypothetical Fishing Platform User exposure to current surface soil concentrations had a RME risk of 3×10^{-4} and HI of 1.8. At the Sandblast Area AOPC, cancer risks also exceeded the USEPA acceptable risk range. Non-cancer hazards were at or slightly higher than 1. COCs contributing to risk were similar to Landfill soils, consisting primarily of cPAHs, but with minor contributions from more chemicals, including arsenic, total PCBs, and DEHP. No individual COPC had a HQ greater than 1, but the cumulative evaluation resulted in a total HI of 2 due to arsenic, nickel, Total PCBs, PCE, and TCE. Similar to the noncancer chemicals at the Landfill, the target organs and effects associated with these chemicals are also variable. In addition to the previously described arsenic and PCBs, nickel may result in generalized decreased body weight, PCE may affect the nervous system (neurotoxicity), and TCE may affect the thymus gland (decreased weight), vascular system (decreased plaque cellforming response), and heart (cardiac malformation) (USEPA 2015b). Based on the low potential for toxicity to individual target organs, the marginal HI of 2 is considered acceptable.

The Nursing Infant receptor was evaluated for this receptor using the ODEQ (2010a) IRAF, as detailed in Section 2.4.7. As presented in Table 2-20, the Nursing Infant cancer risk was acceptable at 1 x 10-6 and had an HI of 2, slightly greater than the threshold of 1, based on PCB exposure.

CTE

As presented in Table 2-21, the Hypothetical Fishing Platform User exposure to surface soil had a CTE risk of x 10^{-5} and HI of 0.7. The cancer risk is within the USEPA acceptable risk range of 1 x 10^{-4} to 1 x 10^{-6} , and the HI was acceptable. The primary contributors to risk, similar to RME, were cPAHs. The Nursing Infant CTE HI was acceptable at 0.3.

Lead

Based on the lead evaluation, the probability of exceeding the target blood lead levels for the Hypothetical Fishing Platform User who may be exposed to Sandblast Area AOPC soils was 0.075% for adult residential-like exposure and 1.5% for child residential-like exposure for the evaluated data sets (Table 2-25). See Appendix B for model details.

2.5.2.2 Sandblast Area AOPC Groundwater

Excavation/Trench Worker

As presented in Table 2-22, the Excavation/Trench Worker RME exposure to contaminants in groundwater yielded cancer risk of 1×10^{-7} and noncancer health hazard of 0.7. Both are below ODEQ and USEPA risk thresholds. This pathway considered incidental ingestion and dermal contact to COPCs in groundwater and inhalation of VOCs, which volatilized from the groundwater into trench air. CTE calculations were not performed due the low RME results and because this is a minor exposure pathway.



2.5.2.3 Sandblast Area AOPC Soil Gas

The vapor intrusion pathway was considered for both current and potential future indoor exposures in an occupational setting. The approach included risk estimation based on modeled estimates of indoor air concentrations using soil gas data as well as a review of other site data and information.

2.5.2.3.1 Sandblast Area – Modeled Risk Estimates

Indoor Office Worker

The indoor office worker exposure to VOCs from soil gas was evaluated using multiple lines of evidence to account for the dynamic and highly variable vapor intrusion pathway. PCE and TCE were the two COPCs identified in soil gas. The USEPA JEM was used to derive AFs for each chemical for the various vapor intrusion exposure scenarios. The JEM input and output details are presented in Appendix B.

2.5.2.3.2 Sandblast Area – Current Existing Buildings Scenario

The current scenario used dimensions of the existing buildings and site-based soil type (i.e., sand) and detected COPC concentrations from the closest soil gas sampling locations (typically adjacent to the building).

As presented in Table 2-23, the indoor office worker at the Equipment Building (represented by data from soil gas location SB-14) has a RME risk of 3×10^{-8} and HI of 0.008, both of which are well below the ODEQ and USEPA risk thresholds. The CTE risk was 6×10^{-9} , with an HI of 0.008 (Table 2-24). The office building is noted as being a portable unit. The JEM does not account for buildings without a foundation which increases the uncertainties related to the derivation of the AFs for this receptor and are discussed in the uncertainty assessment section.

The indoor office worker at the Bradford Island Service Building (BISB) (represented by data from soil gas sample location SB-13) has a RME risk of 1×10^{-8} and a HI of 0.003, which are well below the ODEQ and USEPA risk thresholds (Table 2-23). The CTE risk was 3×10^{-9} and HI of 0.003 (Table 2-24). The BISB is a large L-shaped building having two distinct parts. One side is about 25 ft high with large open bay doors and is used as a paint and equipment shop. The other side is single story and houses the offices. Only the office side and its dimensions were used in deriving the attenuation factors.

In summary, risks and hazards were at acceptable levels for occupants of the current buildings.

2.5.2.3.3 Sandblast Area – Future Building Scenario

The future scenario used USEPA JEM conservative residential building dimension defaults, the most conservative soil type (sand), and the maximum detected concentrations of the COPCs in soil gas.

As presented in Table 2-23, the indoor office worker in a future building has a RME risk of 5 x 10^{-7} and HI of 0.1, which are below the ODEQ and USEPA risk thresholds. The CTE risk was 1 x 10^{-7} and HI was 0.14 (Table 2-24).



2.5.2.3.4 Sandblast Area – VOC Concentrations in All Media

In general, the highest concentrations of VOCs in all media are noted in samples taken from the vicinity of the Former Sandblast Building and the Current HMSA. This area also corresponds to the approximate location of a former aboveground storage tank (AST).

Soil Gas (See Figure 9-8 of the RI) had 5 sample locations (collected in 2009) within the Sandblast Area AOPC: two locations (SB-11 and SB-12) adjacent to the old Sandblast Building (no longer present), one location (SB-10) adjacent to the Current HMSA, one location (SB-14) adjacent to the Equipment Building, and one location (SB-13) adjacent to the BISB. PCE and TCE have been detected in all five samples with exceedances of screening levels at SB-12 and SB-10. The low number of soil gas samples limits the spatial analysis and shows fluctuating concentrations for PCE. The highest PCE concentration was reported as 34,000 μ g/m³ at SB-12, decreases to 800 μ g/m³ at SB-11 and 610 μ g/m³ at SB-10 to the south, increases to 1,800 μ g/m³ at SB-13 further southwest, and decreases to 730 μ g/m³ at SB-14 further south.

Surface Soil (See Figure 9-5j of the RI) had a relatively low number of PCE/TCE detections throughout the Sandblast Area AOPC. Two locations exceeding the SLV, HA4 and SBB18, are adjacent to the Current HMSA.

Deep Soil (See Figure 9-5k) sample locations appear to target the Erodible Unit Sampling Area and the Current HMSA. Although most samples had detections of PCE/TCE and some cis-DCE, none exceeded SLVs and were relatively low. The highest detections were locations (DP11 and DP12) near the Current HMSA.

Groundwater (See Figure 9-7b) data includes both monitoring well (MW) (collected in 2008 and 2009) and direct push samples (collected in 2004). Groundwater flows down northward towards the shore. There were extensive detections of PCE/TCE (and their degradation products) exceeding SLVs throughout almost all the sample locations. Concentrations were generally higher in direct-push samples than in monitoring well samples. The highest PCE/TCE detection location was at DP11 (2004), adjacent to the Current HMSA, and yet the nearby MW-11 location showed ND in years 2008 and 2009. The overall lateral trend shows decreasing concentrations moving away from the Current HMSA area, suggesting a historic source in groundwater near DP11.

There appears to be an association between soil (especially deep soil), groundwater, and soil gas data based on high detections at the Current HMSA and extending towards the former Sandblast Building. In all media, the highest concentrations were noted in this area with decreasing concentrations growing outward and away from the Current HMSA and former Sandblast Building.

Historical Site Activity

These VOC findings also align with historical activity in this area as described in the RI:

The former HMSA was located approximately 200 feet to the south of the former sandblast building and is sometimes also referred to as the 'former drum storage area.'... Before the construction of the current HMSA, an approximately 300-gallon AST was formerly located in the vicinity. Waste paints were temporarily stored in this AST until the late 1990s at which time the tank was removed... Analytical chemistry results for the soil sample identified the presence of several VOCs. From these results, it has been



inferred that there was a historical release from the AST formerly located in the vicinity of the current HMSA. (Section 3.5.1.2 of the Final RI [URS 2012])

The historical record indicates VOC contamination due to spills from containers storing hazardous materials near the Current HMSA. The data also suggest that the locations of the soil gas samples are well situated to capture the highest concentrations of VOC contamination for vapor intrusion concerns. There is no data to imply that VOC concentrations may increase in the future; therefore, the low risks estimated for the indoor office worker, both current and future, are expected to decrease further. There are no soil or groundwater data in the vicinity of soil gas locations SB-13 and SB-14. However, these locations are situated at a distance from the inferred AST location, and soil gas concentrations at these two locations were below all screening levels.

In summary, the multiple lines of evidence suggest a localized historic source of VOCs in the vicinity of the Current HMSA. The source area and the associated plumes appear to be undergoing degradation, and vapor intrusion-related risks are at acceptable levels both for occupants of current buildings and for future buildings.

2.5.3 Pistol Range AOPC

Lead was the only COPC identified at the Pistol Range AOPC and only for the Hypothetical Fishing Platform User receptor.

The ALM predicted that a typical fetus of an adult receptor exposed to lead in soil at the site would have 0.033% of the chance of PbB exceeding 10 μ g/dL (See Table 2-25 and Appendix B). The PbB prediction is well below USEPA's target for women of child-bearing age to limit the risk to a typical developing fetus to no more than a 5% chance of exceeding the 10 μ g/dL PbB level of concern (USEPA 2003b). Therefore, lead does not pose a threat to fetuses of adult receptors exposed to soil or sediment in the Pistol Range AOPC.

The IEUBK predicted that a typical child receptor exposed to lead in soil at the site would have an approximate 0.34% chance of PbBs exceeding 10 μ g/dL (Table 2-25 and Appendix B). The PbB prediction is well below USEPA's target to limit the risk to a typical child to no more than a 5% chance of exceeding the 10 μ g/dL PbB level of concern (USEPA 1994). Therefore, lead does not pose a threat to child receptors exposed to soil in the Pistol Range AOPC.

2.5.4 Bulb Slope AOPC

Lead was the only COPC identified at the Bulb Slope AOPC and only for the Hypothetical Fishing Platform User receptor.

The ALM predicted that a typical fetus of an adult receptor exposed to lead in soil at the site would have 0.038% of the chance of PbB exceeding $10 \mu g/dL$ (See Table 2-25 and Appendix B). Therefore, lead does not pose a threat to fetuses of adult receptors exposed to soil in the Bulb Slope AOPC.

The IEUBK predicted that a typical child receptor exposed to lead in soil at the site would have an approximate 0.45% chance of PbBs exceeding 10 μ g/dL (Table 2-25 and Appendix B). The PbB prediction is well below USEPA's target to limit the risk to a typical child to no more than a 5% chance of exceeding the 10 μ g/dL PbB level of concern (USEPA 1994). Therefore, lead does not pose a threat to child receptors exposed to soil in the Bulb Slope AOPC.



2.6 Uncertainty Assessment

Uncertainties are inherent in any risk-based approach to evaluation and decision making for potentially contaminated sites. The uncertainties may be general and systemic as well as specific to the site. The objective of the uncertainty assessment is to identify the sources of uncertainty in the RA process, understand their potential to contribute to either underestimation or overestimation of risk for the selected receptors and pathways, and describe how the uncertainty is addressed. By describing the nature and magnitude of the uncertainties, the findings and conclusions of the RA can be better understood and used as a tool for decision making.

The following discussion supplements the extensive uncertainty analysis performed in the RI (Appendix O, URS 2012). The sources of uncertainties discussed in this section will involve those related to the risk characterization presented in the BHHRA.

2.6.1 COPC Selection Process

COPCs for the BHHRA were selected using ODEQ's rigorous methodology (ODEQ 2010a), which included selection of all chemicals that exceeded their individual screening levels, as well as chemicals whose exceedances across multiple media exceeded their screening levels. Additionally, noncancer chemicals that were lower than their individual screening levels were also retained as COPCs if the sum of the noncancer exceedance ratio exceeded a value of 1. No consideration of whether chemicals could be related to site activities was applied. Chemicals without screening levels were also retained as COPCs. Degradation products of PCE and TCE were also retained as COPCs even though many of them were reported as not detected in the actual data.

Given this exhaustive and comprehensive COPC selection process, risks are more likely to be overestimated due to the inclusion of COPCs that were nondetect (e.g., degradation products) and COPCs that may not have been site related, which results in low potential for underestimation of risk.

2.6.2 Adequacy of the Analytical Data Used for Site Characterization

The quality of the analytical data used for site characterization was reviewed thoroughly in the RI (URS 2012) and was deemed to be acceptable and useable for RA purposes. The inclusion of total benzofluoranthenes from the earlier data introduced some uncertainty. Since these data represented unique locations not covered in later data, the total benzofluoranthenes data were retained for the BHHRA. Their toxicity was evaluated using the methodology described in Section 2.6.4.

2.6.3 Exposure Assessment

2.6.3.1 EPC for RME and CTE

The 95% UCL values for soil were used as the EPC for both RME and CTE scenarios for occupational exposures. This is consistent with USEPA recommendations (USEPA 1992b), but is more conservative than required by ODEQ guidance (2010a), which recommends the arithmetic mean to represent the CTE. The potential for overestimation of risk is increased by using the 95% UCL to represent both the RME and CTE.



2.6.3.2 Hypothetical Fishing Platform Exposure Assumptions

The exposure factor values assumed for the Hypothetical Fishing Platform receptor are likely to substantially overestimate risk. At this time, there is neither the access nor the usage of the Island for such long-term, high intensity use. Although this evaluation assumes unrestricted access to the Island, there are obstructions and barriers of terrain, vegetation, and industrial use that currently limit access to certain parts of the Island. Even if such uses were undertaken in the future, it is highly unlikely for any one receptor to spend 100% of their time (24 hours/day, 365 days/year, for 26 years) solely at any one AOPC area. The Bulb Slope AOPC, in particular, is a small area with steep slopes that affords very poor access. Overall, the exposure assumptions and risk estimates associated with the Hypothetical Fishing Platform User using current soil exposure concentrations should be interpreted only as a guide to assist in future decision making and not as a literal estimation of risks and potential adverse health effects.

2.6.3.3 Volatilization Factor Calculation for Outdoor Exposure from VOCs in Soil

The USEPA equation (2002b) for deriving an outdoor air concentration from VOC concentrations in soil is highly uncertain when deriving a VF for the BHHRA because of arithmetical uncertainties. Since this pathway was included in the RME calculation and yielded insignificant risk, this pathway was omitted in the risk calculation for the CTE because exposure to VOCs outdoors is a very minor pathway for all receptors. The effect of this uncertainty is minimal with a negligible underestimation of CTE risk.

2.6.3.4 Non-Detect COPCs (Degradation Products of PCE)

As reported in the RI (URS 2012), PCE was detected but its degradation products (i.e., TCE, DCE(s) and vinyl chloride) were not detected but included as COPCs to meet ODEQ recommendations. The potential risk contribution from the degradation compound COPCs were quantitatively evaluated by including them as COPCs for the main exposure pathways for the receptors of both Landfill and Sandblast Area AOPCs.

EPC Selection Process

The process of estimating an EPC for the degradation products considered key questions, which are listed below and addressed in the discussion that follows.

- What were the concentrations of the parent compound(s)?
 - If the parent compound was detected at a relatively high concentration which may have resulted in masking the detections of the degradation products; were there other samples further away with lower detections and therefore less likely to mask the degradation products in the sample?
- Were the degradation products detected in another media (i.e., groundwater or soil gas)?
- What were the MDLs (see Section 7.0 of the RI) for the degradation products and how do they compare to the screening levels? (see RI Appendix I, Tables I-12 to I-15)?
 - If MDLs are used, how will the EPC be derived (i.e., maximum or averaged)?
- Can the current or future concentrations of the degradation products be modeled?

The issue of a high PCE detection masking concentrations of degradation products in the same soil sample has occurred in a sample collected in the Landfill AOPC. The contamination profile



of the Landfill AOPC is heterogeneous due to buried sources from past disposal practices; therefore, concentrations of contaminants in soil may be variable rather than show a gradient range of concentrations. At the Landfill AOPC, the analytical data for PCE from soil samples demonstrate this with only one high detection at 403,000 μ g/kg (Sample 011015BIL04TPG [Table 5-2b of the Final RI]) and relatively low detections elsewhere (ND to 11.0 μ g/kg). Subsequently, for the soil sample containing the high PCE concentration, none of the degradation products were detected (i.e., they were "U-qualified" as non-detect). They are listed with their respective MDLs (μ g/kg) as follows:

- 1,1-DCE: 2300 U
- cis-1,2-DCE: 2300 U
- trans-1,2-DCE: 2300 U
- TCE: 2300 U
- Vinyl Chloride: 2300 U

It should be noted that the MDLs were elevated. The MDLs for other soil samples from the site were typically three orders of magnitude lower (i.e., $0.2300 \ \mu g/kg$).

There were multiple detections of PCE in soil ranging from 0.605 μ g/kg to 65.0 μ g/kg (Appendix I Table I-1 of the Final RI [URS 2012]), but there were no corresponding detections of degradation products.

Amongst groundwater samples, there were multiple detections of PCE ranging from 0.23 μ g/L to 8.78 μ g/L) (Table I-1 of the RI), but none had corresponding detections of degradation product. There were two detections of vinyl chloride (0.531 and 0.507 μ g/L) but these low detections were not co-located with PCE detections.

These findings show that there is no evidence of PCE degrading to its degradation products at detectable levels either in the soil or groundwater at the Landfill. Any risk calculated for these degradation products should be considered with high uncertainty and an overestimation of risk. Nevertheless, the lone high detection of PCE in soil resulted in elevated MDLs for the degradation product. The Landfill AOPC EPC concentrations were conservatively estimated at the maximum MDL of 2300 μ g/kg for all PCE degradation products. The EPCs for degradation products in soil in the Sandblast Area AOPC were based on low MDLs.

PCE was identified as a soil COPC for the Construction Worker at the Landfill AOPC and for both Outdoor Maintenance Worker and Construction Worker at the Sandblast Area AOPC. The degradation products of PCE were added to the list of COPCs for these receptors, whether or not they were detected. The degradation products included TCE, 1,1-dichloroethene, cis-1-2 dichloroethene, trans-1,-2-dichloroethene, and vinyl chloride. More degradation products were detected in the Sandblast Area AOPC soils than in Landfill AOPC soils (Tables 9-1, 9-2 of the RI). The risk calculations are presented in Appendix A.

Landfill: Construction Worker

The RME cancer risk was $5 \ge 10^{-6}$, and HI was 2 (Table A-3.5 of Appendix A). The degradation products of PCE contributed negligibly to cancer risk and slightly elevated the noncancer HI to 2, which would be considered a negligible increase over the threshold value of 1. The increase of



the HI was primarily due to inhalation of TCE and the elevated MDL used for the EPC, which resulted in a HQ of 0.5.

Sandblast Area: Outdoor Maintenance Worker

The RME cancer risk was $2 \ge 10^{-5}$, and HI was 0.1 (Table A-3.10 of Appendix A). The degradation products of PCE contributed negligibly to both the cancer risk and noncancer HI.

Sandblast Area: Construction Worker

The RME cancer risk was 2×10^{-6} , and HI was 1. (Table A-3.15 of Appendix A) The degradation products of PCE contributed negligibly to the cancer risk and the noncancer HI.

The inclusion of non-detected degradation products of PCE generally had a negligible effect on the calculated cancer risk and noncancer health hazard for all the receptors except for noncancer hazards for the Landfill Construction Worker. This was due to using the elevated MDL for the EPC for which inhalation of TCE contributed a HQ of 0.5. Other soil samples and groundwater data from the Landfill AOPC did not show the presence of degradation products; therefore, using the elevated MDL was likely overly conservative. In general, identifying the proper EPC to use for degradation products that were not detected is challenging since theoretically, the concentrations may increase over time, yet there are no established methods to derive future concentrations. Using the detection limits for the EPC showed that the uncertainty of excluding degradation products are expected to only negligibly underestimate risk.

2.6.3.5 Oral and Dermal Bioavailability of Arsenic and PAHs

This BHHRA assumes that the relative oral bioavailability of arsenic 60% in soil and 100% for PAHs. While the 60% value is consistent with USEPA's current default estimates (USEPA 2012a, 2012b), this represents an upper-bound value, and the actual bioavailability may be substantially lower since arsenic at the site is not associated with any particular source type and may be soil-related. The range of soils tested for bioavailability of arsenic included mining and smelter soils, volcanic soils, pesticide-treated soils, and soils with manufacturing and electrical waste (USEPA 2012a, 2012b). A total of 103 estimates based on swine, monkey, and rat bioassays were available. Relative to the bioavailability of arsenic in water, less than 5% of the values exceeded a relative bioavailability of 60%. USEPA selected 60% as a default value from the upper percentile range that would be unlikely to be exceeded or result in an underestimation of risk.

While USEPA does not currently endorse any oral bioavailability factors for PAHs, several studies have noted that the relative bioavailability of cPAHs in soil may range from as low as 0.1% to 29% (Turkall et al. 2014, Harris et al. 2013, Magee et al. 1996). Similarly, dermal penetration of PAHs is currently assumed to be 13% based on USEPA default assumptions (USEPA 2004c). However, studies point to a reasonable estimate of 0.02 for dermal penetration (Magee et al. 1996, MADEP 2013). Therefore, risk estimates for cPAHs and the RBCs for cPAHs are likely to include a high level of overestimation of risk.

2.6.4 Toxicity Assessment

Benzofluoranthenes, Total

Older soil data (collected in 2001 and 2004) reported benzofluoranthenes as total instead of individual chemicals. Due to the lack of chemical physical parameters and toxicity values for



total benzofluoranthenes, the values for benzo(b)fluoranthene were used as a surrogate. The USEPA RSL tables list three benzofluoranthenes: benzo(b)fluoranthene, benzo(j)fluoranthene, and benzo(k)fluoranthene. They all share the same inhalation unit risk (IUR) but different SFo. The SFo for benzo(b)fluoranthene is the middle value. It should be noted that the most toxic benzo(j)fluoranthene has not been detected in Landfill soil. The uncertainty of using benzo(b)fluoranthene as a surrogate for total benzofluoranthenes in the risk screening in the RI intentionally overestimated risk. Use of a scaling factor of 60% of total benzofluoranthenes estimated as benzo(b)fluoranthene is based on a robust average of 30 samples and represents a more realistic and less uncertain method of treating the older data.

cPAHs were assumed to operate by a mutagenic mode of action, following current approaches suggested by ODEQ (2010a) and USEPA (2015a). This is likely to add to the potential for overestimation of risk for children since it is highly unlikely that exposure durations approaching residential levels would actually occur at the Island.

Chromium: Trivalent or Hexavalent Forms

As discussed in Section 2.4.5, the trivalent form of chromium was assumed for the RA for both the Landfill and Sandblast Area. Section 9 of the Final RI (URS 2012) notes the potential presence of nickel-chromium equipment coating that may have been sandblasted. Site conditions (i.e., high moisture and organic matter in soil and non-detect or very low dissolved chromium levels in groundwater) favor the occurrence of stable, insoluble trivalent chromium (Alloway 1990, Brookins 1988, ATSDR 2012a). Therefore, it is unlikely that any of the site chromium would be in the hexavalent form, and the potential for underestimation of risk is considered insignificant.

Petroleum Hydrocarbons

Consistent with USEPA's CERCLA guidance, TPH chemicals were not included as COPCs, although they were included in the initial COPC screening tables in the Final RI and for the Hypothetical Fishing Platform User (Tables 2-1 to 2-4). This is not expected to significantly underestimate risk. The only TPH fraction that exceeded the residential SLV was residual range organics in landfill soils. The residual fraction is a highly immobile and non-volatile fraction whose primary constituents of health concern are PAHs. Both the cancer risks and noncancer hazards associated with PAHs have already been evaluated using PAH-specific data, and cPAHs have been identified as the dominant COCs in both Landfill and Sandblast Area soils. Therefore, it is unlikely that the exclusion of residual range organics would result in any significant underestimation of noncancer hazard at the Landfill.

Thallium: PPRTV Screening Level Values

Thallium was selected as a COPC in groundwater for incidental contact for construction/ excavation workers. It was reported at low levels in groundwater at a maximum concentration of 0.3 ug/l. There were no peer-reviewed toxicity values available from within the USEPA preferred hierarchy of sources for thallium. The PPRTV RfD_o for thallium is 1 x 10⁻⁵ mg/kg-day (USEPA 2015a). However, in the PPRTV Derivation Source Document, due to various critical limitations in the study, USEPA presents this RfD_o as a provisional screening value in Appendix A of the PPRTV document, with even more uncertainty than a PPRTV (USEPA 2012c) and does not endorse this value as part of the recommended hierarchy of values. Therefore, this screening level value should be used and interpreted with great caution. In the absence of reliable human



toxicity data, the provisional screening value is based on a 1988 rat study with hair follicle atrophy as the critical effect. An uncertainty factor of 3,000 was applied. Even with the high degree of uncertainty and poor quality of the toxicity value, the noncancer hazard represented by thallium for the wader scenario was less than 1 and warranted no further consideration. Overall, the use of the PPRTV-screening value for thallium has a high potential to overestimate risk but is useful in eliminating thallium as a health concern at the site.

Vanadium

Vanadium was selected as a COPC for groundwater at the Sandblast AOPC with a maximum value of 0.22 ug/L. Elemental vanadium does not occur in nature, but vanadium may occur in six oxidation states in 65 different mineral ores and in association with fossil fuels (ATSDR 2012b). There were no peer-reviewed toxicity values available from within the USEPA preferred hierarchy of sources for vanadium compounds other than vanadium pentoxide. Values are provided in IRIS for vanadium pentoxide (V₂O₅) which is one of the more toxic forms of vanadium. The RfD_o for vanadium pentoxide is 9×10^{-3} mg/kg-day (USEPA 2015b) and the reference concentration (Rf_c) is 7 x 10^{-6} mg/m³ (ORNL 2015). However, exposure to vanadium pentoxide for humans mostly occurs as ingestion through dietary pathways and inhalation of dust in occupational settings (ATSDR 2012b). Vanadium in environmental media is typically not present as vanadium pentoxide. Therefore, it is unlikely to be the form in which vanadium occurs in site sediments or the pathway by which exposure might occur. To evaluate vanadium without using the vanadium pentoxide values, the HHRA adopted the approach used in the USEPA RSLs (2015a). The RfD₀ for vanadium pentoxide was adjusted by the molecular weight of vanadium only (56% of total molecular weight) and applied to the RfD_0 , resulting in a vanadium-specific RfD_0 of 5.04 x 10⁻³ mg/kg-day. The inhalation Rf_c for vanadium was selected as 1 x 10⁻⁴ mg/m³ based on ATSDR (2012b), which is based on chronic exposure to pentoxide. Overall, the use of the modified oral toxicity values and pentoxide-based inhalation values for vanadium has the potential to overestimate risk but is useful in eliminating vanadium as a health concern at the site.

n-Isopropylbenzene

n-Isopropylbenzene was also selected as a COPC for groundwater in the Landfill AOPC with an EPC of 2 ug/L. There were no readily available toxicity values that corresponded with USEPA's recommended hierarchy of values for n-isopropylbenzene. Therefore, the screening-level PPRTV values with RfD_o of 0.1 mg/kg-day and inhalation Rf_c of 1 mg/m³ were used (USEPA 2009c). Based on structural and metabolic similarities, the PPRTV-screening values for n-isopropylbenzene are based on ethylbenzene as a surrogate. The use of these values for quantitative risk assessment does not fall within USEPA's recommended hierarchy of sources and is accompanied by great uncertainty. Even with the high degree of uncertainty and poor quality of the toxicity value, the noncancer hazard represented by n-isopropylbenzene for the groundwater pathway was less than 1 and warranted no further consideration. Overall, the use of the PPRTV-screening value for n-isopropylbenzene has a high potential to overestimate risk but is useful in eliminating n-isopropylbenzene as a health concern at the site.

2.6.5 Risk and Hazard Estimates

The risk and hazard estimates are presented in the context of USEPA's acceptable risk range of 1 x 10^{-4} to 1 x 10^{-6} and a noncancer HI of 1. Individual chemicals with risks exceeding 1 x 10^{-6} or



HQ of 1 are noted as COPCs, as well as receptors and pathways with cumulative risks exceeding 1×10^{-6} or HI of 1.

Only the soil-related exposure pathway estimates exceeded the *de minimis* risk levels. Even these exceedances include several factors that are likely to overestimate risk. Risks were at acceptable levels for pathways related to groundwater and soil vapor.

Cumulative risks and hazards are also presented in the context of ODEQ's acceptable multicarcinogen risk level of 1×10^{-5} . However, even if the cumulative multi-chemical risk level does not exceed 1×10^{-5} , individual chemicals are still identified as a COC if they exceed the individual risk level of 1×10^{-6} (i.e., the RME Construction Worker at both the Landfill and Sandblast Area AOPCs have cumulative cancer risks less than ODEQ's threshold of 1×10^{-5} but benzo(a)pyrene is identified as a COC at both AOPCs due to exceeding the individual risk level of 1×10^{-6}).

2.6.6 Landfill Heterogeneity

For approximately 40 years (early 1940s until the early 1980s), USACE managed, stored and disposed of waste materials at the landfill in excavated pits or existing depressions. By 1982, the surface of the Landfill AOPC had been capped with soil cover. In 1989, approximately 8-inches of additional soil cover was placed on the Landfill site by the USACE (Hibbs, personnel comm. 2001). Because the waste was buried in separate pits, rather than one continuous pit, the heterogeneity of the landfill composition makes contamination characterize difficult. Although sampling targeted areas known (via historical aerial review or previously exposed areas) or suspected (via electrical resistivity data and seismic refraction data) as having the greatest landfilling activity, risk may be underestimated due to the heterogeneous nature of the landfill.

2.7 Calculation of Site-Specific Risk-Based Concentrations

| АОРС | Outdoor Maintenance Worker | Construction Worker | Excavation/ Trench Worker | Indoor Office Worker | Hypothetical Fishing Platform User |
|----------------|----------------------------------|------------------------|---------------------------------|----------------------------|--|
| Landfill | cPAHs | Benzo(a)pyrene | None | None | Arsenic, PCBs, cPAHs |
| Sandblast Area | cPAHs | Benzo(a)pyrene | None | None | Arsenic, PCBs, cPAHs, DEHP |
| Pistol Range | NA | NA | NA | NA | None |
| Bulb Slope | NA | NA | NA | NA | None |

The COCs for the four Upland AOPCs are summarized below:

To support the FS, RBCs were calculated for the COCs for each receptor, corresponding to a target cancer risk level of 1×10^{-6} (Table 2-27). The Reference Area UPL was also considered for inorganic chemicals (Table 1-9). The recommended RBC for each COC was selected as the higher of the RBC and UPL for inorganics. For organics, the risk-based value was selected as the recommended RBC.



The equation for calculating RBCs is as follows:

RBC = EPC x Target Risk / Calculated Risk

where,

RBC = Risk Based Concentration EPC = EPC for Chemical from BHHRA Target Risk for individual PAHs and cumulative cPAHs = 1×10^{-6} Calculated Risk = Chemical Specific from BHHRA

The RBC for arsenic for the Hypothetical Fishing Platform User is 0.68 mg/kg, while the Reference Area UPL is 5.4 mg/kg. Consistent with USEPA guidance that RBCs lower than ambient or background levels are not feasible for naturally occurring inorganics (USEPA 2002c, 2002d), the Reference Area UPL was selected as the recommended RBC for arsenic and was used to identify soil sampling locations where arsenic exceedances might occur.

The RBCs for Total PCBs (as Aroclors) and DEHP are based only on risk and did not take Reference Area concentrations into account.

The outdoor worker RBC for cPAHs was estimated at 0.3 mg/kg BaPeq. The RBC for the Construction Worker was 2 mg/kg BaPeq. The Outdoor worker RBC and the Construction worker RBC are higher than the Reference Area UPL (0.052 mg/kg), while the Hypothetical Fishing Platform User RBC (0.015 mg/kg) is less than the Reference Area UPL. Although the recommended cPAHs RBC for this receptor is the Reference Area UPL (Table 3-31), the exceedances shown in the figures discussed in Section 2.8 are based only on the risk-based value, in recognition of the fact that the naturally occurring background concept is not always applied to organic compounds. If ODEQ's acceptable risk level of 1 x 10⁻⁵ for COCs with a common and additive mode of action were taken into account, the RBCs for cPAHs may be adjusted upwards by a factor of 10.

2.8 Risk Interpretation

In this section, the results of the risk estimation, uncertainty assessment, and site-specific information are considered together to come up with a final assessment of risks. The RBCs for cPAHs, which are the primary COCs in Landfill and Sandblast AOPC soils, are highly influenced by certain conservative assumptions, including 100% oral bioavailability for all receptors, as well as residential-type exposures and mutagenic mode of action for Hypothetical Fishing Platform Users.

Figures 2-5 and 2-6 show all the sample locations at the Landfill and Sandblast AOPCs that exceeded the recommended occupational RBCs for the Outdoor Maintenance Worker (0.3 mg/kg BaPeq). Locations in deeper soils (3-10 ft bgs) which exceeded the Construction Worker RBC for cPAHs (2 mg/kg BaPeq) area also shown. Figures 2-7 and 2-8 show the locations where the RBCs for the Hypothetical Fishing Platform receptor were exceeded under current exposure conditions.

2.8.1 Landfill AOPC

cPAHs were identified as the only COC for occupational exposures (Section 2.7). A total of 30 locations (25 excluding composite sample locations) exceeded the occupational RBC for cPAHs



at the Landfill in the 0-3 ft and 3-10 ft bgs depths (Figure 2-5). The locations were at the northwestern section of the Landfill, primarily within and around the Gully Test Pit and Lead Hot Spot Test Pit #1, Mercury Vapor-Lamp Test Pit, far eastern portion of the Landfill, and one isolated spot northeast of the Pesticide/Herbicide Wash Area.

For the Hypothetical Fishing Platform User under current conditions, arsenic, Total PCBs, and cPAHs were identified as the COCs (Section 2.7). The RBCs for the Hypothetical Fishing Platform receptor were exceeded at 26 locations in the shallow soils at the Landfill, primarily for cPAHs (Figure 2-7). The exceeded locations for the Hypothetical Fishing Platform receptor (Figure 2-7) were almost identical with the exceeded locations for the occupational workers (Figure 2-5). Exceedances of the RBCs for arsenic and PCBs were co-located with the cPAH exceedances.

Arsenic exceedances were noted at five locations ranging, with four locations lower than approximately twice the Reference Area UPL (i.e., 12 mg/kg) (BIL11SS1, BIL22, BIL02USE, and BIL03USE) and one outlier location at 30.1 mg/kg (BIL05SS1). Overall, arsenic appears to be a relatively minor COC. Total PCBs were exceeded at only two locations (0.728 mg/kg at BIL04SS1 and 0.488 mg/kg at BIL02SS1). These are both estimated (J-qualified) values and are approximately two to three times higher than the RBC. Thus, the values represent a risk level of 3 x 10^{-6} , which is at the low end of the USEPA acceptable risk range. Considering the low frequency and magnitude of exceedance and the hypothetical and overly conservative nature of the Hypothetical Fishing Platform User scenario, risks related to PCBs may be considered to be very minor and represent an insignificant contribution.

Overall, risks related to exposure to shallow soils at the Landfill AOPC are primarily due to cPAHs for both occupational receptors and Hypothetical Fishing Platform Users. Arsenic and Total PCBs contribute very minor risks to the Hypothetical Fishing Platform User under current conditions. A subset of nine locations in deeper soil at the Gully Test Pit area and one location in the Mercury Vapor Lamp Test Pit Area also exceed Construction Worker RBCs for cPAHs. There are no unacceptable risks associated with groundwater at the Landfill.

2.8.2 Sandblast Area AOPC

cPAHs were identified as the only COC for occupational exposures to soil at the Sandblast Area AOPC (Section 2.7). A total of 16 locations exceeded the occupational RBCs at the Sandblast Area AOPC in the shallow and deeper soils (Figure 2-6). The locations are at the Erodible Unit, Current HMSA, Equipment Laydown Area, and the Former HMSA. All exceedances were in the shallow soils (0-3 ft bgs) with the exception of a single location (HA4, near the Current HMSA) that exceeded both the Outdoor Worker and Construction Worker RBC.

For the Hypothetical Fishing Platform User under current conditions, the COCs included arsenic, cPAHs, DEHP, and Total PCBs (Section 2.7). A total of 42 locations exceeded the Hypothetical Fishing Platform receptor RBCs (Figure 2-8). In comparing the figures for the occupational exceedances (Figure 2-6) with the Hypothetical Fishing Platform receptor exceedances (Figure 2-8), more locations exceeded for the Hypothetical Fishing Platform receptor including near the Catch Basin #1 and the Former HMSA. The organic COCs (PCBs, DEHP, and cPAHs) are generally co-located in their exceedances. Arsenic exceedances are sometimes co-located with the organic COCs and sometimes occur on their own. While cPAH and arsenic exceedances are dispersed throughout the Sandblast AOPC, DEHP and PCBs showed localized patterns of



exceedance. DEHP exceedances were limited to two samples in the erodible unit (SB-EUA and SB-EUB). These samples represent composited samples from the area. DEHP exceeded the RBCs by approximately a factor of 1.6 (risk level 1.6×10^{-6}) at SB-EUA and a factor of 7 (risk level 7 x 10^{-6}) at SB-EUB, representing risk levels that fall within the USEPA acceptable risk range of 1 x 10^{-4} to 1 x 10^{-6} . PCB exceedances were limited to seven locations near the Equipment Laydown Area (LD-01, LD-02, LD-03, LD-05, LD-07, LD-10, and LD-11). The magnitude of exceedance ranged from approximately 1.7 times at LD-02 to approximately 9 times at LD-10. The risk levels associated with these exceedances also fall within the USEPA acceptable risk range.

Overall, risks related to exposure to shallow and deeper soils at the Sandblast Area AOPC are primarily due to cPAHs for both occupational receptors and Hypothetical Fishing Platform Users. Arsenic and Total PCBs contribute minor risks to the Hypothetical Fishing Platform User under current conditions. There are no unacceptable risks associated with groundwater or soil gas at the Sandblast Area AOPC.

Although all the locations are shown, not all the locations would need to be remediated to meet the risk target since an AOPC-wide statistical average exposure concentration is more realistic to actual receptor exposure.

2.8.3 ODEQ's Acceptable Risk Level for Multi-Chemical Exposure

If ODEQ's cumulative acceptable risk level of 1×10^{-5} were used as the basis for RBCs, the occupational RBC for cPAHs would be increased by a factor of 10 and would result in a value of 3 mg/kg BaPeq. The number of locations at both AOPCs where cPAHs exceed this RBC would be fewer than the number of locations that exceed the RBC based on 1×10^{-6} risk level (shown in orange on Figures 2-5 and 2-6).

At the Landfill AOPC, there were 16 locations that also exceeded the ODEQ multi-chemical cancer risk threshold of 1 x 10^{-5} . All of these locations were at and around the Gully Test Pit, Mercury Vapor-Lamp Test Pit, and Lead Hot Spot Test Pit #1 and at location BIL05SSI (Figure 2-5).

At the Sandblast Area AOPC, there were three locations that also exceeded the ODEQ multichemical cancer risk threshold of 1×10^{-5} . They were located in the northern Equipment Laydown Area and the Current HMSA (Figure 2-6).

It is apparent that cPAHs are a widespread COC in Landfill and Sandblast AOPC soils. As noted in the Uncertainty Assessment, the current risk estimates and RBCs assume that the oral bioavailability of cPAHs is 100% and the dermal absorption is 13%, based on current USEPA assumptions (USEPA 2015a). However, the actual oral bioavailability and dermal absorption of PAHs in soils are likely to be much lower (Magee et al. 1996, Turkall et al. 2014, Harris et al. 2013). Therefore, the risks are likely to be overestimated and the RBCs are likely to be overly stringent in this BHHRA. The point estimate of the oral-to-soil absorption adjustment factor (AAF) of 0.29 and the dermal absorption factor of 0.02 that have been suggested by Magee et al. (1996) have been incorporated into several approaches with regulatory approval. For example, Massachusetts Department of Environmental Protection uses these factors in developing numerical standards for soil under the Massachusetts Contingency Plan (MADEP 2015). The



oral AAF of 0.29 has also been accepted by USEPA Region 5 in developing remedial goals for PAHs at a Superfund Site (AECOM 2009 for Solutia).

During the FS, if the human health-based RBCs for occupational and Hypothetical Fishing Platform User were modified by using the oral and dermal factors suggested above to develop alternative remedial action levels (RALs), it would result in RBCs approximately 3-fold higher that would still likely be protective of human health. Site-specific considerations for PAH bioavailability could also be explored if appropriate methods are available (ITRC 2015).

2.9 Conclusion and Recommendations

COCs are identified based on RME results exceeding the ODEQ cancer risk threshold of 1×10^{-6} cancer risk. Table 2-27 lists the RBCs and Reference Area UPLs (if available) for the COCs for the Landfill and Sandblast Area AOPCs that are recommended for further evaluation in the Upland OU FS.

Media evaluated for the Landfill AOPC and Sandblast Area AOPC included soil and groundwater. In addition, soil gas data were also evaluated at the Sandblast Area AOPC. Risk assessment findings were similar for both AOPCs. Exposure to shallow soil for the outdoor maintenance worker and to deeper soil for the construction worker showed RME and CTE cancer risks within the USEPA acceptable risk range of 1×10^{-6} to 1×10^{-4} . RME risks from exposure to deeper soils for construction workers were also within the USEPA acceptable risk range. RME risks related to trench worker exposure to groundwater were *de minimis* (less than 1 x 10^{-6}). At the Sandblast Area AOPC, RME and CTE risks related to soil gas were also *de minimis*.

At both the Landfill AOPC and the Sandblast Area AOPC, RME and CTE cancer risks exceeded the USEPA acceptable risk range for the Hypothetical Fishing Platform receptors from exposure to shallow soils under current conditions. Risks and noncancer hazards for nursing infants were generally close to or less than 1×10^{-6} for PCBs under current conditions.

At the Pistol Range and Bulb Slope AOPCs, which were evaluated only for the newly added Hypothetical Fishing Platform Users, lead concentrations were at acceptable levels under current conditions for exposure to soil and lagoon sediment at the Pistol Range and soil at the Bulb Slope.

At both the Landfill and Sandblast Area AOPCs, cPAHs were the primary COCs for occupational exposures in surface soil at the 0-3 ft bgs depth range. Benzo(a)pyrene (also a cPAH) was the only COC for deep soil at the 0-10 ft bgs depth range. Arsenic, PCBs, and DEHP (Sandblast Area AOPC only) were minor additional contributors to risk for Hypothetical Fishing Platform Users under current conditions.



3.0 BASELINE ECOLOGICAL RISK ASSESSMENT

This section presents the findings of the BERA that was conducted for the Upland OU. The methodology followed for the BERA process is described in the Upland OU WP Update (URS 2014) and further details are provided in the RI/FS MP (URS 2007).

The purpose of this BERA is to further evaluate the CPECs identified for the four Upland OU AOPCs during the SLERA performed as part of the RI (URS 2012). The findings of this BERA will be used to identify CECs, receptors of concern, and pathways that should be retained for the Upland OU FS.

3.1 AOPCs, Media, and CPECs

All four AOPCs in the Upland OU were retained for evaluation in the BERA. Surface soil (0-1 ft bgs) and shallow soil (0-3 ft bgs) were identified as media of concern for terrestrial ecological receptors. The CPECs carried into the BERA include metals, total high-molecular-weight PAHs (HPAHs), tributyltin, organochlorine pesticides, VOCs, and SVOCs for the Landfill and Sandblast Area AOPCs (Tables 1-1 and 1-2), lead for the Pistol Range AOPC (Table 1-3), and lead and mercury for the Bulb Slope AOPC (Table 1-4). All of these CPECs were included in the Upland OU-wide evaluation in which wide-ranging receptors were assumed to forage in all four AOPCs combined (Table 1-5). Risk estimates were calculated for each CPEC for all receptors potential present at a given AOPC.

3.2 Receptors and Exposure Pathways

The following list of receptors and exposure pathways identified in the WP Update were included in the Upland BERA:

- Terrestrial plants and soil invertebrates exposed through direct contact with surface and shallow soil.
- Canada goose (*Branta canadensis*) exposed through incidental ingestion of surface and shallow soil, prey (100% plants), and water.
- American robin (*Turdus migratorius*) exposed through incidental ingestion of surface and shallow soil, prey (100% soil invertebrates), and water.
- American kestrel (*Falco sparverius*) exposed through incidental ingestion of surface and shallow soil, prey (100% small mammals), and water.
- Vagrant shrew (*Sorex vagrans*) exposed through incidental ingestion of surface and shallow soil, prey (100% soil invertebrates), and water.
- American mink exposed through incidental ingestion of surface soil, upland prey (15% small mammals), and water.

The assessment and measurement endpoints originally presented in the RI/FS MP (URS 2007), with the recent addition of the mink, are shown in Table 3-2 of the WP Update (Appendix C).

Based on the available data, surface soil is defined as 0 to 1 ft bgs for all AOPCs, except for the Pistol Range, for which surface soil is defined as 0 to 1.5 ft bgs based on the available depth



interval data. Shallow soil is defined as 0 to 3 ft bgs, and this depth interval only applies to the Landfill and Sandblast Area AOPCs.

For the water ingestion pathway, highly mobile wildlife receptors (i.e., kestrel and mink) were assumed to ingest water from the river, while less mobile receptors (i.e., robin and shrew) and the Canada goose were assumed to ingest water that puddles in the Upland OU (via a simple equilibrium partitioning calculation). The soil to water equilibrium partitioning method was also used for the kestrel and mink for soil COIs lacking river surface water data. Ingestion of surface water from the river is a minor exposure pathway given the very low concentrations detected in surface water samples collected from the River OU during the RI (URS 2012) and because the water ingestion pathway typically provides a minor contribution to the overall dose for birds and mammals. Therefore, using RI surface water data from the River OU in the dose estimates for the kestrel and mink will not affect any risk management decisions in the FS for the Upland OU. This approach was used for purposes of dose calculation completeness and because measured data provide the most representative concentrations for this medium.

As described in detail in the WP Update, the mink is the identified receptor that could feasibly forage in the Upland OU as well as in the River OU. According to the studies reviewed during the development of the WP Update, it is likely that a mink would primarily be attracted to foraging resources in the River OU (e.g., crayfish and fish), but could sometimes supplement its aquatic prey diet with terrestrial prey from the Upland OU. For this reason, a dietary composition of 15% small mammals from the Upland OU was conservatively assumed as a first step in the evaluation for the mink, even though a lower proportion may actually be more realistic due to the presence of an immediately available permanent water source (i.e., the riverine habitat). The remaining 85% of the mink's diet was assumed to be comprised of prey from the River OU (i.e., fish and crayfish), which was not evaluated in this Upland OU BERA (see separate River OU BERA report). An estimation of total risk for the mink that reflects the potential for adverse effects from exposure to both Upland and River OU media will be provided in the Upland FS or the River FS.

3.3 Exposure Assessment

The foodweb model for the Upland OU and the ecological CEMs for each AOPC are provided on Figures 3-1 through 3-5. These figures illustrate the most current understanding of potentially complete and significant ecological exposure pathways for the Upland OU, including those associated with transport to the River OU (not considered herein).

The SLERA conducted in the RI concluded that CPECs in Upland OU groundwater potentially discharging to the River OU did not pose a risk to aquatic receptors .Slope stability will be considered in the Upland OU FS as part of the evaluation of available remedial technologies. Therefore, Upland OU to River OU pathways are not included in this Upland OU BERA.

3.3.1 Exposure Factors

Tables 3-1 through 3-5 present the life history parameters for each terrestrial receptor evaluated in the Upland BERA that were input into the dose equation (see Section 3.3.3), which were presented in the approved Upland OU WP Update (URS 2014). For the animal receptors with small home ranges (i.e., shrew and robin), it is feasible for an individual to forage solely within



one specific AOPC. Although the Canada goose has a potentially large home range, part of Bradford Island is managed as goose habitat, and therefore this receptor could spend most of its time on the Island. For the animal receptors with large home ranges (i.e., kestrel and mink), it was assumed that these receptors could forage at each individual AOPC (albeit unlikely) and within all four AOPCs combined. Receptor-specific area use factors were calculated as the AOPC (and/or combined AOPC) site size divided by the size of the home range.

The home range size of 1.85 kilometers for the mink that was originally developed in the RI/FS MP (URS 2007) for the River OU of Bradford Island is based on this receptor's inclination to forage along the riverbanks. The shape of mink home ranges depends on the type of habitat available. In riverine environments, home ranges are linear, whereas those in marsh habitats tend to be more circular (USEPA 1993). In addition, home ranges for adult males are generally larger than adult female home ranges, especially during the mating season. To be conservative, the smallest home range for an adult female in riverine habitat expressed as an area (7.8 hectare, or approximately 19 acres) was selected for use in the Upland BERA (USEPA 1993).

The study cited in USEPA (1993) from which this mink home range was derived (Mitchell 1961) is the same study presented in the document provided by Jeremy Buck of the USFWS during the Upland OU TAG meeting on April 14, 2014. The USFWS recommended that the information in this document be considered during the selection of a home range for the mink for the Upland OU BERA. The selected home range for the mink in the USFWS's document of 23 acres is slightly higher (less conservative) than the selected home range for this BERA. The home range for the mink of 19 acres was used to calculate AOPC-specific area use factors for the mink.

3.3.2 Exposure Point Concentrations

Soil. The lower of the maximum detected concentrations and 95% UCLs calculated for soil in the RI were used as the EPCs for all terrestrial receptors in this BERA to provide more realistic, site-specific estimates of risk (Tables 1-1 through 1-5). There are no protected plant and invertebrate species in the Upland OU and the goal is estimating risk to these receptor groups at the community level (URS 2012).

Surface Water. For the water ingestion pathway, the kestrel and mink were assumed to ingest water from the river, and the robin, shrew, and Canada goose were assumed to ingest water that puddles in the Upland OU. The surface water EPCs for the kestrel and mink are the maximum detected concentration in river water of available data (Table 3-6). Otherwise, the surface water concentrations were derived based on surface soil concentrations and the simple equilibrium partitioning model, as described below.

For the robin, shrew, and goose, surface water concentrations were derived from the lower of the maximum detected concentration and 95% UCLs for surface soil through a simple equilibrium partitioning model (Tables 3-7a through 3-7e). Partitioning coefficients (K_d for inorganics and K_{oc} for organics) were applied to surface soil concentrations in the following equations. The fraction of organic carbon (foc) is a relevant parameter for the estimation of surface water concentrations of organic compounds and was assumed to be 1%:



| Inorganics | $C_{surface water} (mg/L) = \underline{C_{soil} (mg/kg dry weight)}$ |
|-----------------|--|
| | $K_{d}(L/kg)$ |
| Organics | $C_{surface water} (mg/L) = \underline{C_{soil} (mg/kg dry weight)}$ |
| | $K_{oc} (L/kg) \times foc$ |

The K_d and K_{oc} values are presented in Tables 3-7a through 3-7e along with the estimated surface water concentrations. Dilution of porewater that enters surface water and mixes with shallower depths in the water column could result in lower EPCs for surface water; however, no dilution factors were applied to the estimated water (soil-porewater) concentrations.

3.3.3 Bioaccumulation Factors

A combination of regression-derived bioaccumulation factors (BAFs), median BAFs, and octanol-water partition coefficient ($\log K_{ow}$)-based BAFs from the literature were used to predict tissue concentrations in the BERA. The regression-based approach is typically preferred because it provides a more site-specific prediction of a CPEC concentration in a certain dietary tissue type, as it incorporates the site soil EPCs.

Because most environmental data sets are log-normally distributed rather than normally distributed, a linear relationship between the soil and tissue concentrations is often revealed when a data set is transformed into lognormal values. Regression equations are usually presented in the following form:

$$\ln(y) = a + b \times \ln(x)$$

where,

y = chemical concentration in tissue (mg/kg dry weight)

- x = chemical concentration in soil/sediment (mg/kg dry weight)
- a = log-transformed y intercept
- b = slope

In some instances, calculated regression equations do not reliably predict concentrations of chemicals that bioaccumulate (or bioconcentrate) in tissue. This type of unreliability can occur when the soil-tissue concentration relationship for a particular chemical is not linearized by log transformation, when the data set used to generate the regression equation is too limited, or when the chemical-specific lognormal logK_{ow} is very high (e.g., steric hindrance) or very low (low persistence, readily metabolized) and does not fall within the logK_{ow} range used in the model.

In the absence of empirically derived BAFs, equations to estimate BAFs for non-ionic organic compounds are typically based on the $\log K_{ow}$ of the chemical and the organic carbon content of the soil. Caution is recommended when using this approach, as unrealistic BAFs can be generated for SVOCs and VOCs, e.g., phthalates that have high $\log K_{ow}$ values.

The following hierarchy of sources was consulted for soil-to-plant, soil-to-soil invertebrate, and soil-to-small mammal BAFs:



- USEPA's Ecological Soil Screening Levels (Eco-SSLs) guidance document, Attachment 4-1 (2005a, last updated 2007).
- Oak Ridge National Laboratory (ORNL) publications (Sample et al. 1998a and 1998b, Bechtel-Jacobs 1998) and Risk Assessment Information System (RAIS) database.
- LogK_{ow}-based algorithms presented in Jager 1998 (soil invertebrate BAFs) and Travis and Arms 1988 (small mammal BAFs) converted to dry weight tissue concentrations (consistent with Eco-SSL and ORNL BAFs).
- USEPA's SLERA Protocol for Hazardous Waste Combustion Facilities, Peer Review Draft (1999); converted to dry weight tissue concentrations (consistent with Eco-SSL and ORNL BAFs).

Surrogate chemicals were used when appropriate for CPECs lacking BAFs in this hierarchy of sources. In addition, BAFs from the primary literature based on empirical data were used when available (i.e., Briggs et al. 1982 for di-n-butyl phthalate in plants).

Finally, a default BAF of 1.0 was applied for CPECs as a last resort, primarily for SVOCs and VOCs. This assumption is expected to overestimate the level of accumulation for CPECs lacking literature-based BAFs or reasonable surrogates.

The selected BAFs for each of the three dietary types are presented in Tables 3-8 through 3-10 by AOPC as follows:

- Tables 3-8a, b, c, and d for terrestrial plant BAFs for Landfill, Sandblast Area, Pistol Range, and Bulb Slope AOPCs, respectively.
- Tables 3-9a, b, c, and d for soil invertebrate BAFs for Landfill, Sandblast Area, Pistol Range, and Bulb Slope AOPCs, respectively.
- Tables 3-10a, b, c, d, and e for mammalian BAFs for Landfill AOPC, Sandblast Area AOPC, Pistol Range AOPC, Bulb Slope AOPC, and Combined-AOPCs, respectively.

3.3.4 Dose Estimation

For the birds and mammals, site-specific daily dose estimates were developed to estimate chemical intake from food resources, incidental soil ingestion, and water ingestion. The food chain transfer mechanisms via the ingestion pathway for each bird and mammal receptor group are summarized as follows:

- Soil \rightarrow terrestrial plants \rightarrow Canada goose
- Soil \rightarrow soil invertebrates \rightarrow American robin
- Soil \rightarrow small mammals \rightarrow American kestrel
- Soil \rightarrow soil invertebrates \rightarrow vagrant shrew
- Soil \rightarrow small mammals \rightarrow American mink

Site-specific daily dose estimates were calculated using the following general equation:



$$Dose_{total} = \frac{([IR_{food} \times C_{food}] + [IR_{soil} \times C_{soil}] + [IR_{water} \times C_{water}]) \times AUF}{BW}$$

where,

| Dose _{total} | = | Estimated dose from ingestion (mg/kg BW per day dry weight [mg/kg-day]) |
|-----------------------|---|---|
| IR_{food} | = | Ingestion rate of food (kg/day dry weight) |
| C_{food} | = | Concentration in dry weight of CPEC in food (mg/kg) |
| IR _{soil} | = | Ingestion rate of soil (kg/day) |
| C _{soil} | = | Concentration in dry weight of CPEC in soil (mg/kg) |
| IR _{water} | = | Ingestion rate of water (L/day) |
| C _{water} | = | Concentration of CPEC in surface water (mg/L) |
| AUF | = | Area use factor (unitless) |
| BW | = | Adult body weight (kg) |

Exposure parameters for the bird and mammal receptors listed above were discussed in Section 3.3.1 and are presented in Tables 3-1 through 3-5. The methods used to estimate concentrations in food items are described in Section 3.3.3, and the resulting daily dose estimates for each receptor are presented in Tables 3-11a through 3-11e (Landfill), 3-12a through 3-12e (Sandblast Area), 3-13a through 3-13e (Pistol Range), 3-14a through 3-14e (Bulb Slope), and 3-15a and 3-15b for the kestrel and mink (Combined AOPCs).

3.4 Toxicity Values

Screening levels, or SLVs, are expressed as concentrations in media (i.e., mg of chemical/kg of soil). Although "screening levels" are typically associated with exposure via direct contact, and are commonly referred to as direct toxicity benchmarks, limited sources of generic media-based screening levels address both direct contact and dietary exposure for birds and mammals. In contrast, diet-based toxicity reference values (TRVs) protective of birds and mammals are expressed as a daily dose normalized to body weight (mg of chemical/kg of BW/day).

Both low SLVs/no observable adverse effect levels (NOAELs) and high SLVs/lowest observable adverse effect levels (LOAELs) were selected for each receptor group in order to develop a range of HQs for consideration by risk managers. An uncertainty factor of 5 was used to adjust chronic high SLVs/LOAELs to low SLVs/NOAELs and vice-versa, when necessary (ODEQ 2001). Additionally, in a few cases, chronic NOAEL and LOAEL TRVs were extrapolated from subchronic, acute, or lethal dose to 50% of test organisms (LD50) test results using uncertainty factors based on U.S. Army Center for Health Promotion and Preventive Medicine (USACHPPM) Technical Guide 254 - Standard Practice for Wildlife TRVs (USACHPPM 2000). The final selected SLVs and TRVs are described below.

3.4.1 Screening Levels for Terrestrial Plants and Soil Invertebrates

The following hierarchy of sources of terrestrial plant and soil invertebrate SLVs was consulted for the BERA:

• USEPA's Eco-SSLs (USEPA 2005-2008)



- ODEQ's Level II SLVs (ODEQ 2001)
- ORNL's RAIS database and ORNL guidance documents (Efroymson 1997a, 1997b)
- USEPA TRVs (USEPA 1999)

Lower-bound and upper-bound SLVs were selected for each receptor group and CPEC. The benchmarks are shortened to low and high SLVs in the text and tables. In the absence of SLVs from the hierarchy listed above, SLVs were drawn from other sources of the literature (Table 3-16). As shown in Table 3-16, the chromium SLVs for both plants and invertebrates were below the site-specific background UPL and, therefore, the chromium UPL was defaulted to as the SLV for these receptors.

3.4.2 Toxicity Reference Values for Birds and Mammals

The following hierarchy of sources of TRVs for birds and mammals were consulted for the BERA:

- USEPA's Eco-SSLs (USEPA 2005-2008)
- Final Portland Harbor RI/FS (Lower Willamette Group 2013)
- ORNL guidance (Sample et al. 1996)

Both low (NOAEL-based) and high (LOAEL-based) TRVs were selected for each receptor group and CPEC. In the absence of SLVs from the hierarchy listed above, TRVs were drawn from other sources of the literature (Table 3-17). As shown in Table 3-17, chemical surrogates were used when appropriate in the absence of CPEC-specific TRVs.

3.5 Risk Characterization

Risk characterization is the process of integrating the previous elements of the RA into quantitative or semi-quantitative estimates of risk. Risk characterization consists of risk estimation and uncertainty assessment. Risk estimation or the quantification of risk is then used as an integral component in remedial decision making and selection of potential remedies or actions. Uncertainty assessment describes the level of confidence in the risk estimation.

The two types of SLVs/TRVs listed in Section 3.4 were incorporated into the analysis: one based on a low SLV/NOAEL and a second based on an observed adverse effect in a test species (high SLV/LOAEL). For HQs based on the low SLV and NOAEL TRV that are less than 1, adverse effects are unlikely because of the inherent conservatism (protectiveness) built into the exposure and effects assessments. HQs based on the high SLV and LOAEL TRV (upper-bound risk estimates) that are greater than 1 indicate that exposure exceeds a known effect concentration for a test organism. In this case, potential risk management measures in the FS may be warranted for these receptors and exposure pathways.

For estimated exposures that exceeded the low SLV/NOAEL TRV (i.e., the low SLV/NOAEL TRV-based HQ is >1.0) but were less than the high SLV/LOAEL TRV (i.e., the high SLV/LOAEL TRV-based HQ is <1.0), the associated complete exposure pathways were considered in greater detail to develop conclusions about the likelihood that a risk or hazard is present. For non-listed (common) terrestrial species known to be present on Bradford Island, more emphasis



was placed on CPECs that have high SLV/LOAEL TRV-based HQs above 1. However, the range of HQs developed for each receptor and CPEC was presented in the risk characterization.

It is most appropriate to calculate HIs (i.e., a summation of HQs) for CPEC groups when multiple chemicals demonstrate similar modes of toxicity or affect the same target organ. The implications of HQs greater than or less than 1.0 discussed above were also applied to HIs. Due to a lack of data regarding additive effects associated with exposure to multiple chemicals for nonhuman receptors, professional judgment was used in the development of HIs. For the BERA, HIs were calculated based on the following chemical groupings (as cited in the RI/FS MP, URS 2007): inorganics (including butyltins), organochlorine pesticides, phthalates, HPAHs (evaluated as total HPAHs), and VOCs.

3.5.1 Summary of Hazard Quotients and Hazard Indices

This section summarizes the HQs and HIs for each AOPC and receptor and the combined AOPCs for the kestrel and mink. HIs were not calculated for the Pistol Range because only one CPEC (lead) was identified for this AOPC. The significance of CPECs with HQs greater 1.0 based on the low SLV/NOAEL or high SLV/LOAEL are discussed in the risk interpretation section relative to their potential to elicit adverse effects in the receptor populations for each AOPCs.

3.5.1.1 Landfill AOPC

Terrestrial Plants (Table 3-18a). In surface soil (0-1 ft bgs), chromium, mercury, chlordane (technical mixture and metabolites; hereafter simply referred to as "chlordane"), and Total HPAHs have HQs greater than 1.0 based on the low SLVs and high SLVs. In addition to these CPECs, copper, lead, and nickel also have HQs greater than 1.0 based on the low SLVs.

The same seven CPECs listed above for surface soil have low SLV-based HQs greater than one for shallow soil (0-3 ft bgs). Chromium, nickel, chlordane, and Total HPAHs have high SLV-based HQs greater than 1.0 for shallow soil.

For surface soil, the low SLV- and high SLV-based HIs for metals are 24 and 12, respectively. For shallow soil, the low SLV- and high SLV-based HIs for metals are 46 and 26, respectively. For pesticides for both surface and shallow soil, the low SLV- and high SLV-based HIs are 7.0 and 1.4.

Soil Invertebrates (Table 3-19a). In surface soil (0-1 ft bgs), chromium, mercury, and Total HPAHs have HQs greater than 1.0 based on the low SLVs and high SLVs. In addition to these CPECs, copper also has an HQ greater than 1.0 based on the low SLV.

The same four CPECs listed above for surface soil have low SLV-based HQs greater than 1.0 for shallow soil (0-3 ft bgs), with the addition of nickel. Chromium, mercury, and Total HPAHs have high SLV-based HQs greater than 1.0 for shallow soil.

For surface soil, the low SLV- and high SLV-based HIs for metals are 27 and 12, respectively. For shallow soil, the low SLV- and high SLV-based HIs for metals are 40 and 25, respectively. No soil invertebrate SLVs are available for the pesticides; therefore, these CPECs are addressed qualitatively in the uncertainty assessment.



Canada Goose (Table 3-11a). All LOAEL-based HQs are less than 1.0 for the goose. In surface soil, lead and mercury have HQs greater than 1.0 based on the NOAELs. In addition to these CPECs, chromium also has an NOAEL-based HQ greater than 1.0 for shallow soils.

For surface soil, the NOAEL- and LOAEL-based HIs for metals (including tributyltin) are 12 and 1.8, respectively. For shallow soil, the NOAEL- and LOAEL-based HIs for metals are 13 and 2.2, respectively. The remaining HIs are less than 1.0.

American Robin (Table 3-11b). In surface and shallow soil, chromium, copper, lead, mercury, nickel, and chlordane have HQs greater than 1.0 based on the NOAELs and LOAELs. In addition to these CPECs, DEHP also has an HQ greater than 1.0 based on the NOAEL in both surface and shallow soil.

For surface soil, the NOAEL- and LOAEL-based HIs for metals are 100 and 17, respectively. For shallow soil, the NOAEL- and LOAEL-based HIs for metals are 120 and 23, respectively. The NOAEL- based HIs for phthalates in surface and shallow soils are 2.0 and 1.4, respectively. The LOAEL-based HIs for phthalates are less than 1.0. The NOAEL- and LOAEL-based HIs for pesticides for both surface and shallow soils are 6.0 and 1.2, respectively.

American Kestrel (Table 3-11c). All NOAEL- and LOAEL-based HQs and HIs are less than 1.0 for the kestrel.

Vagrant Shrew (**Table 3-11d**). In surface soil, chromium, copper, lead, mercury, nickel, chlordane, and Total HPAHs have HQs greater than 1.0 based on the NOAELs and LOAELs. In addition to these CPECs, antimony also has an HQ greater than 1.0 based on the NOAEL.

The same CPECs listed above for surface soil also have HQs greater than 1.0 for shallow soil.

For surface soil, the NOAEL- and LOAEL-based HIs for metals are 69 and 25, respectively. For shallow soil, the NOAEL- and LOAEL-based HIs for metals are 120 and 47, respectively. All HIs for phthalates are less than 1.0. The NOAEL- and LOAEL-based HIs for pesticides for both surface and shallow soils are 3.2 and 1.6, respectively.

American Mink (Table 3-11e). All NOAEL- and LOAEL-based HQs and HIs are less than 1.0 for the mink.

3.5.1.2 Sandblast Area AOPC

Terrestrial Plants (Table 3-18b). In surface soil (0-1 ft bgs), chromium, nickel, and Total HPAHs have HQs greater than 1.0 based on the low SLVs and high SLVs. In addition to these CPECs, lead, and gamma-benzene hexachloride (BHC; lindane) also have HQs greater than 1.0 based on the low SLVs.

The same six CPECs listed above for surface soil have low SLV-based HQs greater than one for shallow soil (0-3 ft bgs), with the exception of gamma-BHC for which the HQ is less than 1.0. Chromium, nickel, and Total HPAHs have high SLV-based HQs greater than 1.0 for shallow soil.

Soil Invertebrates (Table 3-19b). In surface soil, chromium has HQs greater than 1.0 based on the low SLVs and high SLVs. In addition to chromium, mercury, nickel, and Total HPAHs also have HQs greater than 1.0 based on the low SLV.

In shallow soil, chromium has HQs greater than 1.0 based on the low SLVs and high SLVs, and the low SLV-based HQ for Total HPAHs is also greater than 1.0.



For surface soil, the low SLV- and high SLV-based HIs for metals are 29 and 26, respectively. For shallow soil, the low SLV- and high SLV-based HIs for metals are 23 and 21, respectively. No soil invertebrate SLVs are available for the pesticides; therefore, these CPECs are addressed qualitatively in the uncertainty assessment.

Canada Goose (Table 3-12a). All LOAEL-based HQs are less than 1.0 for the goose. In surface and shallow soil, chromium and lead have HQs greater than 1.0 based on the NOAELs.

For surface soil, the NOAEL- and LOAEL-based HIs for metals (including tributyltin) are 4.2 and 1.2, respectively. For shallow soil, the NOAEL- and LOAEL-based HIs for metals are 3.9 and 1.2, respectively.

American Robin (Table 3-12b). In surface soil, chromium, lead, nickel, and DEHP have HQs greater than 1.0 based on the NOAELs and LOAELs. In addition to these CPECs, cadmium, mercury, and endrin aldehyde also have HQs greater than 1.0 based on the NOAEL.

The same CPECs listed above for surface soil also have HQs greater than 1.0 for shallow soil.

For surface soil, the NOAEL- and LOAEL-based HIs for metals are 52 and 15, respectively, and the NOAEL-based HI for pesticides is 3.6. The LOAEL-based HI for pesticides is less than 1.0. In addition, the NOAEL- and LOAEL-based HIs for phthalates in surface soil are 16 and 1.6, respectively. For shallow soil, the NOAEL- and LOAEL-based HIs for metals are 46 and 14, respectively, and the NOAEL-based HI for pesticides is 3.1. The LOAEL-based HI for pesticides is less than 1.0. In addition, the NOAEL-based HI for pesticides is 3.1. The LOAEL-based HI for pesticides is less than 1.0. In addition, the NOAEL- and LOAEL-based HIs for phthalates in shallow soil are 13 and 1.3, respectively.

American Kestrel (Table 3-12c). All NOAEL- and LOAEL-based HQs and HIs are less than 1.0 for the kestrel.

Vagrant Shrew (**Table 3-12d**). In surface soil, antimony, chromium, lead, nickel, and Total HPAHs have HQs greater than 1.0 based on the NOAELs and LOAELs. In addition to these CPECs, cadmium, mercury, and DEHP also have HQs greater than 1.0 based on the NOAEL.

The same CPECs listed above for surface soil also have NOAEL-based HQs greater than 1.0 for shallow soil, with the exception of DEHP for which the NOAEL-based HQ is less than 1.0. Chromium, lead, nickel, and Total HPAHs in shallow soil also have LOAEL-based HQs greater than 1.0

For surface soil, the NOAEL- and LOAEL-based HIs for metals are 89 and 32, respectively. In addition, the NOAEL-based HI for phthalates is 1.1 for surface soil, but the LOAEL-based HI for phthalates is less than 1.0. For shallow soil, the NOAEL- and LOAEL-based HIs for metals are 70 and 25, respectively. All remaining HIs are less than 1.0.

American Mink (Table 3-12e). All NOAEL- and LOAEL-based HQs and HIs are less than 1.0 for the mink.

3.5.1.3 Pistol Range AOPC

Terrestrial Plants (Table 3-18c). Lead has a HQ of 3.0 based on the low SLV, and the high SLV-based HQ is less than 1.0.

Soil Invertebrates (Table 3-19c). The low SLV- and high SLV-based HQs for lead are less than 1.0 for soil invertebrates.



Canada Goose (Table 3-13a). Lead has a HQ of 1.2 based on the NOAEL, and the LOAEL-based HQ is less than 1.0.

American Robin (Table 3-13b). Lead has HQs of 9.2 and 4.6 based on the NOAEL and LOAEL, respectively.

American Kestrel (Table 3-13c). The NOAEL- and LOAEL-based HQs for lead are less than 1.0 for the kestrel.

Vagrant Shrew (**Table 3-13d**). Lead has HQs of 4.5 and 2.4 based on the NOAEL and LOAEL, respectively.

American Mink (Table 3-13e). The NOAEL- and LOAEL-based HQs for lead are less than 1.0 for the mink.

3.5.1.4 Bulb Slope AOPC

Terrestrial Plants (Table 3-18d). All high SLV-based HQs and the HI are less than 1.0 for the terrestrial plants. Lead and mercury have HQs of 2.6 and 2.4, respectively, based on the low SLVs. The low SLV-based HI is 5.0.

Soil Invertebrates (Table 3-19d). Mercury has HQs of 7.2 and 1.4 based on the low SLV and high SLV, respectively. The low SLV- and high SLV-based HIs are 7.2 and 1.4, respectively.

Canada Goose (Table 3-14a). All LOAEL-based HQs and the HI are less than 1.0 for the goose. Lead and mercury have HQs of 1.0 and 4.3, respectively, based on the NOAELs. The NOAEL-based HI is 5.3.

American Robin (Table 3-14b). All LOAEL-based HQs are less than 1.0 for the robin. Lead and mercury have HQs of 1.7 and 4.8, respectively, based on the NOAELs. The NOAEL- and LOAEL-based HIs are 6.5 and 1.3, respectively.

American Kestrel (Table 3-14c). All NOAEL- and LOAEL-based HQs and the HI are less than 1.0 for the kestrel.

Vagrant Shrew (Table 3-14d). All LOAEL-based HQs are less than 1.0 for the shrew. Mercury has a HQ of 2.6 based on the NOAEL. The NOAEL- and LOAEL-based HIs are 3.4 and 1.2, respectively.

American Mink (Table 3-14e). All NOAEL- and LOAEL-based HQs and the HI are less than 1.0 for the mink.

3.5.1.5 All Four AOPCs Combined

American Kestrel (Table 3-15a). All NOAEL- and LOAEL-based HQs and HIs are less than 1.0 for the kestrel.

American Mink (Table 3-15b). All NOAEL- and LOAEL-based HQs and HIs are less than 1.0 for the mink.

3.5.2 Uncertainty Assessment

Uncertainties are inherent in any risk-based approach to evaluation and decision making for potentially contaminated sites. The uncertainties may be general and systemic as well as specific to the site. The objective of the uncertainty assessment is to identify the sources of uncertainty in



the RA process, understand their potential to contribute to either underestimation or overestimation of risk for the selected receptors and pathways and describe how the uncertainty is addressed. By describing the nature and magnitude of the uncertainties, the findings and conclusions of the RA can be better understood and used as a tool for decision making.

Some uncertainties are common to both the SLERA and the BERA. Those uncertainties inherent to both were previously described in Appendix O of the Final RI (URS 2012). These included:

- Data Adequacy and Data Quality (0.1.1.1)
- Exposure Point Concentrations (0.1.1.3)
- Selection and Use of SLVs (0.1.1.4 and 0.3.1.6)
- Data Sensitivity (0.3.1.1)
- Calculation of Total HPAHs (0.3.1.4)
- Bioavailability, Absorption, and Metabolism of CPECs (0.3.1.8)

The following sections discuss uncertainties that apply specifically to the BERA.

3.5.2.1 Exposure Depth

The BERA assumed all receptors were exposed to surface soil (0-1 ft bgs), except for the Pistol Range where surface soil was defined as 0 to 1.5 ft. Shallow soil (0-3 ft bgs) exposure was also considered for the Landfill and Sandblast Areas. However, the most biologically active zone is represented by surface soil (e.g., bulk root zone for annual plants ingested by trophic receptors and depth at which earthworms are most active and preyed upon); therefore, risk estimates based on surface soil are likely to more realistically represent actual risk exposure potential for wildlife, with the exception of the burrowing shrew.

3.5.2.2 Selection of Toxicity Data

The studies upon which terrestrial plant SLVs were derived typically use crops as the test species, and sensitivity levels of undomesticated plant species are likely to be different from crops species. Measured endpoints for the phytotoxicity SLVs generally consist of growth or yield (biomass). Growth and yield are ecologically significant responses because they directly impact plant populations and the ability of the vegetation to support higher trophic levels. These phytotoxicity SLVs are conservative, and if a constituent is present in soil at a concentration greater than the SLV, yet there is a vegetative community present, then that chemical may not necessarily be phytotoxic to the plant species at the AOPCs. Furthermore, the types of plants found at the four AOPCs (ruderal vegetation and some remaining ornamental plants) are not considered sensitive species. Based on these considerations, the conservative SLVs used to assess risk to plants are likely to overestimate risk to the plant community.

Preferred wildlife TRVs were chronic NOAEL and/or LOAEL TRVs from the established hierarchy. If only a NOAEL or LOAEL TRV was provided from the preferred source, then, as recommended in ODEQ's comments on the Level II Screening Assessment for the Landfill (ODEQ 2004), TRVs for CPECs that lack a NOAEL or LOAEL were generated by either multiplying or dividing the available TRV by a factor of five, depending on the desired TRV. TRVs extrapolated from these sources represent generally conservative values drawn from a review of the toxicological literature. For some CPECs, additional sources were consulted in



order to obtain TRVs not provided for in the established hierarchy, and in some of these cases, uncertainty factors were used to calculate the final NOAEL and LOAEL TRVs for extrapolation of subchronic, acute, or LD50 test result to a chronic duration test result. The uncertainty factors used to adjust to chronic duration were based on USACHPPM Technical Guide 254 - Standard Practice for Wildlife TRVs (USACHPPM 2000). Although the use of these factors is standard practice in RAs, it is a source of uncertainty because linear adjustments to risk are unlikely for all CPECs and actual risk may be over- or under-predicted.

TRVs for avian and mammalian species were selected to assess risk to the selected representative receptors. However, these TRVs were not species-specific to the selected representative receptors, and species respond differently to exposures to toxicants. Responses to CPEC exposure by selected indicator species may be different from species for which toxicity data were reported. Direction and magnitude of this uncertainty are not measurable, although the choice of conservative TRVs was conducted in an attempt to skew the evaluation toward more protective conclusions.

3.5.2.3 Lack of Toxicity Data

Toxicity data were not available for a limited number of CEPCs. Those constituents for which toxicity data (SLVs) were not available for plants, soil invertebrates, or both are summarized below and identified in Table 3-16:

- Tributyltin
- Pesticides (limited to endosulfans and methoxyclor for plants and all 13 pesticides for soil invertebrates)
- SVOCs (limited to four SVOCs for plants, and all six SVOCs for soil invertebrates)
- Three VOCs

Those constituents for which toxicity data were not available for birds are summarized below and identified in Table 3-17:

- HPAHs
- Four SVOCs,
- Two VOCs

Unavailable toxicity data could cause underestimation of risk for those CPECs where surrogate toxicity data are not assigned. However, the presence of ruderal and ornamental plants in these AOPCs suggests that direct toxicity is not occurring in the plant species typically found in disturbed habitats.

The persistence of the pesticides in soil will be affected by microbial degradation and chemical degradation and are likely to decrease over time. Pesticide resistance by invertebrates is a well-established phenomenon, whereby individuals with higher resistance are the ones that have offspring and pass the resistance on to their offspring (Bellinger 1996). Due to this, it is unlikely that the residual pesticides detected in soil are highly toxic to the soil invertebrates present in the soil. Additionally, since the avian TRVs are very conservative (i.e., based on effects to eggs for receptors with known sensitivity), any future activities to address the risk to birds for chlordane



will address the locations with the highest pesticide concentrations and thereby decrease the risk to soil invertebrate communities.

Although several SVOCs and VOCs lack plant/invertebrate and bird toxicity data, these CPECs volatilize, and the concentration of these CPECs in soil is expected to continue to decrease over time (it is likely the concentrations currently present are lower than the concentrations used in the BERA due to the age [collected in 2008] of the data). The assumption that these SVOCs and VOCs bioaccumulate in wildlife is based on the logK_{ow}s, which are predicted values based on a model (not empirically derived) and many of which are only marginally above 3.5 (see J-6 in the Final RI, URS 2012). Whereas, the empirically modeled BAFs for these SVOCs and VOCs from plants are actually low (<0.5, see Table 3-8b). Since several SVOCs and VOCs with established toxicity were evaluated, continued volatilization, and considering the questionable bioaccumulation of these CPECs, the lack of toxicity data is unlikely to significantly impact the BERA findings or impact any risk management decisions in the Upland OU FS.

In the absence of reliable HPAH toxicity data for birds, the potential hazards to birds exposed to PAHs were addressed qualitatively in Section O.3.1.6 in the Final RI (URS 2012), and it was concluded that the lack of PAH TRVs for birds would not significantly impact the RA findings.

3.5.2.4 Selection of Mink as Terrestrial Large Mammal

According to Section 3.5 of the Level I Scoping Assessment that was performed for the Landfill (URS 2002), which included a thorough biological characterization of the Landfill and all habitats on the island, "large mammalian predators do not occur on the island." The only mammals on the island that are mentioned in the Scoping Assessment are small mammals (rodents) and feral cats: "Although the island harbors small mammals, feral cats, Canadian geese, and other bird species, the minimal amount of available habitat (~12 acres) makes it unsuitable for supporting viable populations of wildlife species with larger home ranges."

For this reason, large mammals were not included in the assessment endpoints described in the approved RI MP (URS 2007). However, as discussed in Section 3.2 and described in detail in the WP Update, mink were evaluated at the request of ODEQ to include a large terrestrial mammal (see Appendix P of Final RI [URS 2012]) and because mink are present in the area, are sensitive to environmental contaminants, and could feasibly access the island and forage there, exposure by this species through consumption of rodents was included in the BERA.

A dietary composition of 15% small mammals from the Upland OU was assumed in the BERA; however, this is a conservative estimate because small mammals are likely to comprise a lower proportion of the mink's actual diet based on the presence of an immediately available permanent water source (i.e., riverine habitat) and according to the studies reviewed during the development of the WP Update (i.e., it is likely that a mink would primarily be attracted to foraging resources, e.g., crayfish and fish) in the River OU.

3.5.2.5 Use of Input Exposure Parameters

A detailed explanation of the selection of the input exposure parameters selected for the BERA can be found in the RI MP (URS 2007) and the Upland OU Update to RA WP Technical Memorandum (URS 2014). Resources used to determine input exposure parameter values included USEPA's Wildlife Exposure Factors Handbook (USEPA 1993), Food Requirements of Wild Animals: Predictive Equations for Free-living Mammals, Reptiles, and Birds (Nagy 2001), Estimates of Soil Ingestion by Wildlife (Beyer et al. 1994), California Wildlife Biology,



Exposure Factor, and Toxicity Database (Cal/ECOTOX, 2002), and California's Wildlife: Volume II, Mammals (Zeiner et al. 1990).

There can be variability in input rates among individuals within a species, between species, between soil types, and with the type and quantity of food items available. The direction and magnitude of the uncertainty associated with these variables is not measurable. For most receptors, a lower-end average body weight was paired with an upper-end average ingestion rate, resulting in exposure equations. These assumptions may overestimate or underestimate actual "real world" intake since lower-end body weights would be protective of small adult receptors, but may underestimate risk for young individuals and overestimate risk for larger adult individuals.

3.5.2.6 Dietary Item Assumptions

For the BERA, the terrestrial receptors that typically consume more than one type of food item (i.e., the American robin and the American kestrel) were conservatively assumed to consume only the food item that comprises the majority of their diet. However, this assumption of exclusive intake of a single prey item to estimate risk to a category of receptor (e.g., invertebrate birds) may over- or underestimate risk in those species that consume more than one type of dietary item.

3.5.2.7 Landfill Heterogeneity

For approximately 40 years (early 1940s until the early 1980s), the USACE managed, stored and disposed of waste materials at the landfill in excavated pits or existing depressions. By 1982, the surface of the Landfill AOPC had been capped with soil cover. In 1989, approximately 8-inches of additional soil cover was placed on the Landfill site by the USACE (Hibbs, personnel comm. 2001). Because the waste was buried in separate pits, rather than one continuous pit, the heterogeneity of the landfill composition makes contamination characterize difficult. Although sampling targeted areas known (via historical aerial review or previously exposed areas) or suspected (via electrical resistivity data and seismic refraction data) as having the greatest landfilling activity, risk may be underestimated due to the heterogeneous nature of the landfill.

3.5.3 Calculation of Site-Specific Risk-Based Concentrations

To aid in risk interpretation, site-specific RBCs were derived for each AOPC and wildlife receptor for the CECs with HQs greater than 1.0 based on LOAEL TRVs. All LOAEL-based HQs were less than 1.0 for the kestrel and mink; therefore, RBCs were only calculated for the robin and shrew. By selecting the most sensitive receptors to derive the RBCs, it is likely that the other receptor groups would also be protected. The following equation was used to calculate the RBCs for the robin and shrew:

Site - Specific RBC =
$$\frac{TRV \times BW}{\left[\left(IR_{food} \times BAF\right) + IR_{soil}\right]}$$

where,

RBC = Site-specific risk-based concentration for soil (mg chemical per kg soil dry weight)



| TRV | = LOAEL-based TRV (mg chemical ingested per kg BW per day) |
|--------------------|---|
| BW | = Body weight (kg) |
| IR_{food} | = Ingestion rate of food, as represented by soil invertebrate tissue (kg dry |
| | weight food per day) |
| BAF | = Soil to terrestrial invertebrate bioaccumulation factor (kg soil per kg tissue) |
| IR _{soil} | = Ingestion rate of soil (kg dry weight soil per day) |
| | |

The exposure factors for the robin and shrew used in this equation are presented in Tables 3-2 and 3-4, respectively, and the LOAEL TRVs are presented in Table 3-17. Due to the low contribution of water ingestion to the overall dose for these receptors (less than 5%, but varies by CPEC), this pathway was not included in the RBC calculations. The final site-specific RBCs for each receptor are presented in Table 3-20. Additionally, Table 3-20 presents the high SLVs for plants and/or soil invertebrates per AOPC for the CECs with high-SLV based HQs greater than 1.0. In cases where LOAEL RBCs/high SLVs for metals are below the site-specific background UPLs (e.g., chromium and nickel), the UPL was selected as the default RBC.

3.5.4 Risk Interpretation

In this final phase of the risk characterization process, the quantitative and qualitative components of the BERA are evaluated to characterize the potential for ecological risk. The actual risk drivers at each AOPC and the extent of impacts for these CECs are identified to develop supportable recommendations for risk managers to review. The outcome of the risk characterization will constitute the basis of remedial decisions for the protection of ecological receptors and risk-driving receptors and exposure pathways.

For each AOPC, the CECs with exceedances of the lowest RBC (based on the high SLV/LOAEL TRVs) for any receptor group are plotted in Figures 3-6 through 3-9. The locations with exceedances of the lowest site-specific RBC for each CEC from Table 3-20 are identified with a data posting box that provides the concentrations of the CECs that exceed the RBCs at that particular location. For chromium at the Landfill and Sandblast Areas, the lowest RBC between the robin and shrew was selected for purposes of the mapping exercise because these two receptors drove risk for all other CECs, and the chromium SLVs for plants and invertebrates are below background; the background UPL was the default SLV for these two receptor groups. Sample locations without a data posting box indicate that CEC concentrations are below the RBCs at that particular location.

The spatial distribution of CEC concentrations relative to the RBCs was assessed through a review of this information on figures and the findings are described in the following subsections.

3.5.4.1 Landfill AOPC

The CECs identified for additional evaluation at the Landfill due to the potential for adverse effects to populations of invertivorous birds or mammals, represented by the robin and shrew, or community-level impacts to plants and invertebrates include the following: chromium, copper, lead, mercury, nickel, chlordane, and Total HPAHs. Figure 3-6 illustrates the locations with exceedances of the lowest site-specific RBC for each CEC and locations of the five subareas: Gully Test Pit, Lead Hotspot Test Pits #1 and #2, the Mercury Vapor Lamp Test Pit, and the Pesticide/Herbicide Wash Area.



Maximum concentrations of chromium, lead, and nickel were detected below 1 ft bgs inside northern boundary of the Gully Test Pit (BIL18), while the maximum concentrations of copper, mercury, most Total HPAHs, and chlordane were detected at the surface. Elevated detections of Total HPAHs also occur below the surface, as the mean and median Total HPAH concentrations are associated with the deeper depth interval (Table 1-1). The maximum concentration of Total HPAHs was detected within Lead Hot Spot Test Pit #1 (BIL04SSI).

The maximum concentration of chlordane, and the only detection above the RBC, was also detected within the northern boundary of the Gully Test Pit (BIL17). No pesticides were retained in the pesticide washing area (i.e., chlordane concentrations < LOAEL-based RBCs). The maximum concentration of mercury was detected in the center of Lead Hot Spot Test Pit #2 (BIL06SSI), and the maximum concentration of copper was detected outside of the eastern boundary of Lead Hot Spot Test Pit #1 (BIL05SSI).

In addition to the maximum detection, concentrations of chromium are elevated above the RBC in two more locations along northern boundary of the Gully Test Pit, and at one more location outside of the eastern boundary of Lead Hot Spot Test Pit #1. Lead concentrations are elevated at 21 locations in addition to the maximum detection. These samples are distributed evenly across the Landfill, with much higher concentrations present along northern boundary of the Gully Test Pit, within Lead Hot Spot Test Pit #1, and at one location northeast of the Pesticide/Herbicide Wash Area.

Concentrations of Total HPAHs are elevated at 16 locations in addition to the maximum detection. Similarly to lead, these samples are distributed evenly across the Landfill, with much higher concentrations within and around the Gully Test Pit and in a sample collected from the Mercury Vapor-Lamp Test Pit.

The Reference Area UPL is the default RBC for nickel (Table 3-20), and seven locations have concentrations above this RBC in addition to the maximum detection of nickel. Like most of the other CECs, the highest concentrations of nickel were detected along the northern boundary of the Gully Test Pit and within Lead Hot Spot Test Pit #1. Concentrations of copper were elevated above the RBC at three locations in addition to the maximum detection. One of these samples was collected from the Gully Test Pit, another from the Mercury Vapor-Lamp Test Pit (co-located with the maximum mercury detection), and the third just west of the Lead Hot Spot Test Pit #1.

Concentrations of mercury are elevated at seven locations in addition to the maximum detection. Unlike the other CECs discussed above, mercury concentrations are most elevated in the western portion of the Landfill where seven of the eight elevated locations are located within and just west of the Mercury Vapor-Lamp Test Pit. The remaining single elevated sample location is northeast of the Pesticide/Herbicide Wash Area.

In summary, the northwestern section of the Landfill, primarily within and around the Gully Test Pit and Lead Hot Spot Test Pit #1, is the most impacted area of this AOPC with co-occurrences of most CECs. In addition, the most impacted area of the Landfill for mercury is the Mercury Vapor-Lamp Test Pit and one isolated spot northeast of the Pesticide/Herbicide Wash Area. These areas of the Landfill should be further evaluated in the Upland OU FS.



3.5.4.2 Sandblast Area AOPC

The CECs identified for additional evaluation at the Sandblast Area due to the potential for adverse effects to populations of invertivorous birds or mammals, represented by the robin and shrew, or community-level impacts to plants and invertebrates include the following: antimony, chromium, lead, nickel, DEHP, and total HPAHs. Figure 3-7 illustrates the locations with exceedances of the lowest site-specific RBC for each CEC and the locations of the subareas: Equipment Laydown, Sandblast Grit Disposal Area, Current and Former HMSA, Catch Basins #1 through #4, and the Erodible Unit Sampling Area.

Maximum concentrations of most Total HPAHs were detected below 1 ft bgs, while the maximum concentrations of the remaining CECs were detected at the surface. Elevated detections of some Total HPAHs also occur at the surface. The maximum concentration of Total HPAHs was detected below the surface in a sample collected adjacent to the southwestern corner of the Current HMSA (SBB18) and within the approximate boundary of the Spent Sandblast Grit Disposal Area.

Concentrations of Total HPAHs are elevated at 11 locations in addition to the maximum detection. Most of these samples are located in the northeastern corner of the Sandblast Area within the Equipment Laydown Area, near the Former and Current HMSAs, and in a composite sample collected from the Erodible Unit Sampling Area.

The maximum concentrations of antimony and nickel were also detected within the Spent Sandblast Grit Disposal Area: antimony in a sample adjacent to the southeast corner of the Sandblast Building (SBB12) and nickel in a sample south of the Equipment Laydown Area (SBB23). The maximum concentration of chromium was detected on the northwestern boundary of the Spent Sandblast Grit Disposal Area (HA6), and the maximum concentration of lead was detected within the Equipment Laydown Area near the northern boundary of the Sandblast Area AOPC (HA3). The maximum concentration of DEHP, and the only detection above the RBC, was detected in a composite sample collected from the Erodible Unit Sampling Area (SB-EUB).

In addition to the maximum detection, concentrations of antimony are elevated above the RBC at two more locations (one near the Former and one near the Current HMSA), and at one more location northeast of the Sandblast Building. Chromium concentrations are elevated at 21 locations in addition to the maximum detection, with the highest concentrations present within the Spent Sandblast Grit Disposal Area, and in localized areas of the Erodible Unit Sampling Area, Catch Basin #1, and Former HMSA.

Lead and nickel concentrations are elevated in 51 sample locations and 36 sample locations, respectively, in addition to the maximum detections. These samples are distributed fairly evenly across the Sandblast Area, with much higher concentrations present within the Spent Sandblast Grit Disposal Area, around Catch Basin #1, within the Equipment Laydown Area, and in one sample from the Former HMSA. More elevated lead detections are prevalent within the Equipment Laydown Area than elevated nickel detections. The Reference Area UPL is the default RBC for nickel (Table 3-20).

In summary, most co-occurring exceedances of the RBCs occur within the Spent Sandblast Grit Disposal Area, which basically encompasses the Current HMSA, the northern boundary of the Former HMSA, and around Catch Basin #1. In addition, concentrations of chromium and nickel greater than background, lead, and HPAHs are elevated within the Equipment Laydown Area.



These areas of the Sandblast Area should be further evaluated in the Upland OU FS. Soil removal in these areas would likely reduce concentrations of the CECs identified and result in an AOPC-wide average concentration below the lowest site-specific RBCs.

For DEHP, the LOAEL-based HQs for all receptors, with the exception of the robin, are below 1.0, and the LOAEL-based HQs for the robin are low (1.6 for 0-1 ft bgs and 1.3 for 0-3 ft bgs). In addition, only one sample exceeds the RBC for the robin. For these reasons, no further evaluation is recommended for DEHP, with the exception of the potential erodibility issue that will be considered in the Upland FS or River FS.

3.5.4.3 Pistol Range AOPC

Lead was identified for additional evaluation at the Pistol Range due to the potential for adverse effects to populations of invertivorous birds or mammals, represented by the robin and shrew. Figure 3-8 illustrates the locations with exceedances of the lowest RBC for lead.

The highest concentrations of lead, including the maximum concentration that was detected below the surface (PFR50), were detected at and behind the approximate location of the backstop. Twenty additional locations have concentrations above the RBC in addition to the location with the maximum detection. In addition to the area surrounding the backstop, these locations also occur north and east of the approximate location of the Former Firing Shed. These areas of the Pistol Range should be further evaluated in the Upland OU FS. Limited soil removal in these areas would likely reduce concentrations of the CECs identified and result in an AOPC-wide average concentration below the lowest site-specific RBC.

3.5.4.4 Bulb Slope AOPC

Mercury was identified for additional evaluation at the Bulb Slope due to the potential for community-level impacts to soil invertebrates. Figure 3-9 illustrates the locations with exceedances of the mercury RBC for invertebrates.

The maximum concentration of mercury was detected in a sample collected from Pile #3, Bank #4, which is located near the southwestern boundary of the Bulb Slope, i.e., upgradient and away from the boundary closest to the Columbia River. The remaining two samples with exceedances of the RBC for soil invertebrates were also collected from the southern boundary of this AOPC. All three samples with elevated mercury are bound by samples with concentrations below the RBC collected downgradient of the slope, closest to the river.

The high SLV/LOAEL-based HQs for all receptors, with the exception of soil invertebrates, are below 1.0, and the high SLV-based HQ for invertebrates is low (1.4). The topography of the Bulb Slope is steep and the substrate consists of a mixture of soils, rock that may have been placed in some areas, and what appear to be natural rock outcrops, all of which are underlain by siltstone bedrock. The majority of the Bulb Slope AOPC is herbaceously vegetated and/or covered with organic debris.

For these reasons, no further upland evaluation is recommended for the Bulb Slope.

3.5.4.5 All Four AOPCs Combined

As all of the NOAEL- and LOAEL-based HQs and HIs are less than 1.0 for the wide-ranging receptors (kestrel and mink), no additional evaluation is needed for the combined AOPC exposure area.



3.6 Conclusions and Recommendations

Table 3-21 summarizes the AOPCs and CECs identified through the BERA that are recommended for further evaluation in the Upland OU FS. No further upland evaluation of the Bulb Slope is recommended based on the low risk estimates. Further evaluation in the FS is recommended for the three remaining AOPCs based on multiple exceedances of RBCs for one or more ecological receptor.

The site-specific RBCs were used in Section 3.5.4 to identify the areas within each AOPC that warrant specific attention in the Upland OU FS, which are summarizes as follows:

- Landfill northwestern section of the Landfill, primarily within and around the Gully Test Pit and Lead Hot Spot Test Pit #1, Mercury Vapor-Lamp Test Pit, and two locations within and northeast of the Pesticide/Herbicide Wash Area.
- Sandblast Area Spent Sandblast Grit Disposal Area, northern boundary of the Former HMSA, Catch Basin #1, localized areas within the Erodible Unit Sampling Area, and Equipment Laydown Area (north and south of the northern roadway).
- Pistol Range around the approximate location of the backstop, and north and east of the approximate location of the Former Firing Shed.

To protect ecological receptors, AOPC-wide average concentrations for each AOPC should be below the lowest site-specific RBCs. When the 95% UCL concentrations for the identified CECs are below the RBCs for each AOPC, acceptable levels of ecological risk will be achieved.



4.0 SUMMARY

This report presents the Upland OU BHHRA and BERA for the Bradford Island, Bonneville Dam Complex. The Portland District of the USACE has characterized and evaluated the contamination arising from historical practices at Bradford Island, a multipurpose facility located at RM 146.1 that consists of the First and Second Powerhouses, the old and new navigation locks, and a spillway with a capacity of 1.6 million cubic feet per second.

USACE and their contractors have performed numerous investigations since 1997, focusing on the Upland and River OUs. The Final RI (URS 2012) documented the investigation, identified source areas at Bradford Island, defined the nature and extent of the environmental contamination, and identified the COPCs for human health and CPECs in the media from the two OUs. Based on the screening level RAs for each of the AOPCs that were completed as part of the RI and the recent addition of the Hypothetical Platform Fisher receptor, site-specific BHHRAs and BERAs were conducted for all four AOPCs.

The Upland OU BHHRA and BERA build upon on the data and findings of the Final RI and will be attached as an appendix to the Upland OU FS. This RA document goes beyond the traditional assessment of the presence/absence and magnitude of baseline risk. To maximize use of these site datasets and develop an RA that is most beneficial to the Upland OU FS, site-specific RBCs were calculated for the chemicals recommended for further evaluation in the FS. Exceedances of these RBCs were illustrated for purposes of risk interpretation and to allow for general observations of the spatial distribution of potentially impacted areas.

Only upland exposure pathways were addressed in these baseline RAs. The Upland OU to River OU pathways (i.e., potential mass wasting, soil erosion) that were evaluated at a screening level in the Final RI (URS 2012) were not addressed, as these possible pathways will be considered in the Upland FS or River FS.

4.1 Baseline Human Health Risk Assessment

The Landfill and Sandblast Area AOPCs were evaluated in the BHHRA for occupational receptors exposed to soil, groundwater, and/or soil gas while engaged in outdoor maintenance, construction activities, or indoor office work (Sandblast Area AOPC only). All four AOPCs were evaluated for future Hypothetical Fishing Platform Users exposed to soil while engaged in associated camping activities under conditions representing current concentrations in soil.

Both RME and CTE scenarios were evaluated for most of the receptors, and ELCR and noncancer HI were evaluated with respect to whether they were:

- 1) less than the USEPA's risk level of $1 \ge 10^{-06}$ or HI of 1, whereby risk at or below this threshold has an insignificant contribution to risk (i.e., *de minimis*);
- 2) within USEPA's acceptable risk range of 1×10^{-4} to 1×10^{-6} for ELCR; or
- 3) exceeding the USEPA acceptable risk range, i.e., greater than 1×10^{-04} .

In addition, ODEQ's acceptable risk thresholds of 1×10^{-6} for individual carcinogens and 1×10^{-5} for multiple carcinogens were also considered. Due to this, individual chemicals associated with risk levels greater than 1×10^{-6} or noncancer HQ greater than 1 were identified as COCs, except for arsenic, for which background levels were also considered. Since the individual



compounds that make up the cPAHs all act by the same mode of action, risks for PAHs were calculated by individual compound and also summed to present risks for cPAHs as a group. It is important to note that, since ELCRs are only expressions of likelihood of cancer incidence and HIs are estimated ratios to safe doses, exceedance of the ODEQ risk thresholds or USEPA acceptable risk range does not automatically mean that adverse effects may have occurred or will occur (e.g., does not automatically mean that the COC will be recommended for further evaluation in the FS).

As shown in Tables 1-1 to 1-8 and Tables 2-1 to 2-4, media and COPCs evaluated at the Landfill included metals (arsenic, chromium, lead), cPAHs, and CVOCs including PCE, TCE, and associated degradation products in soil and groundwater. Media and COPCs for the Sandblast Area included similar COPCs in soil, groundwater, and soil gas. The receptors evaluated included Outdoor Maintenance Worker (exposed to shallow soil of 0-3 ft bgs), Construction Worker (exposed to deeper soils of 0-10 ft bgs), Excavation/Trench Worker (incidental exposure to groundwater in a trench), and Indoor Worker exposed by vapor intrusion (Sandblast Area AOPC only). An additional Hypothetical Fishing Platform User consisting of adults and children who might camp on the island while utilizing fishing platforms on the Island in accordance with tribal treaty rights was also evaluated (exposed to shallow soils at 0-3 ft bgs) for the longer list of COPCs. Nursing infants who may be exposed by ingestion of maternal milk were also evaluated for PCBs, which are bioaccumulative COPCs.

Table 2-26 presents a summary of the risks and noncancer hazards for the media and receptors evaluated at all the AOPCs. For occupational receptors, estimated risks were either below ODEQ's threshold of 1 x 10^{-6} or within the USEPA acceptable risk range at both AOPCs for all exposure media. At both the Landfill and Sandblast Area AOPCs, RME exposure to surface and deeper soils resulted in risks that fell within the USEPA acceptable risk range for Outdoor Maintenance Workers and Construction Workers. Risks for Construction Workers were lower than ODEQ's acceptable risk level of 1 x 10^{-5} for multiple carcinogens. Incidental exposure to groundwater for Excavation/Trench Workers was *de minimis* (less than 1 x 10^{-6}) at both AOPCs. Risks related to vapor intrusion and indoor inhalation were also *de minimis* for both current and future Indoor Workers at the Sandblast Area AOPC. Only Outdoor Maintenance Workers at the Landfill AOPC had a CTE risk within the USEPA acceptable risk range and for all other receptors, CTE risks were insignificant contributors to risk.

For the Hypothetical Fishing Platform User, estimated risks exceeded the USEPA acceptable risk range under current conditions. Risks and noncancer hazards for nursing infants were close to or less than the ODEQ threshold of 1×10^{-6} for selected bioaccumulative COPCs.

Table 2-27 summarizes the COCs identified for further evaluation in the Upland OU FS in surface and shallow soils at the Landfill and Sandblast Area AOPCs. Within each AOPC, many of the COCs are spatially co-located.

No further upland evaluation or action is warranted to address lead at the Bulb Slope and Pistol Range and the sediments in the Pistol Range Lagoon. No further evaluation or action is warranted for groundwater at either the Bulb Slope or Pistol Range AOPC, or for soil gas at the Sandblast Area AOPC.

Areas that are recommended for further evaluation or potential risk management in the Upland OU FS based on exceedances of receptor-specific RBCs include:



- Landfill northwestern section of the Landfill, primarily within and around the Gully Test Pit and Lead Hot Spot Test Pit #1, Mercury Vapor-Lamp Test Pit, far eastern portion of the Landfill, and one isolated spot northeast of the Pesticide/Herbicide Wash Area (Figures 2-5, 2-7).
- Sandblast Area the Erodible Unit, Current HMSA, Equipment Laydown Area, and the Former HMSA (Figures 2-6, 2-8).

To protect human receptors, the overall 95% UCL for each AOPC should fall below the relevant RBC. When this is achieved, acceptable levels of risk to human health will be achieved.

4.2 Baseline Ecological Risk Assessment

Surface and shallow soil at all four Upland OU AOPCs were evaluated in the BERA. The CPECs evaluated in the BERA include metals, total HPAHs, tributyltin, organochlorine pesticides, VOCs, and SVOCs for the Landfill and Sandblast Area, lead for the Pistol Range AOPC, and lead and mercury for the Bulb Slope AOPC (Tables 1-1 through 1-5). All of these CPECs were included in the Upland OU-wide evaluation in which wide-ranging receptors were assumed to forage in all four AOPCs combined. Risk estimates were calculated for each CPEC for all selected receptors (terrestrial plants, soil invertebrates, Canada goose, American robin, American kestrel, vagrant shrew, and American mink) potentially present at a given AOPC.

Both low SLVs/NOAELs and high SLVs/LOAELs were selected for each receptor in order to develop a range of HQs for consideration by risk managers. Table 3-21 summarizes the AOPCs and CECs identified in surface and/or shallow soil through the BERA that are recommended for further evaluation in the Upland OU FS.

No further upland evaluation of the Bulb Slope is recommended based on the low risk estimates. Further evaluation in the Upland OU FS is recommended for the three remaining AOPCs based on multiple exceedances of site-specific RBCs for one or more ecological receptor. All of the NOAEL- and LOAEL-based HQs and HIs are less than 1.0 for the wide-ranging receptors (kestrel and mink), and no additional evaluation is needed for these receptors on an individual AOPC basis or on a combined AOPC/Upland OU-wide basis.

Site-specific ecological RBCs were used to identify the areas within each AOPC that warrant specific attention in the Upland OU FS, which are summarized as follows:

- Landfill northwestern section of the Landfill, primarily within and around the Gully Test Pit and Lead Hot Spot Test Pit #1, Mercury Vapor-Lamp Test Pit, and two locations within and northeast of the Pesticide/Herbicide Wash Area.
- Sandblast Area Spent Sandblast Grit Disposal Area, northern boundary of the Former HMSA, Catch Basin #1, localized areas within the Erodible Unit Sampling Area, and Equipment Laydown Area (north and south of the northern roadway).
- Pistol Range around the approximate location of the backstop, and north and east of the approximate location of the Former Firing Shed.

To protect ecological receptors, AOPC-wide average concentrations for each AOPC should be at or below the lowest site-specific RBCs. When the 95% UCL concentrations for the identified CECs are below the RBCs for each AOPC, acceptable levels of ecological risk will be achieved.



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TABLES

Table 1-1Statistical Summary for Soil - Landfill AOPC

| | | | | | | | | | | | | | St | tatistics - ND | censored at the MDL |
|--------|------------------|---------------|-----------|-------------------------------------|----------|-------|-----------|------------|----------|-----------|---------------------|----------|-----------|----------------|----------------------------------|
| | | | Total/ | | Depth | | Number of | Number of | Mean of | Median of | Maximum Detected | | | Final 95% | |
| Medium | Retention Reason | Analyte Group | Dissolved | Analyte | Category | Unit | Samples | Detections | Detects | Detects | Value | KM-Mean | KM-SD | UCL | UCL Type |
| Soil | | Metals | Total | Antimony | 0-1 ft | mg/kg | 20 | 4 | 2.80 | 3.35 | 4.20 | 0.815 | 1.21 | 1.36 | 95% KM (t) UCL |
| Soil | | Metals | Total | Antimony | 0-3 ft | mg/kg | 22 | 6 | 2.44 | 2.71 | 4.20 | 0.958 | 1.23 | 1.48 | 95% KM (t) UCL |
| Soil | | Metals | Total | Arsenic | 0-3 ft | mg/kg | 23 | 23 | 5.30 | 3.70 | 30.1 | - | - | 10.5 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | Metals | Total | Chromium | 0-1 ft | mg/kg | 21 | 21 | 77.2 | 23.0 | 801 | - | - | 242 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | Metals | Total | Chromium | 0-3 ft | mg/kg | 23 | 23 | 190 | 23.1 | 1,950 | - | - | 594 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | Metals | Total | Copper | 0-1 ft | mg/kg | 21 | 21 | 73.6 | 45.4 | 494 | - | - | 170 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | Metals | Total | Copper | 0-3 ft | mg/kg | 23 | 23 | 85.2 | 45.4 | 494 | - | - | 191 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | Metals | Total | Lead | 0-1 ft | mg/kg | 21 | 21 | 211 | 131 | 741 | - | - | 332 | Use 95% Approximate Gamma UCL |
| Soil | | Metals | Total | Lead | 0-3 ft | mg/kg | 29 | 29 | 342 | 147 | 1,660 | - | - | 511 | Use 95% Approximate Gamma UCL |
| Soil | | Metals | Total | Mercury | 0-1 ft | mg/kg | 23 | 18 | 0.524 | 0.160 | 4.15 | 0.427 | 0.851 | 1.57 | 97.5% KM (Chebyshev) UCL |
| Soil | | Metals | Total | Mercury | 0-3 ft | mg/kg | 25 | 20 | 0.478 | 0.126 | 4.15 | 0.397 | 0.823 | 1.45 | 97.5% KM (Chebyshev) UCL |
| Soil | | Metals | Total | Nickel | 0-1 ft | mg/kg | 21 | 21 | 58.4 | 20.0 | 570 | - | - | 175 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | Metals | Total | Nickel | 0-3 ft | mg/kg | 23 | 23 | 148 | 20.2 | 1,610 | - | - | 472 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | Butyltins | Total | Tributyltin | 0-1 ft | ug/kg | 7 | 3 | 59.7 | 9.01 | 165 | 28.5 | 55.7 | - | - |
| Soil | | Butyltins | Total | Tributyltin | 0-3 ft | ug/kg | 9 | 4 | 48.6 | 12.2 | 165 | 24.5 | 49.8 | 60.1 | 95% KM (t) UCL |
| Soil | | PCB Aroclors | Total | Total PCBs as Aroclors (NDs at MDL) | 0-3 ft | ug/kg | 29 | 24 | 174.7 | 103.9 | 996 | 153.7 | 215.9 | 409.9 | 97.5% KM (Chebyshev) UCL |
| Soil | | Herbicides | Total | MCPP | 0-3 ft | ug/kg | 14 | 2 | 11800.00 | 11800.00 | 14000.00 | 10000.00 | 1410.0000 | 11200.00 | 95% KM (t) UCL |
| Soil | | Pesticides | Total | Chlordane (technical) | 0-1 ft | ug/kg | 1 | 1 | 1,560 | 1,560 | 1,560 | - | - | - | - |
| Soil | | Pesticides | Total | Heptachlor | 0-1 ft | ug/kg | 21 | 1 | 2.83 | 2.83 | 2.83 | - | - | - | - |
| Soil | | Pesticides | Total | Chlordane (technical) | 0-3 ft | ug/kg | 3 | 3 | 573 | 92.7 | 1,560 | - | - | - | - |
| Soil | | Pesticides | Total | Heptachlor | 0-3 ft | ug/kg | 23 | 2 | 2.95 | 2.95 | 3.07 | 2.84 | 0.0489 | 2.86 | 95% KM (t) UCL |
| Soil | | SVOCs | Total | Bis(2-ethylhexyl) Phthalate | 0-1 ft | ug/kg | 27 | 24 | 2,010 | 425 | 21,000 | 1,800 | 4,030 | 9,680 | 99% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Bis(2-ethylhexyl) Phthalate | 0-3 ft | ug/kg | 29 | 26 | 1,970 | 595 | 21,000 | 1,780 | 3,890 | 6,380 | 97.5% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Butyl Benzyl Phthalate | 0-1 ft | ug/kg | 26 | 3 | 52.6 | 67.2 | 68.7 | 28.1 | 15.6 | 36.6 | 95% KM (t) UCL |
| Soil | | SVOCs | Total | Butyl Benzyl Phthalate | 0-3 ft | ug/kg | 28 | 3 | 52.6 | 67.2 | 68.7 | 27.4 | 14.8 | 34.9 | 95% KM (t) UCL |
| Soil | | SVOCs | Total | Carbazole | 0-1 ft | ug/kg | 27 | 19 | 384 | 144 | 2,650 | 279 | 547 | 751 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Carbazole | 0-3 ft | ug/kg | 29 | 21 | 510 | 210 | 2,840 | 378 | 706 | 964 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Dibenzofuran | 0-1 ft | ug/kg | 27 | 11 | 160 | 67.7 | 810 | 84.0 | 159 | 152 | 95% KM (BCA) UCL |
| Soil | | SVOCs | Total | Dibenzofuran | 0-3 ft | ug/kg | 29 | 13 | 174 | 76.0 | 810 | 96.2 | 165 | 164 | 95% KM (BCA) UCL |
| Soil | | SVOCs | Total | Di-n-butyl Phthalate | 0-1 ft | ug/kg | 26 | 4 | 496 | 84.5 | 1,800 | 98.5 | 342 | 232 | 95% KM (t) UCL |
| Soil | | SVOCs | Total | Di-n-butyl Phthalate | 0-3 ft | ug/kg | 28 | 4 | 496 | 84.5 | 1,800 | 93.6 | 330 | 217 | 95% KM (t) UCL |
| | | SVOCs | Total | Benzo(a)anthracene | 0-1 ft | ug/kg | 27 | 25 | 3,200 | 960 | 32,000 | 2,970 | 6,280 | 8,340 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(a)anthracene | 0-3 ft | ug/kg | 33 | 31 | 4,600 | 1,400 | 32,000 | 4,330 | 7,300 | 9,960 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(a)pyrene | 0-1 ft | ug/kg | 27 | 25 | 3,440 | 1,200 | 33,000 | 3,180 | 6,450 | 8,700 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(a)pyrene | 0-3 ft | ug/kg | 33 | 31 | 5,200 | 1,600 | 34,000 | 4,890 | 8,130 | 11,200 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(a)pyrene | 0-10 ft | ug/kg | 44 | 42 | 5,890 | 3,240 | 34,000 | 5,620 | 7,740 | 10,800 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(b)fluoranthene | 0-1 ft | ug/kg | 24 | 23 | 4,520 | 680 | 65,000 | 4,330 | 12,900 | 16,000 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(b)fluoranthene | 0-3 ft | ug/kg | 28 | 27 | 5,210 | 1,200 | 65,000 | 5,030 | 12,100 | 15,200 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(g,h,i)perylene | 0-1 ft | ug/kg | 27 | 24 | 1,780 | 470 | 18,000 | 1,580 | 3,510 | 4,590 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(g,h,i)perylene | 0-3 ft | ug/kg | 33 | 30 | 2,740 | 830 | 18,000 | 2,490 | 4,320 | 5,820 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(k)fluoranthene | 0-1 ft | ug/kg | 24 | 23 | 3,610 | 340 | 65000 | 3,460 | 12,900 | 30,200 | 99% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(k)fluoranthene | 0-3 ft | ug/kg | 28 | 27 | 3,560 | 990 | 65000 | 3,440 | 11,900 | 26,300 | 99% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzofluoranthenes, Total | 0-1 ft | ug/kg | 3 | 3 | 6,960 | 5,100 | 14,700 | - | - | - | - |
| Soil | | SVOCs | Total | Benzofluoranthenes, Total | 0-3 ft | ug/kg | 5 | 5 | 12,100 | 8,490 | 31,300 | - | - | - | - |
| | | SVOCs | Total | Chrysene | 0-1 ft | ug/kg | 27 | 26 | 3,230 | 1,250 | 32,000 | 3,110 | 6,180 | 8,400 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Chrysene | 0-3 ft | ug/kg | 33 | 32 | 4,930 | 1,620 | 35,300 | 4,780 | 8,050 | 11,000 | 95% KM (Chebyshev) UCL |
| | | SVOCs | Total | Dibenz(a,h)anthracene | 0-1 ft | ug/kg | | 22 | 732 | 150 | 9,900 | 612 | 1,850 | 4,230 | 99% KM (Chebyshev) UCL |
| Soil | COPC/CPEC | SVOCs | Total | Dibenz(a,h)anthracene | 0-3 ft | ug/kg | 33 | 28 | 854 | 345 | 9,900 | 739 | 1,710 | 2,060 | 95% KM (Chebyshev) UCL |

 Table 1-1

 Statistical Summary for Soil - Landfill AOPC

| | | | | | | | | | | | | | S | tatistics - NI | D censored at the MDL |
|--------|------------------|---------------|-----------|---------------------------------------|----------|-------|-----------|------------|---------|-----------|----------|---------|--------|----------------|--------------------------|
| | | | | | | | | | | | Maximum | | | | |
| | | | Total/ | | Depth | | Number of | Number of | Mean of | Median of | Detected | | | Final 95% | |
| Medium | Retention Reason | Analyte Group | Dissolved | Analyte | Category | Unit | Samples | Detections | Detects | Detects | Value | KM-Mean | KM-SD | UCL | UCL Type |
| Soil | CPEC | SVOCs | Total | Fluoranthene | 0-1 ft | ug/kg | 27 | 25 | 5,740 | 1,300 | 54,000 | 5,320 | 10,800 | 14,500 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Fluoranthene | 0-3 ft | ug/kg | 33 | 31 | 8,240 | 2,700 | 54,000 | 7,750 | 12,600 | 17,500 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Indeno(1,2,3-cd)pyrene | 0-1 ft | ug/kg | 27 | 24 | 2,010 | 520 | 19,000 | 1,790 | 3,790 | 5,040 | 95% KM (Chebyshev) UCL |
| Soil | COPC/CPEC | SVOCs | Total | Indeno(1,2,3-cd)pyrene | 0-3 ft | ug/kg | 33 | 30 | 3,230 | 800 | 20,000 | 2,940 | 5,000 | 6,800 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Pyrene | 0-1 ft | ug/kg | 27 | 26 | 5,130 | 1,900 | 40,000 | 4,940 | 8,640 | 12,300 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Pyrene | 0-3 ft | ug/kg | 33 | 32 | 8,470 | 2,720 | 67,100 | 8,220 | 13,700 | 18,800 | 95% KM (Chebyshev) UCL |
| Soil | COPC | SVOCs | Total | cPAHs as BaPEQ (KM-capped, MDL-based) | 0-3 ft | ug/kg | 33 | 32 | 7,239 | 2,203 | 55,362 | 7,022 | 12,051 | 13,416 | 90% KM (Chebyshev) UCL |
| Soil | COPC | SVOCs | Total | cPAHs as BaPEQ (KM-capped, MDL-based) | 0-10 ft | ug/kg | 44 | 43 | 8,086 | 4,558 | 55,362 | 7,905 | 11,200 | 13,030 | 90% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Total HPAHs (KM, capped; NDs at MDL) | 0-1 ft | ug/kg | 27 | 26 | 37,211 | 13,060 | 367,900 | 30,467 | 69,775 | 90,158 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Total HPAHs (KM, capped; NDs at MDL) | 0-3 ft | ug/kg | 33 | 32 | 47,894 | 16,735 | 367,900 | 43,348 | 75,465 | 101,525 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | VOCs | Total | 1,2,4-Trimethylbenzene | 0-1 ft | ug/kg | 18 | 3 | 5,420 | 1,280 | 14,300 | 1,460 | 3,120 | 3,030 | 95% KM (t) UCL |
| Soil | CPEC | VOCs | Total | 1,2,4-Trimethylbenzene | 0-3 ft | ug/kg | 20 | 3 | 5,420 | 1,280 | 14,300 | 1,380 | 2,970 | 2,790 | 95% KM (t) UCL |
| Soil | COPC | VOCs | Total | Naphthalene | 0-3 ft | ug/kg | 20 | 3 | 3,140 | 542 | 8,360 | 899 | 1,710 | 3,830 | 97.5% KM (Chebyshev) UCL |
| Soil | COPC* | VOCs | Total | Tetrachloroethene (PCE) | 0-10 ft | ug/kg | 35 | 25 | 16,100 | 5.19 | 403,000 | 11,500 | 67,100 | 127,000 | 99% KM (Chebyshev) UCL |

Only COPCs/CPECs identified in the Final RI Report (URS 2012) per relevant depth interval are shown; Table I-12 in Appendix I of the Final RI Report presents the summary statistics for the Landfill AOPC COIs for all media and depth intervals. If not all samples were detected results (i.e., detection rate = <100%), the mean and standard deviation were estimated by the Kaplan-Meier method.

* = PCE and degradation products retained as COPCs.

BHHRA EPCs: 95% UCLs for soil data from 0-3 ft bgs depth interval were used as the EPCs for outdoor worker and fishing platform user scenarios; 95% UCLs for soil data from 0-10 ft bgs depth interval were used as EPCs for construction worker scenario.

BERA EPCs: The lower of the maximum detected concentrations and 95% UCLs calculated for soil in the RI were used as the EPCs for all terrestrial receptors.

% = percentKM = Kaplan-Meier AOPC = Area of Potential Concern MDL = method detection limit mg/kg = milligrams per kilogram BaPEQ = benzo(a)pyrene equivalent(s)BERA = baseline ecological risk assessment ND = non-detectPCB = polychlorinated biphenyls bgs = below ground surface BHHRA = baseline human health risk assessment PCE = tetrachloroethene COI = chemical of interest RI = remedial investigation COPC = contaminant of potential concern SD = standard deviation cPAH = carcinogenic polycyclic aromatic hydrocarbon SVOC = semi-volatile organic carbon CPEC = contaminant of potential ecological concern UCL = upper confidence limit EPC = exposure point concentration ug/kg = micrograms per kilogram ft = foot or feetVOC = volatile organic carbon HPAH = high molecular weight polycyclic aromatic hydrocarbon

Source

Table 1-2Statistical Summary for Soil - Sandblast Area AOPC

| | | | | | | | | | | | | Maximum | | Sta | atistics - ND c | ensored at the MDL |
|--------|------------------|---------------|-----------|-------------------------------------|----------|------------|----------|-----------|------------|---------|-----------|----------|---------|--------|-----------------|----------------------------------|
| | | | Total/ | | Depth | | | Number of | Number of | Mean of | Median of | Detected | | | Final 95% | |
| Medium | Retention Reason | Analyte Group | Dissolved | Analyte | Category | Sieve Size | Unit | Samples | Detections | Detects | Detects | Value | KM-Mean | KM-SD | UCL | UCL Type |
| Soil | | | | Antimony | 0-1 ft | N/A | mg/kg | 58 | 41 | 1.90 | 1.14 | 13.7 | 1.62 | 2.23 | 2.94 | 95% KM (Chebyshev) UCL |
| Soil | | | | Antimony | | N/A | mg/kg | 75 | 55 | 1.57 | 1.07 | 13.7 | 1.38 | 2.02 | 2.42 | 95% KM (Chebyshev) UCL |
| Soil | | | | Arsenic | | N/A | mg/kg | 75 | 74 | 7.24 | 5.02 | 80.9 | 7.15 | 11.4 | 9.71 | 95% KM (BCA) UCL |
| Soil | | | Total | Cadmium | 0-1 ft | N/A | mg/kg | 58 | 53 | 1.82 | 1.04 | 17.3 | 1.66 | 2.65 | 3.20 | 95% KM (Chebyshev) UCL |
| Soil | | | Total | Cadmium | 0-3 ft | N/A | mg/kg | 75 | 67 | 1.58 | 0.963 | 17.3 | 1.42 | 2.38 | 2.63 | 95% KM (Chebyshev) UCL |
| Soil | | | Total | Chromium | 0-1 ft | N/A | mg/kg | 61 | 61 | 383 | 94.9 | 2,650 | - | - | 720 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | | Total | Chromium | 0-3 ft | N/A | mg/kg | 78 | 78 | 306 | 50.6 | 2,650 | - | - | 579 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | | | Lead | 0-1 ft | <250um | mg/kg | 8 | 8 | 300 | 160 | 921 | - | - | 639 | Use 95% Approximate Gamma UCL |
| Soil | | | | Lead | 0-1 ft | <2mm | mg/kg | 8 | 8 | 233 | 133 | 768 | - | - | 551 | Use 95% Approximate Gamma UCL |
| Soil | | | | Lead | 0-1 ft | N/A | mg/kg | 61 | 61 | 362 | 272 | 3,260 | - | - | 465 | Use 95% Approximate Gamma UCL |
| Soil | | | | Lead | 0-3 ft | <250um | mg/kg | 16 | 16 | 202 | 72.2 | 921 | - | - | 418 | Use 95% Approximate Gamma UCL |
| Soil | COPC/CPEC | Metals | Total | Lead | 0-3 ft | <2mm | mg/kg | 16 | 16 | 148 | 34.9 | 768 | - | - | 303 | Use 95% Approximate Gamma UCL |
| Soil | | Metals | Total | Lead | 0-3 ft | N/A | mg/kg | 78 | 78 | 309 | 139 | 3,260 | - | - | 529 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | | Metals | Total | Mercury | 0-1 ft | N/A | mg/kg | 58 | 43 | 0.104 | 0.0440 | 0.723 | 0.0812 | 0.130 | 0.113 | 95% KM (BCA) UCL |
| Soil | | Metals | Total | Mercury | 0-3 ft | N/A | mg/kg | 75 | 58 | 0.0911 | 0.0433 | 0.723 | 0.0741 | 0.116 | 0.0976 | 95% KM (BCA) UCL |
| Soil | | Metals | Total | Nickel | 0-1 ft | N/A | mg/kg | 58 | 58 | 167 | 50.0 | 1,060 | - | - | 353 | Use 95% H-UCL |
| Soil | COPC/CPEC | Metals | Total | Nickel | 0-3 ft | N/A | mg/kg | 75 | 75 | 135 | 33.3 | 1,060 | - | - | 251 | Use 95% Chebyshev (Mean, Sd) UCL |
| Soil | CPEC | Butyltins | Total | Tributyltin | 0-1 ft | N/A | ug/kg | 47 | 16 | 186 | 29.8 | 1,860 | 64.3 | 278 | 481 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | Butyltins | Total | Tributyltin | 0-3 ft | N/A | ug/kg | 62 | 16 | 186 | 29.8 | 1,860 | 49.2 | 243 | 248 | 97.5% KM (Chebyshev) UCL |
| Soil | COPC | PCB Aroclors | Total | Total PCBs as Aroclors (NDs at MDL) | 0-3 ft | N/A | ug/kg | 73 | 61 | 224 | 28.7 | 2,140 | 138.4 | 381 | 419 | 97.5% KM (Chebyshev) UCL |
| Soil | CPEC | Pesticides | Total | BHC (delta) | 0-1 ft | N/A | ug/kg | 31 | 2 | 1.55 | 1.55 | 3.03 | 0.176 | 0.530 | 1.54 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | Pesticides | Total | BHC (delta) | 0-3 ft | N/A | ug/kg | 37 | 2 | 1.55 | 1.55 | 3.03 | 0.160 | 0.485 | 1.30 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | Pesticides | Total | BHC (gamma) Lindane | 0-1 ft | N/A | ug/kg | 31 | 2 | 5.92 | 5.92 | 9.68 | 2.42 | 1.35 | 9.68 | 95% KM (BCA) UCL |
| Soil | CPEC | Pesticides | Total | BHC (gamma) Lindane | 0-3 ft | N/A | ug/kg | 37 | 3 | 4.02 | 2.17 | 9.68 | 0.520 | 1.58 | 1.06 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Chlordane (alpha) | 0-1 ft | N/A | ug/kg | 19 | 3 | 1.01 | 0.860 | 1.50 | 0.718 | 0.195 | 0.815 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Chlordane (alpha) | 0-3 ft | N/A | ug/kg | 25 | 3 | 1.01 | 0.860 | 1.50 | 0.703 | 0.171 | 0.776 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Chlordane (gamma) | 0-1 ft | N/A | ug/kg | 19 | 6 | 43.7 | 29.8 | 97.0 | 19.0 | 26.0 | 30.3 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Chlordane (gamma) | 0-3 ft | N/A | ug/kg | 25 | 10 | 27.4 | 10.6 | 97.0 | 11.3 | 25.2 | 20.4 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endosulfan I | 0-1 ft | N/A | ug/kg | 31 | 2 | 4.32 | 4.32 | 6.45 | 2.35 | 0.775 | 2.69 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endosulfan I | 0-3 ft | N/A | ug/kg | 37 | 5 | 2.21 | 2.00 | 6.45 | 0.447 | 1.13 | 0.809 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endosulfan II | 0-1 ft | N/A | ug/kg | 31 | 3 | 1.06 | 0.960 | 1.99 | 0.341 | 0.367 | 0.495 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endosulfan II | 0-3 ft | N/A | ug/kg | 37 | 3 | 1.06 | 0.960 | 1.99 | 0.322 | 0.333 | 0.447 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endosulfan Sulfate | 0-1 ft | N/A | ug/kg | 31 | 6 | 1.46 | 1.24 | 3.30 | 0.761 | 0.560 | 0.953 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endosulfan Sulfate | 0-3 ft | N/A | ug/kg | 37 | 6 | 1.46 | 1.24 | 3.30 | 0.729 | 0.515 | 0.889 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endrin Aldehyde | 0-1 ft | N/A | ug/kg | 31 | 4 | 8.36 | 7.70 | 16.0 | 2.84 | 2.90 | 3.87 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | | Endrin Aldehyde | 0-3 ft | N/A | ug/kg | 37 | 5 | 7.06 | 4.40 | 16.0 | 2.60 | 2.70 | 3.44 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endrin Ketone | 0-1 ft | N/A | ug/kg | 31 | 4 | 5.25 | 3.07 | 13.0 | 2.35 | 2.04 | 3.09 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endrin Ketone | | N/A | ug/kg | 37 | 4 | 5.25 | 3.07 | 13.0 | 2.26 | 1.87 | 2.88 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endrin | | N/A | ug/kg | 31 | 3 | 11.7 | 15.0 | 17.0 | 3.84 | 3.20 | 5.04 | 95% KM (t) UCL |
| Soil | CPEC | | | Endrin | | N/A | ug/kg | 37 | 4 | 9.28 | 9.00 | 17.0 | 2.88 | 3.15 | 3.89 | 95% KM (t) UCL |
| Soil | | | | Heptachlor | | N/A | ug/kg | 31 | 4 | 0.911 | 0.283 | 2.90 | 0.283 | 0.496 | 0.464 | 95% KM (t) UCL |
| Soil | | | | Heptachlor | | N/A | ug/kg | 37 | 4 | 0.911 | 0.283 | 2.90 | 0.265 | 0.453 | 0.415 | 95% KM (t) UCL |
| Soil | | | | Methoxychlor | | N/A | ug/kg | 31 | 2 | 1.10 | 1.10 | 1.20 | 1.02 | 0.0533 | 1.05 | 95% KM (t) UCL |
| Soil | | | | Methoxychlor | | N/A | ug/kg | 37 | 2 | 1.10 | 1.10 | 1.20 | 1.01 | 0.0447 | 1.04 | 95% KM (t) UCL |
| Soil | | | | Bis(2-ethylhexyl) Phthalate | | N/A | ug/kg | 31 | 21 | 19,200 | 1,500 | 260,000 | 13,100 | 46,400 | 98,000 | 99% KM (Chebyshev) UCL |
| Soil | | | | Bis(2-ethylhexyl) Phthalate | | N/A | ug/kg | 40 | 29 | 15,400 | 1,040 | 260,000 | 11,200 | 41,200 | 77,100 | 99% KM (Chebyshev) UCL |
| Soil | | | | Butyl Benzyl Phthalate | | N/A | ug/kg | 31 | 5 | 38.5 | 15.7 | 124 | 9.59 | 22.1 | 17.1 | 95% KM (t) UCL |
| Soil | | | | Butyl Benzyl Phthalate | | N/A | ug/kg | 39 | 6 | 37.4 | 23.7 | 124 | 9.14 | 20.2 | 15.2 | 95% KM (t) UCL |
| Soil | | | | Carbazole | | N/A | ug/kg | 31 | 20 | 141 | 68.5 | 530 | 91.8 | 144 | 144 | 95% KM (BCA) UCL |
| | | | | | ~ | | ~~B/ 11B | | 20 | | 50.5 | 220 | 21.0 | | | |

Table 1-2Statistical Summary for Soil - Sandblast Area AOPC

| | | | | | | | | | | | | Maximum | | Sta | atistics - ND (| censored at the MDL |
|--------|------------------|---------------|-----------|---------------------------------------|----------|------------|-------|-----------|------------|---------|-----------|----------|---------|--------|-----------------|-----------------------------------|
| | | | Total/ | | Depth | | | Number of | Number of | Mean of | Median of | Detected | | | Final 95% | |
| Medium | Retention Reason | Analyte Group | Dissolved | Analyte | Category | Sieve Size | Unit | Samples | Detections | Detects | Detects | Value | KM-Mean | KM-SD | UCL | UCL Type |
| Soil | CPEC | | Total | Carbazole | 0-3 ft | N/A | ug/kg | 39 | 24 | 118 | 35.0 | 530 | 73.7 | 133 | 112 | 95% KM (BCA) UCL |
| Soil | | | | Dibenzofuran | 0-1 ft | N/A | ug/kg | 31 | 19 | 41.0 | 10.0 | 220 | 26.0 | 47.6 | 42.4 | 95% KM (BCA) UCL |
| Soil | | | | Dibenzofuran | | N/A | ug/kg | 40 | 24 | 58.1 | 8.90 | 485 | 35.5 | 85.3 | 122 | 97.5% KM (Chebyshev) UCL |
| Soil | | | Total | Di-n-butyl Phthalate | 0-1 ft | N/A | ug/kg | 31 | 10 | 107 | 102 | 280 | 44.9 | 66.6 | 80.4 | 95% KM (Percentile Bootstrap) UCL |
| Soil | | | | Di-n-butyl Phthalate | 0-3 ft | N/A | ug/kg | 39 | 14 | 82.0 | 49.0 | 280 | 36.5 | 61.1 | 54.3 | 95% KM (t) UCL |
| Soil | | | Total | Di-n-octyl Phthalate | | N/A | ug/kg | 31 | 5 | 49.4 | 31.0 | 127 | 17.6 | 22.1 | 25.1 | 95% KM (t) UCL |
| Soil | | | Total | Di-n-octyl Phthalate | | N/A | ug/kg | 39 | 6 | 45.0 | 27.0 | 127 | 16.6 | 19.9 | 22.5 | 95% KM (t) UCL |
| Soil | | | Total | Benzo(a)anthracene | | N/A | ug/kg | 31 | 29 | 787 | 190 | 6,440 | 736 | 1,270 | 1,750 | 95% KM (Chebyshev) UCL |
| Soil | | | Total | Benzo(a)anthracene | | N/A | ug/kg | 40 | 35 | 1,070 | 183 | 12,300 | 935 | 2,160 | 2,450 | 95% KM (Chebyshev) UCL |
| Soil | | | | Benzo(a)pyrene | | N/A | ug/kg | 31 | 30 | 771 | 180 | 6,470 | 746 | 1,280 | 1,760 | 95% KM (Chebyshev) UCL |
| Soil | COPC/CPEC | SVOCs | Total | Benzo(a)pyrene | 0-3 ft | N/A | ug/kg | 40 | 36 | 1,030 | 155 | 11,700 | 926 | 2,090 | 2,390 | 95% KM (Chebyshev) UCL |
| Soil | COPC | SVOCs | Total | Benzo(a)pyrene | 0-10 ft | N/A | ug/kg | 43 | 37 | 1,000 | 151 | 11,700 | 862 | 2,030 | 3,980 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(b)fluoranthene | 0-1 ft | N/A | ug/kg | 13 | 13 | 1,350 | 1,100 | 4,100 | - | - | 1,950 | Use 95% Student's-t UCL |
| Soil | COPC/CPEC | SVOCs | Total | Benzo(b)fluoranthene | 0-3 ft | N/A | ug/kg | 18 | 17 | 1,050 | 640 | 4,100 | 991 | 1,150 | 2,210 | 95% KM (Chebyshev) UCL |
| Soil | | SVOCs | Total | Benzo(g,h,i)perylene | 0-1 ft | N/A | ug/kg | 31 | 26 | 534 | 180 | 3,830 | 449 | 774 | 1,070 | 95% KM (Chebyshev) UCL |
| Soil | COPC/CPEC | SVOCs | Total | Benzo(g,h,i)perylene | 0-3 ft | N/A | ug/kg | 40 | 33 | 555 | 120 | 3,830 | 459 | 843 | 1,050 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(k)fluoranthene | 0-1 ft | N/A | ug/kg | 13 | 13 | 469 | 360 | 1,400 | - | - | 676 | Use 95% Student's-t UCL |
| Soil | CPEC | SVOCs | Total | Benzo(k)fluoranthene | 0-3 ft | N/A | ug/kg | 18 | 17 | 365 | 220 | 1,400 | 345 | 396 | 764 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzofluoranthenes, Total | 0-1 ft | N/A | ug/kg | 18 | 17 | 1,100 | 74.2 | 12,100 | 1,040 | 2,780 | 7,760 | 99% KM (Chebyshev) UCL |
| Soil | COPC/CPEC | SVOCs | Total | Benzofluoranthenes, Total | 0-3 ft | N/A | ug/kg | 22 | 20 | 1,900 | 105 | 16,300 | 1,730 | 4,090 | 10,600 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Chrysene | 0-1 ft | N/A | ug/kg | 31 | 30 | 904 | 222 | 7,590 | 875 | 1,500 | 2,070 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Chrysene | 0-3 ft | N/A | ug/kg | 40 | 37 | 1,120 | 129 | 12,000 | 1,030 | 2,230 | 4,590 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Dibenz(a,h)anthracene | 0-1 ft | N/A | ug/kg | 31 | 20 | 232 | 113 | 1,430 | 150 | 280 | 244 | 95% KM (BCA) UCL |
| Soil | COPC/CPEC | SVOCs | Total | Dibenz(a,h)anthracene | 0-3 ft | N/A | ug/kg | 40 | 26 | 235 | 93.2 | 1,430 | 153 | 295 | 238 | 95% KM (BCA) UCL |
| Soil | CPEC | SVOCs | Total | Fluoranthene | 0-1 ft | N/A | ug/kg | 31 | 30 | 1,710 | 284 | 20,700 | 1,660 | 3,740 | 4,640 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Fluoranthene | 0-3 ft | N/A | ug/kg | 40 | 37 | 2,290 | 216 | 28,600 | 2,120 | 5,420 | 10,800 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Indeno(1,2,3-cd)pyrene | 0-1 ft | N/A | ug/kg | 31 | 28 | 530 | 155 | 3,910 | 479 | 811 | 1,130 | 95% KM (Chebyshev) UCL |
| Soil | COPC/CPEC | SVOCs | Total | Indeno(1,2,3-cd)pyrene | 0-3 ft | N/A | ug/kg | 40 | 35 | 575 | 97.8 | 4,170 | 503 | 944 | 1,160 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Pyrene | 0-1 ft | N/A | ug/kg | 31 | 30 | 1,710 | 338 | 21,900 | 1,660 | 3,920 | 4,780 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Pyrene | 0-3 ft | N/A | ug/kg | 40 | 38 | 2,350 | 184 | 32,000 | 2,230 | 5,950 | 11,700 | 99% KM (Chebyshev) UCL |
| Soil | COPC | SVOCs | Total | cPAHs as BaPEQ (KM-capped, MDL-based) | 0-3 ft | N/A | ug/kg | 40 | 38 | 1,441 | 192 | 16,103 | 1,369 | 2,988 | 2,805 | 90% KM (Chebyshev) UCL |
| Soil | COPC | SVOCs | Total | cPAHs as BaPEQ (KM-capped, MDL-based) | 0-10 ft | N/A | ug/kg | 43 | 39 | 1,405 | 173 | 16,103 | 1,274 | 2,903 | 2,620 | 90% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Total HPAHs (KM, capped; NDs at MDL) | 0-1 ft | N/A | ug/kg | 31 | 30 | 7765 | 738 | 72270 | 7516 | 13891 | 18577 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | | Total HPAHs (KM, capped; NDs at MDL) | 0-3 ft | N/A | ug/kg | 40 | 38 | 9430 | 926 | 105200 | 8959 | 19945 | 40758 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | VOCs | Total | 1,2,4-Trimethylbenzene | 0-1 ft | N/A | ug/kg | 37 | 15 | 0.246 | 0.230 | 0.523 | 0.195 | 0.0993 | 0.230 | 95% KM (t) UCL |
| Soil | CPEC | VOCs | Total | 1,2,4-Trimethylbenzene | 0-3 ft | N/A | ug/kg | 46 | 21 | 1,250 | 0.230 | 14,300 | 572 | 2,690 | 4,620 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | VOCs | Total | 4-Isopropyltoluene | 0-1 ft | N/A | ug/kg | 37 | 14 | 1.63 | 0.415 | 12.0 | 0.749 | 1.98 | 1.33 | 95% KM (t) UCL |
| Soil | CPEC | VOCs | Total | 4-Isopropyltoluene | 0-3 ft | N/A | ug/kg | 46 | 17 | 10.9 | 0.900 | 161 | 4.25 | 23.7 | 20.1 | 95% KM (Chebyshev) UCL |
| Soil | COPC* | VOCs | Total | cis-1,2-Dichloroethene | 0-3 ft | N/A | ug/kg | 46 | 2 | 66.0 | 66.0 | 120 | 14.4 | 15.9 | - | - |
| Soil | COPC* | VOCs | Total | cis-1,2-Dichloroethene | 0-10 ft | N/A | ug/kg | | 4 | 48.8 | 35.2 | 120 | 8.61 | 18.0 | 13.6 | 95% KM (t) UCL |
| Soil | CPEC | VOCs | | n-Propylbenzene | | N/A | ug/kg | 37 | 1 | 0.150 | 0.150 | 0.150 | - | - | - | - |
| Soil | CPEC | VOCs | | n-Propylbenzene | | N/A | ug/kg | 46 | 3 | 40.8 | 0.409 | 122 | 2.87 | 18.0 | 35.5 | 99% KM (Chebyshev) UCL |
| Soil | COPC* | | | Tetrachloroethene (PCE) | | N/A | ug/kg | 46 | 5 | 85,900 | 5.84 | 420,000 | 9,340 | 61,200 | 26,300 | 95% KM (t) UCL |
| Soil | COPC* | | | Tetrachloroethene (PCE) | | N/A | ug/kg | 49 | 8 | 53,700 | 10.1 | 420,000 | 8,760 | 59,400 | 99,000 | 99% KM (Chebyshev) UCL |
| Soil | COPC* | | | Trichloroethene (TCE) | | N/A | ug/kg | 37 | 4 | 0.0912 | 0.0661 | 0.171 | 0.0678 | 0.0216 | 0.0766 | 95% KM (t) UCL |
| Soil | | | | Trichloroethene (TCE) | | N/A | ug/kg | 46 | 7 | 1,330 | 0.171 | 6,080 | 202 | 993 | 1,780 | 99% KM (Chebyshev) UCL |

Table 1-2 Statistical Summary for Soil - Sandblast Area AOPC

Notes

Only COPCs/CPECs identified in the Final RI Report (URS 2012) per relevant depth interval are shown; Table I-13 in Appendix I of the Final RI Report presents the summary statistics for the Sandblast Area AOPC COIs for all media and depth intervals. If not all samples were detected results (i.e., detection rate = <100%), the mean and standard deviation were estimated by the Kaplan-Meier method. * = PCE and degradation products retained as COPCs. BHHRA EPCs: 95% UCLs for soil data from 0-3 ft bgs depth interval were used as the EPCs for outdoor worker and fishing platform user scenarios; 95% UCLs for soil data from 0-10 ft bgs depth interval were used as EPCs for construction worker scenario.

Cs for all terrestrial receptors.

| % = percent | KM = Kaplan-Meier |
|--|-------------------------------------|
| AOPC = Area of Potential Concern | MDL = method detection limit |
| BaPEQ = benzo(a)pyrene equivalent(s) | mg/kg = milligrams per kilogram |
| BERA = baseline ecological risk assessment | N/A = not applicable |
| bgs = below ground surface | ND = non-detect |
| BHC = benzene hexachloride | PCB = polychlorinated biphenyls |
| BHHRA = baseline human health risk assessment | PCE = tetrachloroethene |
| COI = chemical of interest | RI = remedial investigation |
| COPC = contaminant of potential concern | SD = standard deviation |
| cPAH = carcinogenic polycyclic aromatic hydrocarbon | SVOC = semi-volatile organic carbon |
| CPEC = contaminant of potential ecological concern | UCL = upper confidence limit |
| EPC = exposure point concentration | ug/kg = micrograms per kilogram |
| ft = foot or feet | VOC = volatile organic carbon |
| HPAH = high molecular weight polycyclic aromatic hydrocarbon | - |

Source

Table 1-3 Statistical Summary for Soil and Lagoon Sediment - Pistol Range AOPC

| | | | | | | | | | | | Maximum | | S | tatistics - ND | censored at the MDL |
|-----------------|------------------|---------------|-----------|---------|----------|-------|-----------|------------|---------|-----------|----------|---------|-------|----------------|----------------------------------|
| | | | Total/ | | Depth | | Number of | Number of | Mean of | Median of | Detected | | | Final 95% | |
| Medium | Retention Reason | Analyte Group | Dissolved | Analyte | Category | Unit | Samples | Detections | Detects | Detects | Value | KM-Mean | KM-SD | UCL | UCL Type |
| Soil | COPC/CPEC | Metals | Total | Lead | 0-1.5 ft | mg/kg | 63 | 63 | 208 | 60 | 1,110 | - | - | 365 | Use 95% Chebyshev (Mean, Sd) UCL |
| Lagoon Sediment | COPC | Metals | Total | Lead | N/A | mg/kg | 5 | 5 | 27 | 28 | 33 | - | - | - | - |

Only COPCs/CPECs identified in the Final RI Report (URS 2012) are shown; Table I-14 in Appendix I of the Final RI Report presents the summary statistics for the Pistol Range AOPC COIs for all media. BHHRA EPCs: 95% UCLs were used as the EPCs for fishing platform user scenario.

BERA EPCs: The lower of the maximum detected concentrations and 95% UCLs calculated for soil in the RI were used as the EPCs for all terrestrial receptors.

| % = percent | KM = Kaplan-Meier |
|--|---------------------------------|
| BERA = baseline ecological risk assessment | MDL = method detection limit |
| BHHRA = baseline human health risk assessment | mg/kg = milligrams per kilogram |
| COPC = contaminant of potential concern | N/A = not applicable |
| CPEC = contaminant of potential ecological concern | ND = non-detect |
| EPC = exposure point concentration | RI = remedial investigation |
| ft = foot or feet | SD = standard deviation |
| | UCL = upper confidence limit |
| | |

Source

Table 1-4Statistical Summary for Soil - Bulb Slope AOPC

| | | | | | | | | | | | Maximum | | Statistics - ND censored at the MDL | | | | | |
|--------|-------------------------|---------------|-----------|---------|----------|-------|-----------|------------|---------|-----------|----------|---------|-------------------------------------|-----------|-------------------------------|--|--|--|
| | | | Total/ | | Depth | | Number of | Number of | Mean of | Median of | Detected | | | Final 95% | | | | |
| Medium | Retention Reason | Analyte Group | Dissolved | Analyte | Category | Unit | Samples | Detections | Detects | Detects | Value | KM-Mean | KM-SD | UCL | UCL Type | | | |
| Soil | COPC/CPEC | Metals | Total | Lead | 0-1 ft | mg/kg | 12 | 12 | 222 | 199 | 597 | - | - | 307 | Use 95% Student's-t UCL | | | |
| Soil | CPEC | Metals | Total | Mercury | 0-1 ft | mg/kg | 12 | 12 | 0.404 | 0.280 | 1.54 | - | - | 0.720 | Use 95% Approximate Gamma UCL | | | |

Only COPCs/CPECs identified in the Final RI Report (URS 2012) are shown; Table I-15 im Appendix I of the Final RI Report presents the summary statistics for the Bulb Slope AOPC COIs. BHHRA EPCs: 95% UCLs were used as the EPCs for fishing platform user scenario.

BERA EPCs: The lower of the maximum detected concentrations and 95% UCLs calculated for soil in the RI were used as the EPCs for all terrestrial receptors.

| % = percent | KM = Kaplan-Meier |
|--|---------------------------------|
| BERA = baseline ecological risk assessment | MDL = method detection limit |
| BHHRA = baseline human health risk assessment | ND = non-detect |
| COPC = contaminant of potential concern | mg/kg = milligrams per kilogram |
| CPEC = contaminant of potential ecological concern | RI = remedial investigation |
| EPC = exposure point concentration | SD = standard deviation |
| ft = foot or feet | UCL = upper confidence limit |

Source

Table 1-5Statistical Summary for Soil - Combined AOPCs

| | | | | | | | | | | | | | Sta | tistics - ND | censored at the MDL |
|--------|---------------------|---------------|------------------|-----------------------------|-------------------|----------------|----------------------|-------------------------|--------------------|----------------------|---------------------------|--------------|--------|------------------|--|
| Medium | Retention Reason | Analyte Group | Total/ Dissolved | Analyte | Depth Category | Unit | Number of Samples | Number of Detections | Mean of Detects | Median of Detects | Maximum Detected Value | KM-Mean | KM-SD | Final 95% UCL | UCL Type |
| | CPEC | Metals | Total | Antimony | 0-1 ft | mg/kg | 83 | 45 | 1.98 | 1.28 | 13.7 | 1.39 | 2.02 | 1.80 | 95% KM (BCA) UCL |
| | CPEC | Metals | Total | Antimony | 0-1 ft | mg/kg | 102 | 61 | 1.65 | 1.10 | 13.7 | 1.26 | 1.86 | 1.64 | 95% KM (BCA) UCL |
| Soil | CPEC | Metals | Total | Cadmium | 0-3 ft 0-1 ft | mg/kg | 79 | 68 | 1.62 | 0.963 | 17.3 | 1.20 | 2.33 | 1.04 | 95% KM (BCA) UCL |
| | CPEC | Metals | Total | Cadmium | 0-1 ft | mg/kg | 98 | 84 | 1.48 | 0.924 | 17.3 | 1.41 | 2.33 | 1.67 | 95% KM (BCA) UCL |
| | CPEC | Metals | Total | Chromium | 0-3 ft | mg/kg | 82 | 82 | 305 | 49.8 | 2,650 | - | - | 567 | 95% Chebyshev (Mean, Sd) UCL |
| Soil | CPEC | Metals | Total | Chromium | 0-1 ft | mg/kg | 101 | 101 | 279 | 35.7 | 2,650 | - | - | 510 | 95% Chebyshev (Mean, Sd) UCL |
| Soil | CPEC | Metals | Total | Copper | 0-3 ft 0-1 ft | mg/kg | 93 | 93 | 62.8 | 43.8 | 494 | _ | - | 95.0 | 95% Chebyshev (Mean, Sd) UCL |
| | CPEC | Metals | Total | Copper | 0-1 ft | mg/kg | 112 | 112 | 64.0 | 45.7 | 494 | _ | - | 93.6 | 95% Chebyshev (Mean, Sd) UCL |
| | CPEC | Metals | Total | Lead | 0-3 ft 0-1 ft | mg/kg | 112 | 112 | 270 | 128 | 3,260 | _ | - | 398 | 95% Chebyshev (Mean, Sd) UCL |
| Soil | CPEC | Metals | Total | Lead | 0-1 ft | mg/kg | 182 | 182 | 270 | 119 | 3,260 | _ | - | 396 | 95% Chebyshev (Mean, Sd) UCL |
| | CPEC | Metals | Total | Mercury | 0-3 ft 0-1 ft | mg/kg | 107 | 73 | 0.257 | 0.0940 | 4.15 | 0.183 | 0.459 | 0.378 | 95% KM (Chebyshev) UCL |
| | CPEC | Metals | Total | Mercury | 0-1 ft | mg/kg | 107 | 90 | 0.219 | 0.0686 | 4.15 | 0.163 | 0.435 | 0.330 | 95% KM (Chebyshev) UCL |
| | CPEC | Metals | Total | Nickel | 0-5 ft | mg/kg | 89 | 89 | 125 | 26.3 | 1,060 | - | - | 227 | 95% Chebyshev (Mean, Sd) UCL |
| Soil | CPEC | Metals | Total | Nickel | 0-1 ft | mg/kg | 108 | 108 | 123 | 26.2 | 1,610 | _ | | 234 | 95% Chebyshev (Mean, Sd) UCL |
| | CPEC | Butyltins | Total | Tributyltin | 0-3 ft | ug/kg | 54 | 100 | 126 | 23.6 | 1,860 | 59.4 | 260 | 422 | 99% KM (Chebyshev) UCL |
| | CPEC | Butyltins | Total | Tributyltin | 0-1 ft | ug/kg | 71 | 20 | 158 | 19.5 | 1,860 | 45.8 | 200 | 219 | 97.5% KM (Chebyshev) UCL |
| | CPEC | Pesticides | Total | BHC (delta) | 0-3 ft | ug/kg | 52 | 20 | 1.55 | 1.55 | 3.03 | 0.136 | 0.409 | - | - |
| Soil | CPEC | Pesticides | Total | BHC (delta) | 0-1 ft | ug/kg | 60 | 2 | 1.55 | 1.55 | 3.03 | 0.130 | 0.381 | _ | |
| | CPEC | Pesticides | Total | BHC (gamma) Lindane | 0-3 ft 0-1 ft | ug/kg | 52 | 2 | 5.92 | 5.92 | 9.68 | 2.32 | 1.04 | _ | - |
| | CPEC | Pesticides | Total | BHC (gamma) Lindane | 0-1 ft | ug/kg | 60 | 3 | 4.02 | 2.17 | 9.68 | 0.395 | 1.04 | _ | |
| Soil | CPEC | Pesticides | Total | Chlordane (alpha) | 0-3 ft 0-1 ft | ug/kg | 39 | 4 | 1.76 | 1.18 | 4.00 | 0.805 | 0.589 | 1.01 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Chlordane (alpha) | 0-1 ft | ug/kg | 45 | 4 | 1.76 | 1.18 | 4.00 | 0.305 | 0.544 | 0.952 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Chlordane (apria) | 0-3 ft 0-1 ft | ug/kg | 39 | 6 | 43.7 | 29.8 | 97.0 | 13.2 | 19.0 | 18.8 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Chlordane (gamma) | 0-1 ft | ug/kg | 45 | 10 | 27.4 | 10.6 | 97.0 | 6.62 | 19.5 | 11.8 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Chlordane (technical) | 0-3 ft 0-1 ft | ug/kg | 13 | 10 | 1,560 | 1,560 | 1,560 | - | - | - | 75% KM (t) OCL |
| Soil | CPEC | Pesticides | Total | Chlordane (technical) | 0-1 ft | ug/kg | 15 | 3 | 573 | 92.7 | 1,560 | - 167 | 372 | 902 | 97.5% KM (Chebyshev) UCL |
| | CPEC | Pesticides | Total | Endosulfan I | 0-3 ft 0-1 ft | ug/kg | 52 | 2 | 4.32 | 4.32 | 6.45 | 2.28 | 0.595 | - | 97.5% KW (Chebyshev) OCL |
| Soil | CPEC | Pesticides | Total | Endosulfan I | 0-1 ft | ug/kg | 60 | 5 | 2.21 | 2.00 | 6.45 | 0.328 | 0.393 | 0.547 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Endosulfan II | 0-3 ft 0-1 ft | ug/kg | 52 | 4 | 3.01 | 1.48 | 8.84 | 0.502 | 1.22 | 0.844 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Endosulfan II | 0-1 ft | ug/kg | 60 | 4 | 3.01 | 1.48 | 8.84 | 0.302 | 1.14 | 0.756 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endosulfan Sulfate | 0-3 ft 0-1 ft | ug/kg | 52 | 7 | 2.11 | 1.48 | 5.97 | 0.403 | 0.887 | 1.08 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endosulfan Sulfate | 0-1 ft | ug/kg | 60 | 7 | 2.11 | 1.70 | 5.97 | 0.800 | 0.825 | 1.00 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Endrin Aldehyde | 0-3 ft 0-1 ft | ug/kg | 52 | 4 | 8.36 | 7.70 | 16.0 | 2.52 | 2.28 | 3.14 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Endrin Aldehyde | 0-1 ft | ug/kg | 60 | 5 | 7.06 | 4.40 | 16.0 | 2.32 | 2.16 | 2.86 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endrin Ketone | 0-3 ft 0-1 ft | ug/kg | 51 | 4 | 5.25 | 3.07 | 13.0 | 2.16 | 1.61 | 2.61 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Endrin Ketone | 0-1 ft | ug/kg | 59 | 4 | 5.25 | 3.07 | 13.0 | 2.10 | 1.49 | 2.50 | 95% KM (t) UCL |
| Soil | CPEC | Pesticides | Total | Endrin | 0-1 ft | | 52 | 3 | 11.7 | 15.0 | 17.0 | 3.50 | 2.51 | 4.21 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Endrin | 0-1 ft | ug/kg ug/kg | 60 | 4 | 9.28 | 9.00 | 17.0 | 2.58 | 2.50 | 3.21 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Heptachlor | 0-5 ft 0-1 ft | ug/kg | 52 | 5 | 1.30 | 0.380 | 2.90 | 0.297 | 0.526 | 0.437 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Heptachlor | 0-1 ft | ug/kg | 60 | 6 | 1.59 | 1.60 | 3.07 | 0.330 | 0.610 | 0.477 | 95% KM (t) UCL |
| | CPEC | Pesticides | Total | Methoxychlor | 0-1 ft | ug/kg | 52 | 2 | 1.10 | 1.10 | 1.20 | 1.02 | 0.0533 | - | - |
| | CPEC | Pesticides | Total | Methoxychlor | 0-3 ft | ug/kg | 60 | 2 | 1.10 | 1.10 | 1.20 | 1.02 | 0.0447 | _ | |
| | CPEC | SVOCs | Total | Bis(2-ethylhexyl) Phthalate | 0-1 ft | ug/kg | 58 | 45 | 10,000 | 895 | 260,000 | 7,820 | 34,500 | 36,400 | 97.5% KM (Chebyshev) UCL |
| | CPEC | SVOCs | Total | Bis(2-ethylhexyl) Phthalate | 0-1 ft | ug/kg | 69 | 55 | 9,040 | 825 | 260,000 | 7,320 | 31,800 | 31,300 | 97.5% KM (Chebyshev) UCL |
| | CPEC | SVOCs | Total | Butyl Benzyl Phthalate | 0-3 ft | ug/kg | 57 | 8 | 43.8 | 30.8 | 124 | 11.3 | 22.2 | 17.2 | 95% KM (t) UCL |
| | CPEC | SVOCs | Total | Butyl Benzyl Phthalate | 0-1 ft | ug/kg | 67 | 9 | 42.5 | 31.7 | 124 | 10.4 | 20.5 | 15.4 | 95% KM (t) UCL |
| | CPEC | SVOCs | Total | Carbazole | 0-3 ft | | 58 | 39 | 259 | 91.0 | 2,650 | 178 | 399 | 272 | 95% KM (BCA) UCL |
| | CPEC | SVOCs | Total | Carbazole | 0-1 ft | ug/kg ug/kg | 68 | 45 | 301 | 78.0 | 2,840 | 202 | 496 | 312 | 95% KM (BCA) UCL |
| | CPEC | SVOCs | Total | Dibenzofuran | 0-3 ft | ug/kg | 58 | 30 | 84.6 | 40.2 | 810 | 48.2 | 119 | 77.1 | 95% KM (BCA) UCL |
| | CPEC | SVOCs | Total | Dibenzofuran | 0-1 ft | ug/kg | 69 | 30 | 98.8 | 40.2 | 810 | 48.2 56.6 | 119 | 83.4 | 95% KM (BCA) UCL |
| | CPEC | SVOCs | Total | Di-n-butyl Phthalate | 0-3 ft 0-1 ft | | | | 218 | 42.4 86.4 | 1,800 | 50.0 68.3 | 238 | 83.4 124 | 95% KM (BCA) UCL 95% KM (t) UCL |
| | CPEC | SVOCs | Total | Di-n-butyl Phthalate | 0-1 ft 0-3 ft | ug/kg | 57 67 | 14 18 | 174 | 61.0 | 1,800 | 58.0 | 238 | 124 | 95% KM (t) UCL |
| | CPEC | SVOCs | | Di-n-octyl Phthalate | | ug/kg | | 5 | 49.4 | 31.0 | 1,800 | 58.0 15.8 | 18.6 | | 95% KM (t) UCL |
| | CPEC | SVOCs | Total | Di-n-octyl Phthalate | 0-1 ft 0-3 ft | ug/kg | 57 67 | | 49.4 | 27.0 | 127 | 15.8 | 18.6 | 21.0 19.4 | 95% KM (t) UCL |
| | CPEC | SVOCs | Total Total | Benzo(a)anthracene | 0-3 ft 0-1 ft | ug/kg | 58 | 6 54 | 45.0 | 428 | 32,000 | 15.2 | 4,520 | 4,390 | 95% KM (t) UCL 95% KM (Chebyshev) UCL |
| 3011 | CLEC | 51003 | I Utal | שלוובט(מ)מונווו מלכוול | 0-1 11 | ug/kg | 50 | 54 | 1,900 | 420 | 52,000 | 1,770 | 4,520 | 4,390 | 7570 KWI (CHEUYSHEV) UCL |

Table 1-5 **Statistical Summary for Soil - Combined AOPCs**

| | | | | | | | | | | | | | Sta | tistics - ND o | ensored at the MDL |
|--------|-----------|---------------|------------------|--------------------------------------|----------|-------|-----------|------------|---------|-----------|----------------|---------|--------|----------------|--------------------------|
| | Retention | | | | Depth | | Number of | Number of | Mean of | Median of | Maximum | | | Final 95% | |
| Medium | Reason | Analyte Group | Total/ Dissolved | Analyte | Category | Unit | Samples | Detections | Detects | Detects | Detected Value | KM-Mean | KM-SD | UCL | UCL Type |
| Soil | CPEC | SVOCs | Total | Benzo(a)anthracene | 0-3 ft | ug/kg | 73 | 66 | 2,730 | 530 | 32,000 | 2,470 | 5,430 | 5,260 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(a)pyrene | 0-1 ft | ug/kg | 58 | 55 | 1,980 | 435 | 33,000 | 1,880 | 4,660 | 4,570 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(a)pyrene | 0-3 ft | ug/kg | 73 | 67 | 2,960 | 450 | 34,000 | 2,720 | 6,020 | 7,150 | 97.5% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(b)fluoranthene | 0-1 ft | ug/kg | 37 | 36 | 3,380 | 835 | 65,000 | 3,280 | 10,500 | 20,700 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(b)fluoranthene | 0-3 ft | ug/kg | 46 | 44 | 3,600 | 835 | 65,000 | 3,450 | 9,690 | 9,750 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(g,h,i)perylene | 0-1 ft | ug/kg | 58 | 50 | 1,130 | 280 | 18,000 | 976 | 2,520 | 2,440 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(g,h,i)perylene | 0-3 ft | ug/kg | 73 | 63 | 1,590 | 290 | 18,000 | 1,380 | 3,140 | 3,690 | 97.5% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(k)fluoranthene | 0-1 ft | ug/kg | 37 | 36 | 2,480 | 350 | 65,000 | 2,410 | 10,500 | 19,800 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzo(k)fluoranthene | 0-3 ft | ug/kg | 46 | 44 | 2,330 | 350 | 65,000 | 2,230 | 9,420 | 11,000 | 97.5% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzofluoranthenes, Total | 0-1 ft | ug/kg | 21 | 20 | 1,980 | 155 | 14,700 | 1,880 | 3,950 | 10,700 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Benzofluoranthenes, Total | 0-3 ft | ug/kg | 27 | 25 | 3,950 | 212 | 31,300 | 3,650 | 7,120 | 17,600 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Chrysene | 0-1 ft | ug/kg | 58 | 56 | 1,980 | 520 | 32,000 | 1,920 | 4,500 | 4,510 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Chrysene | 0-3 ft | ug/kg | 73 | 69 | 2,880 | 620 | 35,300 | 2,730 | 5,960 | 5,790 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Dibenz(a,h)anthracene | 0-1 ft | ug/kg | 58 | 42 | 494 | 140 | 9,900 | 363 | 1,300 | 1,120 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Dibenz(a,h)anthracene | 0-3 ft | ug/kg | 73 | 54 | 556 | 155 | 9,900 | 416 | 1,210 | 1,040 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Fluoranthene | 0-1 ft | ug/kg | 58 | 55 | 3,540 | 795 | 54,000 | 3,360 | 8,060 | 8,020 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Fluoranthene | 0-3 ft | ug/kg | 73 | 68 | 5,000 | 816 | 54,000 | 4,660 | 9,770 | 11,900 | 97.5% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Indeno(1,2,3-cd)pyrene | 0-1 ft | ug/kg | 58 | 52 | 1,210 | 285 | 19,000 | 1,090 | 2,730 | 2,670 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Indeno(1,2,3-cd)pyrene | 0-3 ft | ug/kg | 73 | 65 | 1,800 | 300 | 20,000 | 1,600 | 3,640 | 3,480 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Pyrene | 0-1 ft | ug/kg | 58 | 56 | 3,300 | 810 | 40,000 | 3,190 | 6,750 | 7,090 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Pyrene | 0-3 ft | ug/kg | 73 | 70 | 5,150 | 897 | 67,100 | 4,940 | 10,600 | 10,400 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Total HPAHs (KM, capped; NDs at MDL) | 0-1 ft | ug/kg | 58 | 55.9999999 | 22,488 | 4,976 | 367,900 | 18,198 | 50,007 | 47,078 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | SVOCs | Total | Total HPAHs (KM, capped; NDs at MDL) | 0-3 ft | ug/kg | 73 | 70 | 28,142 | 5,626 | 367,900 | 24,503 | 55,546 | 53,046 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | VOCs | Total | 1,2,4-Trimethylbenzene | 0-1 ft | ug/kg | 55 | 18 | 903 | 0.240 | 14300 | 296 | 1,920 | 2,940 | 99% KM (Chebyshev) UCL |
| Soil | CPEC | VOCs | Total | 1,2,4-Trimethylbenzene | 0-3 ft | ug/kg | 66 | 24 | 1,770 | 0.240 | 14,300 | 645 | 2,830 | 2,860 | 97.5% KM (Chebyshev) UCL |
| Soil | CPEC | VOCs | Total | 4-Isopropyltoluene | 0-1 ft | ug/kg | 55 | 14 | 1.63 | 0.415 | 12.0 | 0.596 | 1.71 | 1.02 | 95% KM (t) UCL |
| Soil | CPEC | VOCs | Total | 4-Isopropyltoluene | 0-3 ft | ug/kg | 66 | 17 | 10.9 | 0.900 | 161 | 3.14 | 20.3 | 14.7 | 95% KM (Chebyshev) UCL |
| Soil | CPEC | VOCs | Total | n-Propylbenzene | 0-1 ft | ug/kg | 55 | 1 | 0.150 | 0.150 | 0.150 | - | - | - | - |
| Soil | CPEC | VOCs | Total | n-Propylbenzene | 0-3 ft | ug/kg | 66 | 3 | 40.8 | 0.409 | 122 | 2.12 | 15.4 | - | - |

Only CPECs identified in the Final RI Report (URS 2012), per relevant depth interval, are shown; Table I-16 in Appendix I of the Final RI Report presents the summary statistics for the COIs in the combined dataset of all Four AOPCs Combined. If not all samples were detected results (i.e., detection rate = <100%), the mean and standard deviation were estimated by the Kaplan-Meier method. BERA EPCs: The lower of the maximum detected concentrations and 95% UCLs calculated for soil in the RI were used as the EPCs for all terrestrial receptors.

| % = percent | KM = Kaplan-Meier |
|--|---------------------------------|
| AOPC = Area of Potential Concern | MDL = method detection limit |
| BERA = baseline ecological risk assessment | mg/kg = milligrams per kilogram |
| BHC = benzene hexachloride | ND = non-detect |
| COI = chemical of interest | RI = remedial investigation |
| CPEC = contaminant of potential ecological concern | SD = standard deviation |
| EPC = exposure point concentration | UCL = upper confidence limit |
| ft = foot or feet | ug/kg = micrograms per kilogram |
| HPAH = high molecular weight polycyclic aromatic hydrocarbon | VOC = volatile organic carbon |

Source

 Table 1-6

 Statistical Summary for Soil Gas - Sandblast Area AOPC

| | | | | | | | | | | | Maximum Statistics - ND censored at the MD | | | the MDL | |
|-------------|------------------|---------------------|-------------------------|-------------------|---------------|-------------------|--|-------------------------|--------------------|----------------------|--|-------------|-------|------------------|----------|
| Medium | Analyte Group | Total/ Dissolved | Analyte | Depth Category | Sieve Size | Unit | Number of Samples | Number of Detections | Mean of Detects | Median of Detects | Detected Value | KM-Mean | KM-SD | Final 95% UCL | UCL Type |
| Witculum | Group | Dissorveu | Analyte | Category | DILC | Cint | Samples | Dettections | Dettetts | Dettets | Value | Kivi-ivican | KM-5D | UCL | OCL Type |
| Soil Gas | VOCs | Total | cis-1,2-Dichloroethene | N/A | N/A | ug/m ³ | 5 | 5 | 115 | 69.0 | 330 | - | - | - | - |
| Soil Gas | VOCs | Total | Tetrachloroethene (PCE) | N/A | N/A | ug/m ³ | 5 | 5 | 7,590 | 800 | 34,000 | - | - | - | - |
| Soil Gas | VOCs | Total | Trichloroethene (TCE) | N/A | N/A | ug/m ³ | 5 | 5 | 214 | 41.0 | 610 | - | - | - | - |
| | | | 1,1-dichloroethene | N/A | N/A | ug/m ³ | NO | . 1.4.1 | 1 1 1 | | | - | - | - | - |
| Degradation | n Products o | f PCE/ TCE | tran-1,2-dichloroethene | N/A | N/A | ug/m ³ | Not Detected Above the Method Detection Limit (MDL) MDL used as theoretical concentration | | | | | - | - | - | - |
| | | | vinyl chloride | N/A | N/A | ug/m ³ | | | | | | - | - | - | - |

Only those analytes with at least 2 detected results are shown on this table.

If not all samples were detected results (i.e., detection rate = <100%), the mean and standard deviation were estimated by the Kaplan-Meier method.

% = percent

KM = Kaplan-Meier

MDL = method detection limit

N/A = not applicable

ND = non-detect

SD = standard deviation

 $ug/m^3 = micrograms$ per cubic meters

UCL = upper confidence limit

VOC = volatile organic carbon

 Table 1-7

 Statistical Summary for Groundwater - Landfill AOPC

| | | | | | | | | | | Maximum | | Statistics - ND censored at the MDL | | |
|-------------|---------------|------------------|-------------------------|----------|------|-----------|------------|---------|-----------|----------|---------|-------------------------------------|-----------|-----------------------------------|
| | | | | Depth | | Number of | Number of | Mean of | Median of | Detected | | | Final 95% | |
| Medium | Analyte Group | Total/ Dissolved | Analyte | Category | Unit | Samples | Detections | Detects | Detects | Value | KM-Mean | KM-SD | UCL | UCL Type |
| Groundwater | Metals | Total | Antimony | N/A | ug/L | 19 | 8 | 1.53 | 1.07 | 3.89 | 1.49 | 1.15 | 2.23 | 95% KM (t) UCL |
| Groundwater | Metals | Total | Iron | N/A | ug/L | 53 | 50 | 11,600 | 3,590 | 42,900 | 10,900 | 13,200 | 29,100 | 99% KM (Chebyshev) UCL |
| Groundwater | Metals | Total | Mercury | N/A | ug/L | 24 | 4 | 0.148 | 0.115 | 0.330 | 0.0496 | 0.0676 | 0.0769 | 95% KM (t) UCL |
| Groundwater | Metals | Total | Thallium | N/A | ug/L | 19 | 9 | 0.202 | 0.197 | 0.323 | 0.202 | 0.0582 | 0.238 | 95% KM (t) UCL |
| Groundwater | Metals | Total | Zinc | N/A | ug/L | 19 | 15 | 416 | 23.7 | 2,660 | 329 | 736 | 2,070 | 99% KM (Chebyshev) UCL |
| Groundwater | Metals | Dissolved | Iron | N/A | ug/L | 36 | 25 | 14,100 | 18,800 | 35,400 | 9,800 | 12,600 | 31,000 | 99% KM (Chebyshev) UCL |
| Groundwater | SVOCs | Total | Di-n-octyl Phthalate | N/A | ug/L | 23 | 3 | 5.07 | 5.29 | 7.08 | 3.59 | 1.45 | 4.61 | 95% KM (t) UCL |
| Groundwater | SVOCs | Total | Naphthalene | N/A | ug/L | 23 | 3 | 0.0993 | 0.101 | 0.157 | 0.0598 | 0.0393 | 0.0873 | 95% KM (t) UCL |
| Groundwater | VOCs | Total | 1,2,4-Trimethylbenzene | N/A | ug/L | 22 | 2 | 3.10 | 3.10 | 5.20 | 1.19 | 0.875 | - | - |
| Groundwater | VOCs | Total | Acetone | N/A | ug/L | 22 | 4 | 13.2 | 14.6 | 15.4 | 9.27 | 2.22 | 10.2 | 95% KM (t) UCL |
| Groundwater | VOCs | Total | Chloroform | N/A | ug/L | 56 | 8 | 1.59 | 1.19 | 3.70 | 0.320 | 0.739 | 0.498 | 95% KM (t) UCL |
| Groundwater | VOCs | Total | Isopropylbenzene | N/A | ug/L | 22 | 2 | 3.90 | 3.90 | 4.60 | 3.26 | 0.292 | 3.42 | 95% KM (t) UCL |
| Groundwater | VOCs | Total | n-Propylbenzene | N/A | ug/L | 22 | 2 | 1.60 | 1.60 | 2.00 | 1.24 | 0.167 | 1.32 | 95% KM (t) UCL |
| Groundwater | VOCs | Total | Tetrachloroethene (PCE) | N/A | ug/L | 56 | 13 | 4.18 | 4.80 | 8.78 | 1.15 | 2.08 | 2.80 | 95% KM (Percentile Bootstrap) UCL |
| Groundwater | VOCs | Total | Vinyl Chloride | N/A | ug/L | 56 | 16 | 0.404 | 0.335 | 0.955 | 0.257 | 0.189 | 0.307 | 95% KM (t) UCL |

Only those analytes with at least two detected results are shown on this table.

If not all samples were detected results (i.e., detection rate = <100%), the mean and standard deviation were estimated by the Kaplan-Meier method.

The potentially mass wasting soil subset is comprised of the following samples: BIL01USE through BIL09USE, BIL13SSI, and L-01 through L-04.

% = percent

KM = Kaplan-Meier

MDL = method detection limit

N/A = not applicable

ND = non-detect

SD = standard deviation

SVOC = semi-volatile organic carbon

ug/L = micrograms per liter

UCL = upper confidence limit

VOC = volatile organic carbon

 Table 1-8

 Statistical Summary for Groundwater - Sandblast Area AOPC

| | | | | | | | | | | | Maximum | | Statistics - ND censored at the MDL | | |
|----------------|---------------|-----------|-----------------------------|----------|------------|------|-----------|------------|---------|-----------|----------|---------|-------------------------------------|-----------|---------------------------|
| | | Total/ | | Depth | | | Number of | Number of | Mean of | Median of | Detected | | | Final 95% | |
| Medium | Analyte Group | Dissolved | Analyte | Category | Sieve Size | Unit | Samples | Detections | Detects | Detects | Value | KM-Mean | KM-SD | UCL | UCL Type |
| DP Groundwater | Metals | Total | Vanadium | N/A | N/A | ug/L | 10 | 10 | 20.3 | 7.32 | 77.6 | - | - | 47.0 | 95% Approximate Gamma UCL |
| DP Groundwater | Metals | Dissolved | Vanadium | N/A | N/A | ug/L | 10 | 6 | 1.35 | 1.24 | 2.61 | 1.24 | 0.648 | 1.74 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | 1,1,1-Trichloroethane (TCA) | N/A | N/A | ug/L | 10 | 7 | 0.467 | 0.127 | 2.22 | 0.348 | 0.638 | 1.71 | 97.5% KM (Chebyshev) UCL |
| DP Groundwater | VOCs | Total | 1,1-Dichloroethane | N/A | N/A | ug/L | 10 | 5 | 0.604 | 0.173 | 2.52 | 0.330 | 0.730 | 0.876 | 95% KM (BCA) UCL |
| DP Groundwater | VOCs | Total | 1,1-Dichloroethene | N/A | N/A | ug/L | 10 | 2 | 0.614 | 0.614 | 1.16 | 0.173 | 0.331 | 1.64 | 99% KM (Chebyshev) UCL |
| DP Groundwater | VOCs | Total | 1,2,4-Trimethylbenzene | N/A | N/A | ug/L | 10 | 3 | 0.0406 | 0.0415 | 0.0485 | 0.0344 | 0.00555 | 0.0383 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | 2,2-Dichloropropane | N/A | N/A | ug/L | 10 | 3 | 0.142 | 0.130 | 0.179 | 0.125 | 0.0186 | 0.138 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | Acetone | N/A | N/A | ug/L | 10 | 3 | 2.23 | 1.59 | 3.88 | 1.52 | 0.793 | 2.09 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | Benzene | N/A | N/A | ug/L | 10 | 6 | 0.0744 | 0.0600 | 0.137 | 0.0664 | 0.0247 | 0.0821 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | Chloroform | N/A | N/A | ug/L | 10 | 2 | 0.132 | 0.132 | 0.174 | 0.0990 | 0.0250 | 0.119 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | cis-1,2-Dichloroethene | N/A | N/A | ug/L | 10 | 9 | 57.3 | 4.54 | 341 | 51.6 | 106 | 206 | 95% KM (Chebyshev) UCL |
| DP Groundwater | VOCs | Total | Ethylbenzene | N/A | N/A | ug/L | 10 | 4 | 0.0402 | 0.0402 | 0.0447 | 0.0377 | 0.00307 | 0.0399 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | m,p-Xylenes | N/A | N/A | ug/L | 10 | 4 | 0.119 | 0.132 | 0.132 | 0.0971 | 0.0226 | 0.112 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | Naphthalene | N/A | N/A | ug/L | 10 | 2 | 0.0406 | 0.0406 | 0.0452 | 0.0375 | 0.00347 | 0.0411 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | o-Xylene | N/A | N/A | ug/L | 10 | 4 | 0.0507 | 0.0495 | 0.0735 | 0.0385 | 0.0139 | 0.0478 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | Tetrachloroethene (PCE) | N/A | N/A | ug/L | 10 | 10 | 7.80 | 1.45 | 54.5 | - | - | 27.6 | 95% Adjusted Gamma UCL |
| DP Groundwater | VOCs | Total | Toluene | N/A | N/A | ug/L | 10 | 6 | 0.205 | 0.199 | 0.299 | 0.189 | 0.0787 | 0.249 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | trans-1,2-Dichloroethene | N/A | N/A | ug/L | 10 | 3 | 0.995 | 1.09 | 1.80 | 0.365 | 0.563 | 0.765 | 95% KM (t) UCL |
| DP Groundwater | VOCs | Total | Trichloroethene (TCE) | N/A | N/A | ug/L | 10 | 9 | 5.83 | 0.597 | 43.7 | 5.26 | 12.8 | 48.1 | 99% KM (Chebyshev) UCL |
| DP Groundwater | VOCs | Total | Vinyl Chloride | N/A | N/A | ug/L | 10 | 2 | 0.372 | 0.372 | 0.611 | 0.180 | 0.144 | - | - |

Only those analytes with at least two detected results are shown on this table.

If not all samples were detected results (i.e., detection rate = <100%), the mean and standard deviation were estimated by the Kaplan-Meier method.

The erodible soil subset is comprised of the following samples: SB-EUA, SB-EUB, SB-EUA-02, SB-EUA-04, SB-EUA-06, SB-EUA-08, SB-EUB-02, SB-EUB-03, SB-EUB-12, SB-EUB-15, and SB-04.

% = percent DP = Direct Push KM = Kaplan-Meier MDL = method detection limit N/A = not applicable ND = non-detect SD = standard deviation SVOC = semi-volatile organic carbon ug/L = micrograms per liter UCL = upper confidence limit VOC = volatile organic carbon

 Table 1-9

 Statistical Summary for Reference Area Samples

| | | | | | | | | | | Maximum | UPL | |
|-------------|---------------|---------------------|---------------------------------------|-------------------|-------|----------------------|-------------------------|--------------------|----------------------|-------------------|----------------|---------|
| Medium | Analyte Group | Total/ Dissolved | Analyte | Depth Category | Unit | Number of Samples | Number of Detections | Mean of Detects | Median of Detects | Detected Value | Distribution | 95% UPL |
| Soil | Metals | Total | Aluminum | 0-1 ft | mg/kg | 14 | 14 | 22,700 | 22,000 | 33,200 | Normal | 31400 |
| Soil | Metals | Total | Antimony | 0-1 ft | mg/kg | 14 | 14 | 0.128 | 0.130 | 0.180 | Normal | 0.176 |
| Soil | Metals | Total | Arsenic | 0-1 ft | mg/kg | 14 | 14 | 3.10 | 3.22 | 5.18 | Normal | 5.40 |
| Soil | Metals | Total | Barium | 0-1 ft | mg/kg | 14 | 14 | 110 | 105 | 182 | Normal | 169 |
| Soil | Metals | Total | Beryllium | 0-1 ft | mg/kg | 14 | 14 | 0.491 | 0.498 | 0.629 | Normal | 0.659 |
| Soil | Metals | Total | Cadmium | 0-1 ft | mg/kg | 14 | 14 | 0.162 | 0.156 | 0.340 | Gamma | 0.271 |
| Soil | Metals | Total | Chromium | 0-1 ft | mg/kg | 14 | 14 | 21.8 | 21.8 | 27.3 | Normal | 28.1 |
| Soil | Metals | Total | Cobalt | 0-1 ft | mg/kg | 14 | 14 | 16.5 | 17.4 | 19.9 | Non-parametric | 19.9 |
| Soil | Metals | Total | Copper | 0-1 ft | mg/kg | 14 | 14 | 39.0 | 36.6 | 58.2 | Normal | 56.7 |
| Soil | Metals | Total | Lead | 0-1 ft | mg/kg | 14 | 14 | 17.7 | 16.4 | 26.5 | Normal | 25.5 |
| Soil | Metals | Total | Manganese | 0-1 ft | mg/kg | 14 | 14 | 627 | 624 | 920 | Normal | 885 |
| Soil | Metals | Total | Mercury | 0-1 ft | mg/kg | 14 | 14 | 0.0494 | 0.0480 | 0.0680 | Normal | 0.0660 |
| Soil | Metals | Total | Nickel | 0-1 ft | mg/kg | 14 | 14 | 18.7 | 19.2 | 26.1 | Normal | 26.5 |
| Soil | Metals | Total | Selenium | 0-1 ft | mg/kg | 14 | 0 | - | - | - | Non-parametric | 0.500 |
| Soil | Metals | Total | Silver | 0-1 ft | mg/kg | 14 | 12 | 0.0635 | 0.0483 | 0.187 | Non-parametric | 0.187 |
| Soil | Metals | Total | Thallium | 0-1 ft | mg/kg | 14 | 8 | 0.158 | 0.150 | 0.203 | Non-parametric | 0.203 |
| Soil | Metals | Total | Vanadium | 0-1 ft | mg/kg | 14 | 14 | 78.8 | 80.2 | 99.3 | Normal | 104 |
| Soil | Metals | Total | Zinc | 0-1 ft | mg/kg | 14 | 14 | 58.8 | 57.6 | 68.5 | Normal | 71.7 |
| Soil | SVOCs | Total | Acenaphthene | 0-1 ft | ug/kg | 14 | 9 | 1.75 | 1.60 | 3.40 | Non-parametric | 3.40 |
| Soil | SVOCs | Total | Acenaphthylene | 0-1 ft | ug/kg | 14 | 1 | - | - | 1.60 | Non-parametric | 1.60 |
| Soil | SVOCs | Total | Anthracene | 0-1 ft | ug/kg | 14 | 12 | 2.47 | 2.05 | 4.90 | Non-parametric | 4.90 |
| Soil | SVOCs | Total | Fluorene | 0-1 ft | ug/kg | 14 | 1 | - | - | 3.20 | Non-parametric | 3.20 |
| Soil | SVOCs | Total | Naphthalene | 0-1 ft | ug/kg | 14 | 6 | 1.70 | 1.65 | 2.20 | Gamma | 29.4 |
| Soil | SVOCs | Total | Phenanthrene | 0-1 ft | ug/kg | 14 | 14 | 13.2 | 12.0 | 34.0 | Non-parametric | 2.20 |
| Soil | SVOCs | Total | Total LPAHs (KM, capped; NDs at MDL) | 0-1 ft | ug/kg | 14 | 14 | 21.2 | 19.8 | 49.3 | Non-parametric | 49.3 |
| Soil | SVOCs | Total | Benzo(a)anthracene | 0-1 ft | ug/kg | 14 | 14 | 14.0 | 12.5 | 34.0 | Gamma | 28.7 |
| Soil | SVOCs | Total | Benzo(a)pyrene | 0-1 ft | ug/kg | 14 | 14 | 18.2 | 16.0 | 45.0 | Gamma | 37.0 |
| Soil | SVOCs | Total | Benzo(b)fluoranthene | 0-1 ft | ug/kg | 14 | 14 | 23.4 | 21.0 | 55.0 | Gamma | 46.4 |
| Soil | SVOCs | Total | Benzo(g,h,i)perylene | 0-1 ft | ug/kg | 14 | 14 | 13.7 | 12.5 | 32.0 | Gamma | 26.5 |
| Soil | SVOCs | Total | Benzo(k)fluoranthene | 0-1 ft | ug/kg | 14 | 14 | 8.00 | 7.45 | 19.0 | Gamma | 16.2 |
| Soil | SVOCs | Total | Chrysene | 0-1 ft | ug/kg | 14 | 14 | 18.3 | 16.0 | 45.0 | Gamma | 37.4 |
| Soil | SVOCs | Total | Dibenz(a,h)anthracene | 0-1 ft | ug/kg | 14 | 7 | 4.36 | 4.30 | 6.90 | Non-parametric | 6.90 |
| Soil | SVOCs | Total | Fluoranthene | 0-1 ft | ug/kg | 14 | 14 | 27.0 | 24.0 | 66.0 | Gamma | 55.1 |
| Soil | SVOCs | Total | Indeno(1,2,3-cd)pyrene | 0-1 ft | ug/kg | 14 | 14 | 14.1 | 12.0 | 34.0 | Gamma | 27.0 |
| Soil | SVOCs | Total | Pyrene | 0-1 ft | ug/kg | 14 | 14 | 26.8 | 25.0 | 64.0 | Gamma | 53.3 |
| Soil | SVOCs | Total | cPAHs as BaPEQ (KM-capped, MDL-based) | 0-1 ft | ug/kg | 14 | 14 | 26.0 | 22.7 | 64.8 | 95% UPL (t) | 51.6 |
| Soil | SVOCs | Total | Total HPAHs (KM, capped; NDs at MDL) | 0-1 ft | ug/kg | 14 | 14 | 167 | 147 | 401 | Non-parametric | 401 |
| Groundwater | Metals | Total | Aluminum | N/A | ug/L | 4 | 4 | 110 | 111 | 210 | - | - |
| Groundwater | Metals | Total | Antimony | N/A | ug/L | 4 | 4 | 0.0515 | 0.0505 | 0.0800 | - | - |
| Groundwater | Metals | Total | Arsenic | N/A | ug/L | 4 | 3 | 1.50 | 1.60 | 1.68 | - | - |

 Table 1-9

 Statistical Summary for Reference Area Samples

| | | | | | | | | | | Maximum | UPL | |
|-------------|---------------|---------------------|-----------|-------------------|------|----------------------|-------------------------|--------------------|----------------------|-------------------|--------------|---------|
| Medium | Analyte Group | Total/ Dissolved | Analyte | Depth Category | Unit | Number of Samples | Number of Detections | Mean of Detects | Median of Detects | Detected Value | Distribution | 95% UPL |
| Groundwater | Metals | Total | Barium | N/A | ug/L | 4 | 4 | 20.8 | 19.6 | 26.5 | - | - |
| Groundwater | Metals | Total | Beryllium | N/A | ug/L | 4 | 1 | 0.0100 | 0.0100 | 0.0100 | - | - |
| Groundwater | Metals | Total | Cadmium | N/A | ug/L | 4 | 1 | 0.0490 | 0.0490 | 0.0490 | - | - |
| Groundwater | Metals | Total | Calcium | N/A | ug/L | 4 | 4 | 30,000 | 30,000 | 31,000 | - | - |
| Groundwater | Metals | Total | Chromium | N/A | ug/L | 4 | 3 | 1.44 | 0.210 | 3.98 | - | - |
| Groundwater | Metals | Total | Cobalt | N/A | ug/L | 4 | 4 | 0.215 | 0.233 | 0.323 | - | - |
| Groundwater | Metals | Total | Copper | N/A | ug/L | 4 | 3 | 0.502 | 0.260 | 1.12 | - | - |
| Groundwater | Metals | Total | Iron | N/A | ug/L | 4 | 2 | 284 | 284 | 451 | - | - |
| Groundwater | Metals | Total | Lead | N/A | ug/L | 4 | 3 | 0.160 | 0.130 | 0.248 | - | - |
| Groundwater | Metals | Total | Magnesium | N/A | ug/L | 4 | 4 | 7,950 | 7,920 | 8,320 | - | - |
| Groundwater | Metals | Total | Manganese | N/A | ug/L | 4 | 4 | 229 | 226 | 258 | - | - |
| Groundwater | Metals | Total | Mercury | N/A | ug/L | 4 | 0 | - | - | - | - | - |
| Groundwater | Metals | Total | Nickel | N/A | ug/L | 4 | 3 | 2.75 | 1.07 | 6.62 | - | - |
| Groundwater | Metals | Total | Potassium | N/A | ug/L | 4 | 4 | 4,850 | 4,870 | 4,960 | - | - |
| Groundwater | Metals | Total | Selenium | N/A | ug/L | 4 | 0 | - | - | - | - | - |
| Groundwater | Metals | Total | Silver | N/A | ug/L | 4 | 1 | 0.00900 | 0.00900 | 0.00900 | - | - |
| Groundwater | Metals | Total | Sodium | N/A | ug/L | 4 | 4 | 5,210 | 5,140 | 5,620 | - | - |
| Groundwater | Metals | Total | Thallium | N/A | ug/L | 4 | 0 | - | - | - | - | - |
| Groundwater | Metals | Total | Vanadium | N/A | ug/L | 4 | 3 | 0.682 | 0.320 | 1.55 | - | - |
| Groundwater | Metals | Total | Zinc | N/A | ug/L | 4 | 4 | 4.61 | 4.20 | 8.28 | - | - |
| Groundwater | Metals | Dissolved | Aluminum | N/A | ug/L | 4 | 4 | 2.30 | 2.30 | 2.60 | - | - |
| Groundwater | Metals | Dissolved | Antimony | N/A | ug/L | 4 | 2 | 0.0320 | 0.0320 | 0.0500 | - | - |
| Groundwater | Metals | Dissolved | Arsenic | N/A | ug/L | 4 | 4 | 1.37 | 1.39 | 1.55 | - | - |
| Groundwater | Metals | Dissolved | Barium | N/A | ug/L | 4 | 4 | 18.7 | 18.4 | 22.8 | - | - |
| Groundwater | Metals | Dissolved | Beryllium | N/A | ug/L | 4 | 0 | - | - | - | - | - |
| Groundwater | Metals | Dissolved | Cadmium | N/A | ug/L | 4 | 2 | 0.0368 | 0.0368 | 0.0425 | - | - |
| Groundwater | Metals | Dissolved | Calcium | N/A | ug/L | 4 | 4 | 29,700 | 29,800 | 30,100 | - | - |
| Groundwater | Metals | Dissolved | Chromium | N/A | ug/L | 4 | 4 | 0.163 | 0.150 | 0.290 | - | - |
| Groundwater | Metals | Dissolved | Cobalt | N/A | ug/L | 4 | 3 | 0.103 | 0.109 | 0.139 | - | - |
| Groundwater | Metals | Dissolved | Copper | N/A | ug/L | 4 | 3 | 0.163 | 0.0950 | 0.315 | - | - |
| Groundwater | Metals | Dissolved | Iron | N/A | ug/L | 4 | 3 | 15.1 | 10.7 | 24.8 | - | - |
| Groundwater | Metals | Dissolved | Lead | N/A | ug/L | 4 | 2 | 0.0140 | 0.0140 | 0.0220 | - | - |
| Groundwater | Metals | Dissolved | Magnesium | N/A | ug/L | 4 | 4 | 7,690 | 7,660 | 8,080 | - | - |
| Groundwater | Metals | Dissolved | Manganese | N/A | ug/L | 4 | 4 | 207 | 208 | 228 | - | - |
| Groundwater | Metals | Dissolved | Mercury | N/A | ug/L | 4 | 0 | - | - | - | - | - |
| Groundwater | Metals | Dissolved | Nickel | N/A | ug/L | 4 | 3 | 1.05 | 1.07 | 1.41 | - | - |
| Groundwater | Metals | Dissolved | Potassium | N/A | ug/L | 4 | 4 | 4,780 | 4,780 | 4,880 | - | - |
| Groundwater | Metals | Dissolved | Selenium | N/A | ug/L | 4 | 0 | - | - | - | - | - |
| Groundwater | Metals | Dissolved | Silver | N/A | ug/L | 4 | 1 | 0.117 | 0.117 | 0.117 | - | - |
| Groundwater | Metals | Dissolved | Sodium | N/A | ug/L | 4 | 3 | 4,890 | 4,880 | 4,980 | - | - |

| Table 1-9 |
|---|
| Statistical Summary for Reference Area Samples |

| | | | | | | | | | | Maximum | UPL | |
|-------------|---------------|---------------------|-------------------------|-------------------|------|----------------------|-------------------------|--------------------|----------------------|-------------------|--------------|---------|
| Medium | Analyte Group | Total/ Dissolved | Analyte | Depth Category | Unit | Number of Samples | Number of Detections | Mean of Detects | Median of Detects | Detected Value | Distribution | 95% UPL |
| Groundwater | Metals | Dissolved | Thallium | N/A | ug/L | 4 | 0 | - | - | - | - | - |
| Groundwater | Metals | Dissolved | Vanadium | N/A | ug/L | 4 | 0 | - | - | - | - | - |
| Groundwater | Metals | Dissolved | Zinc | N/A | ug/L | 4 | 3 | 1.86 | 1.30 | 3.14 | - | - |
| Groundwater | Butyltins | Total | Dibutyltin | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | Butyltins | Total | Monobutyltin | N/A | ug/L | 1 | 1 | 0.0345 | 0.0345 | 0.0345 | - | - |
| Groundwater | NWTPH-Dx | Total | Diesel Range Organics | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | NWTPH-Dx | Total | Residual Range Organics | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | NWTPH-Gx | Total | Gasoline Range Organics | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | SVOCs | Total | 1,4-Dichlorobenzene | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | SVOCs | Total | 4-Nitrophenol | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | SVOCs | Total | Benzo(b)fluoranthene | N/A | ug/L | 1 | 1 | 0.0250 | 0.0250 | 0.0250 | - | - |
| Groundwater | SVOCs | Total | Benzo(k)fluoranthene | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | SVOCs | Total | Phenanthrene | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | SVOCs | Total | Phenol | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | VOCs | Total | Chloroform | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | VOCs | Total | Tetrachloroethene (PCE) | N/A | ug/L | 1 | 0 | - | - | - | - | - |
| Groundwater | VOCs | Total | Vinyl Chloride | N/A | ug/L | 1 | 0 | - | - | - | - | - |

For soil, analytes with less than 100% detection rate, but at least one detection, the maximum detected value was assessed as the non-parametric UPL.

For soil, analytes with no detections (0% detection rate), the maximum MDL is shown as the non-parametric UPL.

UPLs were not calculated when less than 8 samples were available (i.e., no groundwater UPLs).

- = not available/applicable

BaPEQ = benzo(a)pyrene equivalent(s)

cPAH = carcinogenic polycyclic aromatic hydrocarbon

ft = foot or feet

HPAH = high molecular weight polycyclic aromatic hydrocarbon

KM = Kaplan-Meier

LPAH = low molecular weight polycyclic aromatic hydrocarbon

MDL = method detection limit

mg/kg = milligrams per kilogram

ND = non-detect

 $NWTPH\text{-}Dx = northwest \ total \ petroleum \ hydrocarbon-diesel-extended$

NWTPH-Gx =northwest total petroleum hydrocarbon-gasoline-extended

SVOC = semi-volatile organic carbon

ug/kg = micrograms per kilogram

ug/L = micrograms per liter

UPL = upper prediction limit

VOC = volatile organic carbon

| | | | | COPC Selection - Landfill | | | |
|-----------------------------|----------|-----------------------|----------------------|----------------------------|----------|--------------------|-----------------------|
| | | | | Modified* Residential RSLs | - | Screen Result | |
| | | | Lower of the Maximum | (USEPA) | | | COPC: Yes / No |
| Chemicals of Interest | note | Detection Rate | and 95% UCL | SLV | note | SLV _{RSL} | (Selection Rationale) |
| INORGANICS (mg/kg) | | | | | | | |
| Antimony | n | 27% | 1.48 | 30 | | 0.050 | No |
| Arsenic | с | 100% | 10.5 | 0.64 | | 16 | Yes (Csoil > SLV) |
| Cadmium | с | 74% | 1.17 | 67 | | 0.017 | No |
| Chromium | с | 100% | 594 | 0.288 | ** | 2,065 | Yes (Csoil > SLV) |
| Lead | n | 100% | 511 | 384 | | 1.3 | Yes (Pb Special Case) |
| Mercury | n | 80% | 1.45 | 9 | | 0.16 | Yes (NC>0.1) |
| Silver | n | 39% | 0.450 | 374 | | 0.0012 | No |
| Zinc | n | 100% | 417 | 22,057 | | 0.019 | No |
| BUTYLTINS (ug/kg) | | | | | | | |
| Dibutyltin | n | 22% | 20.1 | 15,344 | | 0.0013 | No |
| Monobutyltin | n | 22% | 38.0 | 15,344 | bo | 0.0025 | No |
| Tributyltin | n | 44% | 60.1 | 15,344 | | 0.0039 | No |
| PESTICIDES (ug/kg) | | | | | <u> </u> | | |
| 4,4'-DDT | с | 30% | 11.8 | 1,822 | | 0.0065 | No |
| Chlordane (technical) | с | 100% | 1,560 | 1,726 | | 0.90 | No |
| Heptachlor | с | 9% | 2.86 | 115 | | 0.025 | No |
| PCBs (ug/kg) | ! | | | • | <u> </u> | | |
| Aroclor 1248 | с | 7% | 364 | 230 | | 1.6 | No (see Total PCB) |
| Aroclor 1254 | с | 38% | 31.1 | 230 | | 0.14 | No |
| Aroclor 1260 | с | 83% | 190 | 230 | | 0.83 | No |
| Total PCBs (Aroclors) | с | 55% | 410 | 230 | | 1.8 | Yes (Csoil > SLV) |
| HERBICIDES (ug/kg) | | | | | | | |
| 2,4,5-T | n | 14% | 70.3 | 594,580 | | 0.00012 | No |
| Dichloroprop | n | 14% | 174 | 469,910 | br | 0.00037 | No |
| MCPP | n | 14% | 11,200 | 59,458 | | 0.19 | Yes (NC>0.1) |
| VOCs (ug/kg) | | • | | | <u> </u> | | |
| 1,2,4-Trimethylbenzene | n | 15% | 2,790 | 55,622 | | 0.050 | No |
| 1,3,5-Trimethylbenzene | n | 15% | 1,110 | 748,020 | | 0.0015 | No |
| Naphthalene | с | 15% | 3,830 | 3,644 | | 1.1 | Yes (Csoil > SLV) |
| Tetrachloroethene (PCE) | с | 58% | 16.5 | 23,016 | | 7.2E-04 | No |
| Toluene | n | 42% | 1.64 | 4,699,100 | | 3.5E-07 | No |
| SVOCs (ug/kg) | • | • | | • | | | |
| Benzoic Acid | n | 36% | 218 | 239,750,000 | | 9.09E-07 | No |
| Bis(2-ethylhexyl) Phthalate | с | 90% | 6,380 | 36,442 | | 0.18 | No |
| Butyl Benzyl Phthalate | с | 11% | 34.9 | 268,520 | | 0.00013 | No |
| Carbazole | с | 72% | 964 | 69,048 | | 0.014 | No |
| Dibenzofuran | n | 45% | 164 | 69,048 | bs | 0.0024 | No |

 Table 2-1

 Hypothetical Fishing Platform User: COPC Selection - Landfill AOPC

| | | nypoinctical P | | COPC Selection - Landin | more | | |
|---------------------------|------|----------------|-------------------------------------|--|------|---|---|
| Chemicals of Interest | note | Detection Rate | Lower of the Maximum and 95% UCL | Modified* Residential RSLs (USEPA) SLV | note | Screen Result SLV _{RSL} | COPC: Yes / No (Selection Rationale) |
| Diethyl Phthalate | n | 18% | 38.9 | 46,991,000 | | 8.3E-07 | No |
| Di-n-butyl Phthalate | n | 14% | 217 | 5,945,800 | | 3.6E-05 | No |
| Pentachlorophenol | С | 7% | 98.8 | 949 | | 0.10 | No |
| PAHs (ug/kg) | | | | | | | |
| 2-Methylnaphthalene | n | 31% | 199 | 220,570 | | 0.00090 | No |
| Acenaphthene | n | 76% | 913 | 3,356,500 | | 0.00027 | No |
| Acenaphthylene | n | 39% | 34.9 | 3,644 | g | 0.00958 | No |
| Anthracene | n | 88% | 1,930 | 16,303,000 | | 0.00012 | No |
| Benzo(a)anthracene | с | 94% | 9,960 | 144 | | 69 | Yes (Csoil > SLV) |
| Benzo(a)pyrene | с | 94% | 11,200 | 14 | | 779 | Yes (Csoil > SLV) |
| Benzo(b)fluoranthene | с | 96% | 15,200 | 144 | | 106 | Yes (Csoil > SLV) |
| Benzo(g,h,i)perylene | с | 91% | 5,820 | 1,439 | b | 4.0 | Yes (Csoil > SLV) |
| Benzo(k)fluoranthene | с | 96% | 26,300 | 1,439 | | 18 | Yes (Csoil > SLV) |
| Benzofluoranthenes, Total | с | 100% | 31,300 | 144 | bx | 218 | Yes (Csoil > SLV) |
| Chrysene | с | 97% | 11,000 | 14,385 | | 0.76 | No |
| Dibenz(a,h)anthracene | с | 85% | 2,060 | 14 | | 143 | Yes (Csoil > SLV) |
| Fluoranthene | n | 94% | 17,500 | 2,205,700 | | 0.0079 | No |
| Fluorene | n | 67% | 324 | 2,205,700 | | 1.5E-04 | No |
| Indeno(1,2,3-cd)pyrene | с | 91% | 6,800 | 144 | | 47 | Yes (Csoil > SLV) |
| Naphthalene | С | 52% | 157 | 3,644 | | 0.043 | No |
| Phenanthrene | n | 97% | 5,780 | 16,303,000 | bt | 0.00035 | No |
| Pyrene | n | 97% | 18,800 | 1,630,300 | | 0.012 | No |
| TPH (mg/kg) | | | | | | | |
| Diesel Range Organics | n | 95% | 1,000 | 1,100 | x3 | 0.91 | No*** |
| Gasoline Range Organics | n | 27% | 157 | 1,200 | x3 | 0.13 | No*** |
| Residual Range Organics | n | 95% | 9,450 | 2,800 | x3 | 3.4 | No*** |

 Table 2-1

 Hypothetical Fishing Platform User: COPC Selection - Landfill AOPC

| Hypothetical Fishing Platform User: COPC Selection - Sandblast Area AOPC | | | | | | | | | | | |
|--|------|----------------|----------------------|---------------------------------------|------|---------------------------|-----------------------|--|--|--|--|
| | | | | | | Screen Result | | | | | |
| | | | Lower of the Maximum | Modified* Residential RSLs (USEPA) | | Csoil | COPC: Yes / No | | | | |
| Chemicals of Interest | note | Detection Rate | | (USEFA) SLV | note | SLV _{RSL} | (Selection Rationale) | | | | |
| INORGANICS (mg/kg) | • | | | ~=- | | | <u> </u> | | | | |
| Antimony | n | 73% | 2.42 | 30 | | 0.081 | No | | | | |
| Arsenic | с | 99% | 9.71 | 0.64 | | 15 | Yes (Csoil > SLV) | | | | |
| Cadmium | c | 89% | 2.63 | 67 | | 0.039 | No | | | | |
| Chromium | c | 100% | 579 | 0.288 | ** | 2013 | Yes (Csoil > SLV) | | | | |
| Lead - sieved <250um, 0-3 ft bgs | n | 100% | 418 | 400 | | 1.0 | Yes (Pb Special Case) | | | | |
| Lead - sieved <2mm, 0-3 ft bgs | n | 100% | 303 | 400 | | 0.76 | Yes (Pb Special Case) | | | | |
| Lead - unsieved, 0-3 ft bgs | n | 100% | 529 | 400 | | 1.3 | Yes (Pb Special Case) | | | | |
| Lead - sieved <250um, 0-1 ft bgs | n | 100% | 921 | 400 | | 2.3 | Yes (Pb Special Case) | | | | |
| Lead - sieved <2mm, 0-1 ft bgs | n | 100% | 768 | 400 | | 1.9 | Yes (Pb Special Case) | | | | |
| Lead - unsieved, 0-1 ft bgs | n | 100% | 3,260 | 400 | | 8.2 | Yes (Pb Special Case) | | | | |
| Nickel | с | 100% | 251 | 1,439 | | 0.17 | Yes (NC > 0.1) | | | | |
| Selenium | n | 57% | 0.479 | 374 | | 0.0013 | No | | | | |
| Silver | n | 96% | 0.152 | 374 | | 0.00041 | No | | | | |
| Zinc | n | 100% | 237 | 22,057 | | 0.011 | No | | | | |
| BUTYLTINS (ug/kg) | | | | | | | | | | | |
| Dibutyltin | n | 29% | 26.6 | 15,344 | | 0.0017 | No | | | | |
| Monobutyltin | n | 36% | 8.62 | 15,344 | bo | 0.00056 | No | | | | |
| Tributyltin | n | 26% | 248 | 15,344 | 20 | 0.016 | No | | | | |
| PESTICIDES (ug/kg) | | | | -)- | | | | | | | |
| 4,4'-DDD | с | 5.4% | 0.721 | 2,110 | | 0.00034 | No | | | | |
| 4,4'-DDE | c | 19% | 0.567 | 1,534 | | 0.00037 | No | | | | |
| 4,4'-DDT | c | 60% | 79.4 | 1,822 | | 0.044 | No | | | | |
| BHC (delta) | n | 5.4% | 1.30 | 82 | с | 0.016 | No | | | | |
| BHC (gamma) Lindane | с | 8% | 1.06 | 537 | | 0.0020 | No | | | | |
| Chlordane (alpha) | c | 12% | 0.776 | 1,726 | d | 0.00045 | No | | | | |
| Chlordane (gamma) | с | 40% | 20.4 | 1,726 | d | 0.012 | No | | | | |
| Dieldrin | с | 5.4% | 0.483 | 32 | - | 0.015 | No | | | | |
| Endosulfan I | n | 14% | 0.809 | 354,830 | f | 2.3E-06 | No | | | | |
| Endosulfan II | n | 8% | 0.447 | 354,830 | f | 1.3E-06 | No | | | | |
| Endosulfan Sulfate | n | 16% | 0.889 | 354,830 | f | 2.5E-06 | No | | | | |
| Endrin | n | 11% | 3.89 | 17,262 | | 0.00023 | No | | | | |
| Endrin Aldehyde | n | 14% | 3.44 | 17,262 | t | 0.00020 | No | | | | |
| Endrin Ketone | n | 11% | 2.88 | 17,262 | t | 0.00017 | No | | | | |
| Heptachlor | с | 11% | 0.415 | 115 | | 0.0036 | No | | | | |
| Methoxychlor | n | 5.4% | 1.04 | 297,290 | | 3.5E-06 | No | | | | |
| PCBs (ug/kg) | | | | | | | | | | | |
| Aroclor 1254 | с | 14% | 136 | 230 | | 0.59 | No | | | | |
| Aroclor 1260 | с | 84% | 167 | 230 | | 0.73 | No | | | | |
| Total PCBs (Aroclors) | с | 84% | 419 | 230 | | 1.8 | Yes (Csoil > SLV) | | | | |

 Table 2-2

 Hypothetical Fishing Platform User: COPC Selection - Sandblast Area AOPC

Screen Result **Modified* Residential RSLs** Csoil Lower of the Maximum (USEPA) SLV_{RSL} note **Chemicals of Interest** note **Detection Rate** and 95% UCL SLV VOCs (ug/kg) 1,2,4-Trimethylbenzene 4,620 55,622 0.083 46% n 1,3,5-Trimethylbenzene 17% 1,630 748.020 0.0022 n 2-Butanone (MEK) 41% 13.3 25,893,000 5.1E-07 n 2-Hexanone 7% 5.12 191,800 2.7E-05 n 37% 20.1 1,822,100 1.1E-05 4-Isopropyltoluene n 0.525 4-Methyl-2-pentanone (MIBK) 20% 5,082,700 1.0E-07 n 122 2.1E-06 50% 58,499,000 Acetone n 37% 0.463 0.00040 1,151 Benzene с 17% 1.11 6,521 0.00017 Bromomethane n Carbon Disulfide 52% 1.12 738,430 1.5E-06 n 7% 5.70 Chloroform 307 0.019 с 4% 120 153,440 0.00078 cis-1,2-Dichloroethene n Dichlorodifluoromethane 9% 6.74 83,433 8.1E-05 n Dichloromethane (Methylene Chloride) 37% 56.0 54,663 1.0E-03 с 5,562 Ethylbenzene 22% 4.69 8.4E-04 с 1,820 527,450 3.5E-03 m,p-Xylenes 24% x2 n Naphthalene 24% 1.94 3,644 5.3E-04 с n-Propylbenzene 7% 35.5 3,164,700 1.1E-05 n o-Xylene 11% 200 623,350 x1 3.2E-04 n 26,300 Tetrachloroethene (PCE) 11% 23,016 1.1 с 57% 0.0020 9,360 4,699,100 Toluene n 2% 2.28 1,534,400 1.5E-06 trans-1,2-Dichloroethene n 1,780 Trichloroethene (TCE) 15% 901 2.0 с SVOCs (ug/kg) Benzoic Acid n 31% 132 239,750,000 5.5E-07 2.12 Bis(2-ethylhexyl)Phthalate 77,100 с 73% 36,442 Butyl Benzyl Phthalate 15% 15.2 268,520 5.7E-05 с 112 Carbazole 62% 69,048 bs 0.0016 с Dibenzofuran 60% 122 69,048 0.0018 n Dimethyl Phthalate 8% 31.0 36,442 0.00085 n 1 Di-n-butyl Phthalate 36% 54.3 5,945,800 9.1E-06 n 22.5 Di-n-octyl Phthalate 15% 36,442 1 0.00062 n Pentachlorophenol с 5.1% 13.3 949 0.014 13% 35.0 17,262,000 2.0E-06 Phenol n PAHs (ug/kg) 2-Methylnaphthalene 23.9 220,570 0.00011 45% n 0.00019 Acenaphthene 78% 652 3,356,500 n 40% 29.6 3,644 0.0081 Acenaphthylene n g 772 88% 16,303,000 0.000047 Anthracene n

Table 2-2Hypothetical Fishing Platform User: COPC Selection - Sandblast Area AOPC

| COPC: Yes / No |
|-----------------------|
| (Selection Rationale) |
| |
| No |
| No No |
| No |
| Yes (Csoil > SLV) |
| No |
| No |
| Yes (Csoil > SLV) |
| |
| No |
| Yes (Csoil > SLV) |
| No |
| |
| No |
| No |
| No |
| No |

Screen Result **Modified* Residential RSLs** Csoil Lower of the Maximum (USEPA) SLV_{RSL} note **Chemicals of Interest** note **Detection Rate** and 95% UCL SLV Benzo(a)anthracene 88% 2,450 144 17 с 14 Benzo(a)pyrene 90% 2,390 166 с Benzo(b)fluoranthene 94% 2,210 144 15 с 1.050 14 Benzo(g,h,i)perylene 83% b 72.99 с 1,439 Benzo(k)fluoranthene 94% 764 0.53 с 144 74 Benzofluoranthenes, Total с 91% 10,600 bx 4,590 14,385 0.32 93% Chrysene с Dibenz(a,h)anthracene 65% 238 14 17 с Fluoranthene 93% 10,800 2,205,700 0.0049 n 0.00010 227 2,205,700 Fluorene n 70% 1,160 144 Indeno(1,2,3-cd)pyrene 88% 8.1 с 81.4 3,644 0.022 Naphthalene 60% с 2,910 16,303,000 Phenanthrene 95% bt 0.00018 n 0.0072 Pyrene 95% 11,700 1,630,300 n TPH (mg/kg) 266 1,100 x3 0.24 Diesel Range Organics 85% n 1,030 1,200 0.86 Gasoline Range Organics 23% x3 n Residual Range Organics 89% 577 2,800 x3 0.21 n

Table 2-2Hypothetical Fishing Platform User: COPC Selection - Sandblast Area AOPC

| COPC: Yes / No (Selection Rationale) |
|---|
| Yes (Csoil > SLV) |
| No |
| Yes (Csoil > SLV) |
| No |
| Yes (Csoil > SLV) |
| No |
| No |
| Yes (Csoil > SLV) |
| No |
| No |
| No |
| |
| No*** |
| No*** |
| No*** |

| | | | | ser. cor c selection - r | | | _ |
|-----------------------------------|---------|-------------------|--|--|------|--|---|
| Chemicals of Interest | note | Detection Rate | Lower of the Maximum and 95% UCL | Modified* Residential RSLs (USEPA) SLV | note | Screen Result C _{soil} SLV _{RSL} | |
| SURFACE SOIL (0-1.5 ft bgs): INOI | RGANICS | (mg/kg) | | | | | |
| Lead | n | 100% | 365 | 400 | | 0.91 | Τ |
| Nickel | n | 100% | 25.9 | 1438.5 | | 0.018 | |
| Zinc | n | 100% | 148 | 22057 | | 0.0067 | |
| SEDIMENT: INORGANICS (mg/kg |) | | | | | | |
| Copper | n | 100% | 25.4 | 3100 | | 0.008 | |
| Lead | n | 100% | 33.0 | 400 | | 0.083 | |
| Nickel | с | 100% | 15.4 | 820 | | 0.019 | |
| Zinc | n | 100% | 174 | 23000 | | 0.0076 | |

 Table 2-3

 Hypothetical Fishing Platform User: COPC Selection - Pistol Range AOPC

| COPC: Yes / No (Selection Rationale) |
|---|
| |
| Yes (Pb Special Case) |
| No |
| No |
| |
| No |
| Yes (Pb Special Case) |
| No |
| No |

| Chemicals of Interest | note | Detection Rate | Lower of the Maximum and 95% UCL | Modified* Residential RSLs (USEPA) SLV | note | Screen Result <u>C_{soil}</u> SLV _{RSL} | COPC: Yes / No (Selection Rationale) |
|-------------------------|------|-------------------|--|--|------|--|---|
| INORGANICS (mg/kg) | | | | | | | |
| Lead | n | 100% | 307 | 400 | | 0.77 | Yes (Pb Special Case) |
| Mercury | n | 100% | 0.720 | 9 | | 0.080 | No |
| PCBs (ug/kg) | | | | | | | |
| Aroclor 1260 | с | 67% | 108 | 230 | | 0.47 | No |
| Total PCBs (Aroclors) | с | 67% | 104 | 230 | | 0.45 | No |
| TPH (mg/kg) | | | | | | | |
| Diesel Range Organics | n | 100% | 102 | 1,100 | | 0.093 | No |
| Residual Range Organics | n | 100% | 232 | 2,800 | | 0.083 | No |

Table 2-4Hypothetical Fishing Platform User: COPC Selection - Bulb Slope AOPC

| | COPCs For Fishi | ng Platform Scenario | |
|---------------------------|----------------------------|-------------------------|-----------------|
| Landfill AOPC | Sandblast Area AOPC | Pistol Range AOPC | Bulb Slope AOPC |
| INORGANICS | INORGANICS | SURFACE SOIL INORGANICS | INORGANICS |
| Arsenic | Arsenic | Lead | Lead |
| Chromium | Chromium | | |
| Lead | Lead | SEDIMENT INORGANICS | |
| Mercury | Nickel | Lead | |
| PCBs | PCBs | | |
| Total PCBs (Aroclors) | Total PCBs (Aroclors) | | |
| VOCs | VOCs | | |
| Naphthalene | Tetrachloroethene (PCE) | | |
| Pesticides | Trichloroethene (TCE) | | |
| MCPP | SVOCs | | |
| PAHs | Bis(2-ethylhexyl)Phthalate | | |
| Benzo(a)anthracene | PAHs | | |
| Benzo(a)pyrene | Benzo(a)anthracene | | |
| Benzo(b)fluoranthene | Benzo(a)pyrene | | |
| Benzo(g,h,i)perylene | Benzo(b)fluoranthene | | |
| Benzo(k)fluoranthene | Benzo(g,h,i)perylene | | |
| Benzofluoranthenes, Total | Benzofluoranthenes, Total | | |
| Dibenz(a,h)anthracene | Dibenz(a,h)anthracene | | |
| Indeno(1,2,3-cd)pyrene | Indeno(1,2,3-cd)pyrene | | |

Table 2-5Hypothetical Fishing Platform User: Summary of COPCs in Soil

AOPC = area of potential concern

COPCs = chemicals of potential concern

PAHs = polycyclic aromatic hydrocarbons

PCB = polychlorinated biphenyl

SVOCs = semi-volatile organic compounds

VOCs = volatile organic compounds

| | | Exposure Po | oint Concentr | ations in Soil | (mg/kg) (b) | | Residential (H | ypothetical Fishing | g Platform User) | Commercial (Occupational Exposure) | | | |
|--------------------------|------------|-------------|---------------------|----------------|-----------------|-------------|----------------------------|--------------------------|--|--------------------------------------|------------------------------|--|--|
| Analyte (a) | Landfil | I AOPC | Sandblast Area AOPC | | Bulb Slope AOPC | | Residential RSL (mg/kg) | ODEQ IRAF (c) | | Commercial/Industrial RSL (mg/kg) | DEQ IRAF (c) | | |
| | 0-3 ft bgs | 0-10 ft bgs | 0-3 ft bgs | 0-10 ft bgs | 0-3 ft bgs | 0-10 ft bgs | Noncancer | Residential Noncancer | Modified Nursing Infant RSL (Noncancer) mg/kg | Noncancer | Non-Residential Noncancer | Modified Nursing Infant RSL (Noncancer) mg/kg | |
| Total PCBs (As Aroclors) | 0.4 | 0.28 | 0.419 | 0.419 | 0.10 | | 1.2 | 4 | 0.3 | 15 | 25 | 0.6 | |
| DDx | 0.012 | 0.01 | 0.08 | 0.079 | | | 37 | 0.3 | 123 | 520 | 2 | 260 | |

Table 2-6Nursing Infant Pathway: Soil Screening

Shading represents a soil concentration exceeding a modified RSL.

RSLs are modified using the following equation: RSL / IRAF = modified Nursing Infant RSL.

-- = not applicable

AOPC = area of potential concern

DDD = dichloro-diphenyl-dichloroethane

DDE = dichloro-diphenyl-dichloroethylene

DDT = dichloro-diphenyl-trichloroethane

DDx = sum of DDD, DDE, and DDT

EPC = exposure point concentration

ft bgs = feet below ground surface

mg/kg = milligram per kilogram

ODEQ IRAF = Oregon Department of Environmental Quality Infant Risk Adjustment Factor

PCBs = polychlorinated biphenyls

RI = Remedial Investigation

RSL= regional screening levels (USEPA 2015)

UCL = upper confidence limit

(a) - ODEQ (2010) lists bioaccumulative compounds IRAFs, of which PCBs and DDx (DDT, DDE, DDD) are relevant for this report. Only noncancer is presented as it is the endpoint that may result in higher health hazard to the nursing infant than the mother. These analytes were not detected in the Pistol Range AOPC and therefore not evaluated. See Section 2.1.

(b) - Soil concentrations are lower of maximum and 95% UCLs (EPCs). See Appendix M of the RI (URS 2012)

(c) - From Table D-3, ODEQ (2010)

Sources

ODEQ. 2010. Human Health Risk Assessment Guidance. Final. October. URS. 2012. Upland and River Operable Units Remedial Investigation Report. June. USEPA. 2015. Regional Screening Levels (RSL) for Chemical Contaminants at Superfund Sites. RSL Table Update. May.

Table 2-7aOccupational Worker: Exposure Factors

| Exposure Factors | Abbreviations | Units | Indoor Worker (IW) | Outdoor Maintenance Worker (OW) | Construction Worker (CW) | Excavation / Trench Worker | References |
|---|------------------------|---|--------------------|------------------------------------|-----------------------------|-------------------------------|------------------------|
| Averaging Time, Carcinogens | ATc | yrs | 70 | 70 | 70 | 70 | USEPA 2015 |
| Averaging Time, Noncarcinogens, adult | ATnc,a | yrs | 25 | 25 | 1 | 1 | USEPA 2015 |
| Body Weight, adult | BW _{adult} | kg | 80 | 80 | 80 | 80 | USEPA 2015 |
| Exposure Duration | ED_w | yrs | 25 | 25 | 1 | 1 | USEPA 2015 |
| Exposure Frequency | EF_w | days/yr | 250 | 225 | 250 | 9 (a) | USEPA 2015 |
| Exposure Time | ET_w | hours/day | 8 | 8 | 8 | 4 (b) | USEPA 2015 |
| Event Frequency (contact with groundwater) | EV_{w} | events/day | NA | NA | NA | 2 (b) | Professional Judgement |
| Event Time (contact with groundwater) | t _{event} | hr/event | NA | NA | NA | 2 (b) | Professional Judgement |
| Soil Ingestion Rate | IR _{soil,w} | mg/day | NA | 100 | 330 | NA | USEPA 2015 |
| Particulate Emission Factor (non-VOCs), Salem, Oregon | PEF | m ³ /kg | NA | $1.36 \ge 10^{+09}$ | $1.0 \ge 10^{+06}$ | NA | USEPA 2015; DTSC 2014 |
| Soil Adherence Factor | AF_{w} | mg/cm ² -event | NA | 0.12 | 0.3 | 0.3 | USEPA 2015 |
| Exposed Body Surface Area (soil) | SAs | cm^2 | NA | 3527 | 3527 | 3300 (ODEQ 2010) | USEPA 2015 |
| Exposed Body Surface Area (groundwater/seep) | SA_w | cm^2 | NA | NA | NA | 5700 | USEPA 2015 |
| Duration of event | t _{event} | hr/event | NA | NA | NA | 2 (b) | Professional Judgement |
| Conversion Factor - dermal | CF _d | kg/mg | NA | 1.0 x 10 ⁻⁰⁶ | $1.0 \ge 10^{-06}$ | $1.0 \ge 10^{-06}$ | NA |
| Conversion Factor - inhalation | CF _i | m^2/cm^2 | 0.0001 | 0.0001 | 0.0001 | 0.0001 | NA |
| Conversion Factor - ingestion | CF _o | kg/mg | NA | 1.0 x 10 ⁻⁰⁶ | $1.0 \ge 10^{-06}$ | 1.0 x 10 ⁻⁰⁶ | NA |
| Inverse of the mean concentration at the center of a 0.5-acre | | 2 | | | | | |
| square source in Salem, Oregon (Site Specific) | Q/C | g/m ² -s per kg/m ³ | 73.4 | 73.4 | 73.4 | 73.4 | USEPA 2002 |
| Relative Bioavailability Percentage | RBA _{arsenic} | % | NA | 60 | 60 | 60 | USEPA 2012 |

Reasonable Maximum Exposure (RME)

Central Tendency Exposure (CTE)

| Exposure Factors | Variable | Units | Indoor Worker (IW) | Outdoor Maintenance Worker (OW) | Construction Worker (CW) | References |
|---|----------------------------|---|--------------------|------------------------------------|-----------------------------|-----------------------|
| Averaging Time, Carcinogens | ATc | yrs | 70 | 70 | 70 | ODEQ 2010, 2013 |
| Averaging Time, Noncarcinogens, adult | ATnc,a | yrs | 6 | 6 | 0.5 | ODEQ 2010, 2013 |
| Body Weight, adult | BW_{adult} | kg | 80 | 80 | 80 | ODEQ 2010, 2013 |
| Exposure Duration | ED_{w} | yrs | 6 | 6 | 0.5 | ODEQ 2010, 2013 |
| Exposure Frequency | EF_{w} | days/yr | 250 | 225 (RME) | 250 | ODEQ 2010, 2013 |
| Exposure Time | ET_w | hours/day | 8 (RME) | 8 (RME) | 8 (RME) | ODEQ 2010, 2013 |
| Event Frequency (contact with groundwater) | EV_w | events/day | NA | NA | NA | NA |
| Event Time (contact with groundwater) | t _{event} | hr/event | NA | NA | NA | NA |
| Soil Ingestion Rate | IR _{soil,w} | mg/day | NA | 50 | 100 | ODEQ 2010, 2013 |
| Particulate Emission Factor (non-VOCs), Salem, Oregon | PEF | m ³ /kg | NA | 1.36 x 10 ⁺⁰⁹ | $1.0 \ge 10^{+06}$ | USEPA 2015; DTSC 2014 |
| Soil Adherence Factor | AF_{w} | mg/cm ² -event | NA | 0.02 | 0.1 | ODEQ 2010, 2013 |
| Exposed Body Surface Area (soil) | SAs | cm ² | NA | 3300 | 3300 | ODEQ 2010, 2013 |
| Exposed Body Surface Area (groundwater/seep) | SA_w | cm^2 | NA | NA | 5700 | ODEQ 2010, 2013 |
| Duration of event | t _{event} | hr/event | NA | NA | NA | ODEQ 2010, 2013 |
| Conversion Factor - dermal | CF _d | kg/mg | NA | NA | NA | NA |
| Conversion Factor - inhalation | CF _i | m ² /cm ² | 0.0001 | NA | NA | NA |
| Conversion Factor - ingestion | CF _o | kg/mg | NA | NA | NA | NA |
| Inverse of the mean concentration at the center of a 0.5-acre square source in Salem, Oregon (Site Specific [USEPA 2002]) | Q/C | g/m ² -s per kg/m ³ | 73.4 | 73.4 | 73.4 | USEPA 2002 |
| Relative Bioavailability Percentage | RBA _{arsenic} | % | NA | 60 | 60 | USEPA 2012 |

Table 2-7aOccupational Worker: Exposure Factors

Notes % = percent cm^2 = square centimeter CTE = central tendency exposure DTSC = California Department of Toxic Substances Control g = gram hr = hour kg = kilogram m² = square meter m³ = cubic meter mg = milligram NA = Not Applicable RME = reasonable maximum exposure VOC = volatile organic chemical yr = year

(RME) CTE value was not available (or higher than the RME) and therefore the RME value was used.

a = The value of 9 days per year was based on standard residential excavation site from ODEQ (2010).

b = Assumes that direct soil contact or groundwater contact activities occur twice a day (morning, afternoon) for a total of 4 hours per day.

Sources

DTSC, 2014. Human Health Risk Assessment (HHRA) Note 2. Available at https://www.dtsc.ca.gov/assessingrisk/humanrisk2.cfm

USEPA. 2002. Supplemental Guidance for Developing Soil Screening levels for Superfund Sites. OSWER 9355.4-24. Solid Waste and Emergency Response. December.

USEPA. 2012. Recommendations for Default Value for Relative Bioavailability of Arsenic in Soil. OSER 9200.1-113.

USEPA. 2015. Regional Screening Levels (RSL) for Chemical Contaminants at Superfund Sites. RSL Table Update. May.

ODEQ. 2010. Human Health Risk Assessment Guidance. Final. October.

ODEQ. 2013. Risk-Based Concentrations for Individual Chemicals - Exposure Factors.

 Table 2-7b

 Hypothetical Fishing platform User: RME & CTE Exposure Factors

| | | | RM | [E ^{a,b} | CTE ^{a,c} | | | |
|---|----------------------------|---|--------------------------|--------------------------|--------------------------|--------------------------|--|--|
| | | | USEP | A 2015 | ODEQ 2010 | | | |
| Definitions | Variable | Units | Tribal Fisher (Adult) | Tribal Fisher (Child) | Tribal Fisher (Adult) | Tribal Fisher (Child) | | |
| Averaging Time, Carcinogens | - | yrs | 70 | 70 | 70 | 70 | | |
| Averaging Time, Noncarcinogens, adult | ATnc,a | yrs | 20 | 6 | 3 | 6 | | |
| Body Weight, adult | BW _{adult} | kg | 80 | 15 | 80 (d) | 15 | | |
| Exposure Duration | ED_w | yrs | 20 | 6 | 3 | 6 | | |
| Exposure Frequency | EF_w | days/yr | 365 (b) | 365 (b) | 152 (c) | 152 (c) | | |
| Soil Ingestion Rate | IR _{soil,w} | mg/day | 100 | 200 | 50 | 100 | | |
| Particulate Emission Factor (non-VOCs), Salem, Oregon | | m ³ /kg | 1.360E+09 | 1.360E+09 | 1.360E+09 | 1.360E+09 | | |
| Soil Adherence Factor | AF_w | mg/cm ² -event | 0.07 | 0.2 | 0.01 | 0.04 | | |
| Exposed Body Surface Area (soil) | SAs | cm ² | 6032 | 2373 | 6032 | 2373 (e) | | |
| Conversion Factor - dermal | CF _d | kg/mg | 1.0E-06 | 1.0E-06 | 1.0E-06 | 1.0E-06 | | |
| Conversion Factor - inhalation | CF _i | m^2/cm^2 | 1.0E-04 | 1.0E-04 | 1.0E-04 | 1.0E-04 | | |
| Conversion Factor - ingestion | CFo | kg/mg | 1.0E-06 | 1.0E-06 | 1.0E-06 | 1.0E-06 | | |
| Inverse of the mean concentration at the center of a 0.5-acre | | | | | | | | |
| square source in Salem, Oregon (Site Specific [USEPA 2002]) | Q/C | g/m ² -s per kg/m ³ | 7.34E+01 | 7.34E+01 | 7.34E+01 | 7.34E+01 | | |
| Relative Bioavailability Percentage | RBA _{arsenic} | % | 60 | 60 | 60 | 60 | | |

% = percent cm^2 = square centimeter CTE = central tendency exposure g = gram HHRA = human health risk assessment hr = hour kg = kilogram m^2 = square meter m^3 = cubic meter mg = milligram NA = Not Applicable OU = operable unit RME = reasonable maximum exposure VOC = volatile organic chemical yr = year

Table 2-7b Hypothetical Fishing platform User: RME & CTE Exposure Factors

(a) This scenario is based on assumed exposure to surface soils only. No exposure to groundwater or soil vapor from the Upland OU. Exposures through fish consumption, swimming, and wading in sediments are covered in River OU HHRA and not included here.

(b) Project specific = For RME, assume year-round usage and residential exposure assumptions, per direction from USACE (E-mail from Mike Gross NWP sent June 24,

(c) Based on estimated frequency of subsistence fishing 3 days a week for 12 months (see email from Michael Gross, USACE, sent June 24, 2014 8:45 am)

(d) USEPA 2015 updated value. Updates to ODEQ (2010) are expected in the near future.

(e) Due to the ODEQ 2010 CTE value being higher than USEPA 2015 RME values, the USEPA 2015 value was used. Updates to ODEQ (2010) are expected in the near

Sources

ODEQ. 2010. Human Health Risk Assessment Guidance. Final. October.

USEPA. 2015. Regional Screening (RSL) User's Guide. Available at: http://www.epa.gov/reg3hscd/risk/human/rb-concentration_table/usersguide.htm (Accessed July 14,

| | | | | | | | v | 2-8 red in the HHRA | | | | | | |
|---|--------------------------|--|--------------------------|----------------------------------|--------------------------------|--|--------------------------|--|---------------------------------------|--|---|---|-------------------------------------|--|
| | Re | Noncancer Health-H ference Dose - Oral | Re | eference centration | 5 | Cancer Oral Slope Factor | -Risk Values | Inhalation Unit Risk | Ab | Oral psorption Factor | <u>Dermal</u> Reference Slope Dose Factor | | A | Dermal bsorption Factor |
| Analyte | RfDo (mg/kg-d) | Reference | RfC $(\mu g/m^3)$ | Reference | SFo (mg/kg-d) ⁻¹ | Reference | $URF_i (\mu g/m^3)^{-1}$ | Reference | OAF (dimensionless) | Reference | RfDd ^{\a} (mg/kg-d) | $\mathbf{SF_D}^{\mathbf{b}}$ (mg/kg-d) ⁻¹ | ABS _d (dimensionless) |) Reference |
| Inorganic Constituents | | | | | | | | | · · · · · · · · · · · · · · · · · · · | | | | · · · | |
| Arsenic | 3.00E-04 | IRIS | 1.50E-02 | CalEPA | 1.50E+00 | IRIS | 4.30E-03 | IRIS | 1.00E+00 | USEPA 2004 (RAGS Part E) | 3.00E-04 | 1.50E+00 | 3.00E-02 | RSL 2013-05 |
| Lead ^{\c} | NV | | NV | | NV | | NV | | NV | | | | NV | |
| Chromium (iii) | 1.50E+00 | IRIS | NV | | NV | | NV | | 0.00E+00 | USEPA 2004 (RAGS Part E) | 0.00E+00 | | 1.00E-02 | DTSC 1994 |
| Antimony | 4.00E-04 | IRIS | NV | | NV | | 1.10E-04 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | 4.00E-04 | | 1.30E-01 | DTSC 1994 |
| Iron | 7.00E-01 | PPRTV | NV | | NV | | 1.10E-03 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | 7.00E-01 | | 1.30E-01 | RSL 2013-05 |
| Mercury | 1.60E-04 | CalEPA | 3.00E-01 | IRIS | NV | | 1.10E-04 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | 1.60E-04 | | 1.30E-01 | DTSC 1994 |
| Nickel | 2.00E-02 | IRIS | 9.00E-02 | ATSDR | NV | | 2.60E-04 | CalEPA | 4.00E-02 | USEPA 2004 (RAGS Part E) | | 8.00E-04 | 1.00E-02 | DTSC 1994 |
| Thallium | 1.00E-05 | Screening PPRTV | NV | | NV | | 1.20E-03 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | 1.00E-05 | | 1.30E-01 | DTSC 1994 |
| Zinc | 3.00E-01 | IRIS | NV | | NV | | 1.10E-04 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | 3.00E-01 | | 1.30E-01 | DTSC 1994 |
| Vanadium | 5.00E-03 | MW-adjusted from vanadium pentoxide; (USEPA RSL table) | 1.00E-01 | ATSDR | NV | | NV | | 2.60E-02 | USEPA 2004 (RAGS Part E) | 1.30E-04 | | 1.00E-02 | DTSC 1994 |
| Polychlorinated Biphenyls (PO | | | 1.00E-01 | AISDK | IN V | | IN V | | 2.00E-02 | USEPA 2004 (RAOS Part E) | 1.30E-04 | | 1.00E-02 | D13C 1994 |
| | | 5 | | | | PCB mixture (CAS# | | PCB mixture (CAS# 1336-36- 3), IRIS upper-bound slope | | | | | | |
| Total PCBs | 2.00E-05 | IRIS | NV | | 2.00E+00 | 1336-36-3), IRIS high- | 5.70E-04 | factor, high risk and persistence, converted to IUR | 1.00E+00 | USEPA 2004 (RAGS Part E) | 2.00E+00 | 2.00E-05 | 1.40E-01 | RSL 2013-05 |
| Polycyclic Aromatic Hydroc | | IKIS | IN V | | 2.00E+00 | risk upper estimate | 5.70E-04 | persistence, converted to IUR | 1.00E+00 | USEPA 2004 (RAGS Part E) | 2.00E+00 | 2.00E-05 | 1.40E-01 | KSL 2013-05 |
| | и | | | | | Environmental Criteria and Assessment Office | | | | | | | | |
| Benzo[a]anthracene | NV | | NV | | 7.30E-01 | (USEPA RSL table) | 1.10E-04 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 7.30E-01 | 1.30E-01 | RSL 2013-05 |
| Benzo[a]pyrene | NV | | NV | | 7.30E+00 | IRIS Environmental Criteria and Assessment Office | 1.10E-03 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 7.30E+00 | 1.30E-01 | RSL 2013-05 |
| Benzo[b]fluoranthene | NV | | NV | | 7.30E-01 | (USEPA RSL table) | 1.10E-04 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 7.30E-01 | 1.30E-01 | RSL 2013-05 |
| Benzo[g,h,i]perylene | 3.00E-02 | pyrene surrogate (IRIS) | NV | | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 3.00E-02 | 1.30E-01 | USEPA 2004 (RAGS Part E) |
| Benzo[k]fluoranthene | NV | | NV | | 7.30E-02 | Environmental Criteria and Assessment Office (USEPA RSL table) Environmental Criteria | 1.10E-04 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 7.30E-02 | 1.30E-01 | RSL 2013-05 |
| Dibenz[a,h]anthracene | NV | | NV | | 7.30E+00 | and Assessment Office (USEPA RSL table) Environmental Criteria | 1.20E-03 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 7.30E+00 | 1.30E-01 | RSL 2013-05 |
| Indeno[1,2,3-cd]pyrene | NV | | NV | | 7.30E-01 | and Assessment Office (USEPA RSL table) Environmental Criteria and Assessment Office | 1.10E-04 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 7.30E-01 | 1.30E-01 | RSL 2013-05 |
| Benzofluoranthenes, total | NV | | NV | | 7.30E-01 | (USEPA RSL table) | 1.10E-04 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 7.30E-01 | 1.30E-01 | RSL 2013-05 |
| Volatile Organic Compound | | | | | | | | | | | | | | |
| Acetone | 9.00E-01 | IRIS | 3.10E+04 | ATSDR | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 9.00E-01 | | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| Benzene | 4.00E-03 | IRIS | 3.00E+01 | IRIS | 5.50E-02 | IRIS | 7.80E-06 | IRIS | 1.00E+00 | USEPA 2004 (RAGS Part E) | 4.00E-03 | 5.50E-02 | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| 1,1-Dichloroethane 1,2-Dichloropropane | 2.00E-01 9.00E-02 | PPRTV ATSDR | NV 4.00E+00 | IRIS | 5.70E-03 3.60E-02 | OEHHA OEHHA | 1.60E-06 1.00E-05 | OEHHA OEHHA | 1.00E+00 1.00E+00 | USEPA 2004 (RAGS Part E) USEPA 2004 (RAGS Part E) | 2.00E-01 9.00E-02 | 5.70E-03 3.60E-02 | 0.00E+00 0.00E+00 | USEPA 2004 (RAGS Part E) USEPA 2004 (RAGS Part E) |
| Ethylbenzene | 9.00E-02 1.00E-01 | IRIS | 1.00E+03 | IRIS | 1.10E-02 | OEHHA | 2.50E-06 | OEHHA | 1.00E+00 | USEPA 2004 (RAGS Part E) | 1.00E-02 | 1.10E-02 | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| Toluene | 8.00E-02 | IRIS | 5.00E+03 | IRIS | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 8.00E-02 | | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| 1,1,1-Trichloroethane | 2.00E+00 | IRIS | 5.00E+03 | IRIS | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 2.00E+00 | | 0.00E+00 | USEPA 2004 (RAGS Part E |
| Xylenes | 2.00E-01 | IRIS | 1.00E+02 | IRIS | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 2.00E-01 | | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| o-xylene | | Xylenes surrogate; IRIS | 1.00E+02 | mixed xylenes surrogate; IRIS | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 2.00E-01 | | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| Chloroform Isopropylbenzene | 1.00E-02 1.00E-01 | IRIS IRIS | 9.80E+01 4.00E+02 | ATSDR IRIS | 3.10E-02 NV | CalEPA | 2.30E-05 NV | IRIS | 1.00E+00 1.00E+00 | USEPA 2004 (RAGS Part E) USEPA 2004 (RAGS Part E) | 1.00E-02 1.00E-01 | 3.10E-02 | 0.00E+00 0.00E+00 | USEPA 2004 (RAGS Part E) USEPA 2004 (RAGS Part E) |
| n-propylbenzene | 1.00E-01 1.00E-01 | Screening PPRTV | | Screening PPRTV | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 1.00E-01 | | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| 1,2,4-trimethylbenzene | NV | | 7.00E+00 | PPRTV | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | | | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| Tetrachloroethene | 6.00E-03 | IRIS | 4.00E+01 | IRIS | 2.10E-03 | IRIS | 2.60E-07 | IRIS | 1.00E+00 | USEPA 2004 (RAGS Part E) | 6.00E-03 | 2.10E-03 | 0.00E+00 | USEPA 2004 (RAGS Part E |
| Naphthalene | 2.00E-02 | IRIS | 3.00E+00 | IRIS | NV | | 3.40E-05 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 2.00E-02 | 0.00E+00 | USEPA 2004 (RAGS Part E |
| 1,1-Dichloroethene | 5.00E-02 | IRIS | 2.00E+02 | IRIS | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 5.00E-02 | | 0.00E+00 | USEPA 2004 (RAGS Part E |
| cis-1,2-Dichloroethene | 2.00E-03 | IRIS | NV | | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 2.00E-03 | | 0.00E+00 | USEPA 2004 (RAGS Part E |

| | | | | | | Toxicit | Table ty Values Us | 2-8 sed in the HHRA | | | | | | |
|-----------------------------|--------------|----------------------------------|---------------|-----------|------------------------------------|---|-----------------------|-------------------------------|-----------------|--------------------------|----------------------------------|-------------------------|------------------|--------------------------|
| | Refe | Noncancer Health rence Dose - | | ference | Cancer-Risk Values Oral Inhalation | | | | | Oral | <u>Dermal</u> Reference Slope | | | Dermal |
| | | Oral | | entration | 8 | Slope Factor | | Unit Risk | At | osorption Factor | Dose | Factor | Ab | osorption Factor |
| | RfDo | | RfC | | SFo | | URF _i | | OAF | | RfDd \a | $\mathbf{SF_D}^{\ \ b}$ | ABS _d | |
| Analyte | (mg/kg-d) | Reference | $(\mu g/m^3)$ | Reference | $(mg/kg-d)^{-1}$ | Reference | $(\mu g/m^3)^{-1}$ | Reference | (dimensionless) | Reference | (mg/kg-d) | $(mg/kg-d)^{-1}$ | (dimensionless) | Reference |
| trans-1,2-Dichloroethene | 2.00E-02 | IRIS | 6.00E+01 | PPRTV | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 2.00E-02 | | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| | | | | | | IRIS; adult | | | | | | | | |
| | | | | | | approximation, no age- | | IRIS; adult approximation, no | | | | | | |
| Trichloroethene | 5.00E-04 | IRIS | 2.00E+00 | IRIS | 4.60E-02 | adjustment | 4.10E-06 | age-adjustment | 1.00E+00 | USEPA 2004 (RAGS Part E) | 5.00E-04 | 4.60E-02 | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| | | | | | | IRIS; continuous lifetime exposure during | | IRIS; continuous lifetime | | | | | | |
| Vinyl Chloride | 3.00E-03 | IRIS | 1.00E+02 | IRIS | 7.20E-01 | adulthood | 4.40E-06 | exposure during adulthood | 1.00E+00 | USEPA 2004 (RAGS Part E) | 3.00E-03 | 7.20E-01 | 0.00E+00 | USEPA 2004 (RAGS Part E) |
| Pesticides | | | | | | | | | | | | | | |
| MCPP | 1.00E-03 | IRIS | NV | | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | | 1.00E-03 | 1.00E-01 | RSL 2013-05 |
| Semivolatile Organic Compo | unds (SVOCs) | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 2.00E-02 | IRIS | NV | | 1.40E-02 | IRIS | 2.40E-06 | CalEPA | 1.00E+00 | USEPA 2004 (RAGS Part E) | 1.40E-02 | 2.00E-02 | 1.00E-01 | RSL 2013-05 |
| Phthalates | | | | | | | | | | | | | | |
| Di-n-octyl phthalate | 1.00E-02 | PPRTV | NV | | NV | | NV | | 1.00E+00 | USEPA 2004 (RAGS Part E) | 1.00E-02 | | 1.00E-01 | RSL 2013-05 |

 a Reference dose adjusted for oral absorption: RfDd = RfDo × OAF (USEPA 2004. RAGS Part E. EPA/540/R/99/005).

^{\b} Slope factor adjusted for oral absorption: SFd = SFo \div OAF (USEPA 2004. RAGS Part E. EPA/540/R/99/005).

¹/^c Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B.

"--" = quantitative toxicity values are not available

 $\mu g/m3 =$ microgram per cubic meter ABS = dermal absorption factor

HHRA = human health risk assessment IUR = inhalation unit risk mg/kg-d = milligram per kilogram day MW = molecular weight NV = No Value OAF = oral absorption factor PCB = polychlorinated biphenyl RfC = reference concentration RfDo = oral reference dose RSL = Regional Screening Level SFo = oral slope factor SVOC = semivolatile organic compound URF = unit risk factor VOC = volatile organic chemical

Sources

ATSDR = Agency for Toxic Substances and Disease Registry, Minimal Risk Level (MRL). Available at http://www.atsdr.cdc.gov/mrls/.

DTSC 1994 = CalEPA Department of Toxic Substances and Control Preliminary Endangerment Assessment (PEA) Guidance Manual January 1994.

HEAST = USEPA Health Effects Assessment Summary Tables, FY 1997 Update (EPA-540-R-97-036).

IRIS = USEPA Integrated Risk Information System. Available at: http://www.epa.gov/iris/.

CalEPA = California Environmental Protection Agency Office of Environmental Health hazard Assessment (OEHHA) Toxicity Database (Website). Available at: http://oehha.ca.gov/tcdb/index.asp.

PPRTV = Provisional Peer-Reviewed Reference Toxicity Value, as cited in USEPA's RSL tables. Available at: http://www.epa.gov/region09/superfund/prg/.

RSL 2013-05 = USEPA Regional Screening Level May 2013.

USEPA 2004 = Risk Assessment Guidance for Superfund: Human Health Evaluation Manual. Part E: Supplemental Guidance for Dermal Evaluation. Final. PB99-963312. July.

| | | | | ancer-Risk Est | | i y – Lanum AC | | Non | cancer-Haza | ard Estimate | | | |
|------------------------------------|----------|-----------|------------|----------------|---------------|-----------------|-----------|----------------------------|-------------|---------------|-----------------|--|--|
| | Soil EPC | | | | | | | | | | | | |
| Analyte | (mg/kg) | Ingestion | Inhalation | Dermal | Multi-Pathway | % of Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | % of Cumulative | | |
| Inorganic Constituents | | | | | | | | | | | | | |
| Arsenic | 1.05E+01 | 2.60E-06 | 1.46E-09 | 3.30E-07 | 2.9E-06 | 5% | 1.62E-02 | 6.35E-05 | 2.05E-03 | 1.8E-02 | 93% | | |
| Chromium (III) | 5.94E+02 | | | | | | 3.05E-04 | | 9.93E-04 | 1.3E-03 | 7% | | |
| Lead ¹ | 5.11E+02 | | | | | | | | | | | | |
| Polycyclic Aromatic Hydrocarbons (| PAHs) | | | | | | | | | | | | |
| Benzo[a]anthracene | 9.96E+00 | 2.00E-06 | 5.91E-11 | 1.10E-06 | 3.1E-06 | 5% | | | | | | | |
| Benzo[a]pyrene | 1.12E+01 | 2.25E-05 | 6.65E-10 | 1.24E-05 | 3.5E-05 | 58% | | | | | | | |
| Benzo[b]fluoranthene | 1.52E+01 | 3.05E-06 | 9.02E-11 | 1.68E-06 | 4.7E-06 | 8% | | | | | | | |
| Dibenz[a,h]anthracene | 2.06E+00 | 4.14E-06 | 1.33E-10 | 2.28E-06 | 6.4E-06 | 11% | | | | | | | |
| Indeno[1,2,3-cd]pyrene | 6.80E+00 | 1.37E-06 | 4.04E-11 | 7.52E-07 | 2.1E-06 | 4% | | | | | | | |
| Benzofluoranthenes, total | 1.88E+01 | 3.77E-06 | 1.11E-10 | 2.08E-06 | 5.8E-06 | 10% | | | | | | | |
| | | | cPA | Hs Cancer Risk | 5.1E-05 | | | | | | | | |
| | | | Cu | mulative Risk: | 6.0E-05 | | | Cumulative Hazard: 2.0E-02 | | | | | |

Table 2-9 **Outdoor Maintenance Worker: RME Summary – Landfill AOPC**

Notes

EPCs from Table 1-1.

Toxicity values from Table 2-8.

Exposure Factors from Table 2-7a.

Bolded values indicate a risk estimate > 1E-06 or a hazard estimate >1.

(1) Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B.

% = percent

"--" = data not available or not calculated AOPC = area of potential concern cPAH = carcinogenic PAHs EPC = exposure point concentration mg/kg = milligrams per kilogram PAH = polycyclic aromatic hydrocarbon RME = reasonable maximum exposure

| Outdoor Wannehance Work: CTE Summary – Landin AOFC | | | | | | | | | | | | | |
|--|----------|-----------|------------|---------------|----------------|-----------------|-----------|------------|-------------|---------------|-----------------|--|--|
| | | | <u>Ca</u> | ancer-Risk E | <u>stimate</u> | | | Non | cancer-Haza | ard Estimate | | | |
| | Soil EPC | | | | | | | | | | | | |
| Analyte | (mg/kg) | Ingestion | Inhalation | Dermal | Multi-Pathway | % of Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | % of Cumulative | | |
| Inorganic Constituents | | | | | | | | | | | | | |
| Arsenic | 1.05E+01 | 3.12E-07 | 3.51E-10 | 1.24E-08 | 3.2E-07 | 6% | 8.09E-03 | 6.35E-05 | 3.20E-04 | 8.5E-03 | 96% | | |
| Chromium (III) | 5.94E+02 | | | | | | 1.53E-04 | | 1.55E-04 | 3.1E-04 | 4% | | |
| Lead ¹ | 5.11E+02 | | | | | | | | | | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | |
| Benzo[a]anthracene | 9.96E+00 | 2.40E-07 | 1.42E-11 | 4.12E-08 | 2.8E-07 | 5% | | | | | | | |
| Benzo[a]pyrene | 1.12E+01 | 2.70E-06 | 1.60E-10 | 4.63E-07 | 3.2E-06 | 57% | | | | | | | |
| Benzo[b]fluoranthene | 1.52E+01 | 3.66E-07 | 2.17E-11 | 6.29E-08 | 4.3E-07 | 8% | | | | | | | |
| Dibenz[a,h]anthracene | 2.06E+00 | 4.97E-07 | 3.20E-11 | 8.52E-08 | 5.8E-07 | 11% | | | | | | | |
| Indeno[1,2,3-cd]pyrene | 6.80E+00 | 1.64E-07 | 9.69E-12 | 2.81E-08 | 1.9E-07 | 3% | | | | | | | |
| Benzofluoranthenes, total | 1.91E+01 | 4.60E-07 | 2.72E-11 | 7.90E-08 | 5.4E-07 | 10% | | | | | | | |
| | | | cPAHs | s Cancer Risk | 5.2E-06 | | | | | | | | |
| | | | Cum | ulative Risk: | 5.5E-06 | | | Cumula | tive Hazard | 8.8E-03 | | | |

 Table 2-10

 Outdoor Maintenance Worker: CTE Summary – Landfill AOPC

EPCs from Table 1-1. Toxicity values from Table 2-8.

Exposure Factors from Table 2-7b.

Bolded values indicate a risk estimate > 1E-06 or a hazard estimate > 1.

(1) Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B.

"--" = data not available or not calculated

% = percent

AOPC = area of potential concern

cPAH = carcinogenic polycyclic aromatic hydrocarbons

CTE = central tendency exposure

EPC = exposure point concentration

mg/kg = milligrams per kilogram

PAH = polycyclic aromatic hydrocarbons

| Construction Worker. KME Summary – Landim AOT C | | | | | | | | | | | | | |
|---|-----------------|-----------|------------|---------------|---------------|------------|-----------|------------|---------------------------|---------------|------------|--|--|
| | | | Ca | ncer-Risk Es | stimate | | | Noncai | Noncancer-Hazard Estimate | | | | |
| | Soil EPC | | | | | % of | | | | | % of | | |
| Analyte | (mg/kg) | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | | |
| Volatile Organic Compou | nds (VOCs) | | | | | | | | | | | | |
| Tetrachloroethene | 1.27E+02 | 1.1E-08 | 1.8E-07 | 0.0E+00 | 1.9E-07 | 4% | 6.0E-02 | 1.2E+00 | 0.0E+00 | 1.2E+00 | 100% | | |
| Polycyclic Aromatic Hydr | ocarbons (PAHs) | | | | | | | | | | | | |
| Benzo[a]pyrene | 1.08E+01 | 3.2E-06 | 3.9E-08 | 1.3E-06 | 4.5E-06 | 96% | | | | | | | |
| | | | Cum | ulative Risk: | 4.7E-06 | | | Cumula | tive Hazard: | 1.2E+00 | | | |

| | Table | 2-11 |
|-----------------------------|-------|-------------------------|
| Construction Worker: | RME | Summary – Landfill AOPC |

EPCs from Table 1-1.

Toxicity values from Table 2-8.

Exposure Factors from Table 2-7a.

Bolded values indicate a risk estimate > 1E-06 or a hazard estimate >1.

% = percent

"--" = data not available or not calculated AOPC = area of potential concern EPC = exposure point concentration mg/kg = milligrams per kilogram RME = reasonable maximum exposure

| Construction worker: CTE Summary – Landin AOFC | | | | | | | | | | | | | |
|--|-----------------|-----------|------------|---------------|---------------|------------|-----------|------------|---------------------------|---------------|------------|--|--|
| | | | Ca | ncer-Risk Es | stimate | | | Nonca | Noncancer-Hazard Estimate | | | | |
| | Soil EPC | | | | | % of | | | | | % of | | |
| Analyte | (mg/kg) | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | | |
| Volatile Organic Compou | nds (VOCs) | | | | | | | | | | | | |
| Tetrachloroethene | 1.27E+02 | 1.63E-09 | | 0.00E+00 | 1.6E-09 | < 1% | 1.81E-02 | | 0.00E+00 | 1.8E-02 | 100% | | |
| Polycyclic Aromatic Hydr | ocarbons (PAHs) | | | | | | | | | | | | |
| Benzo[a]pyrene | 1.08E+01 | 4.82E-07 | 1.94E-08 | 2.07E-07 | 7.1E-07 | 100% | | | | | | | |
| | | | Cum | ulative Risk: | 7.1E-07 | | | Cumula | tive Hazard: | 1.8E-02 | | | |

| Table 2-12 |
|--|
| Construction Worker: CTE Summary – Landfill AOPC |

The VF equation is not adequate for exposure durations of less than 1 year; therefore, the inhalation of VOCs in soil is not evaluated. EPCs from Table 1-1. Toxicity values from Table 2-8. Exposure Factors from Table 2-7a. **Bolded** values indicate a risk estimate > 1E-06 or a hazard estimate >1.

"--" = data not available or not calculated % = percent AOPC = area of potential concern CTE = central tendency exposure EPC = exposure point concentration mg/kg = milligrams per kilogram PAHs = polycyclic aromatic hydrocarbons VF = volatilization factor VOCs = volatile organic chemicals

| | <u>Cancer-Risk Estimate</u> <u>Noncancer Hazard Estimate (Chile</u> | | | | | | | | | | | | |
|-------------------------------|---|----------|-----------|------------|----------------|-----------------|------------|-----------|-------------|-------------------------|----------------|--|--|
| | Soil EPC | | | | meet-misk Est | unate | % of | 1101 | | u Estimate (| <u>Ciniu</u> | | |
| Analyte | (mg/kg) | Mutagen? | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | | |
| Inorganic Constituents | × 8 8/ | | ingestion | Innution | Definition | iniani i adimay | | ingestion | Innutation | Domini | inani i aanvay | | |
| Arsenic | 10.5 | 0 | 1.42E-05 | 7.40E-09 | 1.20E-06 | 1.5E-05 | 1% | 2.80E-01 | 3.09E-04 | 1.99E-02 | 3.0E-01 | | |
| Chromium (III) | 594 | 0 | | | | 1.51-05 | 170 | 5.28E-03 | | 9.64E-03 | 1.5E-02 | | |
| | | | | | | | | 5.201 05 | | 7.04E 05 | 1.52 02 | | |
| Lead ¹ | 0 | 0 | | | | | | | | | | | |
| Mercury | 1.45 | 0 | | | | | | 1.21E-01 | 1.47E-01 | 2.87E-03 | 2.7E-01 | | |
| Polychlorinated Biphenyls (PC | | 0 | 1.000 07 | C 20E 11 | 4.045.07 | 1.50.00 | . 10/ | 2 725 01 | | 0.005.02 | 2 (E 01 | | |
| Total PCBs | 0.41 | 0 | 1.23E-06 | 6.38E-11 | 4.84E-07 | 1.7E-06 | < 1% | 2.73E-01 | | 9.08E-02 | 3.6E-01 | | |
| Volatile Organic Compounds (| | | | 0.105.06 | | 0.15.07 | 10/ | 0.555.00 | 2.405.01 | 7 00 F 04 | 2 (E 01 | | |
| Naphthalene | 3.83 | 0 | | 9.10E-06 | | 9.1E-06 | < 1% | 2.55E-03 | 2.40E-01 | 7.88E-04 | 2.4E-01 | | |
| Pesticides | 11.0 | | | | | | | 4.405.04 | | 0.545.00 | 1.05.01 | | |
| MCPP | 11.2 | 0 | | | | | | 1.49E-01 | | 3.54E-02 | 1.8E-01 | | |
| Polycyclic Aromatic Hydrocar | | | | | | | | | | | | | |
| Benzo[a]anthracene | 9.96 | М | 4.95E-05 | 8.29E-10 | 1.65E-05 | 6.6E-05 | 5% | | | | | | |
| Benzo[a]pyrene | 11.2 | М | 5.57E-04 | 9.32E-09 | 1.86E-04 | 7.4E-04 | 59% | | | | | | |
| Benzo[b]fluoranthene | 15.2 | М | 7.56E-05 | 1.26E-09 | 2.52E-05 | 1.0E-04 | 8% | | | | | | |
| Benzo[g,h,i]perylene | 5.82 | 0 | | | | | | 2.59E-03 | | 7.98E-04 | 3.4E-03 | | |
| Benzo[k]fluoranthene | 26.3 | М | 1.31E-05 | 2.19E-09 | 4.36E-06 | 1.7E-05 | 1% | | | | | | |
| Benzofluoranthenes, Total | 18.78 | М | 9.34E-05 | 1.56E-09 | 3.12E-05 | 1.2E-04 | 10% | | | | | | |
| Dibenz[a,h]anthracene | 2.06 | М | 1.02E-04 | 1.87E-09 | 3.42E-05 | 1.4E-04 | 11% | | | | | | |
| Indeno[1,2,3-cd]pyrene | 6.8 | М | 3.38E-05 | 5.66E-10 | 1.13E-05 | 4.5E-05 | 4% | | | | | | |
| | | | | | cPAH Risk: | 1.2E-03 | | | | | | | |
| | | | | Cur | nulative Risk: | 1.3E-03 | | | Hazard In | dex (Child): | 1.4 | | |
| | | | | Nursin | g Infant Risk: | 1.0E-06 | | | Nursing Inj | fant Hazard: | 1.5 | | |

Table 2-13 Hypothetical Fishing Platform User: RME Summary – Landfill AOPC

Cancer risks are presented for time-integrated adult+child exposure. Noncancer hazards are presented separately for exposure during adulthood and exposure during childhood. Nursing infant's risks and hazards are presented as a function of Mother's risks and hazards, assuming an infant exposure duration of 1 year. EPCs from Table 1-1.

Toxicity values from Table 2-8.

Exposure Factors from Table 2-7b.

Italics = Nursing Infant cancer risk and noncancer hazard based on ODEQ (2010) approach.

Bolded values indicate a risk estimate > 1E-06 or a hazard estimate >1.

(1) Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B.

% = percent

- "--" = data not available or not calculated
- AOPC = area of potential concern
- cPAH = carcinogenic PAH
- EPC = exposure point concentration
- M= mutagenic compound
- mg/kg = milligrams per kilogram
- PAH = polycyclic aromatic hydrocarbon
- PCB = polychlorinated biphenyl
- RME = reasonable maximum exposure

| Hypothetical Fishing Platform User: CTE Summary – Landfill AOPC Cancer-Risk Estimate Noncancer Hazard Estimate (Child) Noncancer Hazard Estimate (Adult) | | | | | | | | | | | | | | | |
|--|----------------|----------|-----------|------------|-----------------|---------------|------------|-----------|-----------------|----------------|---------------|-----------|----------------|----------------|---------------|
| | | | | Car | ncer-Risk Estim | ate | | Ne | oncancer Hazaro | d Estimate (Ch | uild) | No | ncancer Hazard | l Estimate (Ad | <u>dult)</u> |
| | Soil EPC | | | | | | % of | | | | | | | | |
| Analyte | (mg/kg) | Mutagen? | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | Ingestion | Inhalation | Dermal | Multi-Pathway |
| Inorganic Constituents | | | | | | | | | | | | | | | |
| Arsenic | 10.5 | 0 | 2.35E-06 | 1.07E-09 | 6.79E-08 | 2.4E-06 | 1% | 5.83E-02 | 1.29E-04 | 1.66E-03 | 6.0E-02 | 5.47E-03 | 1.29E-04 | 1.98E-04 | 5.8E-03 |
| Chromium (III) | 594 | 0 | | | | | | 1.10E-03 | | 8.03E-04 | 1.9E-03 | 1.03E-04 | | 9.56E-05 | 2.0E-04 |
| Lead ¹ | 0 | 0 | | | | | | | | | | | | | |
| Mercury | 1.45 | 0 | | | | | | 2.52E-02 | 1.05E-01 | 2.39E-04 | 1.3E-01 | 2.36E-03 | 1.05E-01 | 2.85E-05 | 1.1E-01 |
| Polychlorinated Biphenyls (PC | Bs) - Mixtures | | | | | | | | | | | | | | |
| Total PCBs | 0.41 | 0 | 2.04E-07 | 9.20E-12 | 2.75E-08 | 2.3E-07 | < 1% | 5.69E-02 | | 7.56E-03 | 6.4E-02 | 5.34E-03 | | 9.01E-04 | 6.2E-03 |
| Volatile Organic Compounds (| VOCs) | | | | | | | | | | | | | | |
| Naphthalene | 3.83 | 0 | | 2.24E-06 | | 2.2E-06 | 1% | 5.32E-04 | 1.71E-01 | 6.56E-05 | 1.7E-01 | 4.98E-05 | 1.71E-01 | 7.82E-06 | 1.7E-01 |
| Pesticides | | | | | | | | | | | | | | | |
| MCPP | 11.2 | 0 | | | | | | 3.11E-02 | | 2.95E-03 | 3.4E-02 | 2.92E-03 | | 3.52E-04 | 3.3E-03 |
| Polycyclic Aromatic Hydrocart | oons (PAHs) | | | | | | | | | | | | | | |
| Benzo[a]anthracene | 9.96 | М | 1.03E-05 | 3.45E-10 | 1.31E-06 | 1.2E-05 | 5% | | | | | | | | |
| Benzo[a]pyrene | 11.2 | М | 1.16E-04 | 3.88E-09 | 1.47E-05 | 1.3E-04 | 59% | | | | | | | | |
| Benzo[b]fluoranthene | 15.2 | М | 1.57E-05 | 5.27E-10 | 2.00E-06 | 1.8E-05 | 8% | | | | | | | | |
| Benzo[g,h,i]perylene | 5.82 | 0 | | | | | | 5.39E-04 | | 6.65E-05 | 6.1E-04 | 5.05E-05 | | 7.92E-06 | 5.8E-05 |
| Benzo[k]fluoranthene | 26.3 | М | 2.72E-06 | 9.11E-10 | 3.45E-07 | 3.1E-06 | 1% | | | | | | | | |
| Benzofluoranthenes, Total | 18.78 | М | 1.94E-05 | 6.51E-10 | 2.47E-06 | 2.2E-05 | 10% | | | | | | | | |
| Dibenz[a,h]anthracene | 2.06 | М | 2.13E-05 | 7.79E-10 | 2.71E-06 | 2.4E-05 | 11% | | | | | | | | |
| Indeno[1,2,3-cd]pyrene | 6.8 | М | 7.04E-06 | 2.36E-10 | 8.93E-07 | 7.9E-06 | 4% | | | | | | | | |
| | | | | | cPAH Risk: | 2.2E-04 | | | | | | | | | |
| | | | | Cu | mulative Risk: | 2.2E-04 | | | Hazard | Index (Child) | : 0.46 | | Hazard | Index (Adult) | 0.29 |
| | | | | Nursin | ng Infant Risk: | 1.4E-07 | | | Nursing | Infant Hazard: | 0.26 | | | | |

 Table 2-14

 Hypothetical Fishing Platform User: CTE Summary – Landfill AOPC

Cancer risks are presented for time-integrated adult+child exposure. Noncancer hazards are presented separately for exposure during adulthood and exposure during childhood. Nursing infant's risks and hazards are presented as a function of Mother's risks and hazards, assuming an infant exposure duration of 1 year.

EPCs from Table 1-1.

Toxicity values from Table 2-8.

Exposure Factors from Table 2-7b.

Italics = Nursing Infant cancer risk and noncancer hazard based on ODEQ (2010) approach.

Bolded values indicate a risk estimate > 1E-06 or a hazard estimate >1.

(1) Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B.

% = percent

- "--" = data not available or not calculated
- AOPC = area of potential concern
- cPAH = carcinogenic PAH
- CTE = central tendency exposure
- EPC = exposure point concentration
- M= mutagenic compound
- mg/kg = milligrams per kilogram
- PAH = polycyclic aromatic hydrocarbon
- PCB = polychlorinated biphenyl
- VOC = volatile organic chemical

Source

ODEQ. 2010. Human Health Risk Assessment Guidance. Final. October.

| | LACUVU | | , | | to Groundwa | | ininary De | | | | |
|--------------------------|-----------------|-----------|------------|-------------|----------------|------------|------------|------------|--------------|-------------------|------------|
| | | | Ca | ncer-Risk E | <u>stimate</u> | 0/ C | | Nonca | ancer-Hazar | <u>d Estimate</u> | 0/ C |
| | Water EPC | | | | | % of | | | | | % of |
| Analyte | $(\mu g/L)$ | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative |
| Inorganic Constituents | | | | | | | | | | | |
| Antimony | 2.23E-01 | | | | | | 3.44E-06 | | 2.61E-05 | 3.0E-05 | < 1% |
| Iron | 2.91E+04 | | | | | | 2.56E-04 | | 2.92E-04 | 5.5E-04 | 2% |
| Mercury | 7.69E-02 | | | | | | 2.96E-06 | | 3.38E-06 | 6.3E-06 | < 1% |
| Thallium | 2.38E-01 | | | | | | 1.47E-04 | | 1.67E-04 | 3.1E-04 | < 1% |
| Zinc | 2.07E+03 | | | | | | 4.25E-05 | | 2.91E-05 | 7.2E-05 | < 1% |
| Phthalates | | | | | | | | | | | |
| di-n-octyl phthalate | 4.61E+00 | | | | | | 2.84E-06 | | | 2.8E-06 | < 1% |
| Volatile Organic Compoun | ıdı | | | | | | | | | | |
| Acetone | 1.54E+01 | | | | | | 1.05E-07 | 8.41E-06 | 7.64E-08 | 8.6E-06 | < 1% |
| Chloroform | 3.70E+00 | 1.01E-11 | 4.67E-08 | 1.15E-10 | 4.7E-08 | 92% | 2.28E-06 | 1.45E-03 | 2.61E-05 | 1.5E-03 | 4% |
| Isopropylbenzene | 4.60E+00 | | | | | | 2.84E-07 | 4.51E-04 | 3.95E-05 | 4.9E-04 | 1% |
| Tetrachloroethene | 8.78E+00 | 1.62E-12 | 1.01E-09 | 1.12E-10 | 1.1E-09 | 2% | 9.02E-06 | 6.79E-03 | 6.22E-04 | 7.4E-03 | 20% |
| Vinyl Chloride | 9.55E-01 | 6.06E-11 | 2.05E-09 | 6.97E-10 | 2.8E-09 | 6% | 1.96E-06 | 3.26E-04 | 2.26E-05 | 3.5E-04 | < 1% |
| n-propylbenzene | 2.00E+00 | | | | | | 1.23E-07 | 6.96E-05 | 1.80E-05 | 8.8E-05 | < 1% |
| 1,2,4-trimethylbenzene | 5.20E+00 | | | | | | | 2.55E-02 | | 2.6E-02 | 70% |
| | Cumulative Risk | | | | | | | Cumula | tive Hazard: | 3.6E-02 | |

| Table 2-15 |
|--|
| Excavation/Trench Worker, Exposure to Groundwater: RME Summary – Landfill AOPC |

EPCs from Table 1-7. Toxicity values from Table 2-8. Exposure Factors from Table 2-7a. **Bolded** values indicate a risk estimate > 1E-06 or a hazard estimate >1.

"--" = data not available or not calculated % = percent AOPC = area of potential concern EPC = exposure point concentration RME = reasonable maximum exposure ug/L = micrograms per liter VOC = volatile organic chemical

| Cancer-Risk Estimate Noncancer-Hazard Estimate | | | | | | | | | | | | |
|--|----------------|-----------|------------|----------------|---------------|------|-----------|------------|-------------|---------------|-----|--|
| | Soil EPC | | | IICET-KISK ESU | mate | % of | | Nonca | ncer-mazaro | w of | | |
| Analyte | (mg/kg) | Ingestion | Inhalation | Dermal | Multi-Pathway | | Ingestion | Inhalation | Dermal | Multi-Pathway | | |
| Inorganic Constituents | | | | | | | | | | | | |
| Arsenic | 9.71E+00 | 2.40E-06 | 1.35E-09 | 3.05E-07 | 2.7E-06 | 21% | 1.50E-02 | 5.87E-05 | 1.90E-03 | 1.7E-02 | 25% | |
| Chromium (III) | 5.79E+02 | | | | | | 2.97E-04 | | 9.68E-04 | 1.3E-03 | 2% | |
| Lead ¹ | 0.00E+00 | | | | | | | | | | | |
| Polycyclic Aromatic Hydrocarbo | ns (PAHs) | | | | | | | | | | | |
| Benzo[a]pyrene | 2.39E+00 | 4.80E-06 | 1.42E-10 | 2.64E-06 | 7.4E-06 | 57% | | | | | | |
| Benzofluoranthenes Total | 6.36E+00 | 1.28E-06 | 3.78E-11 | 7.03E-07 | 2.0E-06 | 15% | | | | | | |
| Dibenz[a,h]anthracene | 2.38E-01 | 4.78E-07 | 1.54E-11 | 2.63E-07 | 7.4E-07 | 6% | | | | | | |
| Volatile Organic Compounds (| VOCs) | | | | | | | | | | | |
| Tetrachloroethene | 2.73E+01 | 1.58E-08 | 1.71E-07 | 0.00E+00 | 1.9E-07 | 1% | 3.51E-03 | 4.60E-02 | 0.00E+00 | 5.0E-02 | 73% | |
| | | | | | | | | | | | | |
| | | | cPAI | Is Cancer Risk | 1.0E-05 | 78% | | | | | | |
| | Cumulative Ris | | | | | | | Cumulat | ive Hazard: | 6.8E-02 | | |

 Table 2-16

 Outdoor Maintenance Worker – RME Summary - Sandblast Area AOPC

EPCs from Table 1-2.

Toxicity values from Table 2-8.

Exposure Factors from Table 2-7a.

Bolded values indicate a risk estimate > 1E-06 or a hazard estimate >1.

(1) Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B.

"--" = data not available or not calculated

% = percent

AOPC = area of potential concern

cPAH = carcinogenic polycyclic aromatic hydrocarbon

EPC = exposure point concentration

mg/kg = milligrams per kilogram

PAH = polycyclic aromatic hydrocarbon

RME = reasonable maximum exposure

| | | | Ca | ncer-Risk Es | timate | | | Nonca | ancer-Hazar | d Estimate | |
|------------------------------|-------------|-----------|------------|---------------|---------------|------------|-----------|------------|--------------|---------------|------------|
| | Soil EPC | | | | | % of | | | | | % of |
| Analyte | (mg/kg) | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative |
| Inorganic Constituents | | | | | | | | | | | |
| Arsenic | 9.71E+00 | 2.89E-07 | 3.24E-10 | 1.14E-08 | 3.0E-07 | 24% | 7.48E-03 | 5.87E-05 | 2.96E-04 | 7.8E-03 | 96% |
| Chromium (III) | 5.79E+02 | | | | | | 1.49E-04 | | 1.51E-04 | 3.0E-04 | 4% |
| Lead ¹ | 4.18E+02 | | | | | | | | | | |
| Polycyclic Aromatic Hydrocar | bons (PAHs) | | | | | | | | | | |
| Benzo[a]pyrene | 2.39E+00 | 5.76E-07 | 3.40E-11 | 9.89E-08 | 6.8E-07 | 53% | | | | | |
| Benzofluoranthenes Total | 7.84E+00 | 1.89E-07 | 1.12E-11 | 3.24E-08 | 2.2E-07 | 18% | | | | | |
| Dibenz[a,h]anthracene | 2.38E-01 | 5.74E-08 | 3.70E-12 | 9.85E-09 | 6.7E-08 | 5% | | | | | |
| Volatile Organic Compounds | s (VOCs) | | | | | | | | | | |
| Tetrachloroethene | 2.73E+01 | | | 0.00E+00 | | | | | 0.00E+00 | | |
| | | | cPAH | s Cancer Risk | 1E-06 | 76% | | | | | |
| | | | Cum | ulative Risk: | 1.3E-06 | | | Cumula | tive Hazard: | 8.1E-03 | |

 Table 2-17

 Outdoor Maintenance Worker – CTE Summary - Sandblast Area AOPC

EPCs from Table 1-2. Toxicity values from Table 2-8.

Exposure Factors from Table 2-7a.

Bolded values indicate a risk estimate > 1E-06 or a hazard estimate > 1.

(1) Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B.

"--" = data not available or not calculated

% = percent

AOPC = area of potential concern

cPAH = carcinogenic polycyclic aromatic hydrocarbon

CTE = central tendency exposure

EPC = exposure point concentration

mg/kg = milligrams per kilogram

PAH = polycyclic aromatic hydrocarbon

| | | 00 | isti uction | -1 | | il y - Dallubla | ast Alta AU | 10 | | | |
|------------------------|----------------|-----------|-------------|---------------|---------------|-----------------|-------------|------------|--------------|-------------------|------------|
| | | | Ca | ncer-Risk Es | stimate | | | Nonca | ncer-Hazard | l <u>Estimate</u> | |
| | Soil EPC | | | | | % of | | | | | % of |
| Analyte | (mg/kg) | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative |
| Volatile Organic Comp | ounds (VOCs) | | | | | | | | | | |
| Tetrachloroethene | 9.90E+01 | 8.39E-09 | 1.38E-07 | 0.00E+00 | 1.5E-07 | 8% | 4.66E-02 | 9.27E-01 | 0.00E+00 | 9.7E-01 | 100% |
| Polycyclic Aromatic Hy | drocarbons (PA | Hs) | | | | | | | | | |
| Benzo[a]pyrene | 3.98E+00 | 1.17E-06 | 1.43E-08 | 4.89E-07 | 1.7E-06 | 92% | | | | | |
| | | | Cum | ulative Risk: | 1.8E-06 | | | Cumula | tive Hazard: | . 0.97 | |

| Table 2-18 |
|---|
| Construction Worker - RMF Summary - Sandblast Area AOPC |

EPCs from Table 1-2. Toxicity values from Table 2-8. Exposure Factors from Table 2-7a. **Bolded** values indicate a risk estimate > 1E-06 or a hazard estimate >1.

"--" = data not available or not calculated

% = percent

AOPC = area of potential concern

EPC = exposure point concentration

mg/kg = milligrams per kilogram

PAH = polycyclic aromatic hydrocarbon

 $RfD_0 = oral reference dose$

RME = reasonable maximum exposure

| Table 2-19 |
|---|
| Construction Worker – CTE Summary - Sandblast Area AOPC |

| | | | Ca | ncer-Risk Es | stimate | | Noncancer-Hazard Estimate | | | | | |
|-------------------------|----------------|-----------|------------|---------------|---------------|------------|---------------------------|------------|--------------|---------------|------------|--|
| | Soil EPC | | | | | % of | | | | | % of | |
| Analyte | (mg/kg) | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | |
| Volatile Organic Compo | unds (VOCs) | | | | | | | | | | | |
| Tetrachloroethene | 9.90E+01 | 1.27E-09 | | 0.00E+00 | 1.3E-09 | < 1% | 1.41E-02 | | 0.00E+00 | 1.4E-02 | 100% | |
| Polycyclic Aromatic Hyd | rocarbons (PAI | Hs) | | | | | | | | | | |
| Benzo[a]pyrene | 3.98E+00 | 1.78E-07 | 7.14E-09 | 7.62E-08 | 2.6E-07 | 100% | | | | | | |
| | | | Cum | ulative Risk: | 2.6E-07 | | | Cumula | tive Hazard: | 1.4E-02 | | |

The VF equation does not adequately assess VOCs from soil for exposure durations less than 1 year; therefore, this pathway in soil is not evaluated. EPCs from Table 1-2. Toxicity values from Table 2-8. Exposure Factors from Table 2-7a. **Bolded** values indicate a risk estimate > 1E-06 or a hazard estimate >1.

"--" = data not available or not calculated % = percent AOPC = area of potential concern CTE = central tendency exposure EPC = exposure point concentration mg/kg = milligrams per kilogram

- PAH = polycyclic aromatic hydrocarbon
- $RfD_0 = oral reference dose$

VF = volatilization factor

| Hypothetical Fishing Platform User: RME Summary - Sandblast Area AOPC | | | | | | | | | | | | | | | |
|---|---------------|----------|-----------|------------|----------------|----------------|------------|-----------|---------------|---------------|---------------|-----------|--------------|---------------|---------------|
| | | | | <u>C</u> | ancer-Risk Es | stimate | | Noi | ncancer Hazar | d Estimate (| Child) | No | ncancer Haza | rd Estimate (| Adult) |
| | Soil EPC | | | | | | % of | | | | | | | | |
| Analyte | (mg/kg) | Mutagen? | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | Ingestion | Inhalation | Dermal | Multi-Pathway |
| Inorganic Constituents | | | | | | | | | | | | | | | |
| Arsenic | 9.71 | 0 | 1.31E-05 | 6.84E-09 | 1.11E-06 | 1.4E-05 | 5% | 2.59E-01 | 2.86E-04 | 1.84E-02 | 2.8E-01 | 2.43E-02 | 2.86E-04 | 3.07E-03 | 2.8E-02 |
| Chromium (III) | 579 | 0 | | | | | | 5.15E-03 | | 9.39E-03 | 1.5E-02 | 4.83E-04 | | 1.57E-03 | 2.0E-03 |
| Lead ¹ | 418 | 0 | | | | | | | | | | | | | |
| Nickel | 251 | 0 | | 1.78E-08 | | 1.8E-08 | < 1% | 1.67E-01 | 2.05E-03 | 9.93E-02 | 2.7E-01 | 1.57E-02 | 2.05E-03 | 1.66E-02 | 3.4E-02 |
| Polychlorinated Biphenyls (PCB | s) - Mixtures | | | | | | | | | | | | | | |
| Total PCBs | 0.419 | 0 | 1.26E-06 | 6.52E-11 | 4.95E-07 | 1.8E-06 | < 1% | 2.79E-01 | | 9.28E-02 | 3.7E-01 | 2.62E-02 | | 1.55E-02 | 4.2E-02 |
| Volatile Organic Compounds (V | | | | | | | | | | | | | | | |
| Tetrachloroethene | 26.3 | 0 | 8.28E-08 | 1.83E-06 | 0.00E+00 | 1.9E-06 | < 1% | 5.84E-02 | 4.75E-01 | 0.00E+00 | 5.3E-01 | 5.48E-03 | 4.75E-01 | 0.00E+00 | 4.8E-01 |
| Trichloroethene | 1.78 | 0 | 1.23E-07 | 1.12E-06 | 0.00E+00 | 1.2E-06 | < 1% | 4.75E-02 | 3.69E-01 | 0.00E+00 | 4.2E-01 | 4.45E-03 | 3.69E-01 | 0.00E+00 | 3.7E-01 |
| Semivolatile Organic Compound | ls (SVOCs) | | | | | | | | | | | | | | |
| DEHP | 77.1 | 0 | 1.62E-06 | 5.05E-11 | 4.56E-07 | 2.1E-06 | < 1% | 5.14E-02 | | 1.22E-02 | 6.4E-02 | 4.82E-03 | | 2.03E-03 | 6.9E-03 |
| Polycyclic Aromatic Hydrocarbo | ons (PAHs) | | | | | | | | | | | | | | |
| Benzo[a]anthracene | 2.45 | М | 1.22E-05 | 2.04E-10 | 4.06E-06 | 1.6E-05 | 6% | | | | | | | | |
| Benzo[a]pyrene | 2.39 | М | 1.19E-04 | 1.99E-09 | 3.96E-05 | 1.6E-04 | 57% | | | | | | | | |
| Benzo[b]fluoranthene | 2.21 | М | 1.10E-05 | 1.84E-10 | 3.67E-06 | 1.5E-05 | 5% | | | | | | | | |
| Benzo[g,h,i]perylene | 1.05 | 0 | | | | | | 4.67E-04 | | 1.44E-04 | 6.1E-04 | 4.38E-05 | | 2.40E-05 | 6.8E-05 |
| Benzofluoranthenes, Total | 6.36 | М | 3.16E-05 | 5.29E-10 | 1.06E-05 | 4.2E-05 | 15% | | | | | | | | |
| Dibenz[a,h]anthracene | 0.238 | М | 1.18E-05 | 2.16E-10 | 3.95E-06 | 1.6E-05 | 6% | | | | | | | | |
| Indeno[1,2,3-cd]pyrene | 1.16 | М | 5.77E-06 | 9.65E-11 | 1.92E-06 | 7.7E-06 | 3% | | | | | | | | |
| | | | | | cPAH Risk: | 2.5E-04 | | | | | | | | | |
| | | | | Cun | nulative Risk: | 2.8E-04 | | | Hazard I | ndex (Child): | 1.9 | | Hazard I | ndex (Adult): | 0.97 |
| | | | | Nursing | g Infant Risk: | 1.1E-06 | | | Nursing In | fant Hazard: | 1.5 | | | | |

Table 2-20 Hypothetical Fishing Platform User: RME Summary - Sandblast Area AOPC

Notes

Cancer risks are presented for time-integrated adult+child exposure. Noncancer hazards are presented separately for exposure during adulthood and exposure during childhood. Nursing infant's risks and hazards are presented as a function of Mother's risks and hazards, assuming an infant exposure duration of 1 year.

EPCs from Table 1-2.

Toxicity values from Table 2-8.

Exposure Factors from Table 2-7b.

Italics = Nursing Infant cancer risk and noncancer hazard based on ODEQ (2010) approach.

Bolded values indicate a risk estimate > 1E-06 or a hazard estimate >1.

(1) Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B.

"--" = data not available or not calculated

- % = percent
- 0 = non-mutagenic

AOPC = area of potential concern

cPAH = carcinogenic polycyclic aromatic hydrocarbon

- DEHP = bis(2-ethylhexyl)phthalate
- EPC = exposure point concentration

M = mutagenic compound

mg/kg = milligrams per kilogram

PAH = polycyclic aromatic hydrocarbon

RME = reasonable maximum exposure

SVOC = semivolatile organic compound

VOC = volatile organic compound

Table 2-20Hypothetical Fishing Platform User: RME Summary - Sandblast Area AOPC

Source ODEQ. 2010. Human Health Risk Assessment Guidance. Final. October.

| | | | | Hypothet | ical Fishing Pla | atform User: O | CTE Summary | - Sandblast A | rea AOPC | | | | | | |
|-------------------------------|----------------|----------|-----------|------------|---|--|-------------|---------------|-----------------|----------------------------------|---------------|-----------|---------------|-----------------|---------------|
| | | | | <u>Ca</u> | ncer-Risk Estimat | e | | No | oncancer Hazaro | d Estimate (Ch | ild) | <u>N</u> | Noncancer Haz | ard Estimate (. | Adult) |
| | Soil EPC | | | | | | % of | | | | | | | | |
| Analyte | (mg/kg) | Mutagen? | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative | Ingestion | Inhalation | Dermal | Multi-Pathway | Ingestion | Inhalation | Dermal | Multi-Pathway |
| Inorganic Constituents | | | | | | | | | | | | | | | |
| Arsenic | 9.71 | | 2.18E-06 | 9.86E-10 | 6.27E-08 | 2.2E-06 | 5% | 5.39E-02 | 1.19E-04 | 1.54E-03 | 5.6E-02 | 5.05E-03 | 1.19E-04 | 1.83E-04 | 5.4E-03 |
| Chromium (III) | 579 | | | | | | | 1.07E-03 | | 7.82E-04 | 1.9E-03 | 1.00E-04 | | 9.32E-05 | 1.9E-04 |
| Lead ¹ | 418 | | | | | | | | | | | | | | |
| Nickel | 251 | | | 2.57E-09 | | 2.6E-09 | < 1% | 3.48E-02 | 8.54E-04 | 8.27E-03 | 4.4E-02 | 3.27E-03 | 8.54E-04 | 9.85E-04 | 5.1E-03 |
| Polychlorinated Biphenyls (PC | Bs) - Mixtures | | | | | | | | | | | | | | |
| Total PCBs | 0.419 | | 2.09E-07 | 9.40E-12 | 2.81E-08 | 2.4E-07 | < 1% | 5.82E-02 | | 7.73E-03 | 6.6E-02 | 5.45E-03 | | 9.21E-04 | 6.4E-03 |
| Volatile Organic Compounds (| VOCs) | | | | | | | | | | | | | | |
| Tetrachloroethene | 26.3 | | 1.38E-08 | 2.45E-07 | 0.00E+00 | 2.6E-07 | < 1% | 1.22E-02 | 1.83E-01 | 0.00E+00 | 2.0E-01 | 1.14E-03 | 1.83E-01 | 0.00E+00 | 1.8E-01 |
| Trichloroethene | 1.78 | | 2.04E-08 | 2.77E-07 | 0.00E+00 | 3.0E-07 | < 1% | 9.88E-03 | 2.63E-01 | 0.00E+00 | 2.7E-01 | 9.27E-04 | 2.63E-01 | 0.00E+00 | 2.6E-01 |
| Semivolatile Organic Compour | nds (SVOCs) | | | | | | | | | | | | | | |
| DEHP | 77.1 | | 2.69E-07 | 7.28E-12 | 2.58E-08 | 2.9E-07 | < 1% | 1.07E-02 | | 1.02E-03 | 1.2E-02 | 1.00E-03 | | 1.21E-04 | 1.1E-03 |
| Polycyclic Aromatic Hydrocar | bons (PAHs) | | | | | | | | | | | | | | |
| Benzo[a]anthracene | 2.45 | М | 2.54E-06 | 8.49E-11 | 3.22E-07 | 2.9E-06 | 6% | | | | | | | | |
| Benzo[a]pyrene | 2.39 | М | 2.47E-05 | 8.28E-10 | 3.14E-06 | 2.8E-05 | 58% | | | | | | | | |
| Benzo[b]fluoranthene | 2.21 | М | 2.29E-06 | 7.66E-11 | 2.90E-07 | 2.6E-06 | 5% | | | | | | | | |
| Benzo[g,h,i]perylene | 1.05 | | | | | | | 9.72E-05 | | 1.20E-05 | 1.1E-04 | 9.11E-06 | | 1.43E-06 | 1.1E-05 |
| Benzofluoranthenes, Total | 6.36 | М | 6.58E-06 | 2.20E-10 | 8.35E-07 | 7.4E-06 | 15% | | | | | | | | |
| Dibenz[a,h]anthracene | 0.238 | М | 2.46E-06 | 9.00E-11 | 3.13E-07 | 2.8E-06 | 6% | | | | | | | | |
| Indeno[1,2,3-cd]pyrene | 1.16 | М | 1.20E-06 | 4.02E-11 | 1.52E-07 | 1.4E-06 | 3% | | | | | | | | |
| | | | | | cPAH Risk: Cumulative Risk: sing Infant Risk: | 4.5E-05 4.8E-05 <i>1.4E-07</i> | | | | Index (Child): Infant Hazard: | | | Hazard | Index (Adult): | 0.47 |

Table 2-21

Notes

Cancer risks are presented for time-integrated adult+child exposure. Noncancer hazards are presented separately for exposure during adulthood and exposure during childhood.

Nursing infant's risks and hazards are presented as a function of Mother's risks and hazards, assuming an infant exposure duration of 1 year.

EPCs from Table 1-2.

Toxicity values from Table 2-8.

Exposure Factors from Table 2-7b.

Italics = Nursing Infant cancer risk and noncancer hazard based on ODEQ (2010) approach.

Bolded values indicate a risk estimate > 1E-06 or a hazard estimate >1.

(1) Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B.

"--" = data not available or not calculated

- % = percent
- 0 =non-mutagenic
- AOPC = area of potential concern
- cPAH = carcinogenic polycyclic aromatic hydrocarbon
- CTE = central tendency exposure
- DEHP = bis(2-ethylhexyl)phthalate
- EPC = exposure point concentration
- M = mutagenic compound
- mg/kg = milligrams per kilogram
- PAH = polycyclic aromatic hydrocarbon
- SVOC = semivolatile organic compound VOC = volatile organic compound

Source

ODEQ. 2010. Human Health Risk Assessment Guidance. Final. October.

| | Excavation/1 | | · • | cer-Risk E | | | ary - Bandb | | | d Estimate | |
|---------------------------|--------------|-----------|------------|-------------|---------------|------|-------------|------------|-------------|-------------------|------------|
| | Water EPC | | Can | CEI-RISK E | sumate | % of | | Nonca | ncer-mazar | <u>u Estimate</u> | % of |
| Analyte | $(\mu g/L)$ | Ingestion | Inhalation | Dermal | Multi-Pathway | | Ingestion | Inhalation | Dermal | Multi-Pathway | Cumulative |
| Inorganic Constituents | • • | | | | * | | | | | | |
| Vanadium | 2.23E-01 | | | | | | 2.75E-07 | | 1.21E-05 | 1.2E-05 | < 1% |
| Volatile Organic Compound | ds (VOCs) | | | | | | | | | | |
| Acetone | 3.88E+00 | | | | | | 2.66E-08 | 2.12E-06 | 1.93E-08 | 2.2E-06 | < 1% |
| Benzene | 1.37E-01 | 6.64E-13 | 5.86E-10 | 1.40E-11 | 6.0E-10 | < 1% | 2.11E-07 | 1.75E-04 | 4.45E-06 | 1.8E-04 | < 1% |
| Carbon disulfide | 2.55E-01 | | | | | | 1.57E-08 | 1.43E-05 | 2.55E-07 | 1.5E-05 | < 1% |
| Chloroform | 1.74E-01 | 4.75E-13 | 1.77E-09 | 5.43E-12 | 1.8E-09 | 2% | 1.07E-07 | 5.49E-05 | 1.23E-06 | 5.6E-05 | < 1% |
| 1,1-Dichloroethane | 5.00E+00 | 2.51E-12 | 3.90E-09 | 2.62E-11 | 3.9E-09 | 4% | 1.54E-07 | | 1.61E-06 | 1.8E-06 | < 1% |
| 1,1-Dichloroethene | 1.16E+00 | | | | | | 1.43E-07 | 2.02E-04 | 2.53E-06 | 2.0E-04 | < 1% |
| cis-1,2-Dichloroethene | 6.60E+02 | | | | | | 2.03E-03 | | 2.40E-02 | 2.6E-02 | 4% |
| trans-1,2-Dichloroethene | 1.80E+00 | | | | | | 5.55E-07 | 1.04E-03 | 9.24E-06 | 1.0E-03 | < 1% |
| 1,2-Dichloropropane | 1.79E-01 | 5.67E-13 | 8.09E-10 | 6.95E-12 | 8.2E-10 | < 1% | 1.23E-08 | 1.42E-03 | 1.50E-07 | 1.4E-03 | < 1% |
| Ethylbenzene | 4.47E-02 | 4.33E-14 | 5.28E-11 | 3.17E-12 | 5.6E-11 | < 1% | 2.76E-09 | 1.48E-06 | 2.02E-07 | 1.7E-06 | < 1% |
| Tetrachloroethene | 5.45E+01 | 1.01E-11 | 5.38E-09 | 6.95E-10 | 6.1E-09 | 6% | 5.60E-05 | 3.62E-02 | 3.86E-03 | 4.0E-02 | 5% |
| Toluene | 6.40E-01 | | | | | | 4.93E-08 | 4.54E-06 | 2.21E-06 | 6.8E-06 | < 1% |
| 1,1,1-Trichloroethane | 2.22E+00 | | | | | | 6.84E-09 | 1.32E-05 | 1.53E-07 | 1.3E-05 | < 1% |
| Trichloroethene | 4.37E+01 | 1.77E-10 | 7.62E-08 | 3.61E-09 | 8.0E-08 | 74% | 5.39E-04 | 6.51E-01 | 1.10E-02 | 6.6E-01 | 90% |
| Vinyl Chloride | 4.10E+00 | 2.60E-10 | 1.12E-08 | 2.99E-09 | 1.4E-08 | 13% | 8.42E-06 | 1.78E-03 | 9.70E-05 | 1.9E-03 | < 1% |
| Xylenes | 1.32E-01 | | | | | | 4.07E-09 | 4.35E-05 | 3.03E-07 | 4.4E-05 | < 1% |
| 1,2,4-trimethylbenzene | 4.85E-02 | | | | | | | 2.15E-04 | | 2.1E-04 | < 1% |
| o-xylene | 7.35E-02 | | | | | | 2.27E-09 | 2.42E-05 | 1.59E-07 | 2.4E-05 | < 1% |
| Naphthalene | 4.52E-02 | | 5.96E-10 | | 6.0E-10 | < 1% | 1.39E-08 | 4.09E-04 | 1.08E-06 | 4.1E-04 | < 1% |
| | | | Cumul | ative Risk: | 1.1E-07 | | | Cumulat | ive Hazard: | 0.73 | |

 Table 2-22

 Excavation/Trench Worker, Exposure to Groundwater: RME Summary - Sandblast Area AOPC

EPCs from Table 1-8. Toxicity values from Table 2-8. Exposure Factors from Table 2-7a. **Bolded** values indicate a risk estimate > 1E-06 or a hazard estimate >1.

RME = reasonable maximum exposure

"--" = data not available or not calculated

% = percent

AOPC = area of potential concern

EPC = exposure point concentration

ug/L = micrograms per liter

 Table 2-23

 Indoor Worker, Exposure to Indoor Air: RME Summary - Sandblast Area AOPC

| Definition | Variable | Valu | ue | | Risk and Hazard | l Equations | | | | |
|---|----------------------------|------------------------------------|----------------------|----------------------------------|--------------------|--|------------------|-------------------------------------|---|----------------|
| Attenuation Factor | α | chemical-specif | ic dimensionless | | Cancer Risk: | | | | | |
| Averaging Time, carcinogens | AT _c | - | 70 yrs | | Die | $k = IIIP \times \begin{bmatrix} c \\ c \end{bmatrix}$ | . ~ [| $(\underline{ED_w \times EF})$ | $W_w \times ET_w$ | |
| Averaging Time, noncarcinogens, worker | AT _{nc,w} | | 25 yrs | | ALS: | $sk = IUR \times C_{ind}$ | oor air ^ | $AT_{a} \times 365 \frac{da}{da}$ | $\frac{y}{x} \times 24 \frac{hour}{y}$ | |
| COPC Concentration in indoor air | Cindoor air | chemical-specif | ïc μg/m ³ | | Noncancer Haza | rd: L | | Vince vec | ir ⁽¹ = ¹ day /] | |
| COPC Concentration in soil gas | C _{soil gas} | chemical-specif | ïc μg/m ³ | | | 1 [| | (ED | \times EF., \times ET., | \1 |
| Exposure Duration, worker | ED_{w} | 2 | 25 yrs | | На | $zard = \frac{1}{RfC} \times 0$ | - indoor air | $\times \left(\frac{1}{1} \right)$ | $\frac{\times EF_w \times ET_w}{65 \frac{day}{year} \times 24 \frac{hou}{day}}$ | \overline{r} |
| Exposure Frequency, worker | EF_{w} | 25 | 50 days/yr | | Where: | , 0 | | $\left(AT_{nc,w} \times 3\right)$ | $65 \frac{1}{year} \times 24 \frac{1}{day}$ | , / |
| Exposure Time, worker | ET_{w} | | 8 hours/day | | | L | | , | | /] |
| Inhalation Reference Concentration Inhalation Unit-Risk Factor | RfC IUR | chemical-specif chemical-specif | | | C_i | $ndoorair = C_{soilg}$ | _{as} ×α | | | |
| | | | | | | Cancer | | | Hazard | |
| | | C _{soil gas} | α | $\mathbf{C}_{\text{indoor air}}$ | IUR | Risk | % of | RfC | Quotient | % of |
| Analyte | | $(\mu g/m^3)$ | (dimensionless) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| Hypothetical Future Scenario: Volatile O | rganic Compo | ounds (VOCs) | | | | | | | | |
| tetrachloroethene | | 3.40E+04 | 5.06E-04 | 1.72E+01 | 2.60E-07 | 3.65E-07 | 77% | 4.00E+01 | 9.82E-02 | 73% |
| trichloroethene | | 6.10E+02 | 5.24E-04 | 3.20E-01 | 4.10E-06 | 1.07E-07 | 23% | 2.00E+00 | 3.65E-02 | 27% |
| Current Scenario BISB: Volatile Organic | Compounds | (VOCs) | | | | | | | | |
| tetrachloroethene | | 1.80E+03 | 2.34E-04 | 4.21E-01 | 2.60E-07 | 8.94E-09 | 2% | 4.00E+01 | 2.41E-03 | 2% |
| trichloroethene | | 2.70E+01 | 2.38E-04 | 6.43E-03 | 4.10E-06 | 2.15E-09 | < 1% | 2.00E+00 | 7.34E-04 | < 1% |
| Current Scenario Equipment Building: V | olatile Organi | ic Compounds (V | VOCs) | | | | | | | |
| tetrachloroethene | | 7.30E+02 | 8.88E-04 | 6.48E-01 | 2.60E-07 | 1.37E-08 | 3% | 4.00E+01 | 3.70E-03 | 3% |
| trichloroethene | | 4.10E+01 | 9.41E-04 | 3.86E-02 | 4.10E-06 | 1.29E-08 | 3% | 2.00E+00 | 4.40E-03 | 3% |
| Hypothetical Future Scenario: Totals | | | | | | | | | | |
| | | | | | Pathway Sums: | Cancer Risk 4.7E-07 | | | Hazard Index 1.3E-01 | |
| Current Scenario BISB: Totals | | | | | | | | | | |
| | | | | | Pathway Sums: | Cancer Risk 1.1E-08 | | | Hazard Index 3.1E-03 | |

 Table 2-23

 Indoor Worker, Exposure to Indoor Air: RME Summary - Sandblast Area AOPC

| Definition | Variable | Value | 2 | | Risk and Haza | rd Equations | | | | |
|--|----------------------------|--|--------------------|-------------------------|--------------------|--|-------------|---------------------------------------|---|----------------|
| Attenuation Factor | α | chemical-specific | e dimensionless | | Cancer Risk: | | | | | |
| Averaging Time, carcinogens | AT _c | 70 |) yrs | | D | $isk = IUR \times \left[C_{ind}\right]$ | / | $(ED_w \times EF_w)$ | $_{w} \times ET_{w}$ | |
| Averaging Time, noncarcinogens, worker | $AT_{nc,w}$ | 25 | 5 yrs | | К | $lsk = 10K \land C_{ind}$ | oor air ^ | $AT_{a} \times 365 \frac{day}{day}$ | $\frac{1}{2} \times 24 \frac{hour}{hour}$ | |
| COPC Concentration in indoor air | Cindoor air | chemical-specific | $p \mu g/m^3$ | | Noncancer Haz | zard: | | Vea | $r^{\prime\prime} = r^{\prime} day /]$ | |
| COPC Concentration in soil gas | C _{soil gas} | chemical-specific | $p \mu g/m^3$ | | | | | | | \1 |
| Exposure Duration, worker | ED_{w} | 25 | 5 yrs | | Н | $Tazard = \frac{1}{RfC} \times \left[0 \right]$ | Gindoor air | $\times \left(\frac{22W}{1} \right)$ | $\frac{dav}{dav}$ hou | $\frac{1}{r}$ |
| Exposure Frequency, worker | EF_{w} | 250 |) days/yr | | Where: | Ny O | | $\langle AT_{nc,w} \times 3 \rangle$ | $65 \frac{ddy}{year} \times 24 \frac{ddy}{day}$ | , / |
| Exposure Time, worker | ET_{w} | 8 | 3 hours/day | | | L | | X X | | L / |
| Inhalation Reference Concentration | RfC | chemical-specific | | | (| $C_{indoor\ air} = C_{soil\ a}$ | as ×α | | | |
| Inhalation Unit-Risk Factor | IUR | chemical-specific | $(\mu g/m^3)^{-1}$ | | | 1111001 uti 3011 g | Jus | | | |
| | | | | | | Cancer | | | Hazard | |
| | | $\mathbf{C}_{\mathbf{soil}\ \mathbf{gas}}$ | α | C _{indoor air} | IUR | Risk | % of | RfC | Quotient | % of |
| Analyte | | $(\mu g/m^3)$ | (dimensionless) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| | | | | | | | | | | |
| Current Scenario Equipment Building: | Totals | | | | | | | | | |
| | | | | | | Concer Diels | | | Horond Indon | |
| | | | | | Pathway Sums | s: 2.7E-08 | | | Hazard Index 8.1E-03 | |
| | | | | | | | | | | |
| Notes | | | | | | | | | | |

EPCs from Table 1-6. Toxicity values from Table 2-8. Exposure Factors from Table 2-7a. **Bolded** values indicate a risk estimate > 1E-06 or a hazard estimate >1.

 α = Attenuation factor

See Appendix B for USEPA model.

- "--" = data not available or not calculated
- % = percent
- AOPC = area of potential concern
- BISB = existing building on site
- EPC = exposure point concentration
- $ug/m^3 = micrograms per cubic meter$
- $RME \ = reasonable \ maximum \ exposure$
- VOCs = volatile organic compounds

 Table 2-24

 Indoor Worker, Exposure to Indoor Air: CTE Summary - Sandblast Area AOPC

| Definition | Variable | Valu | ıe | | Risk and Hazard | 1 Equations | | | | |
|---|----------------------------|--|----------------------|-------------------------|--------------------|---|-------------------------|-------------------------------------|---|-------------------------------|
| Attenuation Factor | α | chemical-specif | ic dimensionless | | Cancer Risk: | | | | | |
| Averaging Time, carcinogens | AT _c | 7 | '0 yrs | | Di | $sk = IUR \times \left[C_{ind}\right]$ | ~ / | $\ell ED_w \times EF$ | $W_w \times ET_w$ | |
| Averaging Time, noncarcinogens, worker | AT _{nc,w} | | 6 yrs | | Πι. | $SK = IOK \land C_{ind}$ | oor air ^ | $AT_{e} \times 365 \frac{da}{da}$ | $\frac{y}{x 24}$ hour | |
| COPC Concentration in indoor air | Cindoor air | chemical-specif | ic µg/m ³ | | Noncancer Haza | ard: | (| vec | ir ⁽¹ = ¹ day /] | |
| COPC Concentration in soil gas | C _{soil gas} | chemical-specif | ic µg/m ³ | | | 1 [| | / ED | \times EF., \times ET., | \] |
| Exposure Duration, worker | ED_{w} | | 6 yrs | | На | $azard = \frac{-}{RfC} \times \left[\frac{1}{RfC} \right]$ | C _{indoor} air | $\times \left(\frac{1}{1} \right)$ | $\frac{\times EF_w \times ET_w}{65 \frac{day}{vear} \times 24 \frac{how}{day}}$ | $\left \overline{r} \right $ |
| Exposure Frequency, worker | EF_{w} | 25 | 50 days/yr | | Where: | , 0 | | $\left(AT_{nc,w} \times 3\right)$ | $65 {year} \times 24 {day}$ | <u>,</u> / |
| Exposure Time, worker | ET_{w} | | 8 hours/day | | | - | | , | | / - |
| Inhalation Reference Concentration Inhalation Unit-Risk Factor | RfC IUR | chemical-specifichemical-specific | | | Ci | $indoor air = C_{soil g}$ | _{jas} ×α | | | |
| | | | | | | Cancer | | | Hazard | |
| | | $\mathbf{C}_{\mathbf{soil}\ \mathbf{gas}}$ | α | C _{indoor air} | IUR | Risk | % of | RfC | Quotient | % of |
| Analyte | | $(\mu g/m^3)$ | (dimensionless) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| Hypothetical Future Scenario: Volatile O | rganic Compo | ounds (VOCs) | | | | | | | | |
| tetrachloroethene | | 3.40E+04 | 5.06E-04 | 1.72E+01 | 2.60E-07 | 8.76E-08 | 77% | 4.00E+01 | 9.82E-02 | 73% |
| trichloroethene | | 6.10E+02 | 5.24E-04 | 3.20E-01 | 4.10E-06 | 2.56E-08 | 23% | 2.00E+00 | 3.65E-02 | 27% |
| Current Scenario BISB: Volatile Organic | Compounds | (VOCs) | | | | | | | | |
| tetrachloroethene | | 1.80E+03 | 2.34E-04 | 4.21E-01 | 2.60E-07 | 2.14E-09 | 2% | 4.00E+01 | 2.41E-03 | 2% |
| trichloroethene | | 2.70E+01 | 2.38E-04 | 6.43E-03 | 4.10E-06 | 5.16E-10 | < 1% | 2.00E+00 | 7.34E-04 | < 1% |
| Current Scenario Equipment Building: V | olatile Organ | ic Compounds (V | VOCs) | | | | | | | |
| tetrachloroethene | | 7.30E+02 | 8.88E-04 | 6.48E-01 | 2.60E-07 | 3.30E-09 | 3% | 4.00E+01 | 3.70E-03 | 3% |
| trichloroethene | | 4.10E+01 | 9.41E-04 | 3.86E-02 | 4.10E-06 | 3.10E-09 | 3% | 2.00E+00 | 4.40E-03 | 3% |
| Hypothetical Future Scenario: Volatile O | rganic Compo | ounds (VOCs) | | | | | | | | |
| | | | | | Pathway Sums: | Cancer Risk 1E-07 | | | Hazard Index 1.3E-01 | |
| Current Scenario BISB: Volatile Organic Co | ompounds (VO | Cs) | | | | | | | | |
| | | | | | Pathway Sums: | Cancer Risk 3E-09 | | | Hazard Index 3.1E-03 | |

 Table 2-24

 Indoor Worker, Exposure to Indoor Air: CTE Summary - Sandblast Area AOPC

| Definition | Variable | Valu | e | | Risk and Haza | ard Equations | | | | |
|---|----------------------------|--------------------------------------|---------------------|-------------------------|--------------------|--|-------------------------|---------------------------------------|---|----------------|
| Attenuation Factor | α | chemical-specifi | c dimensionless | | Cancer Risk: | | | | | |
| Averaging Time, carcinogens | AT _c | 7 | 0 yrs | | L | $Risk = IUR \times \left[C_{ino}\right]$ | 1 | $(ED_w \times EF$ | $W_w \times ET_w$ | |
| Averaging Time, noncarcinogens, worker | AT _{nc,w} | | 6 yrs | | 1 | C_{ind} | loor air ^ | $AT_c \times 365 \frac{da}{da}$ | $\frac{y}{x} \times 24 \frac{hour}{1}$ | |
| COPC Concentration in indoor air | Cindoor air | chemical-specifi | c μg/m ³ | | Noncancer Ha | zard: | | yed | ir day /] | |
| COPC Concentration in soil gas | C _{soil gas} | chemical-specifi | c μg/m ³ | | | 1 [| | (ED | \times EF., \times ET., | \1 |
| Exposure Duration, worker | ED_{w} | | 6 yrs | | H | $Hazard = \frac{1}{RfC} \times \left[$ | C _{indoor air} | $\times \left(\frac{1}{1-w} \right)$ | $\frac{dav}{dav}$, hou | \overline{r} |
| Exposure Frequency, worker | EF_{w} | 25 | 0 days/yr | | Where: | ny o | | $\left(AT_{nc,w} \times 3\right)$ | $65 \frac{day}{year} \times 24 \frac{day}{day}$ | <u>,</u> |
| Exposure Time, worker | ET_{w} | | 8 hours/day | | | L | | X | | 1 |
| Inhalation Reference Concentration Inhalation Unit-Risk Factor | RfC IUR | chemical-specifi chemical-specifi | | | | $C_{indoor\ air} = C_{soil}$ | _{gas} ×α | | | |
| | | | | | | Cancer | | | Hazard | |
| | | C _{soil gas} | α | C _{indoor air} | IUR | Risk | % of | RfC | Quotient | % of |
| Analyte | | $(\mu g/m^3)$ | (dimensionless) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| Current Scenario Equipment Building: | Volatile Or | ganic Compound | ds (VOCs) | | | | | | | |
| | | | | | Pathway Sum | Cancer Risk as: 6.4E-09 | | | Hazard Index 8.1E-03 | |
| Notes | | | | | | | | | | |

EPCs from Table 1-6. Toxicity values from Table 2-8. Exposure Factors from Table 2-7a. **Bolded** values indicate a risk estimate > 1E-06 or a hazard estimate >1.

| AOPC | Medium | EPC (mg/kg) | Receptor | Probability ^(a) | Probability Exceeds 5% ^(b) |
|---------------|---------------------------------------|----------------|-------------------------|----------------------------|--|
| | | | Occupational | 0.013% | No |
| | Unsieved soil (0-1 ft bgs) | 211 | Fetus of Adult Resident | 0.034% | No |
| Landfill | (0 1 1 0 2 3) | | Child Resident | 0.36% | No |
| | Unsieved soil (0-3 ft bgs) | 342 | Occupational | 0.032% | No |
| | | | Occupational | 0.025% | No |
| | Soil - sieved <250 um (0-1 ft bgs) | 300 | Fetus of Adult Resident | 0.075% | No |
| Sandblast | (0 1 10 0 50) | | Child Resident | 1.52% | No |
| | Soil - sieved <250 um (0-3 ft bgs) | 202 | Occupational | 0.013% | No |
| Pistol Range | Unsieved soil | 208 | Fetus of Adult Resident | 0.033% | No |
| r istor Kange | Unsieved som | 208 | Child Resident | 0.34% | No |
| Distol Dongo | Unsieved sediment | 27.1 | Fetus of Adult Resident | 0.0033% | No |
| Pistol Range | Unsieved sediment | 27.1 | Child Resident | < 0.075% | No |
| Dulh Slope | Unsieved soil | 222 | Fetus of Adult Resident | 0.038% | No |
| Bulb Slope | Unsieved som | 222 | Child Resident | 0.45% | No |

Table 2-25Summary of Lead Models Results

Model details are available in Appendix B.

(a) Probability that the receptor would have a PbB exceeding the 10 μ g/dL level of concern.

(b) USEPA's target is to limit the risk to a typical child or fetus of an adult exposed at a site to no

more than a 5% chance of exceeding the 10 µg/dL PbB of concern (USEPA 1994, 2003).

PbB was also evaluated for probability of exceeding the CDC recommended level of concern of 5 ug/dl, and all results were less than 5% (CDC 2012, USEPA 2010).

% = percent

µg/dL = micrograms of lead per deciliter of blood AOPC = area of potential concern bgs = below ground surface EPC = exposure point concentration ft = foot mg/kg = milligrams per kilogram PbB = blood lead level um = micrometers

Sources

CDC. 2012. Low Level Lead Exposure Harms Children: A Renewed Call for Primary Prevention. January.

USEPA. 1994. Revised Interim Soil Lead Guidance for CERCLA Sites and RCRA Corrective Action Facilities. Washington, DC: Office of Emergency and Remedial Response. July. OSWER Directive #9355.4-12. July.

USEPA. 2003. Recommendations of the Technical Review Workgroup for Lead for an Approach to Assessing Risks Associated with Adult Exposures to Lead in Soil [EPA-540-R-03-001, OSWER Dir #9285.7-54. January.

USEPA. 2010. Integrated Exposure Uptake Biokinetic Model for Lead in Children, IEUBK win v1.1 Build 11, Software Program. Developed by U.S. EPA. February.

| | | | Reasonable 1 | Maximum Expo | osure (RME) | Central Tendency Exposure (CTE) | | |
|-------------------|---|------------------------------|---------------------|--------------------|--------------------|---------------------------------|--------------------|--------------------|
| AOPC | Receptor (a) | Exposure Media | Cancer Risk ELCR | Adult Noncancer | Child Noncancer | Cancer Risk ELCR | Adult Noncancer | Child Noncancer |
| Landfill | Outdoor Maintenance Worker | Surface Soil (0-3 ft bgs) | 6E-05 | 0.020 | NA | 6E-06 | 0.0088 | NA |
| | Construction Worker | Deep Soil (0-10 ft bgs) | 5E-06 | 1.2 | NA | 7E-07 | 0.018 | NA |
| | Excavation / Trench Worker | Groundwater | 5E-08 | 0.036 | NA | | | NA |
| | Hypothetical Fishing Platform User | Surface Soil (0-3 ft bgs) | 1E-03 | 0.49 | 1.4 | 2E-04 | 0.14 | 0.31 |
| | Nursing Infant: Hypothetical Fishing Platform User | Surface Soil (0-3 ft bgs) | 1E-06 | NA | 1.5 | 1E-07 | NA | 0.26 |
| Sandblast Area | Outdoor Maintenance Worker | Surface Soil (0-3 ft bgs) | 1E-05 | 0.068 | NA | 1E-06 | 0.0081 | NA |
| | Construction Worker | Deep Soil (0-10 ft bgs) | 2E-06 | 0.97 | NA | 3E-07 | 0.014 | NA |
| | Excavation / Trench Worker | Groundwater | 1E-07 | 0.73 | NA | | | NA |
| | Future: Indoor Worker | Indoor Air | 5E-07 | 0.13 | NA | 1E-07 | 0.13 | NA |
| | Current: BISB Indoor Worker | Indoor Air | 1E-08 | 0.0031 | NA | 3E-09 | 0.0031 | NA |
| | Current: Equipment Building Indoor Worke | Indoor Air | 3E-08 | 0.0081 | NA | 6E-09 | 0.0081 | NA |
| | Hypothetical Fishing Platform User | Surface Soil (0-3 ft bgs) | 3E-04 | 0.97 | 1.9 | 5E-05 | 0.47 | 0.65 |
| | Nursing Infant: Hypothetical Fishing Platform User | Surface Soil (0-3 ft bgs) | 1E-06 | NA | 1.5 | 1E-07 | NA | 0.26 |

 Table 2-26

 Summary of Upland OU Cancer Risk and Noncancer Health Hazard

Table 2-26Summary of Upland OU Cancer Risk and Noncancer Health Hazard

Notes

benzofluoranthenes, total = uses benzo(b)fluoranthene for surrogate toxicity values and chemical properties.

(a) The Nursing Infant pathway screened out for the occupational receptors. See Section 2.1.

Cancer risks are presented for time-integrated child + adult exposure.

Noncancer health hazards are presented separately for exposure during adulthood and exposure during childhood. Nursing infant risks and hazards are presented as a function of mother's risks, assuming a 1-year exposure duration **Bold** font indicates greater than the ODEQ risk threshold for cancer risk $(1x10^{-6})$ and/or noncancer HI (1).

--- = not available / not calculated AOPC = area of potential concern BISB = Bradford Island Service Building (Located near Sandblast AOPC) CTE = central tendency exposure ELCR = excess lifetime cancer risk ft bgs = foot/feet below ground surface HI = hazard index NA = not applicable OU = operable unit RME = reasonable maximum exposure

| AOPC | Receptor | Chemical of Concern (COC) | Calculated Soil RBC (mg/kg) | Reference Area UPL (mg/kg) |
|----------------|---------------------------------------|--------------------------------------|-----------------------------------|-------------------------------|
| Landfill | Outdoor Maintenance Worker | cPAHs ^(b) | 0.3 ^(a) | 0.052 |
| | Construction Worker | cPAHs ^(b) | 2.0 | 0.052 |
| | Hypothetical Fishing Platform User | Arsenic | 0.68 | 5.5 |
| | | Total PCBs | 0.24 | |
| | | cPAHs ^(b) | 0.015 | 0.052 |
| Sandblast Area | Outdoor Maintenance Worker | cPAHs ^(b) | 0.3 ^(a) | 0.037 ^(a) |
| | Hypothetical Fishing Platform User | Arsenic | 0.68 | 5.5 |
| | i lationin Osci | Total PCBs | 0.24 | |
| | | Bis(2-ethylhexyl)phthalate (DEHP) | 37 | |
| | | cPAHs ^(b) | 0.015 | 0.052 |

 Table 2-27

 Summary of Upland OU Human Health COCs for Further Evaluation in the FS

(a) Based on the toxicity value for benzo(a)pyrene and expressed as benzo(a)pyrene equivalents (BaPeq).

(b) RBC for cPAH is based on RME and target cancer risk of 1E-06. If ODEQ's acceptable risk level of 1E-05 was used, the RBCs would increase to: Outdoor Maintenance Worker – 3 mg/kg, Construction Worker – 20 mg/kg, and Hypothetical Fishing Platform User – 0.15 mg/kg.

In Section 2.5, arsenic was tentatively identified as a COC for occupational receptors. The calculated risk-based value was 2 mg/kg, which is lower than the Reference Area UPL of 5.4 mg/kg. Arsenic was included in the exceedance Figure 2-3 and Figure 2-4 based on concentrations exceeding the Reference UPL value of 5.4 mg/kg.

Table 2-27 Summary of Upland OU Human Health COCs for Further Evaluation in the FS

-- = not available

AOPC = area of potential concern

COC = constituent of concern

cPAH = carcinogenic polycyclic aromatic hydrocarbon

DEHP = Bis(2-ethylhexyl)phthalate

mg/kg = milligram per kilogram

PCB = polychlorinated biphenyl

RBC = risk-based concentration

RME = reasonable maximum exposure

UPL = upper prediction limit

 Table 3-1

 Exposure Assumptions and Dose Equations for the Canada Goose (*Branta canadensis*)

Equation:

 $D = \frac{AUF x \left[(C_{F1} x P_{F1} x IR_F) + (C_S x IR_S) + (C_W x IR_W) \right]}{BW}$

Where:

 $D = Chemical \ dose \ (mg/kg-bw/day)$

 $C_{Fl} = C_S x$ plant BAF, where C_S is the chemical concentration

 $C_W = C_S/Kd$ (inorganics) and $C_S/(foc * K_{oc})$ (organics), foc assumed 0.01

 $IR_F = 0.638 \text{ x BW}$ in grams ^0.685 *0.001 ^a

 $IR_s = 0.082 \text{ x } IR_F$

 $IR_W = 0.059 \text{ x BW}^0.67$

| Parameter | Definition | Value | Units | Source ^b |
|-----------------|--|-------------------|------------|---|
| C _{F1} | Chemical concentration in food item 1 (plants) | food chain model | mg/kg dry | Food concentrations were estimated on a dry weight basis using uptake models using concentrations at the AOPC or by multiplying concentrations in soil at the site by BAFs for plants. |
| Cs | Surface: 95% UCL in surface soil (0-1 feet) Shallow: 95% UCL in shallow soil (0-3 feet) | chemical-specific | mg/kg dry | Upland OU analytical data. |
| C _W | Concentration in upland water | chemical-specific | mg/L | Upland OU calculated puddled water (via equilibrium partitioning calculation). |
| HR | Home Range | 2,430 | acres | USEPA 1993 |
| AUF | Area Use Factor - All AOPCs | 1 | unitless | Assumed present 100% of the time. Bradford Island is managed as Canada goose habitat. |
| P _{F1} | Proportion of food item 1 - plants | 1 | unitless | Diet assumed 100% plants. |
| IR _F | Ingestion Rate - plants | 0.15 | kg dry/day | Nagy 2001 for all birds. |
| IRs | Incidental Ingestion Rate - soil | 0.013 | kg dry/day | Based on 8.2% of total food ingestion rate (Beyer et al. 1994). |
| IR _w | Ingestion Rate - water | 0.12 | L/day | USEPA 1993 |
| BW | Body weight | 3.0 | kg | Since no breeding female available, average of adult female data used (USEPA 1993). |

Table 3-1

Exposure Assumptions and Dose Equations for the Canada Goose (Branta canadensis)

Notes:

a) Allometric relationships with gram body weight.b) See Table 3-3 of Technical Memorandum, Update to Risk Assessment Work Plan, Upland OU (URS 2014).

% = percent AOPC = area of potential concern BAF = bioaccumulation factor foc = fraction of organic carbon Kd = soil-water partition coefficient (cm^3/g) kg = kilograms kg dry/day = kilograms per day in dry weight K_{oc} = organic carbon partition coefficient (L/kg) L/day = liters per day mg/kg-day = milligrams per kilogram per day mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter OU = operational unit UCL = upper confidence limit USEPA = U.S. Environmental Protection Agency

References:

Beyer, W.N., E.E. Connor, and S. Gerould. 1994. Estimates of soil ingestion by wildlife. J. Wildl. Manage . 58:375-382.

Nagy, K.A. 2001. Food requirements of wild animals: predictive equations for free-living mammals, reptiles, and birds. Nutrition Abstracts and Reviews, Series B 71, 21R-31R.

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 Table 3-2

 Exposure Assumptions and Dose Equations for the American Robin (*Turdus migratorius*)

Equation:

 $D = \underline{AUF x [(C_{F1} x P_{F1} x IR_F) + (C_S x IR_S) + (C_W x IR_W)]}_{BW}$

Where:

$$\begin{split} D &= \text{Chemical dose (mg/kg-bw/day)} \\ C_{F1} &= C_S \text{ x invertebrate BAF, where } C_S \text{ is the chemical concentration} \\ C_W &= C_S/\text{Kd (inorganics) and } C_S/(\text{foc } * \text{K}_{oc}) \text{ (organics), foc assumed 0.01} \\ IR_F &= 0.638 \text{ x BW in grams } ^{0.685 * 0.001 ^{a}} \\ IR_S &= 0.104 \text{ x IR}_F \\ IR_W &= 0.059 \text{ x BW}^{0.67} \end{split}$$

| Parameter | Definition | Value | Units | Source ^b |
|-----------------|---|-------------------|------------|--|
| C _{F1} | Chemical concentration in food item 1 (invertebrates) | food chain model | mg/kg dry | Food concentrations were estimated on a dry weight basis |
| | | | | using uptake models using concentrations at the site or by |
| | | | | multiplying concentrations in soil at the site by BAFs for |
| | | | | soil invertebrates. |
| Cs | Surface: 95% UCL in surface soil (0-1 feet) | chemical-specific | mg/kg dry | Upland OU analytical data. |
| | Shallow: 95% UCL in shallow soil (0-3 feet) | | | |
| Cw | Concentration in upland water | chemical-specific | mg/L | Upland OU calculated puddled water (via equilibrium |
| | | | | partitioning calculation). |
| HR | Home Range | 0.37 | acres | USEPA 1993 |
| AUF | Landfill AOPC (1.36 acres) | 1 | unitless | Area of site divided by area of HR $(1 = 100\%$ site use). |
| | Sandblast Area AOPC (3.10 acres) | 1 | unitless | Area of site divided by area of HR $(1 = 100\%$ site use). |
| | Pistol Range AOPC (0.26 acres) | 0.26 | unitless | Area of site divided by area of HR. |
| | Bulb Slope AOPC (0.054 acres) | 0.15 | unitless | Area of site divided by area of HR. |
| P _{F1} | Proportion of food item 1 - invertebrates | 1 | unitless | Diet assumed 100% invertebrates. |
| IR _F | Ingestion Rate - invertebrates | 0.013 | kg dry/day | Nagy 2001 for all birds. |
| IR _s | Incidental Ingestion Rate - soil | 0.0013 | kg dry/day | Based on 10.4% of total food ingestion for woodcock |
| | | | | (Beyer et al. 1994). |
| IR _W | Ingestion Rate - water | 0.011 | L/day | USEPA 1993 |
| BW | Body weight | 0.0773 | kg | USEPA 1993 |

Table 3-2

Exposure Assumptions and Dose Equations for the American Robin (Turdus migratorius)

Notes:

a) Allometric relationships with gram body weight.

b) See Table 3-3 of Technical Memorandum, Update to Risk Assessment Work Plan, Upland OU (URS 2014).

% = percentAOPC = area of potential concern AUF = area use factorBAF = bioaccumulation factor foc = fraction of organic carbon Kd = soil-water partition coefficient (cm^3/g) kg = kilograms kg dry/day = kilograms per day in dry weight K_{oc} = organic carbon partition coefficient (L/kg) L/day = liters per daymg/kg-bw/day = milligrams per kilogram body weight per day mg/kg dry = milligrams per kilogram in dry weight mg/L = milligrams per liter OU = operational unit UCL = upper confidence limit USEPA = U.S. Environmental Protection Agency

References:

Beyer, W.N., E.E. Connor, and S. Gerould. 1994. Estimates of soil ingestion by wildlife. J. Wildl. Manage. 58:375-382.

Nagy, K.A. 2001. Food requirements of wild animals: predictive equations for free-living mammals, reptiles, and birds. Nutrition Abstracts and Reviews, Series B 71, 21R-31R.

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U.S. Environmental Protection Agency (USEPA). 1993. "Wildlife Exposure Factors Handbook." December. 1993.

 Table 3-3

 Exposure Assumptions and Dose Equations for the American Kestrel (Falco sparverius)

Equation:

 $D = \underline{AUF x [(C_{F1} x P_{F1} x IR_F) + (C_S x IR_S) + (C_W x IR_W)]}_{BW}$

Where:

D = Chemical dose (mg/kg-bw/day)

 $C_{F1} = C_S x$ small mammal BAF, where C_S is the chemical concentration

 $IR_F = 0.638 \text{ x BW}$ in grams ^0.685 *0.001 ^a

 $IR_s = 0.02 \text{ x } IR_F$

 $IR_W = 0.059 \text{ x BW}^{0.67}$

| Parameter | Definition | Value | Units | Source ^b |
|-----------------|---|-------------------|-------------|--|
| C _{F1} | Chemical concentration in food | food chain model | mg/kg dry | Food concentrations were estimated on a dry weight basis using uptake models using concentrations at the site or by |
| | | | | multiplying concentrations in soil at the site by BAFs for small mammals. |
| Cs | Surface: 95% UCL in surface soil (0-1 feet) | chemical-specific | mg/kg dry | Upland OU analytical data. |
| | Shallow: 95% UCL in shallow soil (0-3 feet) | | | |
| Cw | Concentration in upland water | chemical-specific | mg/L | River OU analytical data. |
| HR | Home Range | 270 | acres | Cal/ECOTOX 2002 |
| AUF | Landfill AOPC (1.36 acres) | 0.0050 | unitless | Area of site divided by area of HR. |
| | Sandblast Area AOPC (3.10 acres) | 0.0115 | unitless | Area of site divided by area of HR $(1 = 100\%$ site use). |
| | Pistol Range AOPC (0.26 acres) | 0.00020 | unitless | Area of site divided by area of HR. |
| | Bulb Slope AOPC (0.054 acres) | 0.00096 | unitless | Area of site divided by area of HR. |
| | Combined AOPCs (4.77 acres) | 0.0177 | unitless | Area of site divided by area of HR. |
| P _{F1} | Proportion of food item 1 - small mammal | 1 | unitless | 100% small mammal diet. |
| IR _F | Food ingestion rate - small mammal | 0.017 | kg dry /day | Nagy 2001 for all birds. |
| IR _S | Incidental Ingestion Rate - soil | 0.00033 | kg dry/day | Default 2% of total food ingestion rate (Beyer et al. 1994). |
| IR _W | Ingestion Rate - water | 0.014 | L/day | USEPA 1993 |
| BW | Body weight | 0.116 | kg | USEPA 1993 |

Table 3-3

Exposure Assumptions and Dose Equations for the American Kestrel (Falco sparverius)

Notes:

a) Allometric relationships with gram body weight.b) See Table 3-3 of Technical Memorandum, Update to Risk Assessment Work Plan, Upland OU (URS 2014).

% = percent AOPC = area of potential concern AUF = area use factor BAF = bioaccumulation factor kg = kilograms kg dry/day = kilograms per day in dry weight L/day = liters per day mg/kg-bw/day = milligrams per kilogram body weight per day mg/kg dry = milligrams per kilogram in dry weight mg/L = milligrams per liter OU = operational unit UCL = upper confidence limit USEPA = U.S. Environmental Protection Agency

References:

Beyer, W.N., E.E. Connor, and S. Gerould. 1994. Estimates of soil ingestion by wildlife. J. Wildl. Manage . 58:375-382.

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 Table 3-4

 Exposure Assumptions and Dose Equations for the Vagrant Shrew (Sorex vagrans)

Equation:

 $D = \underline{AUF x [(C_{F1} x P_{F1} x IR_F) + (C_S x IR_S) + (C_W x IR_W)]}_{BW}$

Where:

$$\begin{split} D &= \text{Chemical dose (mg/kg-bw/day)} \\ C_{F1} &= C_S \text{ x invertebrate BAF, where } C_S \text{ is the chemical concentration} \\ C_W &= C_S/\text{Kd (inorganics) and } C_S/(\text{foc * } K_{oc}) \text{ (organics), foc assumed } 0.01 \\ \text{IR}_F &= 0.323 \text{ x BW in grams } ^0.744 * 0.001 ^a \\ \text{IR}_S &= 0.04 \text{ x IR}_F \\ \text{IR}_W &= 0.099 \text{ x BW} ^0.9 \end{split}$$

| Parameter | Definition | Value | Units | Source ^b |
|-----------------|--|-------------------|-------------|---|
| C _{F1} | Chemical concentration in food | food chain model | mg/kg dry | Food concentrations were estimated on a dry weight basis using uptake models using concentrations at the site or by multiplying concentrations in soil at the site by BAFs for soil invertebrates. |
| Cs | Surface: 95% UCL in surface soil (0-1 feet) Shallow: 95% UCL in shallow soil (0-3 feet) | chemical specific | mg/kg dry | Upland OU analytical data. |
| C _W | Concentration in upland water | chemical-specific | mg/L | Upland OU calculated puddled water (via equilibrium partitioning calculation). |
| HR | Home Range | 0.26 | acres | Zeiner et al. 1990 |
| AUF | Landfill AOPC (1.36 acres) | 1 | unitless | Area of site divided by area of HR $(1 = 100\%$ site use). |
| | Sandblast Area AOPC (3.10 acres) | 1 | unitless | Area of site divided by area of HR $(1 = 100\%$ site use) |
| | Pistol Range AOPC (0.26 acres) | 1 | unitless | Area of site divided by area of HR. |
| | Bulb Slope AOPC (0.054 acres) | 0.21 | unitless | Area of site divided by area of HR. |
| P _{F1} | Proportion of food item 1 - invertebrates | 1 | unitless | Diet assumed 100% invertebrates. |
| IR _F | Food ingestion rate - invertebrates | 0.0014 | kg dry /day | Nagy 2001 for all mammals. |
| IR _s | Incidental Ingestion Rate - soil | 0.000055 | kg dry /day | Based on 4% of total food ingestion rate (double white- footed mouse; Beyer et al. 1994). |
| IR _W | Ingestion Rate - water | 0.0011 | L/day | USEPA 1993 |
| BW | Body weight | 0.007 | kg | USEPA 1993 |

Table 3-4

Exposure Assumptions and Dose Equations for the Vagrant Shrew (Sorex vagrans)

Notes:

a) Allometric relationships with gram body weight.

b) See Table 3-3 of Technical Memorandum, Update to Risk Assessment Work Plan, Upland OU (URS 2014).

% = percent AOPC = area of potential concern AUF = area use factor BAF = bioaccumulation factor kg = kilograms kg dry/day = kilograms per day in dry weight L/day = liters per day mg/kg-bw/day = milligrams per kilogram body weight per day mg/kg dry = milligrams per kilogram in dry weight mg/L = milligrams per liter OU = operational unit UCL = upper confidence limit USEPA = U.S. Environmental Protection Agency

References:

Beyer, W.N., E.E. Connor, and S. Gerould. 1994. Estimates of soil ingestion by wildlife. *J. Wildl. Manage*. 58:375-382.
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 Table 3-5

 Exposure Assumptions and Dose Equations For the American Mink (Neovison vison)

Equation:

 $D = \underline{AUF x [(C_{F1} x P_{F1} x IR_F) + (C_S x IR_S) + (C_W x IR_W)]}{BW}$

Where:

D = Chemical dose (mg/kg-bw/day)

 $C_{F1} = C_S x$ small mammal BAF, where C_S is the chemical concentration

 $IR_F = 0.323 \text{ x BW}$ in grams ^0.744 *0.001 ^a

 $IR_{S} = 0.094 \text{ x } IR_{F}$

 $IR_W = 0.099 \text{ x BW}^0.9$

| Parameter | Definition | Value | Units | Source ^b |
|-----------------|---|-------------------|-------------|--|
| C _{F1} | Chemical concentration in food | food chain model | mg/kg dry | Food concentrations were estimated on a dry weight basis |
| | | | | using uptake models using concentrations at the site or by |
| | | | | multiplying concentrations in soil at the site by BAFs for |
| | | | | small mammals. |
| Cs | Surface: 95% UCL in surface soil (0-1 feet) | chemical specific | mg/kg dry | Analytical data. |
| | Shallow: 95% UCL in shallow soil (0-3 feet) | | | |
| C _W | Concentration in upland water | chemical-specific | mg/L | River OU analytical data. |
| | | | | |
| HR | Home Range | 19 | acres | USEPA 1993 |
| AUF | Landfill AOPC (1.36 acres) | 0.072 | unitless | Area of site divided by area of HR. |
| | Sandblast Area AOPC (3.10 acres) | 0.163 | unitless | Area of site divided by area of HR. |
| | Pistol Range AOPC (0.26 acres) | 0.014 | unitless | Area of site divided by area of HR. |
| | Bulb Slope AOPC (0.054 acres) | 0.0028 | unitless | Area of site divided by area of HR. |
| | Combined AOPCs (4.77 acres) | 0.251 | unitless | Area of site divided by area of HR. |
| P _{F1} | Proportion of food item 1 - small mammal | 0.15 | unitless | 15% small mammal diet. |
| IR _F | Food ingestion rate - small mammal | 0.054 | kg dry /day | Nagy 2001 for all mammals. |
| IR _s | Incidental Ingestion Rate - soil | 0.0051 | kg dry /day | Based on 9.4% of total food ingestion rate for raccoon |
| | | | | (Beyer et al. 1994). |
| IR _W | Ingestion Rate - water | 0.097 | L/day | USEPA 1993 |
| BW | Body weight | 0.974 | kg | USEPA 1993 |

Table 3-5

Exposure Assumptions and Dose Equations For the American Mink (Neovison vison)

Notes:

a) Allometric relationships with gram body weight.

b) See Table 3-3 of Technical Memorandum, Update to Risk Assessment Work Plan, Upland OU (URS 2014).

% = percent AOPC = area of potential concern AUF = area use factor BAF = bioaccumulation factor kg = kilograms kg dry/day = kilograms per day in dry weight L/day = liters per day mg/kg-bw/day = milligrams per kilogram body weight per day mg/kg dry = milligrams per kilogram in dry weight mg/L = milligrams per liter OU = operational unit UCL = upper confidence limit USEPA = U.S. Environmental Protection Agency

References:

Beyer, W.N., E. E. Connor, and S. Gerould. 1994. Estimates of soil ingestion by wildlife. J. Wildl. Manage . 58:375-382.

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 Table 3-6

 Statistical Summary for Random Forebay Surface Water Samples

| Medium | Analyte Group | Preparation Fraction | Analyte | Basis | Number of Samples | Number of Detections | Mean of Detects (ug/L) | Median of Detects (ug/L) | Maximum Detected Value (ug/L) | EPC (mg/L) |
|---------------|---------------|-------------------------|----------------------|-------|----------------------|-------------------------|---------------------------|--------------------------------|-------------------------------------|---------------|
| Surface Water | Metals | Dissolved | Cadmium | Wet | 5 | 2 | 0.00900 | 0.00900 | 0.0100 | 1.00E-05 |
| Surface Water | Metals | Dissolved | Copper | Wet | 5 | 5 | 0.470 | 0.460 | 0.520 | 5.20E-04 |
| Surface Water | Metals | Dissolved | Lead | Wet | 5 | 3 | 0.0240 | 0.0220 | 0.0360 | 3.60E-05 |
| Surface Water | SVOCs | C+F | Benzo(a)anthracene | Wet | 5 | 3 | 0.0536 | - | 0.0600 | 6.00E-07 |
| Surface Water | SVOCs | C+F | Benzo(b)fluoranthene | Wet | 5 | 2 | 0.0902 | - | 0.0917 | 9.17E-07 |
| Surface Water | SVOCs | C+F | Chrysene | Wet | 5 | 5 | 0.139 | 0.136 | 0.171 | 1.71E-06 |
| Surface Water | SVOCs | C+F | Fluoranthene | Wet | 5 | 5 | 0.706 | 0.674 | 0.784 | 7.84E-06 |

Notes

C+F = column plus filter

EPC = exposure point concentration

mg/L = milligrams per liter

SVOC = semi-volatile organic carbon

ug/L = micrograms per liter

Wet = Wet Weight

 Table 3-7a

 Estimated Surface Water Concentrations – Landfill AOPC

| | 95% UCL in | | | Water |
|--|--------------|---|---------------------|---------------------|
| | Surface Soil | | | Concentration |
| | (0-1 feet) | Kd | K _{oc} | (mg/L) ^b |
| Analyte | (mg/kg dw) | $(\mathrm{cm}^3/\mathrm{g})^{\mathrm{a}}$ | (L/kg) ^a | |
| Inorganics | | | | |
| Antimony | 1.36 | 45 | | 0.0302 |
| Chromium | 242 | 1,800,000 | | 1.34E-04 |
| Copper | 170 | 35 | | 4.86 |
| Lead | 332 | 900 | | 0.37 |
| Mercury | 1.57 | 52 | | 0.0302 |
| Nickel | 175 | 65 | | 2.69 |
| Butyltins | | | | |
| Tributyltin | 0.165 | | 12,100 | 0.00136 |
| Pesticides | | | | |
| Chlordane (technical) | 1.56 | | 33,800 | 0.00462 |
| Heptachlor | 0.00283 | | 41,300 | 6.85E-06 |
| HPAHs | | | - | |
| Benzo(a)anthracene | 8.34 | | 177,000 | 0.00471 |
| Benzo(a)pyrene | 8.70 | | 587,000 | 0.00148 |
| Benzo(b)fluoranthene | 16.0 | | 599,000 | 0.00267 |
| Benzo(g,h,i)perylene | 4.59 | | 1,950,000 | 2.35E-04 |
| Benzo(k)fluoranthene | 30.2 | | 587,000 | 0.00514 |
| Benzofluoranthenes, Total ^c | 14.7 | | 587,000 | 0.00250 |
| Chrysene | 8.40 | | 181,000 | 0.00464 |
| Dibenz(a,h)anthracene | 4.23 | | 1,910,000 | 2.21E-04 |
| Fluoranthene | 14.5 | | 55,500 | 0.0261 |
| Indeno(1,2,3-cd)pyrene | 5.04 | | 3,470,000 | 1.45E-04 |
| Pyrene | 12.3 | | 54,300 | 0.0227 |
| Total HPAHs ^d | 90.2 | | 437,607 | 0.0206 |
| SVOCs | | | | |
| Bis(2-ethylhexyl) Phthalate | 9.68 | | 120,000 | 0.00807 |
| Butyl Benzyl Phthalate | 0.0366 | | 7,160 | 5.11E-04 |
| Carbazole | 0.751 | | 9,160 | 0.00820 |
| Dibenzofuran | 0.152 | | 9,160 | 0.00166 |
| Di-n-butyl Phthalate | 0.232 | | 1,160 | 0.0200 |
| VOCs | | | | |
| 1,2,4-Trimethylbenzene | 3.03 | | 614 | 0.493 |

Notes:

a) ORNL RAIS Database Chemical Parameters.

b) COC concentrations in water, used for the goose, robin, and shrew dose calculations, were calculated using equilibrium partitioning calculations:

Concentration in water (inorganics) = Concentration soil/ Kd

Concentration in water (organics) = Concentration soil/ (foc $* K_{oc}$), where foc assumed = 0.01

c) Benzo(k)fluoranthene Koc used as surrogate.

d) Geometric mean Koc of individual HPAHs used for Total HPAH.

-- = no data

% = percent

AOPC = area of potential concern

 $cm^3 = cubic centimeter$

COC = chemical of concern

dw = dry weight

g = gram

HPAHs = high molecular weight polycyclic aromatic hydrocarbons

Table 3-7a Estimated Surface Water Concentrations – Landfill AOPC

foc = fraction of organic carbon

Kd = soil-water partition coefficient (cm^3/g)

kg = kilogram

 K_{oc} = organic carbon partition coefficient (L/kg)

L = liter

mg = milligrams

SVOC = semivolatile organic compound

UCL = upper confidence limit

VOC = volatile organic compound

Table 3-7b Estimated Surface Water Concentrations – Sandblast Area AOPC

| | 95% UCL in | | | |
|--|---------------------|---|---------------------|---------------|
| | Surface Soil | | | Water |
| | (0-1 feet) | Kd | K _{oc} | Concentration |
| Analyte | (mg/kg dw) | $(\mathrm{cm}^3/\mathrm{g})^{\mathrm{a}}$ | (L/kg) ^a | $(mg/L)^{b}$ |
| Inorganics | | | | - |
| Antimony | 2.94 | 45 | | 0.0653 |
| Cadmium | 3.20 | 75 | | 0.0427 |
| Chromium | 720 | 1,800,000 | | 4.00E-04 |
| Lead | 465 | 900 | | 0.517 |
| Mercury | 0.113 | 52 | | 0.00217 |
| Nickel | 353 | 65 | | 5.43 |
| Butyltins | | | | |
| Tributyltin | 0.481 | | 12,100 | 0.00398 |
| Pesticides | | | • | |
| BHC (delta) | 0.00154 | | 2,810 | 5.48E-05 |
| BHC (gamma) Lindane | 0.00968 | | 2,810 | 3.44E-04 |
| Chlordane (alpha) | 0.000815 | | 67,500 | 1.21E-06 |
| Chlordane (gamma) | 0.0303 | | 67,500 | 4.49E-05 |
| Endosulfan I | 0.00269 | | 6,760 | 3.98E-05 |
| Endosulfan II | 0.000495 | | 6,760 | 7.32E-06 |
| Endosulfan Sulfate | 0.000953 | | 9,850 | 9.68E-06 |
| Endrin Aldehyde | 0.00387 | | 3,270 | 1.18E-04 |
| Endrin Ketone | 0.00309 | | 9,720 | 3.18E-05 |
| Endrin | 0.00504 | | 20,100 | 2.51E-05 |
| Heptachlor | 0.000464 | | 41,300 | 1.12E-06 |
| Methoxychlor | 0.00105 | | 41,300 | 2.54E-06 |
| HPAHs | | | • | |
| Benzo(a)anthracene | 1.75 | | 177,000 | 9.89E-04 |
| Benzo(a)pyrene | 1.76 | | 587,000 | 3.00E-04 |
| Benzo(b)fluoranthene | 1.95 | | 599,000 | 3.26E-04 |
| Benzo(g,h,i)perylene | 1.07 | | 1,950,000 | 5.49E-05 |
| Benzo(k)fluoranthene | 0.676 | | 587,000 | 1.15E-04 |
| Benzofluoranthenes, Total ^c | 7.76 | | 587,000 | 0.00132 |
| Chrysene | 2.07 | | 181,000 | 0.00114 |
| Dibenz(a,h)anthracene | 0.244 | | 1,910,000 | 1.28E-05 |
| Fluoranthene | 4.64 | | 55,500 | 0.00836 |
| Indeno(1,2,3-cd)pyrene | 1.13 | | 3,470,000 | 3.26E-05 |
| Pyrene | 4.78 | | 54,300 | 0.00880 |
| Total HPAHs ^d | 18.6 | | 437,607 | 0.00425 |
| SVOCs | | | 7 | |
| Bis(2-ethylhexyl) Phthalate | 98.0 | | 120,000 | 0.0817 |
| Butyl Benzyl Phthalate | 0.0171 | | 7,160 | 2.39E-04 |
| Carbazole | 0.144 | | 9,160 | 0.00157 |
| Dibenzofuran | 0.0424 | | 9,160 | 4.63E-04 |
| Di-n-butyl Phthalate | 0.0804 | | 1,160 | 0.00693 |
| Di-n-octyl Phthalate | 0.0251 | | 141,000 | 1.78E-05 |

Table 3-7b Estimated Surface Water Concentrations – Sandblast Area AOPC

| Analyte | 95% UCL in Surface Soil (0-1 feet) (mg/kg dw) | Kd (cm ³ /g) ^a | K _{oc} (L/kg) ^a | Water Concentration (mg/L) ^b |
|------------------------|--|---|--|---|
| VOCs | | | | |
| 1,2,4-Trimethylbenzene | 0.000230 | | 614 | 3.75E-05 |
| 4-Isopropyltoluene | 0.00133 | | 1,120 | 1.19E-04 |
| n-Propylbenzene | 0.000150 | | 813 | 1.85E-05 |

Notes:

a) ORNL RAIS Database Chemical Parameters.

b) COC concentrations in water, used for the goose, robin, and shrew dose calculations, were calculated using equilibrium partitioning calculations:

Concentration in water (inorganics) = Concentration soil/ Kd

Concentration in water (organics) = Concentration soil/ (foc $* K_{oc}$), where foc assumed = 0.01

c) Benzo(k)fluoranthene Koc used as surrogate.

d) Geometric mean Koc of individual HPAHs used for Total HPAH.

-- = no data

% = percent

AOPC = area of potential concern

BHC = benzene hexachloride

 $cm^3 = cubic centimeter$

COC = chemical of concern

dw = dry weight

g = gram

HPAHs = high molecular weight polycyclic aromatic hydrocarbons

foc = fraction of organic carbon

Kd = soil-water partition coefficient (cm³/g)

kg = kilogram

 K_{oc} = organic carbon partition coefficient (L/kg)

L = liter

mg = milligrams

SVOC = semivolatile organic compound

UCL = upper confidence limit

VOC = volatile organic compound

Table 3-7c Estimated Surface Water Concentrations – Pistol Range AOPC

| CPEC Inorganics | 95% UCL in Surface Soil (0-1.5 feet) (mg/kg dw) | Kd (cm ³ /g) ^a | K _{oc} (L/kg) ^a | Water Concentration (kg/L) ^b |
|--------------------|--|---|--|---|
| Lead | 365 | 900 | | 0.406 |

Notes:

a) ORNL RAIS Database Chemical Parameters.

b) COC concentrations in water, used for the goose, robin, and shrew dose calculations, were

calculated using equilibrium partitioning calculations:

Concentration in water (inorganics) = Concentration soil/ Kd

-- = no data

% = percent

AOPC = area of potential concern

 $cm^3 = cubic centimeter$

COC = chemical of concern

CPEC = chemical of potential ecological concern

dw = dry weight

g = gram

Kd = soil-water partition coefficient (cm^3/g)

kg = kilogram

 K_{oc} = organic carbon partition coefficient (L/kg)

L = liter

mg = milligrams

UCL = upper confidence limit

Table 3-7d Estimated Surface Water Concentrations – Bulb Slope AOPC

| CPEC Inorganics | 95% UCL in Surface Soil (0-1 feet) (mg/kg dw) | Kd (cm ³ /g) ^a | K _{oc} (L/kg) ^a | Water Concentration (kg/L) ^b |
|---------------------------|--|---|--|---|
| Lead | 307 | 900 | | 0.341 |
| Mercury | 0.720 | 52 | | 0.0138 |

Notes:

a) ORNL RAIS Database Chemical Parameters.

b) COC concentrations in water, used for the goose, robin, and shrew dose calculations, were calculated using equilibrium partitioning calculations:

Concentration in water (inorganics) = Concentration soil/ Kd

-- = no data % = percent

AOPC = area of potential concern

 $cm^3 = cubic centimeter$

CPEC = chemical of potential ecological concern

dw = dry weight

g = gram

Kd = soil-water partition coefficient (cm^3/g)

kg = kilogram

 K_{oc} = organic carbon partition coefficient (L/kg)

L = liter

mg = milligrams

UCL = upper confidence limit

Table 3-8aSoil to Plant Tissue Bioaccumulation Factors - Landfill AOPC

| | 95% UCL in Surface Soil | 95% UCL in Shallow Soil | Surface Soil BAF Plant | Shallow Soil BAF Plant | | | |
|-----------------------------|----------------------------|----------------------------|---------------------------|---------------------------|--|--------------------|------------------------|
| | (0-1 ft bgs) | (0-3 ft bgs) | (kg dw tissue/ kg | (kg dw tissue/ | | Regression | Regression |
| CPEC | (mg/kg dw) | (mg/kg dw) | dw soil) ^a | kg dw soil) ^a | Source ^b | Slope ^c | Intercept ^c |
| Inorganics | | | | | • | • | |
| Antimony | 1.36 | 1.48 | 0.0387 | 0.0385 | EcoSSL regression (USEPA 2005) | 0.938 | -3.233 |
| Chromium | | | 0.0410 | 0.0410 | EcoSSL (USEPA 2005) | | |
| Copper | 170 | 191 | 0.0868 | 0.0809 | EcoSSL regression (USEPA 2005) | 0.394 | 0.668 |
| Lead | 332 | 511 | 0.0207 | 0.0171 | EcoSSL regression (USEPA 2005) | 0.561 | -1.328 |
| Mercury | | | 0.650 | 0.650 | Bechtel-Jacobs, 1998 median (ORNL) | | |
| Nickel | 175 | 472 | 0.0295 | 0.0229 | EcoSSL regression (USEPA 2005) | 0.748 | -2.223 |
| Butyltins | | | | | | | |
| Tributyltin | | | 2.04 | 2.04 | McKone, T. E. 1994 (ORNL RAIS) | | |
| Pesticides | | | | | | | |
| Chlordane (technical) | | | 0.00950 | 0.00950 | McKone 1994 (ORNL RAIS) | | |
| Heptachlor | | | 0.0112 | 0.0112 | McKone 1994 (ORNL RAIS) | | |
| HPAHs | | • | | | • • | | • |
| Benzo(a)anthracene | 8.34 | 10.0 | 0.0282 | 0.0262 | EcoSSL regression (USEPA 2005) | 0.5944 | -2.7078 |
| Benzo(a)pyrene | 8.70 | 11.2 | 0.121 | 0.120 | EcoSSL regression (USEPA 2005) | 0.9750 | -2.0615 |
| Benzo(b)fluoranthene | | | 0.310 | 0.310 | EcoSSL (USEPA 2005) | | |
| Benzo(g,h,i)perylene | 4.59 | 5.82 | 0.521 | 0.544 | EcoSSL regression (USEPA 2005) | 1.1829 | -0.9313 |
| Benzo(k)fluoranthene | 30.2 | 26.3 | 0.0716 | 0.0730 | EcoSSL regression (USEPA 2005) | 0.8595 | -2.1579 |
| Benzofluoranthenes, Total | | | 0.31 | 0.31 | Benzofluoranthene surrogate ^d | | |
| Chrysene | 8.40 | 11.0 | 0.0281 | 0.0252 | EcoSSL regression (USEPA 2005) | 0.5944 | -2.7078 |
| Dibenzo(a,h)anthracene | | | 0.130 | 0.130 | EcoSSL (USEPA 2005) | | |
| Fluoranthene | | | 0.500 | 0.500 | EcoSSL (USEPA 2005) | | |
| Indeno(1,2,3-cd)pyrene | | | 0.110 | 0.110 | EcoSSL (USEPA 2005) | | |
| Pyrene | | | 0.720 | 0.720 | EcoSSL (USEPA 2005) | | |
| Total HPAHs | 90.2 | 102 | 0.143 | 0.143 | EcoSSL regression (USEPA 2005) | 0.9469 | -1.7026 |
| SVOCs | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | | | 0.00150 | 0.00150 | McKone 1994 (ORNL RAIS) | | |
| Butyl Benzyl Phthalate | | | 0.0695 | 0.0695 | McKone 1994 (ORNL RAIS) | | |
| Carbazole | | | 0.268 | 0.268 | McKone 1994 (ORNL RAIS) | | |
| Dibenzofuran | | | 0.157 | 0.157 | McKone 1994 (ORNL RAIS) | | |
| Di-n-butyl Phthalate | | | 0.236 | 0.236 | Briggs et al. 1982 | | |
| VOCs | | | | | | | |
| 1,2,4-Trimethylbenzene | | | 0.302 | 0.302 | McKone 1994 (ORNL RAIS) | | |

Table 3-8a Soil to Plant Tissue Bioaccumulation Factors - Landfill AOPC

Notes:

a) Used for Canada goose dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: ln (tissue concentration) = Y- intercept + slope * (ln [soil concentration]).

d) Most conservative (higher) BAF between the benzo(b)- and benzo(k)fluoranthenes.

-- = No Data % = percent AOPC = area of potential concern BAF = bioaccumulation factor CPEC = chemical of potential ecological concern dw = dry weight EcoSSL = ecological soil screening level ft bgs = feet below ground surface HPAHs = high molecular weight polycyclic aromatic hydrocarbons kg = kilogram mg/kg = milligrams per kilogram SVOC = semivolatile organic compound UCL = upper confidence limit VOC = volatile organic compound

References:

Bechtel-Jacobs, 1998. Empirical Models for the Uptake of Inorganic Chemicals from Soil by Plants, Oak Ridge National Laboratory (ORNL), Oak Ridge TN. 116 pp BJC/OR-13.

Briggs, G., R. Bromilow, and A. Evans. 1982. Relationships Between Lipophilicity and Root Uptake and Translocation of Non-Ionized Chemicals by Barley. Pestic. Sci. 13: 495-504. Equation: $((10^{(0.77*logKow-1.52)+0.82})/(Koc^{*0.05}))/0.222$, where logKow = 4.50 and Koc = 33900 for di-n-butyl phthalate.

McKone, T. E. 1994. Uncertainty and variability in human exposures to soil contaminants through home-grown food: a Monte Carlo assessment. Risk Anal. 14(4):449-463. From ORNL Risk Assessment Information System (RAIS) Database (soil-to-dry plant uptake).

Table 3-8b Soil to Plant Tissue Bioaccumulation Factors – Sandblast Area AOPC

| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | 95% UCL in Shallow Soil (0-3 ft bgs) (mg/kg dw) | Surface Soil BAF Plant (kg dw tissue/ kg dw soil) ^a | Shallow Soil BAF Plant (kg dw tissue/ kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|---------------------------|--|--|---|---|--|----------------------------------|--------------------------------------|
| Inorganics | | | | | | | |
| Antimony | 2.94 | 2.42 | 0.0369 | 0.0373 | EcoSSL regression (USEPA 2005) | 0.938 | -3.233 |
| Cadmium | 3.20 | 2.63 | 0.367 | 0.401 | EcoSSL regression (USEPA 2005) | 0.546 | -0.475 |
| Chromium | | | 0.0410 | 0.0410 | EcoSSL (USEPA 2005) | | |
| Lead | 465 | 529 | 0.0179 | 0.0169 | EcoSSL regression (USEPA 2005) | 0.561 | -1.328 |
| Mercury | | | 0.650 | 0.650 | Bechtel-Jacobs 1998 median (ORNL) | | |
| Nickel | 353 | 251 | 0.0247 | 0.0269 | EcoSSL regression (USEPA 2005) | 0.748 | -2.223 |
| Butyltins | • | | | | | • | • |
| Tributyltin | | | 2.04 | 2.04 | McKone 1994 (ORNL RAIS) | | |
| Pesticides | | | | | | | |
| BHC (delta) | | | 0.153 | 0.153 | McKone 1994 (ORNL RAIS) | | |
| BHC (gamma) Lindane | | | 0.268 | 0.268 | McKone 1994 (ORNL RAIS) | | |
| Chlordane (alpha) | | | 0.0112 | 0.0112 | McKone 1994 (ORNL RAIS) | | |
| Chlordane (gamma) | | | 0.00950 | 0.00950 | McKone 1994 (ORNL RAIS) | | |
| Endosulfan I | | | 0.231 | 0.231 | McKone 1994 (ORNL RAIS) | | |
| Endosulfan II | | | 0.231 | 0.231 | McKone 1994 (ORNL RAIS) | | |
| Endosulfan Sulfate | | | 0.290 | 0.290 | McKone 1994 (ORNL RAIS) | | |
| Endrin Aldehyde | | | 0.0633 | 0.0633 | McKone 1994 (ORNL RAIS) | | |
| Endrin Ketone | | | 0.0491 | 0.0491 | McKone 1994 (ORNL RAIS) | | |
| Endrin | | | 0.0371 | 0.0371 | McKone 1994 (ORNL RAIS) | | |
| Heptachlor | | | 0.0112 | 0.0112 | McKone 1994 (ORNL RAIS) | | |
| Methoxychlor | | | 0.0436 | 0.0436 | McKone 1994 (ORNL RAIS) | | |
| HPAHs | | | | | | | |
| Benzo(a)anthracene | 1.75 | 2.45 | 0.0531 | 0.046 | EcoSSL regression (USEPA 2005) | 0.5944 | -2.7078 |
| Benzo(a)pyrene | 1.76 | 2.39 | 0.125 | 0.125 | EcoSSL regression (USEPA 2005) | 0.9750 | -2.0615 |
| Benzo(b)fluoranthene | | | 0.310 | 0.310 | EcoSSL (USEPA 2005) | | |
| Benzo(g,h,i)perylene | 1.07 | 1.05 | 0.399 | 0.398 | EcoSSL regression (USEPA 2005) | 1.1829 | -0.9313 |
| Benzo(k)fluoranthene | 0.676 | 0.764 | 0.122 | 0.120 | EcoSSL regression (USEPA 2005) | 0.8595 | -2.1579 |
| Benzofluoranthenes, Total | | | 0.310 | 0.310 | Benzofluoranthene surrogate ^d | | |
| Chrysene | 2.07 | 4.59 | 0.0496 | 0.0359 | EcoSSL regression (USEPA 2005) | 0.5944 | -2.7078 |
| Dibenzo(a,h)anthracene | | | 0.13 | 0.13 | EcoSSL (USEPA 2005) | | |
| Fluoranthene | | | 0.50 | 0.50 | EcoSSL (USEPA 2005) | | |
| Indeno(1,2,3-cd)pyrene | | | 0.11 | 0.11 | EcoSSL (USEPA 2005) | | |
| Pyrene | | | 0.72 | 0.72 | EcoSSL (USEPA 2005) | | |
| Total HPAHs | 18.6 | 40.8 | 0.156 | 0.150 | EcoSSL regression (USEPA 2005) | 0.9469 | -1.7026 |

Table 3-8b Soil to Plant Tissue Bioaccumulation Factors – Sandblast Area AOPC

| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | 95% UCL in Shallow Soil (0-3 ft bgs) (mg/kg dw) | Surface Soil BAF Plant (kg dw tissue/ kg dw soil) ^a | Shallow Soil BAF Plant (kg dw tissue/ kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|-----------------------------|--|--|---|---|-------------------------|----------------------------------|--------------------------------------|
| SVOCs | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | | | 0.00150 | 0.00150 | McKone 1994 (ORNL RAIS) | | |
| Butyl Benzyl Phthalate | | | 0.0695 | 0.0695 | McKone 1994 (ORNL RAIS) | | |
| Carbazole | | | 0.268 | 0.268 | McKone 1994 (ORNL RAIS) | | |
| Dibenzofuran | | | 0.157 | 0.157 | McKone 1994 (ORNL RAIS) | | |
| Di-n-butyl Phthalate | | | 0.236 | 0.236 | Briggs et al. 1982 | | |
| Di-n-octyl Phthalate | | | 0.000772 | 0.000772 | McKone 1994 (ORNL RAIS) | | |
| VOCs | | | | | | | |
| 1,2,4-Trimethylbenzene | | | 0.302 | 0.302 | McKone 1994 (ORNL RAIS) | | |
| 4-Isopropyltoluene | | | 0.161 | 0.161 | McKone 1994 (ORNL RAIS) | | |
| n-Propylbenzene | | | 0.279 | 0.279 | McKone 1994 (ORNL RAIS) | | |

Notes:

a) Used for Canada goose dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: ln (tissue concentration) = Y- intercept + slope * (ln [soil concentration]).

d) Most conservative (higher) BAF between the benzo(b)- and benzo(k)fluoranthenes.

| = No Data | ft bgs = feet below ground surface |
|---|--|
| % = percent | kg = kilogram |
| AOPC = area of potential concern | mg/kg = milligrams per kilogram |
| BAF = bioaccumulation factor | SVOC = semivolatile organic compound |
| CPEC = chemical of potential ecological concern | UCL = upper confidence limit |
| dw = dry weight | VOC = volatile organic compound |
| EcoSSL = ecological soil screening level | HPAHs = high molecular weight polycyclic aromatic hydrocarbons |

References:

Bechtel-Jacobs, 1998. Empirical Models for the Uptake of Inorganic Chemicals from Soil by Plants, Oak Ridge National Laboratory (ORNL), Oak Ridge TN. 116 pp BJC/OR-13.

Briggs, G., R. Bromilow, and A. Evans. 1982. Relationships Between Lipophilicity and Root Uptake and Translocation of Non-Ionized Chemicals by Barley. Pestic. Sci. 13: 495-504. Equation: ((10^(0.77*logKow-1.52)+0.82)/(Koc*0.05))/0.222, where logKow = 4.50 and Koc = 33900 for di-n-butyl phthalate.

McKone, T. E. 1994. Uncertainty and variability in human exposures to soil contaminants through home-grown food: a Monte Carlo assessment. Risk Anal. 14(4):449-463. From ORNL Risk Assessment Information System (RAIS) Database (soil-to-dry plant uptake).

Table 3-8c Soil to Plant Tissue Bioaccumulation Factors – Pistol Range AOPC

| CPEC | 95% UCL in Surface Soil (0-1.5 ft bgs) (mg/kg dw) | Surface Soil BAF Plant (kg dw tissue/ kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|------------|--|---|--------------------------------|----------------------------------|--------------------------------------|
| Inorganics | | | | | |
| Lead | 365 | 0.0199 | EcoSSL regression (USEPA 2005) | 0.561 | -1.328 |

Notes:

a) Used for Canada goose dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: ln (tissue concentration) = Y- intercept + slope * (ln [soil concentration]).

% = percent AOPC = area of potential concern BAF = bioaccumulation factor CPEC = chemical of potential ecological concern dw = dry weight EcoSSL = ecological soil screening level ft bgs = feet below ground surface kg = kilogram mg/kg = milligrams per kilogram UCL = upper confidence limit

References:

Table 3-8d Soil to Plant Tissue Bioaccumulation Factors – Bulb Slope AOPC

| СРЕС | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Surface Soil BAF Plant (kg dw tissue/ kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|------------|--|---|-----------------------------------|----------------------------------|--------------------------------------|
| Inorganics | | | | | |
| Lead | 307 | 0.0214 | EcoSSL regression (USEPA 2005) | 0.561 | -1.328 |
| Mercury | | 0.650 | Bechtel-Jacobs 1998 median (ORNL) | | |

Notes:

a) Used for Canada goose dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other

c) Regression Formula: ln (tissue concentration) = Y-intercept + slope * (ln [soil concentration]).

--- = No Data % = percent AOPC = area of potential concern BAF = bioaccumulation factor CPEC = chemical of potential ecological concern dw = dry weight EcoSSL = ecological soil screening level ft bgs = feet below ground surface kg = kilogram mg/kg = milligrams per kilogram UCL = upper confidence limit

References:

Bechtel-Jacobs, 1998. Empirical Models for the Uptake of Inorganic Chemicals from Soil by Plants, Oak Ridge National Laboratory (ORNL), Oak Ridge TN. 116 pp BJC/OR-13.

Table 3-9a Soil to Soil Invertebrate Tissue Bioaccumulation Factors - Landfill AOPC

| ann a | 95% UCL in Surface Soil (0-1 ft bgs) | 95% UCL in Shallow Soil (0-3 ft bgs) | Surface Soil BAF Invertebrate (kg dw tissue/ | Shallow Soil BAF Invertebrate (kg dw tissue/ | a b | Regression | Regression |
|-----------------------------|--|--|--|--|--|--------------------|------------------------|
| СРЕС | (mg/kg dw) | (mg/kg dw) | kg dw soil) ^a | kg dw soil) ^a | Source ^b | Slope ^c | Intercept ^c |
| Inorganics | T | | | | 1 | T | |
| Antimony | | | 1.0 | 1.0 | EcoSSL (USEPA 2005) | | |
| Chromium | | | 0.306 | 0.306 | EcoSSL (USEPA 2005) | | |
| Copper | | | 0.515 | 0.515 | EcoSSL (USEPA 2005) | | |
| Lead | 332 | 511 | 0.262 | 0.241 | EcoSSL regression (USEPA 2005) | 0.807 | -0.218 |
| Mercury | | | 1.69 | 1.69 | Sample et al. 1998b (ORNL) | | |
| Nickel | | | 1.06 | 1.06 | Sample et al. 1998b (ORNL) | | |
| Butyltins | | | | | | | |
| Tributyltin | | | 0.0424 | 0.0424 | Jager 1998 (logKow algorithm) | | |
| Pesticides | | | | | | | |
| Chlordane (technical) | | | 47.7 | 47.7 | Jager 1998 (logKow algorithm) | | |
| Heptachlor | | | 8.75 | 8.75 | USEPA 1999 (converted to dw ^d) | | |
| HPAHs | | | | | | | |
| Benzo(a)anthracene | | | 1.59 | 1.59 | EcoSSL (USEPA 2005) | | |
| Benzo(a)pyrene | | | 1.33 | 1.33 | EcoSSL (USEPA 2005) | | |
| Benzo(b)fluoranthene | | | 2.60 | 2.60 | EcoSSL (USEPA 2005) | | |
| Benzo(g,h,i)perylene | | | 2.94 | 2.94 | EcoSSL (USEPA 2005) | | |
| Benzo(k)fluoranthene | | | 2.60 | 2.60 | EcoSSL (USEPA 2005) | | |
| Benzofluoranthenes, Total | | | 2.60 | 2.60 | EcoSSL (USEPA 2005) | | |
| Chrysene | | | 2.29 | 2.29 | EcoSSL (USEPA 2005) | | |
| Dibenz(a,h)anthracene | | | 2.31 | 2.31 | EcoSSL (USEPA 2005) | | |
| Fluoranthene | | | 3.04 | 3.04 | EcoSSL (USEPA 2005) | | |
| Indeno(1,2,3-cd)pyrene | | | 2.86 | 2.86 | EcoSSL (USEPA 2005) | | |
| Pyrene | | | 1.75 | 1.75 | EcoSSL (USEPA 2005) | | |
| Total HPAHs | | | 2.6 | 2.6 | EcoSSL (USEPA 2005) | | |
| SVOCs | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Butyl Benzyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Carbazole | | | 1 | 1 | Conservative Default Value | | |
| Dibenzofuran | | | 1 | 1 | Conservative Default Value | | |

Table 3-9a Soil to Soil Invertebrate Tissue Bioaccumulation Factors - Landfill AOPC

| СРЕС | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | 95% UCL in Shallow Soil (0-3 ft bgs) (mg/kg dw) | Surface Soil BAF Invertebrate (kg dw tissue/ kg dw soil) ^a | Shallow Soil BAF Invertebrate (kg dw tissue/ kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|------------------------|--|--|--|--|----------------------------|----------------------------------|--------------------------------------|
| Di-n-butyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| VOCs | | | | | | | |
| 1,2,4-Trimethylbenzene | | | 1 | 1 | Conservative Default Value | | |

Notes:

a) Used for the American robin and vagrant shrew dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: ln (tissue concentration) = Y- intercept + slope * (ln [soil concentration]).

d) BAF (dry weight) = BAF (wet weight) / 0.16 (assuming earthworm contains 16% dry solids, USEPA 1993)

-- = no data

% = percent

AOPC = area of potential concern

BAF = bioaccumulation factor in kilograms of soil per kilograms of tissue.

CPEC = chemical of potential ecological concern

dw = dry weight

EcoSSL = ecological soil screening level

ft bgs = feet below ground surface

HPAHs = high molecular weight polycyclic aromatic hydrocarbons

kg = kilogram

mg/kg = milligrams per kilogram

SVOC = semivolatile organic compound

UCL = upper confidence limit

VOC = volatile organic compound

References:

Jager, T. 1998. Mechanistic approach for estimating bioconcentration of organic chemicals in earthworms. Environ. Toxicol. Chem. 17:2080-2090. BAF for non-ionic organic compounds = Kww (L/kg worm dw) / Kd (L/kg soil dw), where Kww = $(10^{(0.87 * \log Kow - 2.0)})/(0.16)$; and Kd = foc (0.01) * Kow

Sample, B.E., J.J. Beauchamp, R.A. Efroymson, G.W. Suter, II, and T.L. Ashwood. 1998b. Development and Validation of Bioaccumulation Models for Earthworms. Oak Ridge National Laboratory, Oak Ridge TN. 93 pp, ES/ER/TM-220. Median BAF.

Table 3-9a Soil to Soil Invertebrate Tissue Bioaccumulation Factors - Landfill AOPC

USEPA. 1999. Screening Level Ecological Risk Assessment Protocol for Hazardous Waste Combustion Facilities. Office of Solid Waste & Emergency Response. August. EPA530-D-99-001.

USEPA. 1993. "Wildlife Exposure Factors Handbook." December. 1993.

| Table 3-9b |
|--|
| Soil to Soil Invertebrate Tissue Bioaccumulation Factors – Sandblast Area AOPC |

| | 95% UCL in Surface Soil (0-1 ft bgs) | 95% UCL in Shallow Soil (0-3 ft bgs) | Surface Soil BAF Invertebrate (kg dw tissue/ | Shallow Soil BAF Invertebrate (kg dw tissue/ | | Regression | Regression |
|---------------------------|--|--|--|--|--|--------------------|------------------------|
| CPEC | (mg/kg dw) | (mg/kg dw) | kg dw soil) ^a | kg dw soil) ^a | Source ^b | Slope ^c | Intercept ^c |
| Inorganics | | | | | | | |
| Antimony | | | 1.0 | 1.0 | EcoSSL (USEPA 2005) | | |
| Cadmium | 3.20 | 2.63 | 6.52 | 6.79 | EcoSSL regression (USEPA 2005) | 0.795 | 2.114 |
| Chromium | | | 0.306 | 0.306 | EcoSSL (USEPA 2005) | | |
| Lead | 465 | 529 | 0.246 | 0.240 | EcoSSL regression (USEPA 2005) | 0.807 | -0.218 |
| Mercury | | | 1.69 | 1.69 | Sample et al. 1998 (ORNL) | | |
| Nickel | | | 1.06 | 1.06 | Sample et al. 1998 (ORNL) | | |
| Butyltins | | | | | | | |
| Tributyltin | | | 0.0424 | 0.0424 | Jager 1998 (logKow algorithm) | | |
| Pesticides | | | | | | | |
| BHC (delta) | | | 8.89 | 8.89 | Jager 1998 (logKow algorithm) | | |
| BHC (gamma) Lindane | | | 3.83 | 3.83 | Jager 1998 (logKow algorithm) | | |
| Chlordane (alpha) | | | 18.8 | 18.8 | Jager 1998 (logKow algorithm) | | |
| Chlordane (gamma) | | | 23.9 | 23.9 | Jager 1998 (logKow algorithm) | | |
| Endosulfan I | | | 1.99 | 1.99 | Jager 1998 (logKow algorithm) | | |
| Endosulfan II | | | 1.99 | 1.99 | Jager 1998 (logKow algorithm) | | |
| Endosulfan Sulfate | | | 0.970 | 0.970 | Jager 1998 (logKow algorithm) | | |
| Endrin Aldehyde | | | 28.7 | 28.7 | Jager 1998 (logKow algorithm) | | |
| Endrin Ketone | | | 14.1 | 14.1 | Jager 1998 (logKow algorithm) | | |
| Endrin | | | 10.4 | 10.4 | Jager 1998 (logKow algorithm) | | |
| Heptachlor | | | 30.7 | 30.7 | USEPA 1999 (converted to dw ^d) | | |
| Methoxychlor | | | 3.98 | 3.98 | Jager 1998 (logKow algorithm) | | |
| HPAHs | | | • | • | | | • |
| Benzo(a)anthracene | | | 1.59 | 1.59 | EcoSSL (USEPA 2005) | | |
| Benzo(a)pyrene | | | 1.33 | 1.33 | EcoSSL (USEPA 2005) | | |
| Benzo(b)fluoranthene | | | 2.60 | 2.60 | EcoSSL (USEPA 2005) | | |
| Benzo(g,h,i)perylene | | | 2.94 | 2.94 | EcoSSL (USEPA 2005) | | |
| Benzo(k)fluoranthene | | | 2.60 | 2.60 | EcoSSL (USEPA 2005) | | |
| Benzofluoranthenes, Total | | | 2.60 | 2.60 | EcoSSL (USEPA 2005) | | |
| Chrysene | | | 2.29 | 2.29 | EcoSSL (USEPA 2005) | | |
| Dibenz(a,h)anthracene | | | 2.31 | 2.31 | EcoSSL (USEPA 2005) | | |
| Fluoranthene | | | 3.04 | 3.04 | EcoSSL (USEPA 2005) | | |
| Indeno(1,2,3-cd)pyrene | | | 2.86 | 2.86 | EcoSSL (USEPA 2005) | | |
| Pyrene | | | 1.75 | 1.75 | EcoSSL (USEPA 2005) | | |
| Total HPAHs | | | 2.6 | 2.6 | EcoSSL (USEPA 2005) | | |

Table 3-9b Soil to Soil Invertebrate Tissue Bioaccumulation Factors – Sandblast Area AOPC

| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | 95% UCL in Shallow Soil (0-3 ft bgs) (mg/kg dw) | Surface Soil BAF Invertebrate (kg dw tissue/ kg dw soil) ^a | Shallow Soil BAF Invertebrate (kg dw tissue/ kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|-----------------------------|--|--|--|--|----------------------------|----------------------------------|--------------------------------------|
| SVOCs | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Butyl Benzyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Carbazole | | | 1 | 1 | Conservative Default Value | | |
| Dibenzofuran | | | 1 | 1 | Conservative Default Value | | |
| Di-n-butyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Di-n-octyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| VOCs | | | | | • | | |
| 1,2,4-Trimethylbenzene | | | 1 | 1 | Conservative Default Value | | |
| 4-Isopropyltoluene | | | 1 | 1 | Conservative Default Value | | |
| n-Propylbenzene | | | 1 | 1 | Conservative Default Value | | |

Notes:

a) Used for the American robin and vagrant shrew dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: ln (tissue concentration) = Y- intercept + slope * (ln [soil concentration]).

d) BAF (dry weight) = BAF (wet weight) / 0.16 (assuming earthworm contains 16% dry solids, USEPA 1993)

-- = no data

% = percent

AOPC = area of potential concern

BAF = bioaccumulation factor in kilograms of soil per kilograms of tissue.

BHC = benzene hexachloride

CPEC = chemical of potential ecological concern

dw = dry weight

EcoSSL = ecological soil screening level

ft bgs = feet below ground surface

HPAHs = high molecular weight polycyclic aromatic hydrocarbons

kg = kilogram

mg/kg = milligrams per kilogram

SVOC = semivolatile organic compound

UCL = upper confidence limit

VOC = volatile organic compound

Table 3-9b

Soil to Soil Invertebrate Tissue Bioaccumulation Factors – Sandblast Area AOPC

References:

Jager, T. 1998. Mechanistic approach for estimating bioconcentration of organic chemicals in earthworms. Environ. Toxicol. Chem. 17:2080-2090. BAF for non-ionic organic compounds = Kww (L/kg worm dw) / Kd (L/kg soil dw), where Kww = $(10^{(0.87 * \log Kow - 2.0)})/0.16$; and Kd = foc (0.01) * Koc

LogKow and Koc values taken from ORNL RAIS Database Chemical Parameters (2014).

Sample, B.E., J.J. Beauchamp, R.A. Efroymson, G.W. Suter, II, and T.L. Ashwood. 1998. Development and Validation of Bioaccumulation Models for Earthworms. Oak Ridge National Laboratory, Oak Ridge TN. 93 pp, ES/ER/TM-220. Median BAF.

USEPA. 1993. "Wildlife Exposure Factors Handbook." December. 1993.

USEPA. 1999. Screening Level Ecological Risk Assessment Protocol for Hazardous Waste Combustion Facilities. Office of Solid Waste & Emergency Response. August. EPA530-D-99-001.

Table 3-9c Soil to Soil Invertebrate Tissue Bioaccumulation Factors – Pistol Range AOPC

| | 95% UCL in Surface Soil (0-1.5 ft bgs) | Surface Soil BAF Invertebrate (kg dw tissue/kg dw | a b | Regression | Regression |
|------------|--|---|--------------------------------|--------------------|------------------------|
| CPEC | (mg/kg dw) | soil) ^a | Source ¹ | Slope ^c | Intercept ^c |
| Inorganics | | | | | |
| Lead | 365 | 0.258 | EcoSSL regression (USEPA 2005) | 0.807 | -0.218 |

Notes:

a) Used for the American robin and vagrant shrew dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other

c) Regression Formula: ln (tissue concentration) = Y-intercept + slope * (ln [soil concentration]).

% = percent

AOPC = area of potential concern

BAF = bioaccumulation factor in kilograms of soil per kilograms of tissue.

CPEC = chemical of potential ecological concern

dw = dry weight

EcoSSL = ecological soil screening level

ft bgs = feet below ground surface

kg = kilogram

mg/kg = milligrams per kilogram

UCL = upper confidence limit

References:

Table 3-9d Soil to Soil Invertebrate Tissue Bioaccumulation Factors – Bulb Slope AOPC

| CPEC Inorganics | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Surface Soil BAF Invertebrate (kg dw tissue/kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|--------------------|--|---|---------------------------------|----------------------------------|--------------------------------------|
| | 7 | | | - | - |
| Lead | 307 | 0.266 | EcoSSL regression (USEPA, 2005) | 0.807 | -0.218 |
| Mercury | | 1.69 | Sample et al. 1998b (ORNL) | | |

Notes:

a) Used for the american robin and vagrant shrew dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: ln (tissue concentration) = Y- intercept + slope * (ln [soil concentration]).

-- = no data

% = percent

AOPC = area of potential concern

BAF = bioaccumulation factor in kilograms of soil per kilograms of tissue.

CPEC = chemical of potential ecological concern

dw = dry weight

EcoSSL = ecological soil screening level

ft bgs = feet below ground surface

kg = kilogram

mg/kg = milligrams per kilogram

UCL = upper confidence limit

References:

Sample, B.E., J.J. Beauchamp, R.A. Efroymson, G.W. Suter, II, and T.L. Ashwood. 1998b. Development and Validation of Bioaccumulation Models for Earthworms. Oak Ridge National Laboratory, Oak Ridge TN. 93 pp, ES/ER/TM-220. Median BAF.

 Table 3-10a

 Soil to Mammal Tissue Bioaccumulation Factors - Landfill AOPC

| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | 95% UCL in Shallow Soil (0-3 ft bgs) (mg/kg dw) | Surface Soil BAF Mammal (kg dw tissue/kg dw soil) ^a | Shallow Soil BAF Mammal (kg dw tissue/kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|-----------------------------|--|--|---|---|-----------------------------------|----------------------------------|--------------------------------------|
| Inorganics | | | | | | | |
| Antimony | | | 0.050 | 0.050 | EcoSSL (USEPA 2005) | | |
| Chromium | 242 | 594 | 0.0539 | 0.0424 | EcoSSL regression (USEPA 2005) | 0.7338 | -1.4599 |
| Copper | 170 | 191 | 0.0952 | 0.0861 | EcoSSL regression (USEPA 2005) | 0.1444 | 2.042 |
| Lead | 332 | 511 | 0.0423 | 0.0333 | EcoSSL regression (USEPA 2005) | 0.4422 | 0.0761 |
| Mercury | | | 0.0543 | 0.0543 | Sample et al. 1998 (ORNL) | | |
| Nickel | 175 | 472 | 0.0495 | 0.0292 | EcoSSL regression (USEPA 2005) | 0.4658 | -0.2462 |
| Butyltins | | | | | | | |
| Tributyltin | | | 0.0000124 | 0.0000124 | Travis and Arms 1988 ^d | | |
| Pesticides | | | | | | - | |
| Chlordane (technical) | | | 0.130 | 0.130 | Travis and Arms 1988 ^d | | |
| Heptachlor | | | 0.0232 | 0.0232 | Travis and Arms 1988 ^d | | |
| HPAHs ^e | | | | | | | |
| Benz(a)anthracene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(a)pyrene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(b)fluoranthene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(g,h,i)perylene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(k)fluoranthene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzofluoranthenes, Total | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Chrysene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Dibenz(a,h)anthracene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Fluoranthene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Indeno(1,2,3-cd)pyrene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Pyrene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Total HPAHs | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| SVOCs | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Butyl Benzyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Carbazole | | | 1 | 1 | Conservative Default Value | | |
| Dibenzofuran | | | 1 | 1 | Conservative Default Value | | |
| Di-n-butyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| VOCs | | | | | | | |
| 1,2,4-Trimethylbenzene | | | 1 | 1 | Conservative Default Value | | |

Table 3-10a Soil to Mammal Tissue Bioaccumulation Factors - Landfill AOPC

Notes:

a) Used for the American kestrel and American mink dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: *ln* (tissue concentration) = *Y*-intercept + slope * (*ln* [soil concentration]).

d) Converted to dry weight: BAF (dry weight) = BAF (wet weight) / 0.32 (assuming mammals contains 32% dry solids, USEPA 1993)

e) The recommended BAF for PAHs is zero because of the rapid metabolism of these compounds after ingestion by birds and mammals.

-- = no data

% = percent AOPC = area of potential concern BAF = bioaccumulation factor CPEC = chemical of potential ecological concern dw = dry weight ft bgs = feet below ground surface HPAHs = high molecular weight polycyclic aromatic hydrocarbons kg = kilogram mg/kg = milligrams per kilogram PAH = polycyclic aromatic hydrocarbon SVOC = semivolatile organic compound UCL = 95% upper confidence limit VOC = volatile organic compound

References:

Sample, B.E., J.J. Beauchamp, R.A. Efroymson, G.W. Suter II, and T.L. Ashwood. 1998. Development and Validation of Bioaccumulation Models for Small Mammals, ES/ER/TM-219, Oak Ridge National Laboratory, Oak Ridge, TN. Median BAF.

Travis, C.C., and A.D. Arms, 1988. Bioconcentration of Organics in Beef, Milk, and Vegetation. Environmental Science & Technology. 22: 271-274. Based on biotransfer equation for mammals: logBa = -7.6 + logKow. LogKow values taken from ORNL RAIS Database Chemical Parameters (2014).

USEPA. 1993. "Wildlife Exposure Factors Handbook." December. 1993.

 Table 3-10b

 Soil to Mammal Tissue Bioaccumulation Factors – Sandblast Area AOPC

| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | 95% UCL in Shallow Soil (0-3 ft bgs) (mg/kg dw) | Surface Soil BAF Mammal (kg dw tissue/ kg dw soil) ^a | Shallow Soil BAF Mammal (kg dw tissue/ kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|-----------------------|--|--|--|--|-----------------------------------|----------------------------------|--------------------------------------|
| Inorganics | | | | | | | |
| Antimony | | | 0.050 | 0.050 | EcoSSL (USEPA 2005) | | |
| Cadmium | 3.20 | 2.63 | 0.154 | 0.171 | EcoSSL regression (USEPA 2005) | 0.4723 | -1.2571 |
| Chromium | 720 | 579 | 0.0403 | 0.0427 | EcoSSL regression (USEPA 2005) | 0.7338 | -1.4599 |
| Lead | 465 | 529 | 0.0351 | 0.0327 | EcoSSL regression (USEPA 2005) | 0.4422 | 0.0761 |
| Mercury | | | 0.0543 | 0.0543 | Sample et al. 1998 (ORNL) | | |
| Nickel | 353 | 251 | 0.0340 | 0.0408 | EcoSSL regression (USEPA 2005) | 0.4658 | -0.2462 |
| Butyltins | | | | | • | 1 | |
| Tributyltin | | | 0.0000124 | 0.0000124 | Travis and Arms 1988 ^d | | |
| Pesticides | | | | | | | |
| BHC (delta) | | | 0.00108 | 0.00108 | Travis and Arms 1988 ^d | | |
| BHC (gamma) Lindane | | | 0.000412 | 0.000412 | Travis and Arms 1988 ^d | | |
| Chlordane (alpha) | | | 0.0988 | 0.0988 | Travis and Arms 1988 ^d | | |
| Chlordane (gamma) | | | 0.130 | 0.130 | Travis and Arms 1988 ^d | | |
| Chlordane (technical) | | | 0.113 | 0.113 | Travis and Arms 1988 ^d | | |
| Endosulfan I | | | 0.000531 | 0.000531 | Travis and Arms 1988 ^d | | |
| Endosulfan II | | | 0.000531 | 0.000531 | Travis and Arms 1988 ^d | | |
| Endosulfan Sulfate | | | 0.000359 | 0.000359 | Travis and Arms 1988 ^d | | |
| Endrin Aldehyde | | | 0.00495 | 0.00495 | Travis and Arms 1988 ^d | | |
| Endrin Ketone | | | 0.00767 | 0.00767 | Travis and Arms 1988 ^d | | |
| Endrin | | | 0.0124 | 0.0124 | Travis and Arms 1988 ^d | | |
| Heptachlor | | | 0.0988 | 0.0988 | Travis and Arms 1988 ^d | | |
| Methoxychlor | | | 0.00944 | 0.00944 | Travis and Arms 1988 ^d | | |
| HPAHs ^e | | | | | | | |
| Benz(a)anthracene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(a)pyrene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |

 Table 3-10b

 Soil to Mammal Tissue Bioaccumulation Factors – Sandblast Area AOPC

| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | 95% UCL in Shallow Soil (0-3 ft bgs) (mg/kg dw) | Surface Soil BAF Mammal (kg dw tissue/ kg dw soil) ^a | Shallow Soil BAF Mammal (kg dw tissue/ kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|-----------------------------|--|--|--|--|----------------------------|----------------------------------|--------------------------------------|
| Benzo(b)fluoranthene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(g,h,i)perylene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(k)fluoranthene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzofluoranthenes, Total | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Chrysene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Dibenz(a,h)anthracene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Fluoranthene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Indeno(1,2,3-cd)pyrene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Pyrene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Total HPAHs | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| SVOCs | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Butyl Benzyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Carbazole | | | 1 | 1 | Conservative Default Value | | |
| Dibenzofuran | | | 1 | 1 | Conservative Default Value | | |
| Di-n-butyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Di-n-octyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| VOCs | | | | | | | |
| 1,2,4-Trimethylbenzene | | | 1 | 1 | Conservative Default Value | | |
| 4-Isopropyltoluene | | | 1 | 1 | Conservative Default Value | | |
| n-Propylbenzene | | | 1 | 1 | Conservative Default Value | | |

Notes:

a) Used for the American kestrel and American mink dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: *ln* (tissue concentration) = *Y*-intercept + slope * (*ln* [soil concentration]).

d) Converted to dry weight: BAF (dry weight) = BAF (wet weight) / 0.32(assuming mammals contains 32% dry solids, USEPA 1993)

e) The recommended BAF for PAHs is zero because of the rapid metabolism of these compounds after ingestion by birds and mammals.

Table 3-10b Soil to Mammal Tissue Bioaccumulation Factors – Sandblast Area AOPC

-- = no data
% = percent
AOPC = area of potential concern
BAF = bioaccumulation factor
BHC = benzene hexachloride
CPEC = chemical of potential ecological concern
dw = dry weight
ft bgs = feet below ground surface
HPAHs = high molecular weight polycyclic aromatic hydrocarbons
kg = kilogram
mg/kg = milligrams per kilogram
SVOC = semivolatile organic compound
UCL = 95% upper confidence limit
VOC = volatile organic compound

References:

Sample, B.E., J.J. Beauchamp, R.A. Efroymson, G W. Suter II, and T.L. Ashwood. 1998. Development and Validation of Bioaccumulation Models for Small Mammals, ES/ER/TM-219, Oak Ridge National Laboratory, Oak Ridge, TN. Median BAF.

Travis, C.C., and A.D. Arms, 1988. Bioconcentration of Organics in Beef, Milk, and Vegetation. Environmental Science & Technology. 22: 271-274. Based on biotransfer equation for mammals: logBa = -7.6 + logKow. LogKow values taken from ORNL RAIS Database Chemical Parameters (2014).

USEPA. 1993. "Wildlife Exposure Factors Handbook." December. 1993.

Table 3-10c Soil to Mammal Tissue Bioaccumulation Factors – Pistol Range AOPC

| CPEC | 95% UCL in Surface Soil (0-1.5 ft bgs) (mg/kg dw) | Soil BAF Mammal bgs) (kg dw tissue/kg dw | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|------------|--|---|--------------------------------|----------------------------------|--------------------------------------|
| Inorganics | | | | | |
| Lead | 365 | 0.0402 | EcoSSL regression (USEPA 2005) | 0.4422 | 0.0761 |

Notes:

a) Used for the American kestrel and American mink dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: ln (tissue concentration) = Y- intercept + slope * (ln [soil concentration]).

The recommended BAF for PAHs is zero because of the rapid metabolism of these compounds after ingestion by birds and mammals.

-- = no data % = percent AOPC = area of potential concern BAF = bioaccumulation factor CPEC = chemical of potential ecological concern dw = dry weight EcoSSL = ecological soil screening level ft bgs = feet below ground surface kg = kilogram mg/kg = milligrams per kilogram PAH = polycyclic aromatic hydrocarbon UCL = 95% upper confidence limit

References:

USEPA. 2005. "Guidance for Developing Ecological Soil Screening Levels." Attachment 4-1. Office of Solid Waste and Emergency Response, Washington, DC. Office of Solid Waste and Emergency Response Directive 9285.7-55. November. Last Updated 2007.

Table 3-10d Soil to Mammal Tissue Bioaccumulation Factors – Bulb Slope AOPC

| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Surface Soil BAF Mammal (kg dw tissue/kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|------------|--|---|--------------------------------|----------------------------------|--------------------------------------|
| Inorganics | | | | | |
| Lead | 307 | 0.0442 | EcoSSL regression (USEPA 2005) | 0.4422 | 0.0761 |
| Mercury | | 0.0543 | Sample et al. 1998a (ORNL) | | |

Notes:

a) Used for the american kestrel and american mink dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: ln (tissue concentration) = Y- intercept + slope * (ln [soil concentration]).

-- = no data

% = percent AOPC = area of potential concern BAF = bioaccumulation factor CPEC = chemical of potential ecological concern dw = dry weight EcoSSL = ecological soil screening level ft bgs = feet below ground surface kg = kilogram mg/kg = milligrams per kilogram UCL = 95% upper confidence limit

References:

Sample, B.E., J.J. Beauchamp, R.A. Efroymson, G W. Suter II, and T.L. Ashwood. 1998a. Development and Validation of Bioaccumulation Models for Small Mammals, ES/ER/TM-219, Oak Ridge National Laboratory, Oak Ridge, TN. Median BAF.

USEPA. 2005. "Guidance for Developing Ecological Soil Screening Levels." Attachment 4-1. Office of Solid Waste and Emergency Response, Washington, DC. Office of Solid Waste and Emergency Response Directive 9285.7-55. November. Last Updated 2007.

 Table 3-10e

 Soil to Mammal Tissue Bioaccumulation Factors – Combined AOPCs

| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | 95% UCL in Shallow Soil (0-3 ft bgs) (mg/kg dw) | Surface Soil BAF Mammal (kg dw tissue/kg dw soil) ^a | Shallow Soil BAF Mammal (kg dw tissue/kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|-----------------------|--|--|---|---|-----------------------------------|----------------------------------|--------------------------------------|
| Inorganics | | | | | | | _ |
| Antimony | | | 0.050 | 0.050 | EcoSSL (USEPA 2005) | | |
| Cadmium | 1.97 | 1.67 | 0.199 | 0.217 | EcoSSL regression (USEPA 2005) | 0.4723 | -1.2571 |
| Chromium | 567 | 510 | 0.0430 | 0.0442 | EcoSSL regression (USEPA 2005) | 0.7338 | -1.4599 |
| Copper | 95.0 | 93.6 | 0.157 | 0.159 | EcoSSL regression (USEPA 2005) | 0.1444 | 2.042 |
| Lead | 398 | 396 | 0.0383 | 0.0384 | EcoSSL regression (USEPA 2005) | 0.4422 | 0.0761 |
| Mercury | | | 0.0543 | 0.0543 | Sample et al. 1998a (ORNL) | | |
| Nickel | 227 | 234 | 0.0431 | 0.0424 | EcoSSL regression (USEPA 2005) | 0.4658 | -0.2462 |
| Butyltins | | - | | | | | • |
| Tributyltin | | | 0.0424 | 0.0424 | Travis and Arms 1988 ^d | | |
| Pesticides | | | | | | | |
| BHC (delta) | | | 8.89 | 8.89 | Travis and Arms 1988 ^d | | |
| BHC (gamma) Lindane | | | 3.83 | 3.83 | Travis and Arms 1988 ^d | | |
| Chlordane (alpha) | | | 18.8 | 18.8 | Travis and Arms 1988 ^d | | |
| Chlordane (gamma) | | | 23.9 | 23.9 | Travis and Arms 1988 ^d | | |
| Chlordane (technical) | | | 47.7 | 47.7 | Travis and Arms 1988 ^d | | |
| Endosulfan I | | | 1.99 | 1.99 | Travis and Arms 1988 ^d | | |
| Endosulfan II | | | 1.99 | 1.99 | Travis and Arms 1988 ^d | | |
| Endosulfan Sulfate | | | 0.970 | 0.970 | Travis and Arms 1988 ^d | | |
| Endrin Aldehyde | | | 28.7 | 28.7 | Travis and Arms 1988 ^d | | |
| Endrin Ketone | | | 14.1 | 14.1 | Travis and Arms 1988 ^d | | |
| Endrin | | | 10.4 | 10.4 | Travis and Arms 1988 ^d | | |
| Heptachlor | | | 8.75 | 8.75 | Travis and Arms 1988 ^d | | |
| Methoxychlor | | | 3.98 | 3.98 | Travis and Arms 1988 ^d | | |
| HPAHs ^e | | | | | | | |
| Benz(a)anthracene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |

 Table 3-10e

 Soil to Mammal Tissue Bioaccumulation Factors – Combined AOPCs

| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | 95% UCL in Shallow Soil (0-3 ft bgs) (mg/kg dw) | Surface Soil BAF Mammal (kg dw tissue/kg dw soil) ^a | Shallow Soil BAF Mammal (kg dw tissue/kg dw soil) ^a | Source ^b | Regression Slope ^c | Regression Intercept ^c |
|-----------------------------|--|--|---|---|----------------------------|----------------------------------|--------------------------------------|
| Benzo(a)pyrene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(b)fluoranthene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(g,h,i)perylene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzo(k)fluoranthene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Benzofluoranthenes, Total | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Chrysene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Dibenz(a,h)anthracene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Fluoranthene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Indeno(1,2,3-cd)pyrene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Pyrene | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| Total HPAHs | | | 0 | 0 | EcoSSL (USEPA 2005) | | |
| SVOCs | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Butyl Benzyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Carbazole | | | 1 | 1 | Conservative Default Value | | |
| Dibenzofuran | | | 1 | 1 | Conservative Default Value | | |
| Di-n-butyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| Di-n-octyl Phthalate | | | 1 | 1 | Conservative Default Value | | |
| VOCs | | | | | | | |
| 1,2,4-Trimethylbenzene | | | 1 | 1 | Conservative Default Value | | |
| 4-Isopropyltoluene | | | 1 | 1 | Conservative Default Value | | |
| n-Propylbenzene | | | 1 | 1 | Conservative Default Value | | |

a) Used for the American kestrel and American mink dose calculations.

b) Hierarchy of sources: USEPA Eco-SSL Attachment 4-1, ORNL RAIS database, other.

c) Regression Formula: ln (tissue concentration) = Y- intercept + slope * (ln [soil concentration]).

d) Converted to dry weight: BAF (dry weight) = BAF (wet weight) / 0.32 (assuming mammals contains 32% dry solids, USEPA 1993)

e) The recommended BAF for PAHs is zero because of the rapid metabolism of these compounds after ingestion by birds and mammals.

Table 3-10e Soil to Mammal Tissue Bioaccumulation Factors – Combined AOPCs

-- = no data
% = percent
AOPC = area of potential concern
BAF = bioaccumulation factor
BHC = benzene hexachloride
CPEC = chemical of potential ecological concern
dw = dry weight
ft bgs = feet below ground surface
HPAHs = high molecular weight polycyclic aromatic hydrocarbons
kg = kilogram
mg/kg = milligrams per kilogram
SVOC = semivolatile organic compound
UCL = 95% upper confidence limit
VOC = volatile organic compound

References:

Sample, B.E., J.J. Beauchamp, R.A. Efroymson, G W. Suter II, and T.L. Ashwood. 1998a. Development and Validation of Bioaccumulation Models for Small Mammals, ES/ER/TM-219, Oak Ridge National Laboratory, Oak Ridge, TN. Median BAF.

Travis, C.C., and A.D. Arms, 1988. Bioconcentration of Organics in Beef, Milk, and Vegetation. Environmental Science & Technology. 22: 271-274. Based on biotransfer equation for mammals: logBa = -7.6 + logKow. LogKow values taken from ORNL RAIS Database Chemical Parameters (2014).

USEPA. 1993. Wildlife Exposure Factors Handbook. Volumes I and II. Washington, D.C.

USEPA. 2005. "Guidance for Developing Ecological Soil Screening Levels." Attachment 4-1. Office of Solid Waste and Emergency Response, Washington, DC. Office of Solid Waste and Emergency Response Directive 9285.7-55. November. Last Updated 2007.

Table 3-11a Calculation of Dose and Hazard Quotient for the Canada Goose - Landfill AOPC

| | | EPCs ^d | | BA | AFs | | Ex | posure Fac | tors | | - | Do | oses | TR | <u>aVs</u> | | Hazard | Quotients | |
|--|--|--|---|--|--|---|---|---------------------------------------|--------------|-----|---------|--|--|------------------------------------|------------------------------------|--|--|--|--|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^c | Surface Soil BAF plant (kg tissue/kg soil) ^a | Shallow Soil BAF plant (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Plants | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) LOAEL HQ | Shallow Soil (0-3 ft bgs) NOAEL HQ | Shallow Soil (0-3 ft bgs) LOAEL HQ |
| Inorganics | - | + | | | ł | + | • | | • | • | | | | | | • | • | • | • |
| Antimony ^b | 1.36 | 1.48 | 0.0302 | 0.0387 | 0.0385 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.010 | 0.010 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Chromium III | 242 | 594 | 0.000134 | 0.0410 | 0.0410 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 1.5 | 3.7 | 2.66 | 14.5 | 5.7E-01 | 1.1E-01 | 1.4E+00 | 2.6E-01 |
| Copper | 170 | 191 | 4.86 | 0.0868 | 0.0809 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 1.7 | 1.8 | 4.05 | 12.1 | 4.1E-01 | 1.4E-01 | 4.4E-01 | 1.5E-01 |
| Lead | 332 | 511 | 0.369 | 0.0207 | 0.0171 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 1.8 | 2.6 | 1.63 | 3.26 | 1.1E+00 | 5.4E-01 | 1.6E+00 | 8.0E-01 |
| Mercury | 1.57 | 1.45 | 0.0302 | 0.650 | 0.650 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.060 | 0.056 | 0.0064 | 0.064 | 9.4E+00 | 9.4E-01 | 8.7E+00 | 8.7E-01 |
| Nickel | 175 | 472 | 2.69 | 0.0295 | 0.0229 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 1.1 | 2.6 | 6.71 | 26.8 | 1.7E-01 | 4.1E-02 | 3.9E-01 | 9.9E-02 |
| Butyltins | | • | - | | | - | • | | | | - | | | | | • | • | | - |
| Tributyltin | 0.165 | 0.0601 | 0.00136 | 2.04 | 2.04 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.018 | 0.0066 | 6.8 | 16.9 | 2.6E-03 | 1.1E-03 | 9.7E-04 | 3.9E-04 |
| Pesticides | I | I | T | | I | T | T | 1 | 1 | 1 | 1 | 1 | 1 | | 1 | r | r | T | I |
| Chlordane (technical) | 1.56 | 1.56 | 0.00462 | 0.00950 | 0.00950 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0075 | 0.0075 | 2.14 | 10.7 | 3.5E-03 | 7.0E-04 | 3.5E-03 | 7.0E-04 |
| Heptachlor | 0.00283 | 0.00286 | 0.00000685 | 0.0112 | 0.0112 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.000014 | 0.000014 | 0.013 | 0.065 | 1.1E-03 | 2.1E-04 | 1.1E-03 | 2.1E-04 |
| HPAHs ^b | | | | | | | | | | | | | | | | | | | |
| SVOCs | | | - | | | | 1 | T | 1 | | • | 1 | 1 | | 1 | | | | |
| Bis(2-ethylhexyl) Phthalate | 9.68 | 6.38 | 0.00807 | 0.00150 | 0.00150 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.042 | 0.028 | 1.1 | 11 | 3.8E-02 | 3.8E-03 | 2.5E-02 | 2.5E-03 |
| Butyl Benzyl Phthalate ^b | 0.0366 | 0.0349 | 0.000511 | 0.0695 | 0.0695 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.00031 | 0.00029 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Carbazole ^b | 0.751 | 0.964 | 0.00820 | 0.268 | 0.268 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.014 | 0.018 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Dibenzofuran ^b | 0.152 | 0.164 | 0.00166 | 0.157 | 0.157 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0019 | 0.0021 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Di-n-butyl Phthalate | 0.232 | 0.217 | 0.0200 | 0.236 | 0.236 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0046 | 0.0044 | 0.11 | 1.1 | 4.2E-02 | 4.2E-03 | 4.0E-02 | 4.0E-03 |
| VOCs | | | | | | | | | | | · | | | | | | | | |
| 1,2,4-Trimethylbenzene ^b | 3.03 | 2.79 | 0.493 | 0.302 | 0.302 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.080 | 0.075 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| | • | • | | | • | • | | | | • | • | | | 36.4.3 | D / 1/1 777 | 1 05 01 | 1 01 00 | | |
| Notes: | | | | | | | | | | | | | | | Butyltin HI | 1.2E+01 | 1.8E+00 | 1.3E+01 | 2.2E+00 |
| a) Sources listed on Table 3-8ab) These chemicals do not have | | | | | | | | | | | | | | P | esticides HI | 4.6E-03 | 9.1E-04 | 4.6E-03 | 9.2E-04 |

c) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7a.

d) The EPC is the 95% UCL, if calculated (see Table 1-1), or maximum detected concentration (if 95% UCL not calculated).

| % = percent | kg = kilogram |
|--|--|
| = not calculated | kg dry/day = kilograms per day in dry weight |
| AOPC = area of potential concern | L/day = liters per day |
| AUF = Area Use Factor | LOAEL = lowest observed adverse effect level |
| BAF = bioaccumulation factor | mg/kg = milligrams per kilogram |
| BW = body weight | mg/kg-bw/day = milligrams per kilogram body weight per day |
| CPEC = chemical of potential ecological concern | mg/L = milligrams per liter |
| BW = body weight | NA = not available |
| Dose = average daily dose (mg/kg-bw/day) | NOAEL = no observed adverse effect level |
| dw = dry weight | PF = portion of food item |
| EPC = exposure point concentration | SVOC = semivolatile organic compound |
| ft bgs = feet below ground surface | TRV = toxicity reference value |
| HI = (cumulative) hazard index | UCL = upper confidence limit |
| HPAHs = high molecular weight polycyclic aromatic hydrocarbons | VOC = volatile organic compound |
| HQ = hazard quotient | |

Table 3-11b Calculation of Dose and Hazard Quotient for the American Robin - Landfill AOPC

| | | EPCs ^d | | BA | AFs | | Ex | posure Fac | tors | | | Do | oses | TR | RVs | | Hazard | Quotients | |
|-------------------------------------|------------------------------|---------------------------------------|--------------------------------|---|---|---------------------------|---------------------------|----------------------------|----------------|------|---------|---------------------------------|---------------------------------|-------------------------|-------------------------|------------------------------|------------------------------|------------------------------|---------------------------------------|
| | Surface Soil (0-1 ft bgs) | Shallow Soil (0-3 ft bgs) | Concen- tration in Water | Surface Soil BAF invertebrates (kg tissue/kg | Shallow Soil BAF invertebrates (kg tissue/kg | Soil Ingestion Rate | Food Ingestion Rate | Water Ingestion Rate | PF Inverte- | | | Surface Soil Dose (mg/kg- | Shallow Soil Dose (mg/kg- | NOAEL TRV (mg/kg- | LOAEL TRV (mg/kg- | Surface Soil (0-1 ft bgs) | Surface Soil (0-1 ft bgs) | Shallow Soil (0-3 ft bgs) | Shallow Soi (0-3 ft bgs) |
| CPEC | (mg/kg dw) | · · · · · · · · · · · · · · · · · · · | (mg/L) ^c | soil) ^a | soil) ^a | | (kg dry/day) | | brates | AUF | BW (kg) | | bw/day) | bw/day) | bw/dav) | | LOAEL HQ | | |
| Inorganics | | | | | | | | (| | | | j, | | | | | | | · · · · · · · · · · · · · · · · · · · |
| Antimony ^b | 1.36 | 1.48 | 0.0302 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 0.25 | 0.3 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Chromium III | 242 | 594 | 0.000134 | 0.306 | 0.306 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 16 | 40 | 2.66 | 14.5 | 6.1E+00 | 1.1E+00 | 1.5E+01 | 2.7E+00 |
| Copper | 170 | 191 | 4.86 | 0.515 | 0.515 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 18 | 20 | 4.05 | 12.1 | 4.4E+00 | 1.5E+00 | 4.9E+00 | 1.6E+00 |
| Lead | 332 | 511 | 0.369 | 0.262 | 0.241 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 20 | 29 | 1.63 | 3.26 | 1.2E+01 | 6.1E+00 | 1.8E+01 | 8.8E+00 |
| Mercury | 1.57 | 1.45 | 0.0302 | 1.69 | 1.69 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 0.46 | 0.43 | 0.0064 | 0.064 | 7.2E+01 | 7.2E+00 | 6.7E+01 | 6.7E+00 |
| Nickel | 175 | 472 | 2.69 | 1.06 | 1.06 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 33 | 89 | 6.71 | 26.8 | 5.0E+00 | 1.2E+00 | 1.3E+01 | 3.3E+00 |
| Butyltins | | | | | | | | | | | | | | | | | | | |
| Tributyltin | 0.165 | 0.0601 | 0.00136 | 0.042 | 0.042 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 0.0041 | 0.0016 | 6.80 | 16.9 | 6.0E-04 | 2.4E-04 | 2.4E-04 | 9.6E-05 |
| Pesticides | | | | | | | | | | | | | | | | | | | |
| Chlordane (technical) | 1.56 | 1.56 | 0.00462 | 47.7 | 47.7 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 12 | 12 | 2.14 | 10.7 | 5.7E+00 | 1.1E+00 | 5.7E+00 | 1.1E+00 |
| Heptachlor | 0.00283 | 0.00286 | 0.00000685 | 8.75 | 8.75 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 0.0041 | 0.0041 | 0.013 | 0.065 | 3.1E-01 | 6.3E-02 | 3.2E-01 | 6.3E-02 |
| HPAHs ^b | | | | | | | | | | | | | | | | | | | |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 9.68 | 6.38 | 0.00807 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 1.7 | 1.1 | 1.1 | 11 | 1.6E+00 | 1.6E-01 | 1.0E+00 | 1.0E-01 |
| Butyl Benzyl Phthalate ^b | 0.0366 | 0.0349 | 0.000511 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 0.0066 | 0.0063 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Carbazole ^b | 0.751 | 0.964 | 0.00820 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 0.14 | 0.17 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Dibenzofuran ^b | 0.152 | 0.164 | 0.00166 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 0.027 | 0.030 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Di-n-butyl Phthalate | 0.232 | 0.217 | 0.0200 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 0.044 | 0.042 | 0.11 | 1.1 | 4.0E-01 | 4.0E-02 | 3.8E-01 | 3.8E-02 |
| VOCs | - | • | • | • | | - | - | • | | • | • | | • | | • | • | • | • | |
| 1,2,4-Trimethylbenzene ^b | 3.03 | 2.79 | 0.493 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1.00 | 0.0773 | 0.61 | 0.57 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| | | | | | | | | | | | | | | | | | | | |
| Notes: | | | | | | | | | | | | | | | Butyltin HI | 1.0E+02 | 1.7E+01 | 1.2E+02 | 2.3E+01 |
| a) Sources listed on Table 3-9a | for invertebrate | BAFs. | | | | | | | | | | | | Pe | esticides HI | 6.0E+00 | 1.2E+00 | 6.0E+00 | 1.2E+00 |

c) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7a.d) The EPC is the 95% UCL, if calculated (see Table 1-1), or maximum detected concentration (if 95% UCL not calculated).

| = not calculated | kg = kilogram |
|--|--|
| %= percent | kg dry/day = kilograms per day in dry weight |
| AOPC = area of potential concern | L/day = liters per day |
| AUF = Area Use Factor | LOAEL = lowest observed adverse effect level |
| BAF = bioaccumulation factor | mg/kg = milligrams per kilogram |
| BW = body weight | mg/kg-bw/day = milligrams per kilogram body weight per day |
| CPEC = chemical of potential ecological concern | mg/L = milligrams per liter |
| Dose = average daily dose (mg/kg-bw/day) | NA = not available |
| dw = dry weight | NOAEL = no observed adverse effect level |
| EPC = exposure point concentration | PF = portion of food item |
| ft bgs = feet below ground surface | SVOC = semivolatile organic compound |
| HI = (cumulative) hazard index | TRV = toxicity reference value |
| HPAHs = high molecular weight polycyclic aromatic hydrocarbons | UCL = upper confidence limit |
| HQ = hazard quotient | VOC = volatile organic compound |

Table 3-11c Calculation of Dose and Hazard Quotient for the American Kestrel - Landfill AOPC

| | | EPCs ^d | | BA | AFs | | E | xposure Fa | ctors | - | | Do | oses | TF | RVs | | Hazard (| Quotients ^b | |
|--|--|--|---|--|--|---|---|---------------------------------------|---------------|---------|---------|---|---|------------------------------------|------------------------------------|--|--|--|--|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^c | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Shallow Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF mammals | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) LOAEL HQ | Shallow Soil (0-3 ft bgs) NOAEL HQ | Shallow Soil (0-3 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | • | | • | • | | | | | | | | |
| Antimony ^b | 1.36 | 1.48 | 0.0302 | 0.0500 | 0.0500 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.000087 | 0.000093 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Chromium III | 242 | 594 | 0.000134 | 0.0539 | 0.0424 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.013 | 0.027 | 2.66 | 14.5 | 4.8E-03 | 8.9E-04 | 1.0E-02 | 1.8E-03 |
| Copper | 170 | 191 | 0.000520 | 0.0952 | 0.0861 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.014 | 0.015 | 4.05 | 12.1 | 3.5E-03 | 1.2E-03 | 3.6E-03 | 1.2E-03 |
| Lead | 332 | 511 | 0.0000360 | 0.0423 | 0.0333 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.015 | 0.020 | 1.63 | 3.26 | 9.1E-03 | 4.6E-03 | 1.2E-02 | 6.0E-03 |
| Mercury | 1.57 | 1.45 | 0.0302 | 0.0543 | 0.0543 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.00010 | 0.000096 | 0.0064 | 0.064 | 1.6E-02 | 1.6E-03 | 1.5E-02 | 1.5E-03 |
| Nickel | 175 | 472 | 2.69 | 0.0495 | 0.0292 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.010 | 0.018 | 6.71 | 26.8 | 1.5E-03 | 3.9E-04 | 2.7E-03 | 6.8E-04 |
| Butyltins | T | 1 | 1 | r | | | r | 1 | 1 | 1 | 1 | T | 1 | | T | 1 | r | F | |
| Tributyltin | 0.165 | 0.0601 | 0.00136 | 0.000012 | 0.000012 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.0000032 | 0.0000017 | 6.8 | 16.9 | 4.7E-07 | 1.9E-07 | 2.5E-07 | 1.0E-07 |
| Pesticides | 1 4 7 4 | 1.5.5 | 0.004.60 | 0.120 | 0.100 | 0.00000 | 0.015 | 0.014 | | 0.00504 | 0.11.6 | 0.00015 | 0.00015 | 2.1.1 | 10.5 | 0.07.05 | 1 60 0 5 | 0.07.07 | 1 62 05 |
| Chlordane (technical) | 1.56 0.00283 | 1.56 0.00286 | 0.00462 | 0.130 0.0232 | 0.130 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | | 0.00017 | 0.00017 | 2.14 | 10.7 0.065 | 8.0E-05 | 1.6E-05 | 8.0E-05 | 1.6E-05 |
| Heptachlor HPAHs ^b | 0.00283 | 0.00286 | 0.00000685 | 0.0232 | 0.0232 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.000000092 | 0.000000093 | 0.013 | 0.065 | 7.1E-06 | 1.4E-06 | 7.1E-06 | 1.4E-06 |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 9.68 | 6.38 | 0.00807 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.0071 | 0.0047 | 1.1 | 11 | 6.5E-03 | 6.5E-04 | 4.3E-03 | 4.3E-04 |
| | | | | 1 | 1 | | | | 1 | | | | | | | | | | |
| Butyl Benzyl Phthalate ^b | 0.0366 | 0.0349 | 0.000511 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.000027 | 0.000026 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Carbazole ^b | 0.751 | 0.964 | 0.00820 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.00056 | 0.00071 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Dibenzofuran ^b | 0.152 | 0.164 | 0.00166 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.00011 | 0.00012 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Di-n-butyl Phthalate | 0.232 | 0.217 | 0.0200 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.00018 | 0.00017 | 0.11 | 1.1 | 1.7E-03 | 1.7E-04 | 1.6E-03 | 1.6E-04 |
| VOCs | -1 | 1 | 1 | | | | 1 | 1 | 1 | 1 | 1 | 1 | 1 | | 1 | 1 | 1 | | |
| 1,2,4-Trimethylbenzene ^b | 3.03 | 2.79 | 0.493 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.00504 | 0.116 | 0.0025 | 0.0023 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Notes: |)o for more 1 D | A Ec | | | | | | | | | | | | | Butyltin HI | 3.5E-02 | 8.6E-03 | 4.3E-02 | 1.1E-02 |
| a) Sources listed on Table 3-10b) These chemicals do not have | | | walnoted and | itativaly | | | | | | | | | | | esticides HI hthalate HI | 8.7E-05 8.1E-03 | 1.7E-05 8.1E-04 | 8.7E-05 5.8E-03 | 1.7E-05 5.8E-04 |
| i) These chemicals up not nav | e colabilisticu TK | v s, but will be t | valuated qual | itatively. | | | | | | | | | | r | ппатате ПІ | 0.1E-03 | 0.11-04 | J.0E-03 | J.0E-04 |

c) Copper and lead (dissolved) maximum detected concentrations in River OU Forebay surface water; see Table 3-6.

All other CPEC concentrations in water calculated using equilibrium partitioning equations due to lack of River OU analytical data; see Table 3-7a.

d) The EPC is the 95% UCL, if calculated (see Table 1-1), or maximum detected concentration (if 95% UCL not calculated).

| = not calculated | kg dry/day = kilograms per day in dry weight |
|--|--|
| % = percent | L/day = liters per day |
| AOPC = area of potential concern | LOAEL = lowest observed adverse effect level |
| AUF = Area Use Factor | mg/kg = milligrams per kilogram |
| BAF = bioaccumulation factor | mg/kg-bw/day = milligrams per kilogram body weight per day |
| BW = body weight | mg/L = milligrams per liter |
| CPEC = chemical of potential ecological concern | NA = not available |
| Dose = average daily dose (mg/kg-bw/day) | NOAEL = no observed adverse effect level |
| dw = dry weight | OU = operable unit |
| EPC = exposure point concentration | PF = portion of food item |
| ft bgs = feet below ground surface | SVOC = semivolatile organic compound |
| HI = (cumulative) hazard index | TRV = toxicity reference value |
| HPAHs = high molecular weight polycyclic aromatic hydrocarbons | UCL = upper confidence limit |
| HQ = hazard quotient | VOC = volatile organic compound |
| kg = kilogram | |
| | |

Table 3-11d Calculation of Dose and Hazard Quotient for the Vagrant Shrew - Landfill AOPC

| | | EPCs ^c | | BA | AFs | | F | xposure Fa | etors | | | De | ses | TR | Vs | | Hozord | Quotients | |
|---------------------------------|--|--|---|---|---|---|---|---------------------------------------|--------------------------|------|----------|--|--|------------------------------------|------------------------------------|--------------|--------------|--|--------------|
| | | EIUS | | | | | E. | The survey of a | | | | D | 565 | 1 K | V 5 | | | | |
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^b | Surface Soil BAF invertebrates (kg tissue/kg soil) ^a | Shallow Soil BAF invertebrates (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Inverte- brates | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | (0-1 ft bgs) | (0-1 ft bgs) | Shallow Soil (0-3 ft bgs) NOAEL HQ | (0-3 ft bgs) |
| Inorganics | • | | | | | | | - | • | | | | | | | • | | • | |
| Antimony | 1.36 | 1.48 | 0.0302 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 0.28 | 0.31 | 0.059 | 0.59 | 4.8E+00 | 4.8E-01 | 5.2E+00 | 5.2E-01 |
| Chromium III | 242 | 594 | 0.000134 | 0.306 | 0.306 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 16 | 40 | 2.40 | 12.0 | 6.8E+00 | 1.4E+00 | 1.7E+01 | 3.4E+00 |
| Copper | 170 | 191 | 4.86 | 0.515 | 0.515 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 19 | 22 | 5.60 | 9.34 | 3.4E+00 | 2.1E+00 | 3.9E+00 | 2.3E+00 |
| Lead | 332 | 511 | 0.369 | 0.262 | 0.241 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 20 | 28 | 4.70 | 8.90 | 4.2E+00 | 2.2E+00 | 6.0E+00 | 3.2E+00 |
| Mercury | 1.57 | 1.45 | 0.0302 | 1.69 | 1.69 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 0.54 | 0.50 | 0.02 | 0.07 | 2.7E+01 | 7.7E+00 | 2.5E+01 | 7.1E+00 |
| Nickel | 175 | 472 | 2.69 | 1.06 | 1.06 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 38 | 102 | 1.70 | 3.40 | 2.2E+01 | 1.1E+01 | 6.0E+01 | 3.0E+01 |
| Butyltins | 1 | | | | | | | <u>.</u> | <u> </u> | | <u> </u> | | <u> </u> | | | 1 | | 1 | |
| Tributyltin | 0.165 | 0.0601 | 0.00136 | 0.042 | 0.042 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 0.0029 | 0.0012 | 23.4 | 35 | 1.2E-04 | 8.3E-05 | 5.1E-05 | 3.4E-05 |
| Pesticides | • | | | | | | | | | | | | | | | • | | • | |
| Chlordane (technical) | 1.56 | 1.56 | 0.00462 | 47.7 | 47.7 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 15 | 15 | 4.6 | 9.2 | 3.2E+00 | 1.6E+00 | 3.2E+00 | 1.6E+00 |
| Heptachlor | 0.00283 | 0.00286 | 0.00000685 | 8.75 | 8.75 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 0.0049 | 0.0049 | 0.1 | 1 | 4.9E-02 | 4.9E-03 | 4.9E-02 | 4.9E-03 |
| HPAHs | • | | | | | | | | • | | | | | | | • | | • | |
| Benzo(a)anthracene | 8.34 | 10.0 | 0.00471 | 1.59 | 1.59 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 2.7 | 3.2 | *** | *** | 4.3E+00 | 8.7E-01 | 5.2E+00 | 1.0E+00 |
| Benzo(a)pyrene | 8.70 | 11.2 | 0.00148 | 1.33 | 1.33 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 2.3 | 3.0 | *** | *** | 3.8E+00 | 7.6E-01 | 4.9E+00 | 9.8E-01 |
| Benzo(b)fluoranthene | 16.0 | 15.2 | 0.00267 | 2.60 | 2.60 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 8.3 | 7.9 | *** | *** | 1.3E+01 | 2.7E+00 | 1.3E+01 | 2.6E+00 |
| Benzo(g,h,i)perylene | 4.59 | 58.2 | 0.000235 | 2.94 | 2.94 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 2.7 | 34 | *** | *** | 4.4E+00 | 8.7E-01 | 5.5E+01 | 1.1E+01 |
| Benzo(k)fluoranthene | 30.2 | 26.3 | 0.00514 | 2.60 | 2.60 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 16 | 14 | *** | *** | 2.5E+01 | 5.1E+00 | 2.2E+01 | 4.4E+00 |
| Benzofluoranthenes, Total | 14.7 | 31.3 | 0.00250 | 2.60 | 2.60 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 7.6 | 16 | *** | *** | 1.2E+01 | 2.5E+00 | 2.6E+01 | 5.3E+00 |
| Chrysene | 8.40 | 11.0 | 0.00464 | 2.29 | 2.29 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 3.8 | 5 | *** | *** | 6.2E+00 | 1.3E+00 | 8.2E+00 | 1.6E+00 |
| Dibenz(a,h)anthracene | 4.23 | 20.6 | 0.000221 | 2.31 | 2.31 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 2.0 | 10 | *** | *** | 3.2E+00 | 6.4E-01 | 1.5E+01 | 3.1E+00 |
| Fluoranthene | 14.5 | 17.5 | 0.0261 | 3.04 | 3.04 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 8.8 | 11 | *** | *** | 1.4E+01 | 2.9E+00 | 1.7E+01 | 3.4E+00 |
| Indeno(1,2,3-cd)pyrene | 5.04 | 68.0 | 0.000145 | 2.86 | 2.86 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 2.9 | 39 | *** | *** | 4.7E+00 | 9.3E-01 | 6.3E+01 | 1.3E+01 |
| Pyrene | 12.3 | 18.8 | 0.0227 | 1.75 | 1.75 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 4.3 | 6.6 | *** | *** | 7.0E+00 | 1.4E+00 | 1.1E+01 | 2.2E+00 |
| Total HPAHs | 90.2 | 102 | 0.0206 | 2.60 | 2.60 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 47 | 53 | 0.615 | 3.07 | 7.6E+01 | 1.5E+01 | 8.6E+01 | 1.7E+01 |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 9.68 | 6.38 | 0.00807 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 2.0 | 1.3 | 18.3 | 183 | 1.1E-01 | 1.1E-02 | 7.1E-02 | 7.1E-03 |
| Butyl Benzyl Phthalate | 0.0366 | 0.0349 | 0.000511 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 0.0076 | 0.0072 | 159 | 470 | 4.8E-05 | 1.6E-05 | 4.5E-05 | 1.5E-05 |
| Carbazole | 0.751 | 0.964 | 0.00820 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 0.15 | 0.20 | 5 | 50 | 3.1E-02 | 3.1E-03 | 4.0E-02 | 4.0E-03 |
| Dibenzofuran | 0.152 | 0.164 | 0.00166 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 0.031 | 0.034 | 6 | 30 | 5.2E-03 | 1.0E-03 | 5.6E-03 | 1.1E-03 |
| Di-n-butyl Phthalate | 0.232 | 0.217 | 0.0200 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 0.051 | 0.048 | 550 | 1,833 | 9.2E-05 | 2.8E-05 | 8.6E-05 | 2.6E-05 |
| VOCs | | | | | | | | | | | | | | | | | | | |
| 1,2,4-Trimethylbenzene | 3.03 | 2.79 | 0.493 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1.00 | 0.007 | 0.7 | 0.6 | 13 | 65 | 5.4E-02 | 1.1E-02 | 5.0E-02 | 1.0E-02 |
| Notes: | | | | | | | | | | | | | | Metals + 1 | Butyltin HI | 6.9E+01 | 2.5E+01 | 1.2E+02 | 4.7E+01 |
| a) Sources listed on Table 3-9a | for invertebrate | BAFs. | | | | | | | | | | | | | sticides HI | 3.2E+00 | 1.6E+00 | 3.2E+00 | 1.6E+00 |
| b) Analyte concentrations in wa | ater calculated u | sing an equilibi | * | ing equation; see | | | | | | | | | | | nthalate HI | 1.1E-01 | 1.1E-02 | 7.1E-02 | 7.2E-03 |

c) The EPC is the 95% UCL, if calculated (see Table 1-1), or maximum detected concentration (if 95% UCL not calculated).

Bold indicates hazard quotient greater than 1.0.

*** = To be evaluated as Total HPAHs, per EcoSSL; however, HQs for individual HPAHs were calculated using the Total HPAH SLV to determine the primary risk drivers.

Table 3-11d Calculation of Dose and Hazard Quotient for the Vagrant Shrew - Landfill AOPC

% = percent AOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BW = body weight CPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weight EcoSSL = ecological soil screening level EPC = exposure point concentration ft bgs = feet below ground surface HI = (cumulative) hazard index HPAHs = high molecular weight polycyclic aromatic hydrocarbons HQ = hazard quotient kg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per day LOAEL = lowest oberserved adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NOAEL = no oberserved adverse effect level PF = portion of food item SLV = screening level value SVOC = semivolatile organic compound TRV = toxicity reference value UCL = upper confidence limit VOC = volatile organic compound

Table 3-11e Calculation of Dose and Hazard Quotient for the American Mink - Landfill AOPC

| | | EPCs ^c | | B | AFs | | E | xposure Fa | actors | | | De | oses | TR | RVs | | Hazard | Quotients | |
|---------------------------------|--|-------------------|---|---|--|---|---------------------------|----------------------------|------------|------------|-------------|-----------|--|------------------------------------|------------------------------------|--|------------------------------|--|--------------|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | | Concen- tration in Water (mg/L) ^b | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Shallow Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate | Water Ingestion Rate | PF | | BW (kg) | Surface | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) | Shallow Soil (0-3 ft bgs) NOAEL HQ | (0-3 ft bgs) |
| Inorganics | <u> </u> | • | | | • | | <u> </u> | <u> </u> | | | <u> </u> | <u> </u> | <u> </u> | | | • | | - | |
| Antimony | 1.36 | 1.48 | 0.0302 | 0.0500 | 0.0500 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.00076 | 0.00081 | 0.059 | 0.59 | 1.3E-02 | 1.3E-03 | 1.4E-02 | 1.4E-03 |
| Chromium III | 242 | 594 | 0.000134 | 0.0539 | 0.0424 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.10 | 0.24 | 2.40 | 12.0 | 4.1E-02 | 8.2E-03 | 9.9E-02 | 2.0E-02 |
| Copper | 170 | 191 | 0.000520 | 0.0952 | 0.0861 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.073 | 0.081 | 5.60 | 9.34 | 1.3E-02 | 7.8E-03 | 1.4E-02 | 8.7E-03 |
| Lead | 332 | 511 | 0.0000360 | 0.0423 | 0.0333 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.13 | 0.20 | 4.70 | 8.90 | 2.8E-02 | 1.5E-02 | 4.3E-02 | 2.3E-02 |
| Mercury | 1.57 | 1.45 | 0.0302 | 0.0543 | 0.0543 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.00085 | 0.00080 | 0.02 | 0.07 | 4.3E-02 | 1.2E-02 | 4.0E-02 | 1.1E-02 |
| Nickel | 175 | 472 | 2.69 | 0.0495 | 0.0292 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.090 | 0.20 | 1.70 | 3.40 | 5.3E-02 | 2.6E-02 | 1.2E-01 | 6.0E-02 |
| Butyltins | | | | | | | | | | | | | | | | | | | |
| Tributyltin | 0.165 | 0.0601 | 0.00136 | 0.0000124 | 0.0000124 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.000071 | 0.000032 | 23.4 | 35 | 3.0E-06 | 2.0E-06 | 1.4E-06 | 9.2E-07 |
| Pesticides | | | | | • | • | • | | | | • | | | • | | • | • | | |
| Chlordane (technical) | 1.56 | 1.56 | 0.00462 | 0.130 | 0.130 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.00074 | 0.00074 | 4.6 | 9.2 | 1.6E-04 | 8.0E-05 | 1.6E-04 | 8.0E-05 |
| Heptachlor | 0.00283 | 0.00286 | 0.00000685 | 0.0232 | 0.0232 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0000011 | 0.0000012 | 0.1 | 1 | 1.1E-05 | 1.1E-06 | 1.2E-05 | 1.2E-06 |
| HPAHs | - | | | | | - | - | | | | | | | - | | | | - | |
| Benzo(a)anthracene | 8.34 | 10.0 | 0.00000060 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0031 | 0.0037 | *** | *** | 5.1E-03 | 1.0E-03 | 6.0E-03 | 1.2E-03 |
| Benzo(a)pyrene | 8.70 | 11.2 | 0.00148 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | | 0.0033 | 0.0042 | *** | *** | 5.3E-03 | 1.1E-03 | 6.8E-03 | 1.4E-03 |
| Benzo(b)fluoranthene | 16.0 | 15.2 | 0.0000092 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0060 | 0.0057 | *** | *** | 9.7E-03 | 1.9E-03 | 9.2E-03 | 1.8E-03 |
| Benzo(g,h,i)perylene | 4.59 | 58.2 | 0.000235 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0017 | 0.022 | *** | *** | 2.8E-03 | 5.6E-04 | 3.5E-02 | 7.1E-03 |
| Benzo(k)fluoranthene | 30.2 | 26.3 | 0.00514 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.011 | 0.010 | *** | *** | 1.8E-02 | 3.7E-03 | 1.6E-02 | 3.2E-03 |
| Benzofluoranthenes, Total | 14.7 | 31.3 | 0.00250 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0055 | 0.012 | *** | *** | 9.0E-03 | 1.8E-03 | 1.9E-02 | 3.8E-03 |
| Chrysene | 8.40 | 11.0 | 0.00000171 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0031 | 0.0041 | *** | *** | 5.1E-03 | 1.0E-03 | 6.7E-03 | 1.3E-03 |
| Dibenz(a,h)anthracene | 4.23 | 20.6 | 0.000221 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0016 | 0.0077 | *** | *** | 2.6E-03 | 5.1E-04 | 1.3E-02 | 2.5E-03 |
| Fluoranthene | 14.5 | 17.5 | 0.00000784 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0054 | 0.0065 | *** | *** | 8.8E-03 | 1.8E-03 | 1.1E-02 | 2.1E-03 |
| Indeno(1,2,3-cd)pyrene | 5.04 | 68.0 | 0.000145 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0019 | 0.025 | *** | *** | 3.1E-03 | 6.1E-04 | 4.1E-02 | 8.3E-03 |
| Pyrene | 12.3 | 18.8 | 0.0227 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0048 | 0.0072 | *** | *** | 7.7E-03 | 1.5E-03 | 1.2E-02 | 2.3E-03 |
| Total HPAHs | 90.2 | 102 | 0.0206 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.034 | 0.038 | 0.615 | 3.07 | 5.5E-02 | 1.1E-02 | 6.2E-02 | 1.2E-02 |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 9.68 | 6.38 | 0.00807 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | | 0.0094 | 0.0062 | 18.3 | 183 | 5.2E-04 | 5.2E-05 | 3.4E-04 | 3.4E-05 |
| Butyl Benzyl Phthalate | 0.0366 | 0.0349 | 0.000511 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.000039 | 0.000037 | 159 | 470 | 2.5E-07 | 8.3E-08 | 2.4E-07 | 8.0E-08 |
| Carbazole | 0.751 | 0.964 | 0.00820 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | | 0.00079 | 0.00099 | 5 | 50 | 1.6E-04 | 1.6E-05 | 2.0E-04 | 2.0E-05 |
| Dibenzofuran | 0.152 | 0.164 | 0.00166 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | | 0.00016 | 0.00017 | 6 | 30 | 2.7E-05 | 5.3E-06 | 2.8E-05 | 5.7E-06 |
| Di-n-butyl Phthalate | 0.232 | 0.217 | 0.0200 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.00037 | 0.00035 | 550 | 1,833 | 6.7E-07 | 2.0E-07 | 6.4E-07 | 1.9E-07 |
| VOCs | | | | | | | | | | | | | | | | | | | |
| 1,2,4-Trimethylbenzene | 3.03 | 2.79 | 0.493 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.07158 | 0.974 | 0.0064 | 0.0062 | 13 | 65 | 5.0E-04 | 9.9E-05 | 4.8E-04 | 9.6E-05 |
| Notes: | | | | | | | | | | | | | | Metals + | Butyltin HI | 1.9E-01 | 7.1E-02 | 3.3E-01 | 1.2E-01 |
| a) Sources listed on Table 3-10 | a for mammal H | BAFs. | | | | | | | | | | | | Pe | esticides HI | 1.7E-04 | 8.1E-05 | 1.7E-04 | 8.1E-05 |
| b) Copper & lead (dissolved), l | | | • | | | | | | OU Forebay | surface wa | ter; see Ta | able 3-6. | | P | hthalate HI | 5.2E-04 | 5.2E-05 | 3.4E-04 | 3.4E-05 |

All other CPEC concentrations in water calculated using equilibrium partitioning equations due to lack of River OU analytical data; see Table 3-7a.

c) The EPC is the 95% UCL, if calculated (see Table 1-1), or maximum detected concentration (if 95% UCL not calculated).

Bold indicates hazard quotient greater than 1.0.

*** = To be evaluated as Total HPAHs, per EcoSSL; however, HQs for individual HPAHs were calculated using the Total HPAH SLV to determine the primary risk drivers.

Table 3-11e Calculation of Dose and Hazard Quotient for the American Mink - Landfill AOPC

% = percent AOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BW = body weight CPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weight EcoSSL = ecological soil screening level EPC = exposure point concentration ft bgs = feet below ground surface HI = (cumulative) hazard index HPAHs = high molecular weight polycyclic aromatic hydrocarbons HQ = hazard quotient kg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per day LOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram body weight per day mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NOAEL = no observed adverse effect level OU = operable unit PF = portion of food item SLV = screening level value SVOC = semivolatile organic compound TRV = toxicity reference value UCL = upper confidence limit VOC = volatile organic compound

Table 3-12a Calculation of Dose and Hazard Quotient for the Canada Goose – Sandblast Area AOPC

| | | EPCs ^d | | BA | AFs | | Ex | posure Fac | ctors | | | Do | ses | TR | Vs | | Hazard | Quotients | |
|-------------------------------------|--|--|---|--|--|---|---|---------------------------------------|--------------|-----|---------|--|--|------------------------------------|------------------------------------|----------------|--|--|--------------|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^c | Surface Soil BAF plant (kg tissue/kg soil) ^a | Shallow Soil BAF plant (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Plants | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | (0-1 ft bgs) | Surface Soil (0-1 ft bgs) LOAEL HQ | Shallow Soil (0-3 ft bgs) NOAEL HQ | (0-3 ft bgs) |
| Inorganics | | | | | | | | | | | | | | | | | | | |
| Antimony ^b | 2.94 | 2.42 | 0.0653 | 0.0369 | 0.0373 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.021 | 0.017 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Cadmium | 3.20 | 2.63 | 0.0427 | 0.367 | 0.401 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.075 | 0.067 | 1.47 | 5.88 | 5.1E-02 | 1.3E-02 | 4.5E-02 | 1.1E-02 |
| Chromium III | 720 | 579 | 0.000400 | 0.0410 | 0.0410 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 4.5 | 3.6 | 2.66 | 14.5 | 1.7E+00 | 3.1E-01 | 1.4E+00 | 2.5E-01 |
| Lead | 465 | 529 | 0.517 | 0.0179 | 0.0169 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 2.4 | 2.7 | 1.63 | 3.26 | 1.5E+00 | 7.4E-01 | 1.7E+00 | 8.3E-01 |
| Mercury | 0.113 | 0.098 | 0.00217 | 0.650 | 0.650 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0043 | 0.0037 | 0.0064 | 0.064 | 6.8E-01 | 6.8E-02 | 5.9E-01 | 5.9E-02 |
| Nickel | 353 | 251 | 5.43 | 0.0247 | 0.0269 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 2.2 | 1.6 | 6.71 | 26.8 | 3.2E-01 | 8.0E-02 | 2.4E-01 | 6.1E-02 |
| Butyltins | | | | | • | | L | | | | | | | | | • | | | |
| Tributyltin | 0.481 | 0.248 | 0.00398 | 2.04 | 2.04 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.052 | 0.027 | 6.8 | 16.9 | 7.7E-03 | 3.1E-03 | 4.0E-03 | 1.6E-03 |
| Pesticides | | | | | | | | 1 | 1 | | 1 | | | | | | | | |
| BHC (delta) | 0.00154 | 0.00130 | 5.48E-05 | 0.153 | 0.153 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.000021 | 0.000018 | 2 | 20 | 1.0E-05 | 1.0E-06 | 9.0E-06 | 9.0E-07 |
| BHC (gamma) Lindane | 0.00968 | 0.00106 | 0.000344 | 0.268 | 0.268 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.00019 | 0.000033 | 2 | 20 | 9.4E-05 | 9.4E-06 | 1.7E-05 | 1.7E-06 |
| Chlordane (alpha) | 0.000815 | 0.000776 | 1.21E-06 | 0.011 | 0.011 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0000039 | 0.0000038 | 2.14 | 10.7 | 1.8E-06 | 3.7E-07 | 1.8E-06 | 3.5E-07 |
| Chlordane (gamma) | 0.0303 | 0.0204 | 4.49E-05 | 0.010 | 0.010 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.00014 | 0.00010 | 2.14 | 10.7 | 6.7E-05 | 1.3E-05 | 4.6E-05 | 9.1E-06 |
| Endosulfan I | 0.00269 | 0.000809 | 3.98E-05 | 0.231 | 0.231 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.000045 | 0.000015 | 10 | 50 | 4.5E-06 | 9.0E-07 | 1.5E-06 | 2.9E-07 |
| Endosulfan II | 0.000495 | 0.000447 | 7.32E-06 | 0.231 | 0.231 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | | 0.0000075 | 10 | 50 | 8.2E-07 | 1.6E-07 | 7.5E-07 | 1.5E-07 |
| Endosulfan Sulfate | 0.000953 | 0.000889 | 9.68E-06 | 0.290 | 0.290 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.000019 | 0.000017 | 10 | 50 | 1.9E-06 | 3.7E-07 | 1.7E-06 | 3.5E-07 |
| Endrin Aldehyde | 0.00387 | 0.00344 | 0.000118 | 0.063 | 0.063 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.000034 | 0.000030 | 0.01 | 0.1 | 3.4E-03 | 3.4E-04 | 3.0E-03 | 3.0E-04 |
| Endrin Ketone | 0.00309 | 0.00288 | 3.18E-05 | 0.049 | 0.049 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.000022 | 0.000021 | 0.01 | 0.1 | 2.2E-03 | 2.2E-04 | 2.1E-03 | 2.1E-04 |
| Endrin | 0.00504 | 0.00389 | 2.51E-05 | 0.037 | 0.037 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.000032 | 0.000025 | 0.01 | 0.1 | 3.2E-03 | 3.2E-04 | 2.5E-03 | 2.5E-04 |
| Heptachlor | 0.000464 | 0.000415 | 1.12E-06 | 0.011 | 0.011 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0000023 | 0.0000020 | 0.013 | 0.065 | 1.7E-04 | 3.5E-05 | 1.6E-04 | 3.1E-05 |
| Methoxychlor | 0.00105 | 0.00104 | 2.54E-06 | 0.044 | 0.044 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0000069 | 0.0000068 | 125 | 625 | 5.5E-08 | 1.1E-08 | 5.4E-08 | 1.1E-08 |
| HPAHs ^b | | | • | | | | • | • | | | | | | | • | | | • | |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 98.0 | 77.1 | 0.0817 | 0.00150 | 0.00150 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.42 | 0.33 | 1.1 | 11 | 3.8E-01 | 3.8E-02 | 3.0E-01 | 3.0E-02 |
| Butyl Benzyl Phthalate ^b | 0.0171 | 0.0152 | 0.000239 | 0.0695 | 0.0695 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.00014 | 0.00013 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Carbazole ^b | 0.144 | 0.112 | 0.00157 | 0.268 | 0.268 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0026 | 0.0021 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Dibenzofuran ^b | 0.0424 | 0.122 | 0.000463 | 0.157 | 0.157 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.00054 | 0.0015 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Di-n-butyl Phthalate | 0.0424 | 0.122 | 0.000403 | 0.137 | 0.137 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.00034 | 0.0013 | 0.11 | 1.1 | 1.4E-02 | 1.4E-03 | 1.1E-02 | 1.1E-03 |
| Di-n-octyl Phthalate ^b | 0.0251 | 0.0225 | 0.0000178 | 0.000772 | 0.000772 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.00011 | 0.00012 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| VOCs | 0.0231 | 0.0223 | 0.0000170 | 0.000772 | 0.000772 | 0.015 | 0.15 | 0.12 | 1 | 1 | 5.0 | 0.00011 | 0.00010 | 1 17 1 | 101 | 110 111 1 | 110 110 1 | nont | |
| 1,2,4-Trimethylbenzene ^b | 0.000230 | 4.62 | 0.0000375 | 0.302 | 0.302 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0000061 | 0.091 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| 4-Isopropyltoluene | 0.000230 | 0.0201 | 0.0000375 | 0.302 | 0.302 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0000081 | 0.001 | NA 3.16 | 15.8 | 6.8E-06 | 1.4E-06 | 8.1E-05 | 1.6E-05 |
| n-Propylbenzene ^b | | | | | | | 0.15 | 0.12 | 1 | 1 | | | | | | | | | |
| п-г юруюендене | 0.000150 | 0.0355 | 0.0000185 | 0.279 | 0.279 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0000035 | 0.00066 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Notes: | | | | | | | | | | | | | | | Butyltin HI | 4.2E+00 | 1.2E+00 | 3.9E+00 | 1.2E+00 |
| a) Sources listed on Table 3-8b | * | | | | | | | | | | | | | | esticides HI | 9.1E-03 | 9.4E-04 | 7.8E-03 | 8.0E-04 |
| b) These chemicals do not have | e established TR | Vs, but will be | | alitatively. | | | | | | | | | | P | hthalate HI | 4.0E-01 | 4.0E-02 | 3.1E-01 | 3.1E-02 |

c) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7b.d) The EPC is the 95% UCL, if calculated (see Table 1-2), or maximum detected concentration (if 95% UCL not calculated).

Table 3-12a Calculation of Dose and Hazard Quotient for the Canada Goose – Sandblast Area AOPC

-- = not calculated % = percentAOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BHC = benzene hexachloride BW = body weightCPEC = chemical of potential ecological concern BW = body weightDose = average daily dose (mg/kg-bw/day) dw = dry weightEPC = exposure point concentration ft bgs = feet below ground surface HI = (cumulative) hazard index HPAHs = high molecular weight polycyclic aromatic hydrocarbons HQ = hazard quotientkg = kilogram

kg dry/day = kilograms per day in dry weight L/day = liters per day LOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NA = not available NOAEL = no observed adverse effect level PF = portion of food item SVOC = semivolatile organic compound TRV = toxicity reference value UCL = upper confidence limit VOC = volatile organic compound

Table 3-12b Calculation of Dose and Hazard Quotient for the American Robin – Sandblast Area AOPC

| | | EPCs ^d | | BA | AFs | | Ex | posure Fac | tors | | | Do | oses | TR | Vs | | Hazard | Quotients | |
|---|--|--|---|---|---|---|---|---------------------------------------|--------------------------|-----|---------|--|--|------------------------------------|------------------------------------|--|--|--|--|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^c | Surface Soil BAF invertebrates (kg tissue/kg soil) ^a | Shallow Soil BAF invertebrates (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Inverte- brates | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) LOAEL HQ | Shallow Soil (0-3 ft bgs) NOAEL HQ | Shallow Soil (0-3 ft bgs) LOAEL HQ |
| Inorganics | - | | | • | | | | | | | | | | | | 1 | • | • | |
| Antimony ^b | 2.94 | 2.42 | 0.0653 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.54 | 0.44 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Cadmium | 3.20 | 2.63 | 0.0427 | 6.52 | 6.79 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 3.4 | 2.9 | 1.47 | 5.88 | 2.3E+00 | 5.9E-01 | 2.0E+00 | 5.0E-01 |
| Chromium III | 720 | 579 | 0.000400 | 0.306 | 0.306 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 48 | 39 | 2.66 | 14.5 | 1.8E+01 | 3.3E+00 | 1.4E+01 | 2.7E+00 |
| Lead | 465 | 529 | 0.517 | 0.246 | 0.240 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 26 | 30 | 1.63 | 3.26 | 1.6E+01 | 8.1E+00 | 1.8E+01 | 9.1E+00 |
| Mercury | 0.113 | 0.098 | 0.00217 | 1.69 | 1.69 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.033 | 0.029 | 0.0064 | 0.064 | 5.2E+00 | 5.2E-01 | 4.5E+00 | 4.5E-01 |
| Nickel | 353 | 251 | 5.43 | 1.06 | 1.06 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 67 | 48 | 6.71 | 26.8 | 1.0E+01 | 2.5E+00 | 7.2E+00 | 1.8E+00 |
| Butyltins | - | - | • | • | | • | - | | | | | • | • | | • | | • | • | • |
| Tributyltin | 0.481 | 0.248 | 0.00398 | 0.0424 | 0.0424 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.012 | 0.0064 | 6.80 | 16.9 | 1.8E-03 | 7.1E-04 | 9.5E-04 | 3.8E-04 |
| Pesticides | • | | | | | | • | • | | | | • | • | | • | | | | |
| BHC (delta) | 0.00154 | 0.00130 | 5.48E-05 | 8.89 | 8.89 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.0023 | 0.0019 | 2 | 20 | 1.1E-03 | 1.1E-04 | 9.5E-04 | 9.5E-05 |
| BHC (gamma) Lindane | 0.00968 | 0.00106 | 0.000344 | 3.83 | 3.83 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.0062 | 0.00072 | 2 | 20 | 3.1E-03 | 3.1E-04 | 3.6E-04 | 3.6E-05 |
| Chlordane (alpha) | 0.000815 | 0.000776 | 1.21E-06 | 18.8 | 18.8 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.0025 | 0.0024 | 2.14 | 10.7 | 1.2E-03 | 2.3E-04 | 1.1E-03 | 2.2E-04 |
| Chlordane (gamma) | 0.0303 | 0.0204 | 4.49E-05 | 23.9 | 23.9 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.12 | 0.079 | 2.14 | 10.7 | 5.5E-02 | 1.1E-02 | 3.7E-02 | 7.4E-03 |
| Endosulfan I | 0.00269 | 0.000809 | 3.98E-05 | 1.99 | 1.99 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.00092 | 0.00028 | 10 | 50 | 9.2E-05 | 1.8E-05 | 2.8E-05 | 5.6E-06 |
| Endosulfan II | 0.000495 | 0.000447 | 7.32E-06 | 1.99 | 1.99 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.00017 | 0.00015 | 10 | 50 | 1.7E-05 | 3.4E-06 | 1.5E-05 | 3.1E-06 |
| Endosulfan Sulfate | 0.000953 | 0.000889 | 9.68E-06 | 0.970 | 0.970 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.00017 | 0.00016 | 10 | 50 | 1.7E-05 | 3.3E-06 | 1.6E-05 | 3.1E-06 |
| Endrin Aldehyde | 0.00387 | 0.00344 | 0.000118 | 28.7 | 28.7 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.018 | 0.016 | 0.01 | 0.1 | 1.8E+00 | 1.8E-01 | 1.6E+00 | 1.6E-01 |
| Endrin Ketone | 0.00309 | 0.00288 | 3.18E-05 | 14.1 | 14.1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.0071 | 0.0066 | 0.01 | 0.1 | 7.1E-01 | 7.1E-02 | 6.6E-01 | 6.6E-02 |
| Endrin | 0.00504 | 0.00389 | 2.51E-05 | 10.4 | 10.4 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.0086 | 0.0066 | 0.01 | 0.1 | 8.6E-01 | 8.6E-02 | 6.6E-01 | 6.6E-02 |
| Heptachlor | 0.000464 | 0.000415 | 1.12E-06 | 30.7 | 30.7 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.0023 | 0.0021 | 0.013 | 0.065 | 1.8E-01 | 3.6E-02 | 1.6E-01 | 3.2E-02 |
| Methoxychlor | 0.00105 | 0.00104 | 2.54E-06 | 3.98 | 3.98 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.00070 | 0.00069 | 125 | 625 | 5.6E-06 | 1.1E-06 | 5.5E-06 | 1.1E-06 |
| HPAHs ^b | | | | | | | | | | | | | | | | | | | |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 98.0 | 77.1 | 0.0817 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 18 | 14 | 1.1 | 11 | 1.6E+01 | 1.6E+00 | 1.3E+01 | 1.3E+00 |
| Butyl Benzyl Phthalate ^b | 0.0171 | 0.0152 | 0.000239 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.0031 | 0.0028 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Carbazole ^b | 0.144 | 0.112 | 0.00157 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.026 | 0.020 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Dibenzofuran ^b | 0.0424 | 0.122 | 0.000463 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.0077 | 0.020 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Dienzoluran Di-n-butyl Phthalate | 0.0424 | 0.122 | 0.000463 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.0077 | 0.022 | 0.11 | 1.1 | 1.4E-01 | 1.4E-02 | 9.7E-02 | 9.7E-03 |
| | | | | 1 | 1 | | | 0.011 | 1 | 1 | | | | 0111 | | | | | |
| Di-n-octyl Phthalate ^b VOCs | 0.0251 | 0.0225 | 0.0000178 | | | 0.0013 | 0.013 | 0.011 | 1 | | 0.0773 | 0.0045 | 0.0040 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| | 1 | 1 | | | | 1 | 1 | 1 | | 1 | 1 | 1 | 1 | | 1 | | 1 | | T |
| 1,2,4-Trimethylbenzene ^b | 0.000230 | 4.62 | 0.0000375 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.000046 | 0.83 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| 4-Isopropyltoluene | 0.00133 | 0.0201 | 0.000119 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.00025 | 0.0036 | 3 | 15.8 | 8.1E-05 | 1.6E-05 | 1.1E-03 | 2.3E-04 |
| n-Propylbenzene ^b | 0.000150 | 0.0355 | 0.0000185 | 1 | 1 | 0.0013 | 0.013 | 0.011 | 1 | 1 | 0.0773 | 0.000029 | 0.0064 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Notes: | | | | | | | | | | | | | | Metals + | Butyltin HI | 5.2E+01 | 1.5E+01 | 4.6E+01 | 1.4E+01 |
| a) Sources listed on Table 3-9h | o for invertebrate | BAFs. | | | | | | | | | | | | | esticides HI | 3.6E+00 | 3.9E-01 | 3.1E+00 | 3.3E-01 |
| b) These chemicals do not have | | | 1 . 1 | 1 1 | | | | | | | | | | | hthalate HI | 1.6E+01 | 1.6E+00 | 1.3E+01 | 1.3E+00 |

c) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7b.

d) The EPC is the 95% UCL, if calculated (see Table 1-2), or maximum detected concentration (if 95% UCL not calculated).

Table 3-12b Calculation of Dose and Hazard Quotient for the American Robin – Sandblast Area AOPC

-- = not calculated %= percent AOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BHC = benzene hexachloride BW = body weight CPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weight EPC = exposure point concentration ft bgs = feet below ground surface HI = (cumulative) hazard index HPAHs = high molecular weight polycyclic aromatic hydrocarbons HQ = hazard quotient

kg = kilogram

kg dry/day = kilograms per day in dry weight L/day = liters per day LOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NA = not available NOAEL = no observed adverse effect level PF = portion of food item SVOC = semivolatile organic compound TRV = toxicity reference value UCL = upper confidence limit VOC = volatile organic compound

Table 3-12c Calculation of Dose and Hazard Quotient for the American Kestrel – Sandblast Area AOPC

| | | EPCs ^d | | B | AFs | | E | xposure Fa | ctors | | | Do | ses | TI | RVs | | Hazard | Quotients | |
|-------------------------------------|--|--|---|---|--|---|---|---------------------------------------|---------------|---------|----------|---|--|------------------------------------|------------------------------------|--|--|--|--------------|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^c | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Shallow Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF mammals | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) LOAEL HQ | Shallow Soil (0-3 ft bgs) NOAEL HQ | (0-3 ft bgs) |
| Inorganics | | | | I | | | | | | | | | | | | | | | |
| Antimony ^b | 2.94 | 2.42 | 0.0653 | 0.050 | 0.050 | 0.00033 | 0.017 | 0.014 | 1 | 0.01148 | 0.116 | 0.00043 | 0.00037 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Cadmium | 3.20 | 2.63 | 0.0000100 | 0.154 | 0.171 | 0.00033 | 0.017 | 0.014 | 1 | 0.01148 | | 0.00091 | 0.0008 | 1.47 | 5.88 | 6.2E-04 | 1.6E-04 | 5.6E-04 | 1.4E-04 |
| Chromium III | 720 | 579 | 0.000400 | 0.0403 | 0.0427 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.071 | 0.060 | 2.66 | 14.5 | 2.7E-02 | 4.9E-03 | 2.2E-02 | 4.1E-03 |
| Lead | 465 | 529 | 0.0000360 | 0.0351 | 0.0327 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.042 | 0.046 | 1.63 | 3.26 | 2.6E-02 | 1.3E-02 | 2.8E-02 | 1.4E-02 |
| Mercury | 0.113 | 0.098 | 0.00217 | 0.0543 | 0.0543 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.000017 | 0.000015 | 0.0064 | 0.064 | 2.6E-03 | 2.6E-04 | 2.3E-03 | 2.3E-04 |
| Nickel | 353 | 251 | 5.43 | 0.0340 | 0.0408 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.039 | 0.033 | 6.71 | 26.8 | 5.8E-03 | 1.4E-03 | 4.8E-03 | 1.2E-03 |
| Butyltins | | | • | | • | | • | • | • | • | ÷ | • | | | • | • | | | |
| Tributyltin | 0.481 | 0.248 | 0.00398 | 0.0000124 | 0.0000124 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.000021 | 0.000014 | 6.8 | 16.9 | 3.1E-06 | 1.3E-06 | 2.0E-06 | 8.1E-07 |
| Pesticides | • | • | • | • | | • | L | | | | <u>.</u> | • | | | | | • | • | |
| BHC (delta) | 0.00154 | 0.00130 | 5.48E-05 | 0.00108 | 0.00108 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.00000013 | 0.00000012 | 2 | 20 | 6.4E-08 | 6.4E-09 | 6.0E-08 | 6.0E-09 |
| BHC (gamma) Lindane | 0.00968 | 0.00106 | 0.000344 | 0.000412 | 0.000412 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | | 0.00000051 | 2 | 20 | 4.0E-07 | 4.0E-08 | 2.6E-07 | 2.6E-08 |
| Chlordane (alpha) | 0.000815 | 0.000776 | 1.21E-06 | 0.0988 | 0.0988 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | | 0.00000015 | 2.14 | 10.7 | 7.5E-08 | 1.5E-08 | 7.1E-08 | 1.4E-08 |
| Chlordane (gamma) | 0.0303 | 0.0204 | 4.49E-05 | 0.130 | 0.130 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.0000075 | 0.0000051 | 2.14 | 10.7 | 3.5E-06 | 7.0E-07 | 2.4E-06 | 4.8E-07 |
| Endosulfan I | 0.00269 | 0.000809 | 3.98E-05 | 0.000531 | 0.000531 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | | 0.00000082 | 10 | 50 | 1.5E-08 | 2.9E-09 | 8.2E-09 | 1.6E-09 |
| Endosulfan II | 0.000495 | 0.000447 | 7.32E-06 | 0.000531 | 0.000531 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.00000027 | | 10 | 50 | 2.7E-09 | 5.4E-10 | 2.5E-09 | 5.0E-10 |
| Endosulfan Sulfate | 0.000953 | 0.000889 | 9.68E-06 | 0.000359 | 0.000359 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.00000045 | | 10 | 50 | 4.5E-09 | 9.0E-10 | 4.3E-09 | 8.6E-10 |
| Endrin Aldehyde | 0.00387 | 0.00344 | 0.000118 | 0.00495 | 0.00495 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.0000032 | | 0.01 | 0.1 | 3.2E-05 | 3.2E-06 | 3.0E-05 | 3.0E-06 |
| Endrin Ketone | 0.00309 | 0.00288 | 3.18E-05 | 0.00767 | 0.00767 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.00000018 | | 0.01 | 0.1 | 1.8E-05 | 1.8E-06 | 1.7E-05 | 1.7E-06 |
| Endrin | 0.00504 | 0.00389 | 2.51E-05 | 0.0124 | 0.0124 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.0000030 | | 0.01 | 0.1 | 3.0E-05 | 3.0E-06 | 2.4E-05 | 2.4E-06 |
| Heptachlor | 0.000464 | 0.000415 | 1.12E-06 | 0.0988 | 0.0988 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.00000092 | 0.00000082 | 0.013 | 0.065 | 7.1E-06 | 1.4E-06 | 6.3E-06 | 1.3E-06 |
| Methoxychlor | 0.00105 | 0.00104 | 2.54E-06 | 0.00944 | 0.00944 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | | 0.00000054 | | 125 | 625 | 4.3E-10 | 8.7E-11 | 4.3E-10 | 8.6E-11 |
| HPAHs ^b | | | <u>.</u> | I | | | I | 1 | | | <u> </u> | | 1 | | <u> </u> | | I. | I. | |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 98.0 | 77.1 | 0.0817 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.16 | 0.13 | 1.1 | 11 | 1.5E-01 | 1.5E-02 | 1.2E-01 | 1.2E-02 |
| Butyl Benzyl Phthalate ^b | 0.0171 | 0.0152 | 0.000239 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.000029 | 0.000026 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Carbazole ^b | 0.144 | 0.112 | 0.00157 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.00024 | 0.00019 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Dibenzofuran ^b | 0.0424 | 0.112 | 0.000463 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.000072 | 0.00019 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Di-n-butyl Phthalate | 0.0424 | 0.122 | 0.000403 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0113 | 0.116 | 0.00012 | 0.000100 | 0.11 | 1.1 | 1.3E-03 | 1.3E-04 | 9.1E-04 | 9.1E-05 |
| Di-n-octyl Phthalate ^b | 0.0304 | 0.0225 | 0.0000178 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | | | 0.00014 | 0.000100 | NA | | No TRV | | No TRV | No TRV |
| VOCs | 0.0231 | 0.0225 | 0.0000178 | 1 | 1 | 0.00055 | 0.017 | 0.014 | | 0.0115 | 0.116 | 0.0000 | 0.0000 | INA | NA | NOTKV | No TRV | NOTKV | NOTKV |
| | 0.000 | 1.10 | | | | | - | 0.044 | | 0.0447 | | | | | | | | | |
| 1,2,4-Trimethylbenzene ^b | 0.000230 | 4.62 | 0.0000375 | 1 | 1 | 0.00033 | 0.017 | 0.014 | | 0.0115 | 0.116 | 0.00000044 | 0.0077 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| 4-Isopropyltoluene | 0.00133 | 0.0201 | 0.000119 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.0000024 | 0.000034 | 3 | 15.8 | 7.6E-07 | 1.5E-07 | 1.1E-05 | 2.1E-06 |
| n-Propylbenzene ^b | 0.000150 | 0.0355 | 0.0000185 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0115 | 0.116 | 0.0000028 | 0.000059 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Notes: | | | | | | | | | | | | | | Metals + | Butyltin HI | 6.2E-02 | 2.0E-02 | 5.8E-02 | 2.0E-02 |
| a) Sources listed on Table 3-10 |)b for mammal H | BAFs. | | | | | | | | | | | | | esticides HI | 9.2E-05 | 1.0E-05 | 8.1E-05 | 9.0E-06 |
| b) These chemicals do not have | e established TR | Vs. but will be | evaluated qua | alitatively. | | | | | | | | | | | hthalate HI | 1.5E-01 | 1.5E-02 | 1.2E-01 | 1.2E-02 |

c) Cadmium and lead (dissolved) maximum detected concentrations in River OU Forebay surface water; see Table 3-6.

All other CPEC concentrations in water calculated using equilibrium partitioning equations due to lack of River OU analytical data; see Table 3-7b.

d) The EPC is the 95% UCL, if calculated (see Table 1-2), or maximum detected concentration (if 95% UCL not calculated).

Table 3-12c

Calculation of Dose and Hazard Quotient for the American Kestrel – Sandblast Area AOPC

-- = not calculated % = percent AOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BHC = benzene hexachloride BW = body weight CPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weight EPC = exposure point concentration ft bgs = feet below ground surface HI = (cumulative) hazard index HPAHs = high molecular weight polycyclic aromatic hydrocarbons HQ = hazard quotient kg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per day LOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NA = not available NOAEL = no observed adverse effect level OU = operable unit PF = portion of food item SVOC = semivolatile organic compound TRV = toxicity reference value UCL = upper confidence limit VOC = volatile organic compound

 Table 3-12d

 Calculation of Dose and Hazard Quotient for the Vagrant Shrew – Sandblast Area AOPC

| | | EPCs ^c | | BA | Fs | | Б | xposure Fa | ctors | | | Do | oses | TR | Vs | | Hazard | Quotients | |
|-----------------------------|--------------|-------------------|---------------------|--------------------|--------------------|--------------|-----------|------------|----------|-----|---------|-----------|-----------|---------|---------|--------------|--------------|--------------|--------------|
| | | | Concen- | Surface Soil | Shallow Soil | Soil | Food | Water | | | | Surface | Shallow | NOAEL | LOAEL | | | | |
| CPEC | | Shallow Soil | tration in | BAF | BAF | Ingestion | Ingestion | Ingestion | PF | | | Soil Doco | Soil Dose | TRV | TRV | | Surface Soil | Shallow Soil | Shallow Soil |
| | (0-1 ft bgs) | (0-3 ft bgs) | Water | invertebrates | invertebrates | Rate | Rate | Rate | Inverte- | AUF | BW (kg) | (mg/kg- | (mg/kg- | (mg/kg- | (mg/kg- | (0-1 ft bgs) | (0-1 ft bgs) | (0-3 ft bgs) | (0-3 ft bgs) |
| | (mg/kg dw) | (mg/kg dw) | (mg/L) ^b | (kg tissue/kg | (kg tissue/kg | (kg dry/day) | | (L/day) | brates | | | bw/day) | bw/day) | bw/day) | bw/day) | NOAEL HQ | LOAEL HQ | NOAEL HQ | LOAEL HQ |
| | | | (g,) | soil) ^a | soil) ^a | | | ~ • / | | | | | • | • | | | | | |
| Inorganics | | | • | | | | | • | • | | | | | | • | | | | |
| Antimony | 2.94 | 2.42 | 0.0653 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.61 | 0.50 | 0.059 | 0.59 | 1.0E+01 | 1.0E+00 | 8.6E+00 | 8.6E-01 |
| Cadmium | 3.20 | 2.63 | 0.0427 | 6.52 | 6.79 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 4.1 | 3.5 | 0.770 | 7.70 | 5.4E+00 | 5.4E-01 | 4.6E+00 | 4.6E-01 |
| Chromium III | 720 | 579 | 0.000400 | 0.306 | 0.306 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 49 | 39 | 2.40 | 12.0 | 2.0E+01 | 4.1E+00 | 1.6E+01 | 3.3E+00 |
| Lead | 465 | 529 | 0.517 | 0.246 | 0.240 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 26 | 29 | 4.70 | 8.90 | 5.6E+00 | 2.9E+00 | 6.2E+00 | 3.3E+00 |
| Mercury | 0.113 | 0.098 | 0.00217 | 1.69 | 1.69 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.039 | 0.033 | 0.02 | 0.07 | 1.9E+00 | 5.5E-01 | 1.7E+00 | 4.8E-01 |
| Nickel | 353 | 251 | 5.43 | 1.06 | 1.06 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 77 | 55 | 1.70 | 3.40 | 4.5E+01 | 2.3E+01 | 3.2E+01 | 1.6E+01 |
| Butyltins | | | • | | | | - | - | - | | | - | | | | - | | - | |
| Tributyltin | 0.481 | 0.248 | 0.00398 | 0.0424 | 0.0424 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.0084 | 0.0047 | 23.4 | 35 | 3.6E-04 | 2.4E-04 | 2.0E-04 | 1.3E-04 |
| Pesticides | | | • | • | | • | | • | | | | | | | • | | • | | • |
| BHC (delta) | 0.00154 | 0.00130 | 5.48E-05 | 8.89 | 8.89 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.0027 | 0.0023 | 8 | 40 | 3.4E-04 | 6.8E-05 | 2.9E-04 | 5.7E-05 |
| BHC (gamma) Lindane | 0.00968 | 0.00106 | 0.000344 | 3.83 | 3.83 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.0074 | 0.00086 | 8 | 40 | 9.3E-04 | 1.9E-04 | 1.1E-04 | 2.2E-05 |
| Chlordane (alpha) | 0.000815 | 0.000776 | 1.21E-06 | 18.8 | 18.8 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.0030 | 0.0029 | 4.6 | 9.2 | 6.5E-04 | 3.3E-04 | 6.2E-04 | 3.1E-04 |
| Chlordane (gamma) | 0.0303 | 0.0204 | 4.49E-05 | 23.9 | 23.9 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.142 | 0.096 | 4.6 | 9.2 | 3.1E-02 | 1.5E-02 | 2.1E-02 | 1.0E-02 |
| Endosulfan I | 0.00269 | 0.000809 | 3.98E-05 | 1.99 | 1.99 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.0011 | 0.00033 | 0.15 | 0.75 | 7.2E-03 | 1.4E-03 | 2.2E-03 | 4.4E-04 |
| Endosulfan II | 0.000495 | 0.000447 | 7.32E-06 | 1.99 | 1.99 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.0002 | 0.0002 | 0.15 | 0.75 | 1.3E-03 | 2.6E-04 | 1.2E-03 | 2.4E-04 |
| Endosulfan Sulfate | 0.000953 | 0.000889 | 9.68E-06 | 0.970 | 0.970 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.00019 | 0.00018 | 0.15 | 0.75 | 1.3E-03 | 2.5E-04 | 1.2E-03 | 2.4E-04 |
| Endrin Aldehyde | 0.00387 | 0.00344 | 0.000118 | 28.7 | 28.7 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.022 | 0.019 | 0.092 | 0.92 | 2.4E-01 | 2.4E-02 | 2.1E-01 | 2.1E-02 |
| Endrin Ketone | 0.00309 | 0.00288 | 3.18E-05 | 14.1 | 14.1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.009 | 0.008 | 0.092 | 0.92 | 9.3E-02 | 9.3E-03 | 8.7E-02 | 8.7E-03 |
| Endrin | 0.00504 | 0.00389 | 2.51E-05 | 10.4 | 10.4 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.010 | 0.008 | 0.092 | 0.92 | 1.1E-01 | 1.1E-02 | 8.7E-02 | 8.7E-03 |
| Heptachlor | 0.000464 | 0.000415 | 1.12E-06 | 30.7 | 30.7 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.00280 | 0.00250 | 0.1 | 1 | 2.8E-02 | 2.8E-03 | 2.5E-02 | 2.5E-03 |
| Methoxychlor | 0.00105 | 0.00104 | 2.54E-06 | 3.98 | 3.98 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.00083 | 0.00082 | 4 | 8 | 2.1E-04 | 1.0E-04 | 2.1E-04 | 1.0E-04 |
| HPAHs | | | | | | | | | | | | | | | | | | | |
| Benzo(a)anthracene | 1.75 | 2.45 | 0.000989 | 1.59 | 1.59 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.56 | 0.78 | *** | *** | 9.1E-01 | 1.8E-01 | 1.3E+00 | 2.6E-01 |
| Benzo(a)pyrene | 1.76 | 2.39 | 0.000300 | 1.33 | 1.33 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.47 | 0.64 | *** | *** | 7.7E-01 | 1.5E-01 | 1.0E+00 | 2.1E-01 |
| Benzo(b)fluoranthene | 1.95 | 2.21 | 0.000326 | 2.60 | 2.60 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 1.0 | 1.1 | *** | *** | 1.6E+00 | 3.3E-01 | 1.9E+00 | 3.7E-01 |
| Benzo(g,h,i)perylene | 1.07 | 1.05 | 5.49E-05 | 2.94 | 2.94 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.63 | 0.61 | *** | *** | 1.0E+00 | 2.0E-01 | 1.0E+00 | 2.0E-01 |
| Benzo(k)fluoranthene | 0.676 | 0.764 | 0.000115 | 2.60 | 2.60 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.35 | 0.40 | *** | *** | 5.7E-01 | 1.1E-01 | 6.4E-01 | 1.3E-01 |
| Benzofluoranthenes, Total | 7.76 | 10.6 | 0.00132 | 2.60 | 2.60 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 4.0 | 5.5 | *** | *** | 6.5E+00 | 1.3E+00 | 8.9E+00 | 1.8E+00 |
| Chrysene | 2.07 | 4.59 | 0.00114 | 2.29 | 2.29 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.95 | 2.1 | *** | *** | 1.5E+00 | 3.1E-01 | 3.4E+00 | 6.8E-01 |
| Dibenz(a,h)anthracene | 0.244 | 0.238 | 1.28E-05 | 2.31 | 2.31 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.11 | 0.11 | *** | *** | 1.8E-01 | 3.7E-02 | 1.8E-01 | 3.6E-02 |
| Fluoranthene | 4.64 | 10.8 | 0.00836 | 3.04 | 3.04 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 2.8 | 6.5 | *** | *** | 4.6E+00 | 9.1E-01 | 1.1E+01 | 2.1E+00 |
| Indeno(1,2,3-cd)pyrene | 1.13 | 1.16 | 3.26E-05 | 2.86 | 2.86 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.64 | 0.66 | *** | *** | 1.0E+00 | 2.1E-01 | 1.1E+00 | 2.2E-01 |
| Pyrene | 4.78 | 11.7 | 0.00880 | 1.75 | 1.75 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 1.7 | 4.1 | *** | *** | 2.7E+00 | 5.5E-01 | 6.7E+00 | 1.3E+00 |
| Total HPAHs | 18.6 | 40.8 | 0.00425 | 2.60 | 2.60 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 9.6 | 21 | 0.615 | 3.07 | 1.6E+01 | 3.1E+00 | 3.4E+01 | 6.9E+00 |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 98.0 | 77.1 | 0.0817 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 20 | 16 | 18.3 | 183 | 1.1E+00 | 1.1E-01 | 8.6E-01 | 8.6E-02 |
| Butyl Benzyl Phthalate | 0.0171 | 0.0152 | 0.000239 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.004 | 0.003 | 159 | 470 | 2.2E-05 | 7.5E-06 | 2.0E-05 | 6.7E-06 |
| Carbazole | 0.144 | 0.112 | 0.00157 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.030 | 0.023 | 5 | 50 | 5.9E-03 | 5.9E-04 | 4.6E-03 | 4.6E-04 |
| Dibenzofuran | 0.0424 | 0.122 | 0.000463 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.0087 | 0.025 | 6 | 30 | 1.5E-03 | 2.9E-04 | 4.2E-03 | 8.3E-04 |
| Di-n-butyl Phthalate | 0.0804 | 0.0543 | 0.00693 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.018 | 0.012 | 550 | 1,833 | 3.2E-05 | 9.6E-06 | 2.2E-05 | 6.7E-06 |
| Di-n-octyl Phthalate | 0.0251 | 0.0225 | 0.0000178 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.0051 | 0.0046 | 7,500 | 37,500 | 6.8E-07 | 1.4E-07 | 6.1E-07 | 1.2E-07 |
| VOCs | | | | | | | | | | | | | | | | | | | |
| 1,2,4-Trimethylbenzene | 0.000230 | 4.62 | 0.0000375 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.000053 | 0.94 | 13 | 65 | 4.1E-06 | 8.2E-07 | 7.3E-02 | 1.5E-02 |

Table 3-12d Calculation of Dose and Hazard Quotient for the Vagrant Shrew – Sandblast Area AOPC

| | | EPCs^c | | BA | AFs | | E | xposure Fa | ctors | | | Do | oses | TR | RVs | | Hazard | Quotients | |
|--|-------------------|--|-----------|---------------|---------------------------------------|---------------------------|---|---------------------------------------|--------------------------|-----|---------|----------|--|------------------------------------|---|--------------|--|---|--|
| CPEC | (0-1 ft bgs) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Water | invertebrates | BAF invertebrates (kg tissne/kg | Soil Ingestion Rote | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Inverte- brates | AUF | BW (kg) | (mg/kg- | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | (0-1 ft bgs) | (0-1 ft bgs) | · · · · · · · · · · · · · · · · · · · | |
| 4-Isopropyltoluene | 0.00133 | 0.0201 | 0.000119 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.00029 | 0.0041 | 47.5 | 237.5 | 6.1E-06 | 1.2E-06 | 8.7E-05 | 1.7E-05 |
| n-Propylbenzene | 0.000150 | 0.0355 | 0.0000185 | 1 | 1 | 0.0000550 | 0.00137 | 0.00114 | 1 | 1 | 0.007 | 0.000034 | 0.0072 | 60.4 | 302 | 5.6E-07 | 1.1E-07 | 1.2E-04 | 2.4E-05 |
| Notes: a) Sources listed on Table 3-9b b) Analyte concentrations in w c) The EPC is the 95% UCL, if | ater calculated u | sing an equilib | | | | not calculated). | | | | | | | | Pe | Butyltin HI esticides HI hthalate HI VOCs HI | | 3.2E+01 6.5E-02 1.1E-01 2.2E-06 | 7.0E+01 4.4E-01 8.6E-01 7.3E-02 | 2.5E+01 5.3E-02 8.6E-02 1.5E-02 |

Bold indicates hazard quotient greater than 1.0.

*** = To be evaluated as Total HPAHs, per EcoSSL; however, HQs for individual HPAHs were calculated using the Total HPAH SLV to determine the primary risk drivers.

% = percent

AOPC = area of potential concernAUF = Area Use Factor BAF = bioaccumulation factor BHC = benzene hexachloride BW = body weightCPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weightEPC = exposure point concentration EcoSSL = ecological soil screening level ft bgs = feet below ground surface HI = (cumulative) hazard index HPAHs = high molecular weight polycyclic aromatic hydrocarbons HQ = hazard quotient

kg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per dayLOAEL = lowest oberserved adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NOAEL = no oberserved adverse effect level PF = portion of food itemSLV =screening level value SVOC = semivolatile organic compound TRV = toxicity reference valueUCL = upper confidence limit VOC = volatile organic compound

 Table 3-12e

 Calculation of Dose and Hazard Quotient for the American Mink – Sandblast Area AOPC

| | | EPCs ^c | | BA | AFs | | E | Exposure Fa | actors | | | Do | oses | TF | RVs | | Hazard | Quotients | |
|-----------------------------|--|--|---|--|---|---|---|---------------------------------------|------------------------|----------|---------|------------|---|------------------------------------|------------------------------------|--------------|--|--------------|--|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^b | Surface Soil BAF mammals (kg tissue/ kg soil) ^a | Shallow Soil BAF mammals (kg tissue/ kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Small Mammals | | BW (kg) | | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | (0-1 ft bgs) | Surface Soil (0-1 ft bgs) LOAEL HQ | (0-3 ft bgs) | Shallow Soil (0-3 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | | | | | | |
| Antimony ^b | 2.94 | 2.42 | 0.0653 | 0.050 | 0.050 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.0038 | 0.0033 | 0.059 | 0.59 | 6.4E-02 | 6.4E-03 | 5.6E-02 | 5.6E-03 |
| Cadmium | 3.20 | 2.63 | 0.0000100 | 0.1540 | 0.1708 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.0034 | 0.0028 | 0.770 | 7.70 | 4.4E-03 | 4.4E-04 | 3.7E-03 | 3.7E-04 |
| Chromium III | 720 | 579 | 0.000400 | 0.0403 | 0.0427 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.65 | 0.53 | 2.40 | 12.0 | 2.7E-01 | 5.4E-02 | 2.2E-01 | 4.4E-02 |
| Lead | 465 | 529 | 0.0000360 | 0.0351 | 0.0327 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.42 | 0.47 | 4.70 | 8.90 | 8.9E-02 | 4.7E-02 | 1.0E-01 | 5.3E-02 |
| Mercury | 0.113 | 0.098 | 0.00217 | 0.0543 | 0.0543 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.00014 | 0.00013 | 0.02 | 0.07 | 7.0E-03 | 2.0E-03 | 6.3E-03 | 1.8E-03 |
| Nickel | 353 | 251 | 5.43 | 0.0340 | 0.0408 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.40 | 0.32 | 1.70 | 3.40 | 2.4E-01 | 1.2E-01 | 1.9E-01 | 9.3E-02 |
| Butyltins | | - | | | | | | | | - | - | | | | | | | • | |
| Tributyltin | 0.481 | 0.248 | 0.00398 | 0.00001 | 0.00001 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.00047 | 0.00028 | 23.4 | 35 | 2.0E-05 | 1.4E-05 | 1.2E-05 | 7.9E-06 |
| Pesticides | | | | • | | • • | | I | | | • | | | | | • | • | | |
| BHC (delta) | 0.00154 | 0.00130 | 5.48E-05 | 0.00108 | 0.00108 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.0000022 | 0.0000020 | 8 | 40 | 2.8E-07 | 5.5E-08 | 2.5E-07 | 5.0E-08 |
| BHC (gamma) Lindane | 0.00968 | 0.00106 | 0.000344 | 0.000412 | 0.000412 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | | 0.000014 | 0.0000065 | 8 | 40 | 1.7E-06 | 3.5E-07 | 8.1E-07 | 1.6E-07 |
| Chlordane (alpha) | 0.000815 | 0.000776 | 1.21E-06 | 0.0988 | 0.0988 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.0000082 | 0.0000078 | 4.6 | 9.2 | 1.8E-07 | 8.9E-08 | 1.7E-07 | 8.5E-08 |
| Chlordane (gamma) | 0.0303 | 0.0204 | 4.49E-05 | 0.1303 | 0.1303 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | | 0.000032 | 0.000022 | 4.6 | 9.2 | 6.9E-06 | 3.5E-06 | 4.7E-06 | 2.4E-06 |
| Endosulfan I | 0.00269 | 0.000809 | 3.98E-05 | 0.000531 | 0.000531 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | | 0.0000029 | 0.0000013 | 0.15 | 0.75 | 2.0E-05 | 3.9E-06 | 8.9E-06 | 1.8E-06 |
| Endosulfan II | 0.000495 | 0.000447 | 7.32E-06 | 0.0005 | 0.0005 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | | 0.00000054 | 0.00000050 | 0.15 | 0.75 | 3.6E-06 | 7.2E-07 | 3.3E-06 | 6.7E-07 |
| Endosulfan Sulfate | 0.000953 | 0.000889 | 9.68E-06 | 0.000359 | 0.000359 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163158 | 0.974 | 0.0000010 | 0.00000091 | 0.15 | 0.75 | 6.5E-06 | 1.3E-06 | 6.1E-06 | 1.2E-06 |
| Endrin Aldehyde | 0.00387 | 0.00344 | 0.000118 | 0.005 | 0.005 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0000052 | 0.0000049 | 0.092 | 0.92 | 5.7E-05 | 5.7E-06 | 5.3E-05 | 5.3E-06 |
| Endrin Ketone | 0.00309 | 0.00288 | 3.18E-05 | 0.0077 | 0.0077 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0000032 | 0.0000030 | 0.092 | 0.92 | 3.5E-05 | 3.5E-06 | 3.3E-05 | 3.3E-06 |
| Endrin | 0.00504 | 0.00389 | 2.51E-05 | 0.0124 | 0.0124 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0000048 | 0.0000038 | 0.092 | 0.92 | 5.2E-05 | 5.2E-06 | 4.1E-05 | 4.1E-06 |
| Heptachlor | 0.000464 | 0.000415 | 1.12E-06 | 0.0988 | 0.0988 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.00000048 | 0.00000043 | 0.1 | 1 | 4.8E-06 | 4.8E-07 | 4.3E-06 | 4.3E-07 |
| Methoxychlor | 0.00105 | 0.00104 | 2.54E-06 | 0.00944 | 0.00944 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.00000095 | 0.0000094 | 4 | 8 | 2.4E-07 | 1.2E-07 | 2.3E-07 | 1.2E-07 |
| HPAHs | | | | | • | | | | | | | • | • | | | | | | • |
| Benzo(a)anthracene | 1.75 | 2.45 | 0.00000060 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0015 | 0.0021 | *** | *** | 2.4E-03 | 4.9E-04 | 3.4E-03 | 6.8E-04 |
| Benzo(a)pyrene | 1.76 | 2.39 | 0.000300 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0015 | 0.0020 | *** | *** | 2.4E-03 | 4.9E-04 | 3.3E-03 | 6.6E-04 |
| Benzo(b)fluoranthene | 1.95 | 2.21 | 0.00000092 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0017 | 0.0019 | *** | *** | 2.7E-03 | 5.4E-04 | 3.1E-03 | 6.1E-04 |
| Benzo(g,h,i)perylene | 1.07 | 1.05 | 5.49E-05 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.00091 | 0.00089 | *** | *** | 1.5E-03 | 3.0E-04 | 1.5E-03 | 2.9E-04 |
| Benzo(k)fluoranthene | 0.676 | 0.764 | 0.000115 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.00058 | 0.00065 | *** | *** | 9.4E-04 | 1.9E-04 | 1.1E-03 | 2.1E-04 |
| Benzofluoranthenes, Total | 7.76 | 10.6 | 0.00132 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0066 | 0.0090 | *** | *** | 1.1E-02 | 2.2E-03 | 1.5E-02 | 2.9E-03 |
| Chrysene | 2.07 | 4.59 | 0.0000017 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0018 | 0.0039 | *** | *** | 2.9E-03 | 5.7E-04 | 6.4E-03 | 1.3E-03 |
| Dibenz(a,h)anthracene | 0.244 | 0.238 | 1.28E-05 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.00021 | 0.00020 | *** | *** | 3.4E-04 | 6.8E-05 | 3.3E-04 | 6.6E-05 |
| Fluoranthene | 4.64 | 10.8 | 0.0000078 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0039 | 0.0092 | *** | *** | 6.4E-03 | 1.3E-03 | 1.5E-02 | 3.0E-03 |
| Indeno(1,2,3-cd)pyrene | 1.13 | 1.16 | 3.26E-05 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0010 | 0.0010 | *** | *** | 1.6E-03 | 3.1E-04 | 1.6E-03 | 3.2E-04 |
| Pyrene | 4.78 | 11.7 | 0.00880 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0042 | 0.010 | *** | *** | 6.8E-03 | 1.4E-03 | 1.6E-02 | 3.3E-03 |
| Total HPAHs | 18.6 | 40.8 | 0.00425 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.016 | 0.035 | 0.615 | 3.07 | 2.6E-02 | 5.2E-03 | 5.7E-02 | 1.1E-02 |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 98.0 | 77.1 | 0.0817 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.22 | 0.17 | 18.3 | 183 | 1.2E-02 | 1.2E-03 | 9.4E-03 | 9.4E-04 |
| Butyl Benzyl Phthalate | 0.0171 | 0.0152 | 0.000239 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.000042 | 0.000037 | 159 | 470 | 2.6E-07 | 8.9E-08 | 2.4E-07 | 8.0E-08 |
| Carbazole | 0.144 | 0.112 | 0.00157 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.00034 | 0.00027 | 5 | 50 | 6.9E-05 | 6.9E-06 | 5.5E-05 | 5.5E-06 |
| Dibenzofuran | 0.0424 | 0.122 | 0.000463 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.00010 | 0.00028 | 6 | 30 | 1.7E-05 | 3.4E-06 | 4.6E-05 | 9.2E-06 |
| Di-n-butyl Phthalate | 0.0804 | 0.0543 | 0.00693 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.00029 | 0.00023 | 550 | 1,833 | 5.3E-07 | 1.6E-07 | 4.2E-07 | 1.3E-07 |
| Di-n-octyl Phthalate | 0.0251 | 0.0225 | 0.0000178 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.00006 | 0.00005 | 7,500 | 37,500 | 7.4E-09 | 1.5E-09 | 6.7E-09 | 1.3E-09 |
| VOCs | | | | | | | | | | | | | | | | | | | |
| 1,2,4-Trimethylbenzene | 0.000230 | 4.62 | 0.0000375 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0000011 | 0.010 | 13 | 65 | 8.6E-08 | 1.7E-08 | 7.8E-04 | 1.6E-04 |
| /// | | | | | - | | | | | | | | | | | | | | |

Table 3-12e Calculation of Dose and Hazard Quotient for the American Mink-Sandblast Area AOPC

| | | EPCs ^c | | B | AFs | | I | Exposure F | actors | | | Do | ses | TR | Vs | | Hazard (| Quotients | |
|--|------------------|--|-----------|--|--|---|---|------------|------------------|-------------|------------|--|----------|------|---|--------------------|--|--|--|
| CPEC | | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Water | Surface Soil BAF mammals (kg tissue/ kg soil) ^a | Shallow Soil BAF mammals (kg tissue/ | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Rate | Small Mammals | AUF | | Surface Soil Dose (mg/kg bw/day) | | | LOAEL TRV (mg/kg- bw/day) | (0-1 ft bgs) | Surface Soil (0-1 ft bgs) LOAEL HQ | (0-3 ft bgs) | (0-3 ft bgs) |
| 4-Isopropyltoluene | 0.00133 | 0.0201 | 0.000119 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0000049 | 0.000046 | 47.5 | 237.5 | 1.0E-07 | 2.0E-08 | 9.8E-07 | 2.0E-07 |
| n-Propylbenzene | 0.000150 | 0.0355 | 0.0000185 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.163 | 0.974 | 0.0000063 | 0.000079 | 60.4 | 302 | 1.0E-08 | 2.1E-09 | 1.3E-06 | 2.6E-07 |
| Notes: a) Sources listed on Table 3-10 b) Cadmium & lead (dissolved All other CPEC concentratio |), benzo(a)anthr | acene, benzo(b | | | | | | | | rebay surfa | ace water; | see Table 3-6. | | Pe | Butyltin HI esticides HI hthalate HI VOCs HI | 1.9E-04 1.2E-02 | 2.3E-01 2.5E-05 1.2E-03 4.0E-08 | 5.7E-01 1.6E-04 9.4E-03 7.9E-04 | 2.0E-01 2.0E-05 9.4E-04 1.6E-04 |

c) The EPC is the 95% UCL, if calculated (see Table 1-2), or maximum detected concentration (if 95% UCL not calculated).

Bold indicates hazard quotient greater than 1.0.

*** = To be evaluated as Total HPAHs, per EcoSSL; however, HQs for individual HPAHs were calculated using the Total HPAH SLV to determine the primary risk drivers.

% = percentAOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BHC = benzene hexachloride BW = body weight CPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weightEcoSSL = ecological soil screening level EPC = exposure point concentration ft bgs = feet below ground surface HI = (cumulative) hazard index HPAHs = high molecular weight polycyclic aromatic hydrocarbons HQ = hazard quotient

kg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per dayLOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NOAEL = no observed adverse effect level OU = operable unitPF = portion of food itemSLV = screening level value SVOC = semivolatile organic compound TRV = toxicity reference value UCL = upper confidence limit VOC = volatile organic compound

URS

Table 3-13aCalculation of Dose and Hazard Quotient for the Canada Goose – Pistol Range AOPC

| | EPO | Cs | BAF | | Ex | posure Fac | tors | | | Dose | TR | Vs | Hazard (| Quotients |
|------------|----------------|-------|--|---|---|---------------------------------------|--------------|-----|------------|--|------------------------------------|------------------------------------|----------|--|
| CPEC | (0-1.5 ft bgs) | | Surface Soil BAF plant (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Plants | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | | Surface Soil (0-1.5 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | |
| Lead | 365 | 0.406 | 0.0199 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 1.9 | 1.63 | 3.26 | 1.2E+00 | 5.9E-01 |

a) Sources listed on Table 3-8c for plant BAFs.

b) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7c. **Bold** indicates hazard quotient greater than 1.0.

| % = percent | kg dry/day = kilograms per day in dry weight |
|--|--|
| AOPC = area of potential concern | L/day = liters per day |
| AUF = Area Use Factor | LOAEL = lowest observed adverse effect level |
| BAF = bioaccumulation factor | mg/kg = milligrams per kilogram |
| BW = body weight | mg/kg-bw/day = milligrams per kilogram body weight per day |
| Dose = average daily dose (mg/kg-bw/day) | mg/L = milligrams per liter |
| dw = dry weight | NOAEL = no observed adverse effect level |
| EPC = exposure point concentration | PF = portion of food item |
| ft bgs = feet below ground surface | TRV = toxicity reference value |
| HQ = hazard quotient | UCL = upper confidence limit |
| kg = kilogram | |

 Table 3-13b

 Calculation of Dose and Hazard Quotient for the American Robin – Pistol Range AOPC

| | EPO | Cs | BAF | | Ex | posure Fac | tors | | | Dose | TR | Vs | Hazard (| Quotients |
|------------|--|-------|---------------|---|---|------------|--------------------------|-------|------------|--|------------------------------------|------------------------------------|----------|--|
| CPEC | 95% UCL in Surface Soil (0-1.5 ft bgs) (mg/kg dw) | | (kg tissue/kg | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Rate | PF Inverte- brates | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | | Surface Soil (0-1.5 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | |
| Lead | 365 | 0.406 | 0.258 | 0.0013 | 0.013 | 0.011 | 1 | 0.703 | 0.0773 | 15 | 1.63 | 3.26 | 9.3E+00 | 4.6E+00 |

a) Sources listed on Table 3-9c for invertebrate BAFs.

b) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7c. **Bold** indicates hazard quotient greater than 1.0.

| % = percent | kg dry/day = kilograms per day in dry weight |
|--|--|
| AOPC = area of potential concern | L/day = liters per day |
| AUF = Area Use Factor | LOAEL = lowest observed adverse effect level |
| BAF = bioaccumulation factor | mg/kg = milligrams per kilogram |
| BW = body weight | mg/kg-bw/day = milligrams per kilogram body weight per day |
| Dose = average daily dose (mg/kg-bw/day) | mg/L = milligrams per liter |
| dw = dry weight | NOAEL = no observed adverse effect level |
| EPC = exposure point concentration | PF = portion of food item |
| ft bgs = feet below ground surface | TRV = toxicity reference value |
| HQ = hazard quotient | UCL = upper confidence limit |
| kg = kilogram | |

 Table 3-13c

 Calculation of Dose and Hazard Quotient for the American Kestrel – Pistol Range AOPC

| | EPO | Cs | BAF | | Ex | xposure Fa | ctors | | | Dose | TR | Vs | Hazard (| Quotients |
|------------|--|------------|---|---|---|------------|---------------|---------|------------|--|------------------------------------|------------------------------------|----------------|--|
| CPEC | 95% UCL in Surface Soil (0-1.5 ft bgs) (mg/kg dw) | tration in | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | | PF mammals | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | (0-1.5 ft bgs) | Surface Soil (0-1.5 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | |
| Lead | 365 | 0.0360 | 0.0402 | 0.00033 | 0.017 | 0.014 | 1 | 0.00096 | 0.116 | 0.0030 | 1.63 | 3.26 | 1.9E-03 | 9.3E-04 |

a) Sources listed on Table 3-10c for mammal BAFs.

b) Lead (dissolved) maximum detected concentrations in River OU Forebay surface water; see Table 3-6. **Bold** indicates hazard quotient greater than 1.0.

| % = percent | kg dry/day = kilograms per day in dry weight |
|--|--|
| AOPC = area of potential concern | L/day = liters per day |
| AUF = Area Use Factor | LOAEL = lowest observed adverse effect level |
| BAF = bioaccumulation factor | mg/kg = milligrams per kilogram |
| BW = body weight | mg/kg-bw/day = milligrams per kilogram body weight per day |
| Dose = average daily dose (mg/kg-bw/day) | mg/L = milligrams per liter |
| dw = dry weight | NOAEL = no observed adverse effect level |
| EPC = exposure point concentration | OU = operable unit |
| ft bgs = feet below ground surface | PF = portion of food item |
| HQ = hazard quotient | TRV = toxicity reference value |
| kg = kilogram | UCL = upper confidence limit |
| | |

 Table 3-13d

 Calculation of Dose and Hazard Quotient for the Vagrant Shrew – Pistol Range AOPC

| | EPO | Cs | BAF | | E | xposure Fa | ctors | | | Dose | TR | Vs | Hazard (| Quotients |
|------------|--|---|--------------------------------|---|---|---------------------------------------|--------------------------|-----|------------|--|------------------------------------|------------------------------------|----------------|--|
| CPEC | 95% UCL in Surface Soil (0-1.5 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^b | invertebrates (kg tissue/kg | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Inverte- brates | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | (0-1.5 ft bgs) | Surface Soil (0-1.5 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | |
| Lead | 365 | 0.406 | 0.258 | 0.000055 | 0.0014 | 0.0011 | 1 | 1 | 0.007 | 21 | 4.70 | 8.90 | 4.5E+00 | 2.4E+00 |

a) Sources listed on Table 3-9c for invertebrate BAFs.

b) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7c. **Bold** indicates hazard quotient greater than 1.0.

| % = percent | kg dry/day = kilograms per day in dry weight |
|--|--|
| AOPC = area of potential concern | L/day = liters per day |
| AUF = Area Use Factor | LOAEL = lowest oberserved adverse effect level |
| BAF = bioaccumulation factor | mg/kg = milligrams per kilogram |
| BW = body weight | mg/kg-bw/day = milligrams per kilogram body weight per day |
| Dose = average daily dose (mg/kg-bw/day) | mg/L = milligrams per liter |
| dw = dry weight | NOAEL = no oberserved adverse effect level |
| EPC = exposure point concentration | PF = portion of food item |
| ft bgs = feet below ground surface | TRV = toxicity reference value |
| HQ = hazard quotient | UCL = upper confidence limit |
| kg = kilogram | |

 Table 3-13e

 Calculation of Dose and Hazard Quotient for the American Mink – Pistol Range AOPC

| | EPO | Cs | BAF | | E | xposure Fa | ctors | | | Dose | TR | RVs | Hazard (| Quotients |
|------------|--|------------|---|---|---|---------------------------------------|------------------------|--------|------------|--|------------------------------------|------------------------------------|----------------|--|
| CPEC | 95% UCL in Surface Soil (0-1.5 ft bgs) (mg/kg dw) | tration in | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Small Mammals | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | (0-1.5 ft bgs) | Surface Soil (0-1.5 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | |
| Lead | 365 | 0.0360 | 0.0402 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.0137 | 0.974 | 0.028 | 4.70 | 8.90 | 5.9E-03 | 3.1E-03 |

a) Sources listed on Table 3-10c for mammal BAFs.

b) Lead (dissolved) maximum detected concentrations in River OU Forebay surface water; see Table 3-6.

| % = percent | kg dry/day = kilograms per day in dry weight |
|--|--|
| AOPC = area of potential concern | L/day = liters per day |
| AUF = Area Use Factor | LOAEL = lowest observed adverse effect level |
| BAF = bioaccumulation factor | mg/kg = milligrams per kilogram |
| BW = body weight | mg/kg-bw/day = milligrams per kilogram body weight per day |
| Dose = average daily dose (mg/kg-bw/day) | mg/L = milligrams per liter |
| dw = dry weight | NOAEL = no observed adverse effect level |
| EPC = exposure point concentration | OU = operable unit |
| ft bgs = feet below ground surface | PF = portion of food item |
| HQ = hazard quotient | TRV = toxicity reference value |
| kg = kilogram | UCL = upper confidence limit |
| | |

 Table 3-14a

 Calculation of Dose and Hazard Quotient for the Canada Goose – Bulb Slope AOPC

| | EPO | Cs | BAF | | Ex | posure Fac | tors | | | Dose | TR | RVs | Hazard | Quotients |
|------------|--|---|--|---|---|---------------------------------------|--------------|-----|------------|--|------------------------------------|------------------------------------|--------------|--|
| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^b | Surface Soil BAF plant (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Plants | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | (0-1 ft bgs) | Surface Soil (0-1 ft bgs) LOAEL HQ |
| Inorganics | | | | • | | | | | • | | | • | | |
| Lead | 307 | 0.341 | 0.0214 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 2 | 1.63 | 3.26 | 1.0E+00 | 5.0E-01 |
| Mercury | 0.720 | 0.0138 | 0.650 | 0.013 | 0.15 | 0.12 | 1 | 1 | 3.0 | 0.0276 | 0.0064 | 0.064 | 4.3E+00 | 4.3E-01 |
| | | | | | | | | | | | | HI | 5.3E+00 | 9.3E-01 |

a) Sources listed on Table 3-8d for plant BAFs.

b) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7d. **Bold** indicates hazard quotient greater than 1.0.

% = percentAOPC = area of potential concernAUF = Area Use Factor BAF = bioaccumulation factor BW = body weightCPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weightEPC = exposure point concentrations ft bgs = feet below ground surface HI = (cumulative) hazard index HQ = hazard quotientkg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per dayLOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NOAEL = no observed adverse effect level PF = portion of food itemTRV = toxicity reference value UCL = upper confidence limit

 Table 3-14b

 Calculation of Dose and Hazard Quotient for the American Robin – Bulb Slope AOPC

| | EPO | Cs | BAF | | Ex | posure Fac | ctors | | | Dose | TR | RVs | Hazard | Quotients |
|------------|--|---|--------------------------------|---|---|------------|--------------------------|-------|------------|--|------------------------------------|------------------------------------|--|--|
| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^b | invertebrates (kg tissue/kg | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Rate | PF Inverte- brates | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | |
| Lead | 307 | 0.341 | 0.266 | 0.0013 | 0.013 | 0.011 | 1 | 0.146 | 0.0773 | 3 | 1.63 | 3.26 | 1.7E+00 | 8.3E-01 |
| Mercury | 0.720 | 0.0138 | 1.69 | 0.0013 | 0.013 | 0.011 | 1 | 0.146 | 0.0773 | 0.0309 | 0.0064 | 0.064 | 4.8E+00 | 4.8E-01 |
| | | | | | | | | | | | | HI | 6.5E+00 | 1.3E+00 |

a) Sources listed on Table 3-9d for invertebrate BAFs.

b) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7d. **Bold** indicates hazard quotient greater than 1.0.

% = percentAOPC = area of potential concernAUF = Area Use Factor BAF = bioaccumulation factor BW = body weightCPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weightEPC = exposure point concentrations ft bgs = feet below ground surface HI = (cumulative) hazard index HQ = hazard quotientkg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per dayLOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NOAEL = no observed adverse effect level PF = portion of food itemTRV = toxicity reference value UCL = upper confidence limit

 Table 3-14c

 Calculation of Dose and Hazard Quotient for the American Kestrel – Bulb Slope AOPC

| | EPO | Cs | BAF | | E | xposure Fa | ctors | | | Dose | TR | RVs | Hazard (| Quotients |
|------------|--|---|---|---|---|------------|---------------|--------|------------|--|------------------------------------|------------------------------------|--|--|
| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^b | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | | PF mammals | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | |
| Lead | 307 | 0.0360 | 0.0442 | 0.00033 | 0.017 | 0.014 | 1 | 0.0002 | 0.116 | 0.00 | 1.63 | 3.26 | 3.5E-04 | 1.7E-04 |
| Mercury | 0.720 | 0.0138 | 0.0543 | 0.00033 | 0.017 | 0.014 | 1 | 0.0002 | 0.116 | 0.0000019 | 0.0064 | 0.064 | 2.9E-04 | 2.9E-05 |
| | - | | | - | - | • | | | | · · · · · · · · · · · · · · · · · · · | | HI | 6.4E-04 | 2.0E-04 |

a) Sources listed on Table 3-10d for mammal BAFs.

b) Lead (dissolved) maximum detected concentration in River OU Forebay surface water; see Table 3-6.

Mercury concentration in water calculated using an equilibrium partitioning equation due to lack of River OU analytical data; see Table 3-7d.

Bold indicates hazard quotient greater than 1.0.

% = percentAOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BW = body weightCPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weightEPC = exposure point concentrations ft bgs = feet below ground surface HI = (cumulative) hazard index HQ = hazard quotientkg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per dayLOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NOAEL = no observed adverse effect level OU = operable unitPF = portion of food itemTRV = toxicity reference value UCL = upper confidence limit

 Table 3-14d

 Calculation of Dose and Hazard Quotient for the Vagrant Shrew – Bulb Slope AOPC

| | EPO | Cs | BAF | | E | xposure Fa | ctors | | | Dose | TR | V s | Hazard (| Quotients |
|------------|--|------------|--------------------------------|---|---|---------------------------------------|--------------------------|-------|------------|--|------------------------------------|------------------------------------|--|--|
| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | tration in | invertebrates (kg tissue/kg | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Inverte- brates | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) LOAEL HQ |
| Inorganics | _ | | | | | | | | | | | | | |
| Lead | 307 | 0.341 | 0.266 | 0.000055 | 0.0014 | 0.0011 | 1 | 0.208 | 0.007 | 4 | 4.7 | 8.9 | 8.2E-01 | 4.3E-01 |
| Mercury | 0.720 | 0.0138 | 1.69 | 0.000055 | 0.0014 | 0.0011 | 1 | 0.208 | 0.007 | 0.0512 | 0.020 | 0.070 | 2.6E+00 | 7.3E-01 |
| | | | | | | | | | | | | HI | 3.4E+00 | 1.2E+00 |

a) Sources listed on Table 3-9d for invertebrate BAFs.

b) Analyte concentrations in water calculated using an equilibrium partitioning equation; see Table 3-7d.

Bold indicates hazard quotient greater than 1.0.

% = percent AOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BW = body weightCPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weightEPC = exposure point concentrations ft bgs = feet below ground surface HI = (cumulative) hazard index HQ = hazard quotient kg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per dayLOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NOAEL = no observed adverse effect level PF = portion of food itemTRV = toxicity reference value UCL = upper confidence limit

 Table 3-14e

 Calculation of Dose and Hazard Quotient for the American Mink – Bulb Slope AOPC

| | EPO | Cs | BAF | | Ε | xposure Fa | ctors | | | Dose | TR | RVs | Hazard | Quotients |
|------------|--|---|---|---|---|---------------------------------------|------------------------|---------|------------|--|------------------------------------|------------------------------------|--|--|
| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^b | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Small Mammals | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | |
| Lead | 307 | 0.0360 | 0.0442 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.00284 | 0.974 | 0.005 | 4.7 | 8.9 | 1.0E-03 | 5.5E-04 |
| Mercury | 0.720 | 0.0138 | 0.0543 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.00284 | 0.974 | 0.0000155 | 0.020 | 0.070 | 7.8E-04 | 2.2E-04 |
| | | | | | | | | | | | | HI | 1.8E-03 | 7.7E-04 |

a) Sources listed on Table 3-10d for mammal BAFs.

b) Lead (dissolved) maximum detected concentration in River OU Forebay surface water; see Table 3-6.

Mercury concentration in water calculated using an equilibrium partitioning equation due to lack of River OU analytical data; see Table 3-7d.

Bold indicates hazard quotient greater than 1.0.

% = percent

AOPC = area of potential concern

AUF = Area Use Factor

BAF = bioaccumulation factor

BW = body weight

CPEC = chemical of potential ecological concern

Dose = average daily dose (mg/kg-bw/day)

dw = dry weight

EPC = exposure point concentrations

ft bgs = feet below ground surface

HI = (cumulative) hazard index

HI = (cumulative) hazard index

- HQ = hazard quotient
- kg = kilogram

kg dry/day = kilograms per day in dry weight

L/day = liters per day

LOAEL = lowest observed adverse effect level

mg/kg = milligrams per kilogram

mg/kg-bw/day = milligrams per kilogram body weight per day

mg/L = milligrams per liter

NOAEL = no observed adverse effect level

OU = operable unit

PF = portion of food item

TRV = toxicity reference value

UCL = upper confidence limit

Table 3-15a Calculation of Dose and Hazard Quotient for the American Kestrel – Combined AOPCs

| | | EPCs ^d | | B | AFs | | E | xposure Fa | ctors | | | Do | oses | TF | RVs | | Hazard | Quotients | |
|---|--|--|---|---|--|---|---|---------------------------------------|---------------|--------|---------|---|--------------------|------------------------------------|------------------------------------|--|--------------------|--|--------------------|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^c | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Shallow Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF mammals | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | (0-1 ft bgs) | Shallow Soil (0-3 ft bgs) NOAEL HQ | (0-3 ft bgs) |
| Inorganics | _ | | | | | | | | | | | | | | | | | | |
| Antimony ^b | 1.80 | 1.64 | 0.0400 | 0.050 | 0.050 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.00040 | 0.00037 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Cadmium | 1.97 | 1.67 | 0.0000100 | 0.199 | 0.217 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.0011 | 0.0010 | 1.47 | 5.88 | 7.4E-04 | 1.9E-04 | 6.8E-04 | 1.7E-04 |
| Chromium III | 567 | 510 | 0.000315 | 0.0430 | 0.0442 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.090 | 0.083 | 2.66 | 14.5 | 3.4E-02 | 6.2E-03 | 3.1E-02 | 5.7E-03 |
| Copper | 95.0 | 93.6 | 0.000520 | 0.157 | 0.159 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.042 | 0.042 | 4.05 | 12.1 | 1.0E-02 | 3.5E-03 | 1.0E-02 | 3.5E-03 |
| Lead | 398 | 396 | 0.0000360 | 0.0383 | 0.0384 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.059 | 0.058 | 1.63 | 3.26 | 3.6E-02 | 1.8E-02 | 3.6E-02 | 1.8E-02 |
| Mercury | 0.378 | 0.330 | 0.00727 | 0.0543 | 0.0543 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000086 | 0.000077 | 0.0064 | 0.064 | 1.3E-02 | 1.3E-03 | 1.2E-02 | 1.2E-03 |
| Nickel | 227 | 234 | 3.49 | 0.0431 | 0.0424 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.044 | 0.044 | 6.71 | 26.8 | 6.5E-03 | 1.6E-03 | 6.6E-03 | 1.7E-03 |
| Butyltins | | | 1 | | | | | | 1 | | | | | | | | | | |
| Tributyltin | 0.422 | 0.219 | 0.00349 | 0.0424 | 0.0424 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000074 | 0.000042 | 6.8 | 16.9 | 1.1E-05 | 4.4E-06 | 6.2E-06 | 2.5E-06 |
| Pesticides | 01.22 | 01212 | 0100015 | 010121 | 010121 | 0.000000 | 01017 | 01011 | - | 010177 | 01110 | 01000071 | 0.000012 | 0.0 | 100 | 1112 00 | 1112 00 | 0.22 00 | |
| BHC (delta) | 0.00303 | 0.00303 | 0.000108 | 8.89 | 8.89 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000068 | 0.000068 | 2 | 20 | 3.4E-05 | 3.4E-06 | 3.4E-05 | 3.4E-06 |
| BHC (gamma) Lindane | 0.00968 | 0.00968 | 0.000344 | 3.83 | 3.83 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000095 | 0.000095 | 2 | 20 | 4.7E-05 | 4.7E-06 | 4.7E-05 | 4.7E-06 |
| Chlordane (alpha) | 0.00101 | 0.000952 | 1.50E-06 | 18.8 | 18.8 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000048 | 0.000045 | 2.14 | 10.7 | 2.2E-05 | 4.5E-06 | 2.1E-05 | 4.2E-06 |
| Chlordane (gamma) | 0.0188 | 0.0118 | 2.79E-05 | 23.9 | 23.9 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.0011 | 0.00071 | 2.14 | 10.7 | 5.3E-04 | 1.1E-04 | 3.3E-04 | 6.7E-05 |
| Chlordane (technical) | 1.56 | 0.902 | 0.00462 | 47.7 | 47.7 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.19 | 0.11 | 2.14 | 10.7 | 8.8E-02 | 1.8E-02 | 5.1E-02 | 1.0E-02 |
| Endosulfan I | 0.00645 | 0.000547 | 9.54E-05 | 1.99 | 1.99 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000033 | 0.0000030 | 10 | 50 | 3.3E-06 | 6.6E-07 | 3.0E-07 | 5.9E-08 |
| Endosulfan II | 0.000844 | 0.000756 | 1.25E-05 | 1.99 | 1.99 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.0000043 | 0.0000039 | 10 | 50 | 4.3E-07 | 8.6E-08 | 3.9E-07 | 7.7E-08 |
| Endosulfan Sulfate | 0.00108 | 0.00101 | 1.10E-05 | 0.970 | 0.970 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.0000027 | 0.0000025 | 10 | 50 | 2.7E-07 | 5.4E-08 | 2.5E-07 | 5.1E-08 |
| Endrin Aldehyde | 0.00314 | 0.00286 | 9.60E-05 | 28.7 | 28.7 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.00023 | 0.00021 | 0.01 | 0.1 | 2.3E-02 | 2.3E-03 | 2.1E-02 | 2.1E-03 |
| Endrin Ketone | 0.00261 | 0.00250 | 2.69E-05 | 14.1 | 14.1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000093 | 0.000089 | 0.01 | 0.1 | 9.3E-03 | 9.3E-04 | 8.9E-03 | 8.9E-04 |
| Endrin | 0.00421 | 0.00321 | 2.09E-05 | 10.4 | 10.4 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.00011 | 0.000084 | 0.01 | 0.1 | 1.1E-02 | 1.1E-03 | 8.4E-03 | 8.4E-04 |
| Heptachlor | 0.000437 | 0.000477 | 1.06E-06 | 8.75 | 8.75 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.0000097 | 0.000011 | 0.013 | 0.065 | 7.4E-04 | 1.5E-04 | 8.1E-04 | 1.6E-04 |
| Methoxychlor | 0.00120 | 0.00120 | 2.91E-06 | 3.98 | 3.98 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000012 | 0.000012 | 125 | 625 | 9.7E-08 | 1.9E-08 | 9.7E-08 | 1.9E-08 |
| HPAHs ^b | 0100120 | 0100120 | 2012 00 | 0.00 | 0.00 | 0100000 | 01017 | 01011 | - | 010177 | 01110 | 01000012 | 01000012 | 120 | 010 | 7112 00 | 102 00 | 7112 00 | |
| SVOCs | | | | | | | | | | | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 36.4 | 31.3 | 0.0303 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.094 | 0.081 | 11 | 11 | 8.5E-02 | 8.5E-03 | 7.3E-02 | 7.3E-03 |
| Butyl Benzyl Phthalate ^b | 0.0172 | 0.0154 | 0.000240 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000045 | 0.000040 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Carbazole ^b | 0.0172 | 0.312 | 0.00240 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.110 | 0.00043 | 0.00040 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| | | | | 1 | 1 | | | | 1 | | | | | | | | | | |
| Dibenzofuran ^b Di-n-butyl Phthalate | 0.0771 | 0.0834 0.105 | 0.000842 0.0107 | 1 | 1 | 0.00033 0.00033 | 0.017 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.00020 0.00034 | 0.00022 0.00029 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| ~ | | | | 1 | 1 | | | 0.014 | 1 | 0.0177 | 0.116 | | | 0.11 | 1.1 | 3.1E-03 | 3.1E-04 | 2.7E-03 | 2.7E-04 |
| Di-n-octyl Phthalate ^b | 0.0210 | 0.0194 | 0.0000149 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000054 | 0.000050 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| VOCs | I | T | T | T | T | 1 | | | 1 | 1 | - | T | T | | | T | 1 | T | |
| 1,2,4-Trimethylbenzene ^b | 2.94 | 2.86 | 0.479 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.0086 | 0.0084 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| 4-Isopropyltoluene | 0.00102 | 0.0147 | 0.0000911 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.000028 | 0.000038 | 3 | 15.8 | 8.9E-07 | 1.8E-07 | 1.2E-05 | 2.4E-06 |
| n-Propylbenzene ^b | 0.000150 | 0.122 | 0.0000185 | 1 | 1 | 0.00033 | 0.017 | 0.014 | 1 | 0.0177 | 0.116 | 0.00000043 | 0.00031 | NA | NA | No TRV | No TRV | No TRV | No TRV |
| Notes: | | | | | | | | | | | | | | Matala | Butyltin HI | 1.0E-01 | 3.1E-02 | 9.7E-02 | 3.0E-02 |
| Bold indicates hazard quotient | graatar than 1 (|) | | | | | | | | | | | | | esticides HI | | 3.1E-02 2.2E-02 | 9.7E-02 9.0E-02 | 3.0E-02 1.4E-02 |
| a) Sources listed on Table 3-10 | | | | | | | | | | | | | | | thalates HI | | 2.2E-02 8.8E-03 | 9.0E-02 7.6E-02 | 7.6E-03 |
| a) Sources listed on Table 3-10 | | | | | | | | | | | | | | PI | maiates HI | 0.0E-U2 | 0.0E-U3 | 7.0E-02 | 7.0E-03 |

b) These chemicals do not have established TRVs, but will be evaluated qualitatively.

c) Cadmium, copper, and lead (dissolved), benzo(a)anthracene, benzo(b)fluoranthene, chrysene, and fluoranthene (C+F) maximum detected concentrations in River OU Forebay surface water; see Table 3-6. d) The EPC is the 95% UCL, if calculated (see Table 1-5), or maximum detected concentration (if 95% UCL not calculated).

All other CPEC concentrations in water calculated using equilibrium partitioning equations due to lack of River OU analytical data; see Table 3-7e.

URS

Table 3-15a Calculation of Dose and Hazard Quotient for the American Kestrel – Combined AOPCs

% = percent AOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BW = body weight C+F = column + filter CPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weight EPC = exposure point concentration ft bgs = feet below ground surface HI = (cumulative) hazard index HPAHs = high molecular weight polycyclic aromatic hydrocarbons HQ = hazard quotient kg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per day LOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram body weight per day mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NA = not available NOAEL = no observed adverse effect level OU = operable unit PF = portion of food item SVOC = semivolatile organic compound TRV = toxicity reference value UCL = upper confidence limit VOC = volatile organic compound

 Table 3-15b

 Calculation of Dose and Hazard Quotient for the American Mink – Combined AOPCs

| | | EPCs ^c | | BA | AFs | | Ex | xposure Fa | actors | | | Do | ses | TF | RVs | | Hazard | Ouotients | |
|--|--|--|---|---|--|---|---|---------------------------------------|------------------------|-------------|----------------|--|--|------------------------------------|------------------------------------|--|--|--|--|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Concen- tration in Water (mg/L) ^b | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Shallow Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Food Ingestion Rate (kg dry/day) | Water Ingestion Rate (L/day) | PF Small Mammals | AUF | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | Surface Soil (0-1 ft bgs) NOAEL HQ | Surface Soil (0-1 ft bgs) LOAEL HQ | Shallow Soil (0-3 ft bgs) NOAEL HQ | Shallow Soil (0-3 ft bgs) LOAEL HQ |
| Inorganics | | | | | | | | | | | | | | | | | | | |
| Antimony | 1.80 | 1.64 | 0.0400 | 0.050 | 0.050 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.0035 | 0.0033 | 0.059 | 0.59 | 6.0E-02 | 6.0E-03 | 5.6E-02 | 5.6E-03 |
| Cadmium | 1.97 | 1.67 | 0.0000100 | 0.199 | 0.217 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.0034 | 0.0029 | 0.770 | 7.70 | 4.4E-03 | 4.4E-04 | 3.8E-03 | 3.8E-04 |
| Chromium III | 567 | 510 | 0.000315 | 0.0430 | 0.0442 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.79 | 0.72 | 2.40 | 12.0 | 3.3E-01 | 6.6E-02 | 3.0E-01 | 6.0E-02 |
| Copper | 95.0 | 93.6 | 0.000520 | 0.157 | 0.159 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.16 | 0.15 | 5.60 | 9.34 | 2.8E-02 | 1.7E-02 | 2.7E-02 | 1.6E-02 |
| Lead | 398 | 396 | 0.0000360 | 0.0383 | 0.0384 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.55 | 0.55 | 4.70 | 8.90 | 1.2E-01 | 6.2E-02 | 1.2E-01 | 6.2E-02 |
| Mercury | 0.378 | 0.330 | 0.00727 | 0.0543 | 0.0543 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.00072 | 0.00065 | 0.02 | 0.07 | 3.6E-02 | 1.0E-02 | 3.3E-02 | 9.3E-03 |
| Nickel | 227 | 234 | 3.49 | 0.0431 | 0.0424 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.41 | 0.41 | 1.70 | 3.40 | 2.4E-01 | 1.2E-01 | 2.4E-01 | 1.2E-01 |
| Butyltins | | | | | | | | | | | | | | | | | | | |
| Tributyltin | 0.422 | 0.219 | 0.00349 | 0.0424 | 0.0424 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.00068 | 0.00039 | 23.4 | 35 | 2.9E-05 | 1.9E-05 | 1.7E-05 | 1.1E-05 |
| Pesticides | | | | | | | | | + | | | | | | + | | | | |
| BHC (delta) | 0.00303 | 0.00303 | 0.000108 | 8.89 | 8.89 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.000063 | 0.000063 | 8 | 40 | 7.9E-06 | 1.6E-06 | 7.9E-06 | 1.6E-06 |
| BHC (gamma) Lindane | 0.00968 | 0.00968 | 0.000344 | 3.83 | 3.83 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.00010 | 0.00010 | 8 | 40 | 1.2E-05 | 2.5E-06 | 1.2E-05 | 2.5E-06 |
| Chlordane (alpha) | 0.00101 | 0.000952 | 1.50E-06 | 18.8 | 18.8 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.000041 | 0.000039 | 4.6 | 9.2 | 8.9E-06 | 4.5E-06 | 8.4E-06 | 4.2E-06 |
| Chlordane (gamma) | 0.0188 | 0.0118 | 2.79E-05 | 23.9 | 23.9 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.00096 | 0.00061 | 4.6 | 9.2 | 2.1E-04 | 1.0E-04 | 1.3E-04 | 6.6E-05 |
| Chlordane (technical) | 1.56 | 0.902 | 0.00462 | 47.7 | 47.7 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.16 | 0.091 | 4.6 | 9.2 | 3.4E-02 | 1.7E-02 | 2.0E-02 | 9.9E-03 |
| Endosulfan I | 0.00645 | 0.000547 | 9.54E-05 | 1.99 | 1.99 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.000038 | 0.0000054 | 0.15 | 0.75 | 2.5E-04 | 5.0E-05 | 3.6E-05 | 7.2E-06 |
| Endosulfan II | 0.000844 | 0.000756 | 1.25E-05 | 1.99 | 1.99 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.0000049 | 0.0000044 | 0.15 | 0.75 | 3.3E-05 | 6.6E-06 | 3.0E-05 | 5.9E-06 |
| Endosulfan Sulfate | 0.00108 | 0.00101 | 1.10E-05 | 0.970 | 0.970 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.0000039 | 0.0000036 | 0.15 | 0.75 | 2.6E-05 | 5.2E-06 | 2.4E-05 | 4.9E-06 |
| Endrin Aldehyde | 0.00314 | 0.00286 | 9.60E-05 | 28.7 | 28.7 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.00019 | 0.00018 | 0.092 | 0.92 | 2.1E-03 | 2.1E-04 | 1.9E-03 | 1.9E-04 |
| Endrin Ketone | 0.00261 | 0.00250 | 2.69E-05 | 14.1 | 14.1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.000081 | 0.000078 | 0.092 | 0.92 | 8.8E-04 | 8.8E-05 | 8.4E-04 | 8.4E-05 |
| Endrin | 0.00421 | 0.00321 | 2.09E-05 | 10.4 | 10.4 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.00010 | 0.000076 | 0.092 | 0.92 | 1.1E-03 | 1.1E-04 | 8.1E-04 | 8.1E-05 |
| Heptachlor | 0.000421 | 0.000477 | 1.06E-06 | 8.75 | 8.75 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.0000086 | 0.0000094 | 0.072 | 1 | 8.6E-05 | 8.6E-06 | 9.4E-05 | 9.4E-06 |
| Methoxychlor | 0.00120 | 0.00120 | 2.91E-06 | 3.98 | 3.98 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.000012 | 0.000012 | 4 | 8 | 2.9E-06 | 1.5E-06 | 2.9E-06 | 1.5E-06 |
| HPAHs | 0.00120 | 0.00120 | 2.712.00 | 5.70 | 5.70 | 0.0001 | 0.034 | 0.077 | 0.15 | 0.231 | 0.774 | 0.000012 | 0.000012 | ļ - | | 2.91 00 | 1.512 00 | 2.712 00 | 1.512-00 |
| Benzo(a)anthracene | 4.39 | 5.26 | 6.00E-07 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 9.4E-03 | 1.9E-03 | 1.1E-02 | 2.2E-03 |
| | 4.57 | 7.15 | 0.000779 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 9.4E-03 | 2.0E-03 | 1.1E-02 1.5E-02 | 3.1E-03 |
| Benzo(a)pyrene Benzo(b)fluoranthene | 20.7 | 9.75 | 9.17E-07 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 4.4E-02 | 2.0E-03 8.8E-03 | 2.1E-02 | 4.2E-03 |
| Benzo(g,h,i)perylene | 2.44 | 3.69 | 0.000125 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 5.2E-03 | 1.0E-03 | 7.9E-02 | 1.6E-03 |
| Benzo(k)fluoranthene | 19.8 | 11.0 | 0.000123 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 4.2E-03 | 8.5E-03 | 2.4E-02 | 4.7E-03 |
| Benzofluoranthenes, Total | 19.8 | 17.6 | 0.00182 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 4.2E-02 2.3E-02 | 4.6E-03 | 3.8E-02 | 7.5E-03 |
| Chrysene | 4.51 | 5.79 | 0.00000171 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 9.6E-03 | 4.0E-03 | 1.2E-02 | 2.5E-03 |
| Dibenz(a,h)anthracene | 1.12 | 1.04 | 5.86E-05 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 2.4E-03 | 4.8E-04 | 2.2E-02 | 4.4E-04 |
| Fluoranthene | 8.02 | 11.9 | 0.0000 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 1.7E-02 | 3.4E-03 | 2.5E-02 | 5.1E-03 |
| Indeno(1,2,3-cd)pyrene | 2.67 | 3.48 | 7.69E-05 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 5.7E-02 | 1.1E-03 | 7.4E-03 | 1.5E-03 |
| Pyrene | 7.09 | 10.4 | 0.0131 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0 | 0 | *** | *** | 1.6E-02 | 3.1E-03 | 2.3E-02 | 4.5E-03 |
| Total HPAHs | 47.1 | 53.0 | 0.0131 | 0 | 0 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.062 | 0.070 | 0.615 | 3.07 | 1.0E-02 1.0E-01 | 2.0E-02 | 2.3E-02 1.1E-01 | 4.3E-03 2.3E-02 |
| SVOCs | +/.1 | 55.0 | 0.0108 | 0 | 0 | 0.0031 | 0.034 | 0.097 | 0.15 | 0.231 | 0.974 | 0.002 | 0.070 | 0.015 | 5.07 | 1.0E-01 | 2.012-02 | 1.112-01 | 2.5E-02 |
| Bis(2-ethylhexyl) Phthalate | 36.4 | 31.3 | 0.0303 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.12 | 0.11 | 18.3 | 183 | 6.8E-03 | 6.8E-04 | 5.9E-03 | 5.9E-04 |
| Butyl Benzyl Phthalate | 0.0172 | 0.0154 | 0.0303 | 1 | 1 | 0.0051 | 0.054 | 0.097 | | 0.251 | 0.974 | 0.12 | 0.11 | 18.3 | 470 | 6.8E-03 4.1E-07 | 0.8E-04 1.4E-07 | 3.7E-07 | 5.9E-04 1.2E-07 |
| ~ ~ | 0.0172 | 0.0154 | 0.000240 | 1 | 1 | 0.0051 | 0.054 | | 0.15 | | | 0.000064 | 0.000058 | | | 4.1E-07 2.0E-04 | 1.4E-07 2.0E-05 | 3.7E-07 2.3E-04 | 1.2E-07 2.3E-05 |
| Carbazole | 0.272 | 0.312 | 0.00297 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 0.974 | 0.0010 | 0.00011 | 5 | 50 | 2.0E-04 4.7E-05 | 2.0E-05 9.4E-06 | | 2.3E-05 1.0E-05 |
| Dibenzofuran | | | | 1 | 1 | 0.0051 | | 0.097 | 0.15 | 0.251 | 1 | 0.00028 | 0.00030 | 6 550 | 30 | 4.7E-05 1.3E-06 | 9.4E-06 3.8E-07 | 5.1E-05 | |
| Di-n-butyl Phthalate Di-n-octyl Phthalate | 0.124 0.0210 | 0.105 0.0194 | 0.0107 1.49E-05 | 1 | 1 | 0.0051 | 0.054 0.054 | 0.097 0.097 | 0.15 | 0.251 0.251 | 0.974 0.974 | 0.00069 | 0.00062 | 550 7,500 | 1,833 37,500 | 9.6E-09 | 3.8E-07 1.9E-09 | 1.1E-06 8.8E-09 | 3.4E-07 1.8E-09 |
| Di-ii-Octyl Filulalate | 0.0210 | 0.0194 | 1.47E-03 | 1 | 1 | 0.0031 | 0.034 | 0.097 | 0.15 | 0.231 | 0.7/4 | 0.000072 | 0.000000 | 7,500 | 57,500 | 7.0E-09 | 1.76-07 | 0.0E-09 | 1.012-09 |

Table 3-15b Calculation of Dose and Hazard Quotient for the American Mink - Combined AOPCs

| | EPCs ^c | | | BAFs | | Exposure Factors | | | | | | Doses | | TRVs | | Hazard Quotients | | | |
|---|-------------------|--|---------------------|---|--|---|-------|---------------------------------------|------------------------|-------|---------|--|--|------------------------------------|------------------------------------|------------------|--------------|--|--------------|
| CPEC | (0-1 ft bgs) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | tration in Water | Surface Soil BAF mammals (kg tissue/kg soil) ^a | Shallow Soil BAF mammals (kg tissue/kg soil) ^a | Soil Ingestion Rate (kg dry/day) | Rate | Water Ingestion Rate (L/day) | PF Small Mammals | | BW (kg) | Surface Soil Dose (mg/kg- bw/day) | Shallow Soil Dose (mg/kg- bw/day) | NOAEL TRV (mg/kg- bw/day) | LOAEL TRV (mg/kg- bw/day) | (0-1 ft bgs) | (0-1 ft bgs) | Shallow Soil (0-3 ft bgs) NOAEL HQ | (0-3 ft bgs) |
| VOCs | · | | | | | | | | | | | | | | | • | | | |
| 1,2,4-Trimethylbenzene | 2.94 | 2.86 | 0.479 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.022 | 0.022 | 13 | 65 | 1.7E-03 | 3.4E-04 | 1.7E-03 | 3.3E-04 |
| 4-Isopropyltoluene | 0.00102 | 0.0147 | 9.11E-05 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.0000057 | 0.000052 | 47.5 | 237.5 | 1.2E-07 | 2.4E-08 | 1.1E-06 | 2.2E-07 |
| n-Propylbenzene | 0.000150 | 0.122 | 1.85E-05 | 1 | 1 | 0.0051 | 0.054 | 0.097 | 0.15 | 0.251 | 0.974 | 0.0000010 | 0.00042 | 60.4 | 302 | 1.6E-08 | 3.2E-09 | 6.9E-06 | 1.4E-06 |
| Notes: | | | | | | | | | | | | | | Metals + | Butyltin HI | 8.2E-01 | 2.8E-01 | 7.8E-01 | 2.8E-01 |
| Bold indicates hazard quotient greater than 1.0. | | | | | | | | | | | | | | Pesticides HI | | 3.9E-02 | 1.8E-02 | 2.4E-02 | 1.0E-02 |
| a) Sources listed on Table 3-10e for mammal BAFs. | | | | | | | | | | | | Ph | thalates HI | 6.8E-03 | 6.8E-04 | 5.9E-03 | 5.9E-04 | | |

b) Cadmium, copper, and lead (dissolved), benzo(a) anthracene, benzo(b) fluoranthene, chrysene, and fluoranthene (C+F) maximum detected concentrations in River OU Forebay surface water; see Table 3-6.

All other CPEC concentrations in water calculated using equilibrium partitioning equations due to lack of River OU analytical data; see Table 3-7e.

c) The EPC is the 95% UCL, if calculated (see Table 1-5), or maximum detected concentration (if 95% UCL not calculated).

*** = To be evaluated as Total HPAHs, per EcoSSL; however, HQs for individual HPAHs were calculated using the Total HPAH SLV to determine the primary risk drivers.

% = percentAOPC = area of potential concern AUF = Area Use Factor BAF = bioaccumulation factor BHC = benzene hexachloride BW = body weightC+F = column + filterCPEC = chemical of potential ecological concern Dose = average daily dose (mg/kg-bw/day) dw = dry weight EcoSSL = ecological soil screening level EPC = exposure point concentration ft bgs = feet below ground surface HI = (cumulative) hazard index HPAHs = high molecular weight polycyclic aromatic hydrocarbons HQ = hazard quotient

kg = kilogram kg dry/day = kilograms per day in dry weight L/day = liters per dayLOAEL = lowest observed adverse effect level mg/kg = milligrams per kilogram mg/kg-bw/day = milligrams per kilogram body weight per day mg/L = milligrams per liter NOAEL = no observed adverse effect level OU = operable unitPF = portion of food itemSLV = screening level value SVOC = semivolatile organic compound TRV = toxicity reference valueUCL = upper confidence limit VOC = volatile organic compound

| Table 3-16 |
|---|
| Terrestrial Plant and Soil Invertebrate Toxicity Benchmarks |

| Analyte | Background 95% UPL ^a (mg/kg) | Terrestrial Plant Low SLV (mg/kg) | Terrestrial Plant High SLV ^b (mg/kg) | Source | Soil Invertebrate Low SLV (mg/kg) | Soil Invertebrate High SLV ^b (mg/kg) | Source |
|--|---|--|--|---------------|--|--|--------|
| Inorganics | | | | | | | |
| Antimony | 0.176 | 5 | 25 | с | 78 | 390 | d |
| Cadmium | 0.271 | 32 | 160 | d | 140 | 700 | d |
| Chromium ^e | 28.1 | 1 | 5 | f | 0.4 | 2 | g |
| Copper | 56.7 | 70 | 350 | d | 80 | 400 | d |
| Lead | 25.5 | 120 | 600 | d | 1,700 | 8,500 | d |
| Mercury | 0.0660 | 0.3 | 1.5 | с | 0.1 | 0.5 | с |
| Nickel | 26.5 | 38 | 190 | d | 280 | 1,400 | d |
| Butyltins | | | | | | , | 1 |
| Tributyltin | NA | NA | NA | | NA | NA | 1 |
| Pesticides | 1111 | 1011 | 1111 | | 1111 | 141 | |
| BHC (delta) | NA | 0.005 | 0.03 | i | NA | NA | |
| BHC (dena) BHC (gamma) Lindane | NA | 0.005 | 0.03 | 1 h | NA | NA | |
| Chlordane (alpha) | NA | 0.005 | 1.12 | n i | NA | NA | |
| Chlordane (apna) Chlordane (gamma) | NA NA | 0.224 | 1.12 | 1 h | NA NA | NA NA | |
| Chlordane (gamma) Chlordane (technical) | NA | 0.224 | 1.12 | n i | NA | NA | |
| Endosulfan I | NA | 0.224 NA | NA | 1 | NA | NA | |
| | NA | NA | NA | | | NA | |
| Endosulfan II Endosulfan Sulfate | NA | NA | NA NA | | NA NA | NA NA | |
| Endosunan Sunate Endrin Aldehyde | NA | 0.011 | 0.054 | : | NA | NA | |
| Endrin Ketone | NA | 0.011 | 0.054 | i | NA | NA | |
| Endrin Ketone | NA | 0.011 | 0.054 | i | NA | NA | |
| Heptachlor | NA | 1 | 5 | J k | NA | NA | |
| Methoxychlor | NA | NA | NA | К | NA | NA | |
| HPAHs | INA | NA | NA | | INA | NA | |
| | | 1.2 | 6 | 1 | *** | *** | 1 |
| Benz(a)anthracene | NA | 1.2 | 6 | k | *** | *** | |
| Benzo(a)pyrene | NA | 1.2 | 6 | k | *** | *** | |
| Benzo(b)fluoranthene | NA | 1.2 | 6 | k | *** | *** | |
| Benzo(g,h,i)perylene | NA | 1.2 | 6 | 1 | *** | *** | |
| Benzo(k)fluoranthene | NA | 1.2 | 6 6 | k | *** | *** | |
| Benzofluoranthenes, Total | NA | 1.2 1.2 | | 1 k | *** | *** | |
| Chrysene Dibenz(a,h)anthracene | NA NA | 1.2 | 6 6 | k k | *** | *** | |
| | NA | 1.2 | 6 | <u>к</u> 1 | *** | *** | |
| Fluoranthene | | 1.2 | - | - | *** | *** | |
| Indeno(1,2,3-cd)pyrene Pyrene | NA NA | 1.2 | 6 6 | k 1 | *** | *** | |
| Total HPAHs | NA | 1.2 | 6 | l k | 18 | 90 | d |
| SVOCs | NA | 1.2 | 0 | ~ | 10 | 90 | u |
| | | 274 | | 1 | | | T |
| Bis(2-ethylhexyl) Phthalate | NA | NA | NA | | NA | NA | |
| Butyl Benzyl Phthalate | NA | NA | NA | | NA | NA | |
| Carbazole | NA | NA | NA | <u> </u> | NA | NA | |
| Dibenzofuran | NA | 19 | 96 | J | NA | NA | |
| Di-n-butyl Phthalate | NA | 200 NA | 1,000 NA | с | NA | NA | |
| Di-n-octyl Phthalate | NA | NA | NA | | NA | NA | |
| VOCs | | | | | 1 | | |
| 1,2,4-Trimethylbenzene | NA | NA | NA | | NA | NA | |
| 4-Isopropyltoluene | NA | NA | NA | | NA | NA | |
| n-Propylbenzene | NA | NA | NA | | NA | NA | |

Table 3-16 Terrestrial Plant and Soil Invertebrate Toxicity Benchmarks

Notes:

a) The Reference Area 95% UPL concentrations are shown in Appendix I, Table I-17 of the Final RI (URS. 2012. Upland and River Operable Units Remedial Investigation Report. June.).

b) High SLVs were calculated by multiplying the Low SLVs by 5.

c) Oregon Department of Environmental Quality. 1998. Guidance for Ecological Risk Assessment: Levels I, II, III, IV. Waste Management and Cleanup Division. Final. April. Updated December 2001.

d) U.S. Environmental Protection Agency (USEPA). 2005-2008. Interim Ecological Soil Screening Levels (Eco-SSLs).

Available on-line at: http://www.epa.gov/ecotox/ecossl/

e) The background concentration is used for screening because it is higher than the SLV(s).

f) Efroymson RA, ME Wil, GW Suter II, and AC Wooten. 1997. Toxicological Benchmarks for Screening Contaminants of Potential Concern for Effects on Terrestrial Plants: 1997 Revision. November. ORNL. ES/ER/TM-85/R3. From Oak Ridge National Laboratory (ORNL) Risk Assessment Information System (RAIS) Database.

g) Efroymson, R.A., M.E. Will, and G.W. Suter II. 1997. Toxicological Benchmarks for Contaminants of Potential Concern for Effects on Soil and Litter Invertebrates and Heterotrophic Process: 1997 Revision. November. ORNL. ES/ER/TM-126/R2. From ORNL RAIS Database.

h) USEPA. 2003. The ESL reference database consists of Region 5 media-specific (soil, water, sediment, and air) Ecological Screening Levels (ESLs) for RCRA Appendix IX hazardous constituents. See the August 2003 revision of the ESLs (formerly EDQLs) at http://www.epa.gov/reg5rcra/ca/ESL.pdf.

i) The following surrogate SLVs were used: Lindane for BHC (gamma), chlordane (gamma) for other chlordanes (alpha and technical), endrin for other endrins (aldehyde and ketone)

j) Los Alamos National Laboratory ECORISK Database Release 3.1, 2012. Geometric mean of no-effect ESLs and low-effect ESLs. Online: http://www.lanl.gov/community-environment/environmental-stewardship/protection/eco-risk-assessment.php

k) USEPA. 1999. Screening Level Ecological Risk Assessment Protocol for Hazardous Waste Combustion Facilities, Peer review Draft. EPA530-D-99-001A. August.

l) The SLV for benzo(a)pyrene was used to screen the individual HPAHs without SLVs.

*** = PAH Eco SSLs for Invertebrates are based on Total LMW PAHs and Total HMW PAHs.

% = percent BHC = benzene hexachloride EcoSSL = ecological soil screening level HPAHs = high molecular weight polycyclic aromatic hydrocarbons mg/kg = milligrams per kilogram NA = not available ORNL = Oak Ridge National Laboratory PAH = polycyclic aromatic hydrocarbons SLV = screening level value SVOC = semivolatile organic compound UPL = 95 % upper prediction limit VOC = volatile organic compound

| Table 3-17 |
|--|
| Toxicity Reference Values for Birds and Mammals |

| CPEC NOAEL/NOAE TRV for Birds (mg/kg-bw/day) | | | LOAEL/LOAEC TRV for Birds (mg/kg-bw/day) | | NOAEL/NOAEC TRV for Mammals (mg/kg-bw/day) | | LOAEL/LOAEC TRV for Mammals (mg/kg-bw/day) | |
|--|--------|------|--|---|---|------|---|------|
| Inorganics | | | | | | | | |
| Antimony | NA | a | NA | а | 0.059 | а | 0.59 | a |
| Cadmium | 1.47 | а | 5.88 | а | 0.770 | а | 7.70 | а |
| Chromium III | 2.66 | a | 14.5 | а | 2.40 | а | 12.0 | a |
| Copper | 4.05 | а | 12.1 | а | 5.60 | а | 9.34 | а |
| Lead | 1.63 | а | 3.26 | а | 4.70 | а | 8.90 | а |
| Mercury | 0.0064 | b | 0.064 | b | 0.02 | b | 0.07 | b |
| Nickel | 6.71 | а | 26.8 | а | 1.70 | а | 3.40 | а |
| Butyltins | | | | | | | | |
| Tributyltin | 6.8 | b | 16.9 | b | 23.4 | c, k | 35 | c, k |
| Pesticides | | | | | • | | | |
| BHC (delta) | 2 | k | 20 | k | 8 | k | 40 | k |
| BHC (gamma) Lindane | 2 | с | 20 | с | 8 | с | 40 | i |
| Chlordane (alpha) | 2.14 | k | 10.7 | k | 4.6 | k | 9.2 | k |
| Chlordane (gamma) | 2.14 | k | 10.7 | k | 4.6 | k | 9.2 | k |
| Chlordane (technical) | 2.14 | с | 10.7 | с | 4.6 | с | 9.2 | с |
| Endosulfan I | 10 | c, k | 50 | i | 0.15 | c, k | 0.75 | i |
| Endosulfan II | 10 | c, k | 50 | i | 0.15 | c, k | 0.75 | i |
| Endosulfan Sulfate | 10 | c, k | 50 | i | 0.15 | c, k | 0.75 | i |
| Endrin Aldehyde | 0.01 | k | 0.1 | k | 0.092 | k | 0.92 | k |
| Endrin Ketone | 0.01 | k | 0.1 | k | 0.092 | k | 0.92 | k |
| Endrin | 0.01 | с | 0.1 | с | 0.092 | с | 0.92 | с |
| Heptachlor | 0.013 | i | 0.065 | е | 0.1 | с | 1 | с |
| Methoxychlor | 125 | f | 625 | i | 4 | с | 8 | с |
| HPAHs | | | | | | | | |
| Benz(a)anthracene | NA | а | NA | а | *** | а | *** | a |
| Benzo(a)pyrene | NA | a | NA | a | *** | a | *** | a |
| Benzo(b)fluoranthene | NA | a | NA | a | *** | a | *** | a |
| Benzo(g,h,i)perylene | NA | a | NA | a | *** | a | *** | a |
| Benzo(k)fluoranthene | NA | a | NA | a | *** | a | *** | a |
| Benzofluoranthenes, Total | NA | а | NA | а | *** | а | *** | а |
| Chrysene | NA | а | NA | а | *** | а | *** | а |
| Dibenz(a,h)anthracene | NA | а | NA | а | *** | а | *** | а |
| Fluoranthene | NA | а | NA | а | *** | а | *** | а |
| Indeno(1,2,3-cd)pyrene | NA | а | NA | а | *** | а | *** | а |
| Pyrene | NA | a | NA | a | *** | a | *** | a |
| Total HPAHs (benzo(a)pyrene) | NA | а | NA | а | 0.615 | а | 3.07 | а |
| SVOCs | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 1.1 | b | 11 | b | 18.3 | с | 183 | с |
| Butyl Benzyl Phthalate | NA | ~ | NA | | 159 | d | 470 | d |
| Carbazole | NA | | NA | | 5 | g | 50 | g |
| Dibenzofuran | NA | | NA | | 6 | i | 30 | i |
| Di-n-butyl Phthalate | 0.11 | с | 1.1 | с | 550 | c | 1,833 | c |
| Di-n-octyl Phthalate | NA | - | NA | - | 7,500 | e | 37,500 | i |

| Table 3-17 |
|--|
| Toxicity Reference Values for Birds and Mammals |

| CPEC | NOAEL/NOAEC TRV for Birds (mg/kg-bw/day) | | LOAEL/LOAEC TRV for Birds (mg/kg-bw/day) | | NOAEL/NOAEC TRV for Mammals (mg/kg-bw/day) | | LOAEL/LOAEC TRV for Mammals (mg/kg-bw/day) | |
|------------------------|--|---|--|---|---|---|---|---|
| VOCs | | | | | | | | |
| 1,2,4-Trimethylbenzene | NA | | NA | | 13 | j | 65 | h |
| 4-Isopropyltoluene | 3.16 | m | 15.8 | m | 47.5 | 1 | 238 | 1 |
| n-Propylbenzene | NA | | NA | | 60.4 | 1 | 302 | 1 |

Notes:

*** = To be evaluated as Total HPAHs, per EcoSSL

BHC = benzene hexachloride

EcoSSL = ecological soil screening level

HPAHs = high molecular weight polycyclic aromatic hydrocarbons

LOAEL = lowest observed adverse effect level

LOAEC = lowest observed adverse effect concentration

mg/kg-bw/day = milligrams per kilogram body weight per day

NA = no available TRV

Sources:

a) Interim Ecological Soil Screening Levels (Eco-SSLs; USEPA 2005-2008).

1) The Eco-SSL NOAEL TRV is a geometric mean of the NOAEL values for reproduction and growth. If the geometric mean NOAEL TRV is lower than the lowest bounded LOAEL for reproduction, growth, or survival, then a geometric mean LOEAL TRV was calculated with the LOAELs from the studies used for the geometric mean NOAEL calculation. If the studies used for the geometric mean NOAEL do not have LOAEL values, then an uncertainty factor of 5 was used to adjust the geometric mean NOAEL to a chronic LOAEL.

2) If the geometric mean NOAEL TRV is higher than the lowest bounded LOAEL for reproduction, growth, or mortality results, then the TRV selected is equal to the highest bounded NOAEL below the lowest bounded LOAEL for reproduction, growth, or survival. The selected LOAEL is the LOAEL from the same study as the highest bounded NOAEL. If no LOAEL is available from that study, then an uncertainty factor of 5 was used to adjust the selected NOAEL to chronic LOAEL.

b) LWG (2011); Appendix G: BERA, Tables 8-9 and 8-10.

<u>Birds</u>

Mercury: Heinz (1975, 1979) Tributyltin: Schlatterer et al. (1993) Benzo(a)pyrene: Hough et al. (1993) Bis(2-ethylhexyl) phthalate: Peakall (1974) <u>Mammals</u>

Mercury: Dansereau et al. (1999)

c) Sample et al. (1996), ORNL

d) USEPA Integrated Risk Information System (IRIS) database

Butyl benzyl phthalate - TRVs based on oral chronic rat toxicity (NTP 1985)

e) USEPA (1999), Appendix E, TRVs

Di-n-octyl phthalate for mammals: Heindel et al. (1989) - TRV based on chronic NOAEL

Heptachlor for birds: Hill and Camardese (1986) - TRV based on acute LOAEL for reduced survival (adjusted to chronic) f) Hunt and Sacho (1969) - Acute five-day NOAEL of 3,750 mg/kg/day for robins fed earthworms, mortality endpoint.

Divided by 30 (USACHPPM 2000) to approximate chronic NOAEL of 125 mg/kg/day.

g) Sax (1984) - TRVs based on chronic toxicity.

h) Maltoni et al. (1997) - TRV based on chronic LOAEL.

NOAEL = no observed adverse effect level NOAEC = no observed adverse effect concentration ORNL = Oak Ridge National Laboratory SVOC = semivolatile organic compound TRV = toxicity reference value VOC = volatile organic compound

Table 3-17Toxicity Reference Values for Birds and Mammals

i) NTP (1989) as cited in ATSDR (1990) - TRV based on chronic toxicity of 2,3-benzofuran (reduced survival)

j) An uncertainty factor of 5 was used to adjust chronic LOAEL/LOEACs to NOAEL/NOAECs and vise-versa.

k) Surrogates used:

BHC (gamma) Lindane TRVs for BHC (delta) for birds and mammals Chlordane TRVs for chlordane (alpha) and chlordane (gamma) for birds and mammals Endosulfan TRVs for endosulfan I, II, and sulfate for birds and mammals Endrin TRVs for endrin aldehyde and ketone for birds and mammals Tin TRV, based on bis(tributyltin)oxide (TBTO) for tributyltin for mammals

1) Hazardous Substances Data Bank (HSDB)

4-Isopropyltoluene (O'Neil ed. 2006): Rat LD50 of 4,750 mg/kg/day adjusted to chronic NOAEL and LOAEL by dividing by 100 and 20, respectively (USACHPPM 2000).

n-Propylbenzene (Budavari ed. 1996): Rat LD50 of 6,040 mg/kg/day adjusted to chronic NOAEL and LOAEL by dividing by 100 and 20, respectively (USACHPPM 2000).

m) Schafer et al. (1983). Red-winged blackbird LD50 of 316 mg/kg/day adjusted to chronic NOAEL and LOAEL by dividing by 100 and 20, respectively (USACHPPM 2000).

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Table 3-17Toxicity Reference Values for Birds and Mammals

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Table 3-18aHazard Quotients for Terrestrial Plants - Landfill AOPC

| | EP | Cs ^d | SL | .Vs | | Hazard (| Quotients ^c | |
|-----------------------------|--|--|--|---|-----------|--|------------------------|--------------|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Terrestrial Plant Low SLV (mg/kg dw) ^a | Terrestrial Plant High SLV (mg/kg dw) ^a | 1 ft bgs) | Surface Soil (0- 1 ft bgs) High SLV HQ | 3 ft bgs) | (0-3 ft bgs) |
| Inorganics | | | | | | | | |
| Antimony | 1.36 | 1.48 | 5 | 25 | 2.7E-01 | 5.4E-02 | 3.0E-01 | 5.9E-02 |
| Chromium ^b | 242 | 594 | 28.1 | 28.1 | 8.6E+00 | 8.6E+00 | 2.1E+01 | 2.1E+01 |
| Copper | 170 | 191 | 70 | 350 | 2.4E+00 | 4.9E-01 | 2.7E+00 | 5.5E-01 |
| Lead | 332 | 511 | 120 | 600 | 2.8E+00 | 5.5E-01 | 4.3E+00 | 8.5E-01 |
| Mercury | 1.57 | 1.45 | 0.3 | 1.5 | 5.2E+00 | 1.0E+00 | 4.8E+00 | 9.7E-01 |
| Nickel | 175 | 472 | 38 | 190 | 4.6E+00 | 9.2E-01 | 1.2E+01 | 2.5E+00 |
| Butyltins | | | | | | | | |
| Tributyltin | 0.165 | 0.0601 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Pesticides | | | | | | - | | |
| Chlordane (technical) | 1.56 | 1.56 | 0.224 | 1.12 | 7.0E+00 | 1.4E+00 | 7.0E+00 | 1.4E+00 |
| Heptachlor | 0.00283 | 0.00286 | 1 | 5 | 2.8E-03 | 5.7E-04 | 2.9E-03 | 5.7E-04 |
| HPAHs | | | | | | - | | |
| Benz(a)anthracene | 8 | 10 | 1.2 | 6 | 7.0E+00 | 1.4E+00 | 8.3E+00 | 1.7E+00 |
| Benzo(a)pyrene | 9 | 11 | 1.2 | 6 | 7.3E+00 | 1.5E+00 | 9.3E+00 | 1.9E+00 |
| Benzo(b)fluoranthene | 16 | 15 | 1.2 | 6 | 1.3E+01 | 2.7E+00 | 1.3E+01 | 2.5E+00 |
| Benzo(g,h,i)perylene | 5 | 58 | 1.2 | 6 | 3.8E+00 | 7.7E-01 | 4.9E+01 | 9.7E+00 |
| Benzo(k)fluoranthene | 30 | 26 | 1.2 | 6 | 2.5E+01 | 5.0E+00 | 2.2E+01 | 4.4E+00 |
| Benzofluoranthenes, Total | 15 | 31 | 1.2 | 6 | 1.2E+01 | 2.5E+00 | 2.6E+01 | 5.2E+00 |
| Chrysene | 8 | 11 | 1.2 | 6 | 7.0E+00 | 1.4E+00 | 9.2E+00 | 1.8E+00 |
| Dibenz(a,h)anthracene | 4 | 21 | 1.2 | 6 | 3.5E+00 | 7.1E-01 | 1.7E+01 | 3.4E+00 |
| Fluoranthene | 15 | 18 | 1.2 | 6 | 1.2E+01 | 2.4E+00 | 1.5E+01 | 2.9E+00 |
| Indeno(1,2,3-cd)pyrene | 5 | 68 | 1.2 | 6 | 4.2E+00 | 8.4E-01 | 5.7E+01 | 1.1E+01 |
| Pyrene | 12 | 19 | 1.2 | 6 | 1.0E+01 | 2.1E+00 | 1.6E+01 | 3.1E+00 |
| Total HPAHs | 90 | 102 | 1.2 | 6 | 7.5E+01 | 1.5E+01 | 8.5E+01 | 1.7E+01 |
| SVOCs | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 9.68 | 6.38 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Butyl Benzyl Phthalate | 0.0366 | 0.0349 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Carbazole | 0.751 | 0.964 | NA | NA | No SLV | No SLV | No SLV | No SLV |

Table 3-18a Hazard Quotients for Terrestrial Plants - Landfill AOPC

| | EP | Cs ^d | SL | ZVs | Hazard Quotients ^c | | | | |
|------------------------|--|--|--|---|-------------------------------|-----------|---|---|--|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Terrestrial Plant Low SLV (mg/kg dw) ^a | Terrestrial Plant High SLV (mg/kg dw) ^a | 1 ft bgs) | 1 ft bgs) | Shallow Soil (0- 3 ft bgs) Low SLV HQ | Shallow Soil (0-3 ft bgs) High SLV HQ | |
| Dibenzofuran | 0.152 | 0.164 | 19 | 96 | 7.9E-03 | 1.6E-03 | 8.5E-03 | 1.7E-03 | |
| Di-n-butyl Phthalate | 0.232 | 0.217 | 200 | 1,000 | 1.2E-03 | 2.3E-04 | 1.1E-03 | 2.2E-04 | |
| VOCs | | | | | | | | | |
| 1,2,4-Trimethylbenzene | 3.03 | 2.79 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| Notes: | | | | Metals HI | 2.4E+01 | 1.2E+01 | 4.6E+01 | 2.6E+01 | |

Pesticides HI

7.0E+00

1.4E+00

Notes:

a) Sources listed on Table 3-16.

b) The background concentration (95% UPL) is higher than the risk-

based SLV and replaced the SLV in this table.

c) Those chemicals that do not have established SLVs will be evaluated qualitatively.

d) The EPC is the 95% UCL, if calculated (see Table 1-1), or maximum detected concentration (if 95% UCL not calculated).

Bold indicates hazard quotient greater than 1.0.

% = percent

- AOPC = area of potential concern
- CPEC = chemical of potential ecological concern

dw = dry weight

EPC = exposure point concentration

ft bgs = feet below ground surface

HI = (cumulative) hazard index

HPAH = high molecular weight polycyclic aromatic hydrocarbons

HQ = hazard quotient

mg/kg = milligrams per kilogram

NA = not available

SLV = screening level value

SVOC = semivolatile organic compound

VOC = volatile organic compound

UCL = upper confidence limit

UPL = upper prediction limit

1.4E+00

7.0E+00

 Table 3-18b

 Hazard Quotients for Terrestrial Plants – Sandblast Area AOPC

| | EP | Cs ^d | SL | Vs | | Hazard (| Juotients ^c | |
|---------------------------|--|--|--|---|---|--|------------------------|---|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Terrestrial Plant Low SLV (mg/kg dw) ^a | Terrestrial Plant High SLV (mg/kg dw) ^a | Surface Soil (0- 1 ft bgs) Low SLV HQ | Surface Soil (0- 1 ft bgs) High SLV HQ | 3 ft bgs) | Shallow Soil (0-3 ft bgs) High SLV HQ |
| Inorganics | | | | | | | | |
| Antimony | 2.94 | 2.42 | 5 | 25 | 5.9E-01 | 1.2E-01 | 4.8E-01 | 9.7E-02 |
| Cadmium | 3.20 | 2.63 | 32 | 160 | 1.0E-01 | 2.0E-02 | 8.2E-02 | 1.6E-02 |
| Chromium ^b | 720 | 579 | 28.1 | 28.1 | 2.6E+01 | 2.6E+01 | 2.1E+01 | 2.1E+01 |
| Lead | 465 | 529 | 120 | 600 | 3.9E+00 | 7.8E-01 | 4.4E+00 | 8.8E-01 |
| Mercury | 0.113 | 0.098 | 0.3 | 1.5 | 3.8E-01 | 7.5E-02 | 3.3E-01 | 6.5E-02 |
| Nickel | 353 | 251 | 38 | 190 | 9.3E+00 | 1.9E+00 | 6.6E+00 | 1.3E+00 |
| Butyltins | | | | | | | | |
| Tributyltin | 0.481 | 0.248 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Pesticides | | | • | | • | • | • | |
| BHC (delta) | 0.00154 | 0.00130 | 0.005 | 0.03 | 3.1E-01 | 6.2E-02 | 2.6E-01 | 5.2E-02 |
| BHC (gamma) Lindane | 0.00968 | 0.00106 | 0.005 | 0.03 | 1.9E+00 | 3.9E-01 | 2.1E-01 | 4.2E-02 |
| Chlordane (alpha) | 0.000815 | 0.000776 | 0.224 | 1.12 | 3.6E-03 | 7.3E-04 | 3.5E-03 | 6.9E-04 |
| Chlordane (gamma) | 0.0303 | 0.0204 | 0.224 | 1.12 | 1.4E-01 | 2.7E-02 | 9.1E-02 | 1.8E-02 |
| Endosulfan I | 0.00269 | 0.000809 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Endosulfan II | 0.000495 | 0.000447 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Endosulfan Sulfate | 0.000953 | 0.000889 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Endrin Aldehyde | 0.00387 | 0.00344 | 0.011 | 0.054 | 3.6E-01 | 7.2E-02 | 3.2E-01 | 6.4E-02 |
| Endrin Ketone | 0.00309 | 0.00288 | 0.011 | 0.054 | 2.9E-01 | 5.7E-02 | 2.7E-01 | 5.4E-02 |
| Endrin | 0.00504 | 0.00389 | 0.011 | 0.054 | 4.7E-01 | 9.4E-02 | 3.6E-01 | 7.2E-02 |
| Heptachlor | 0.000464 | 0.000415 | 1 | 5 | 4.6E-04 | 9.3E-05 | 4.2E-04 | 8.3E-05 |
| Methoxychlor | 0.00105 | 0.00104 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| HPAHs | | | | | | | | |
| Benz(a)anthracene | 1.75 | 2.45 | 1.2 | 6 | 1.5E+00 | 2.9E-01 | 2.0E+00 | 4.1E-01 |
| Benzo(a)pyrene | 1.76 | 2.39 | 1.2 | 6 | 1.5E+00 | 2.9E-01 | 2.0E+00 | 4.0E-01 |
| Benzo(b)fluoranthene | 1.95 | 2.21 | 1.2 | 6 | 1.6E+00 | 3.3E-01 | 1.8E+00 | 3.7E-01 |
| Benzo(g,h,i)perylene | 1.07 | 1.05 | 1.2 | 6 | 8.9E-01 | 1.8E-01 | 8.8E-01 | 1.8E-01 |
| Benzo(k)fluoranthene | 0.676 | 0.764 | 1.2 | 6 | 5.6E-01 | 1.1E-01 | 6.4E-01 | 1.3E-01 |
| Benzofluoranthenes, Total | 7.76 | 10.6 | 1.2 | 6 | 6.5E+00 | 1.3E+00 | 8.8E+00 | 1.8E+00 |
| Chrysene | 2.07 | 4.59 | 1.2 | 6 | 1.7E+00 | 3.5E-01 | 3.8E+00 | 7.7E-01 |
| Dibenz(a,h)anthracene | 0.244 | 0.238 | 1.2 | 6 | 2.0E-01 | 4.1E-02 | 2.0E-01 | 4.0E-02 |
| Fluoranthene | 4.64 | 10.8 | 1.2 | 6 | 3.9E+00 | 7.7E-01 | 9.0E+00 | 1.8E+00 |

 Table 3-18b

 Hazard Quotients for Terrestrial Plants – Sandblast Area AOPC

| | EP | Cs ^d | SL | Vs | Hazard Quotients ^c | | | | |
|----------------------------------|--|--|--|---|---|--|-----------|---|--|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Terrestrial Plant Low SLV (mg/kg dw) ^a | Terrestrial Plant High SLV (mg/kg dw) ^a | Surface Soil (0- 1 ft bgs) Low SLV HQ | Surface Soil (0- 1 ft bgs) High SLV HQ | 3 ft bgs) | Shallow Soil (0-3 ft bgs) High SLV HQ | |
| Indeno(1,2,3-cd)pyrene | 1.13 | 1.16 | 1.2 | 6 | 9.4E-01 | 1.9E-01 | 9.7E-01 | 1.9E-01 | |
| Pyrene | 4.78 | 11.7 | 1.2 | 6 | 4.0E+00 | 8.0E-01 | 9.8E+00 | 2.0E+00 | |
| Total HPAHs | 18.6 | 40.8 | 1.2 | 6 | 1.5E+01 | 3.1E+00 | 3.4E+01 | 6.8E+00 | |
| SVOCs | | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 98.0 | 77.1 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| Butyl Benzyl Phthalate | 0.0171 | 0.0152 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| Carbazole | 0.144 | 0.112 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| Dibenzofuran | 0.0424 | 0.122 | 19 | 96 | 2.2E-03 | 4.4E-04 | 6.3E-03 | 1.3E-03 | |
| Di-n-butyl Phthalate | 0.0804 | 0.0543 | 200 | 1,000 | 4.0E-04 | 8.0E-05 | 2.7E-04 | 5.4E-05 | |
| Di-n-octyl Phthalate | 0.0251 | 0.0225 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| VOCs | | | | | | | | | |
| 1,2,4-Trimethylbenzene | 0.000230 | 4.62 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| 4-Isopropyltoluene | 0.00133 | 0.0201 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| n-Propylbenzene | 0.000150 | 0.0355 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| Notes: | | | | Metals HI | 4.0E+01 | 2.8E+01 | 3.3E+01 | 2.3E+01 | |
| a) Sources listed on Table 3-16. | | | | Pesticides HI | 3.5E+00 | 7.0E-01 | 1.5E+00 | 3.0E-01 | |

b) The background concentration (95% UPL) is higher than the risk-

based SLV and replaced the SLV in this table.

c) Those chemicals that do not have established SLVs will be evaluated qualitatively.

d) The EPC is the 95% UCL, if calculated (see Table 1-2), or maximum detected concentration (if 95% UCL not calculated).

Bold indicates hazard quotient greater than 1.0.

| % = percent | HQ = hazard quotient |
|---|--------------------------------------|
| AOPC = area of potential concern | mg/kg = milligrams per kilogram |
| BHC = benzene hexachloride | NA = not available |
| CPEC = chemical of potential ecological concern | SLV = screening level value |
| dw = dry weight | SVOC = semivolatile organic compound |
| EPC = exposure point concentration | VOC = volatile organic compound |
| ft bgs = feet below ground surface | UCL = upper confidence limit |
| HI = (cumulative) hazard index | UPL = upper prediction limit |
| HPAH = high molecular weight polycyclic aromatic hydrocarbons | |

Table 3-18cHazard Quotients for Terrestrial Plants – Pistol Range AOPC

| | | EPC | EPC SLV | | | Quotients |
|------------|-----|--|--|---|-------------|--|
| CI | PEC | 95% UCL in Surface Soil (0-1.5 ft bgs) (mg/kg dw) | Terrestrial Plant Low SLV (mg/kg dw) ^a | Terrestrial Plant High SLV (mg/kg dw) ^a | 1.5 ft bgs) | Surface Soil (0- 1.5 ft bgs) High SLV HQ |
| Inorganics | | | | | | |
| Lead | | 365 | 120 | 600 | 3.0E+00 | 6.1E-01 |

Notes:

a) Sources listed on Table 3-16.

Bold indicates hazard quotient greater than 1.0.

% = percent

AOPC = area of potential concern

CPEC = chemical of potential ecological concern

dw = dry weight

EPC = exposure point concentration

ft bgs = feet below ground surface

HQ = hazard quotient

mg/kg = milligrams per kilogram

SLV = screening level value

UCL = upper confidence limit

 Table 3-18d

 Hazard Quotients for Terrestrial Plants – Bulb Slope AOPC

| | EPC | SI | SLV Hazard Quotients | | | |
|------------|--|--|---|---|--|--|
| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Terrestrial Plant Low SLV (mg/kg dw) ^a | Terrestrial Plant High SLV (mg/kg dw) ^a | Surface Soil (0- 1 ft bgs) Low SLV HQ | Surface Soil (0- 1 ft bgs) High SLV HQ | |
| Inorganics | | | | | | |
| Lead | 307 | 120 | 600 | 2.6E+00 | 5.12E-01 | |
| Mercury | 0.720 | 0.3 | 1.5 | 2.4E+00 | 4.80E-01 | |
| | | | HI | 5.0E+00 | 9.9E-01 | |

Notes:

a) Sources listed on Table 3-16.

Bold indicates hazard quotient greater than 1.0.

% = percent

AOPC = area of potential concern

CPEC = chemical of potential ecological concern

dw = dry weight

EPC = exposure point concentration

ft bgs = feet below ground surface

HI = (cumulative) hazard index

HQ = hazard quotient

mg/kg = milligrams per kilogram

SLV = screening level value

UCL = upper confidence limit

Table 3-19aHazard Quotients for Soil Invertebrates - Landfill AOPC

| | EP | Cs ^d | SI | LVs | | Hazard | Quotients ^c | |
|-----------------------------|--|--|---|--|-----------|--|------------------------|---|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Soil Invertebrate Low SLV (mg/kg dw) ^a | Soil Invertebrate High SLV (mg/kg dw) ^a | 1 ft bgs) | Surface Soil (0- 1 ft bgs) High SLV HQ | 3 ft bgs) | Shallow Soil (0-3 ft bgs) High SLV HQ |
| Inorganics | | | | | | • | | |
| Antimony | 1.36 | 1.48 | 78 | 390 | 1.7E-02 | 3.5E-03 | 1.9E-02 | 3.8E-03 |
| Chromium ^b | 242 | 594 | 28.1 | 28.1 | 8.6E+00 | 8.6E+00 | 2.1E+01 | 2.1E+01 |
| Copper | 170 | 191 | 80 | 400 | 2.1E+00 | 4.3E-01 | 2.4E+00 | 4.8E-01 |
| Lead | 332 | 511 | 1,700 | 8,500 | 2.0E-01 | 3.9E-02 | 3.0E-01 | 6.0E-02 |
| Mercury | 1.57 | 1.45 | 0.1 | 0.5 | 1.6E+01 | 3.1E+00 | 1.5E+01 | 2.9E+00 |
| Nickel | 175 | 472 | 280 | 1,400 | 6.3E-01 | 1.3E-01 | 1.7E+00 | 3.4E-01 |
| Butyltins | | | | | | | | |
| Tributyltin | 0.165 | 0.0601 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Pesticides | • | | | | | | | |
| Chlordane (technical) | 1.56 | 1.56 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Heptachlor | 0.00283 | 0.00286 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| HPAHs . | • | | | | | | | |
| Benz(a)anthracene | 8 | 10 | *** | *** | 4.6E-01 | 9.3E-02 | 5.5E-01 | 1.1E-01 |
| Benzo(a)pyrene | 9 | 11 | *** | *** | 4.8E-01 | 9.7E-02 | 6.2E-01 | 1.2E-01 |
| Benzo(b)fluoranthene | 16 | 15 | *** | *** | 8.9E-01 | 1.8E-01 | 8.4E-01 | 1.7E-01 |
| Benzo(g,h,i)perylene | 5 | 58 | *** | *** | 2.6E-01 | 5.1E-02 | 3.2E+00 | 6.5E-01 |
| Benzo(k)fluoranthene | 30 | 26 | *** | *** | 1.7E+00 | 3.4E-01 | 1.5E+00 | 2.9E-01 |
| Benzofluoranthenes, Total | 15 | 31 | *** | *** | 8.2E-01 | 1.6E-01 | 1.7E+00 | 3.5E-01 |
| Chrysene | 8 | 11 | *** | *** | 4.7E-01 | 9.3E-02 | 6.1E-01 | 1.2E-01 |
| Dibenz(a,h)anthracene | 4 | 21 | *** | *** | 2.4E-01 | 4.7E-02 | 1.1E+00 | 2.3E-01 |
| Fluoranthene | 15 | 18 | *** | *** | 8.1E-01 | 1.6E-01 | 9.7E-01 | 1.9E-01 |
| Indeno(1,2,3-cd)pyrene | 5 | 68 | *** | *** | 2.8E-01 | 5.6E-02 | 3.8E+00 | 7.6E-01 |
| Pyrene | 12 | 19 | *** | *** | 6.8E-01 | 1.4E-01 | 1.0E+00 | 2.1E-01 |
| Total HPAHs | 90 | 102 | 18 | 90 | 5.0E+00 | 1.0E+00 | 5.6E+00 | 1.1E+00 |
| SVOC | | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 9.68 | 6.38 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Butyl Benzyl Phthalate | 0.0366 | 0.0349 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Carbazole | 0.751 | 0.964 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Dibenzofuran | 0.152 | 0.164 | NA | NA | No SLV | No SLV | No SLV | No SLV |

Table 3-19a Hazard Quotients for Soil Invertebrates - Landfill AOPC

| | EP | Cs ^d | SLVs | | | Hazard Quotients ^c | | | |
|------------------------|--|--|---|--|-----------|--|---|--------------|--|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Soil Invertebrate Low SLV (mg/kg dw) ^a | Soil Invertebrate High SLV (mg/kg dw) ^a | 1 ft bgs) | Surface Soil (0- 1 ft bgs) High SLV HQ | Shallow Soil (0- 3 ft bgs) Low SLV HQ | (0-3 ft bgs) | |
| Di-n-butyl Phthalate | 0.232 | 0.217 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| VOCs | | | | | | | | | |
| 1,2,4-Trimethylbenzene | 3.03 | 2.79 | NA | NA | No SLV | No SLV | No SLV | No SLV | |
| Notes: | | | | Metals HI | 2.7E+01 | 1.2E+01 | 4.0E+01 | 2.5E+01 | |

a) Sources listed on Table 3-16.

b) The background concentration (95% UPL) is higher than the risk-based

SLV and replaced the SLV in this table.

c) Those chemicals that do not have established SLVs will be evaluated qualitatively.

d) The EPC is the 95% UCL, if calculated (see Table 1-1), or maximum detected concentration (if 95% UCL not calculated).

Bold indicates hazard quotient greater than 1.0.

*** = To be evaluated as Total HPAHs, per EcoSSL; however, HQs for individual HPAHs were calculated using the Total HPAH SLV to determine the primary risk drivers.

% = percent

AOPC = area of potential concern CPEC = chemical of potential ecological concern dw = dry weight

EcoSSL = ecological soil screening level

EPC = exposure point concentration

ft bgs = feet below ground surface

HI = (cumulative) hazard index

HPAH = high molecular weight polycyclic aromatic hydrocarbons

HQ = hazard quotient

mg/kg = milligrams per kilogram

NA = not available

SLV = screening level value

SVOC = semivolatile organic compound

VOC = volatile organic compound

UCL = upper confidence limit

 Table 3-19b

 Hazard Quotients for Soil Invertebrates – Sandblast Area AOPC

| | EP | Cs ^d | SI | .Vs | | Hazard (| Quotients ^c | |
|---------------------------|--|--|---|--|---|--|---|---|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Soil Invertebrate Low SLV (mg/kg dw) ^a | Soil Invertebrate High SLV (mg/kg dw) ^a | Surface Soil (0- 1 ft bgs) Low SLV HQ | Surface Soil (0- 1 ft bgs) High SLV HQ | Shallow Soil (0- 3 ft bgs) Low SLV HQ | Shallow Soil (0-3 ft bgs) High SLV HQ |
| Inorganics | | | | | | | | |
| Antimony | 2.94 | 2.42 | 78 | 390 | 3.8E-02 | 7.5E-03 | 3.1E-02 | 6.2E-03 |
| Cadmium | 3.20 | 2.63 | 140 | 700 | 2.3E-02 | 4.6E-03 | 1.9E-02 | 3.8E-03 |
| Chromium ^b | 720 | 579 | 28.1 | 28.1 | 2.6E+01 | 2.6E+01 | 2.1E+01 | 2.1E+01 |
| Lead | 465 | 529 | 1,700 | 8,500 | 2.7E-01 | 5.5E-02 | 3.1E-01 | 6.2E-02 |
| Mercury | 0.113 | 0.098 | 0.1 | 0.5 | 1.1E+00 | 2.3E-01 | 9.8E-01 | 2.0E-01 |
| Nickel | 353 | 251 | 280 | 1,400 | 1.3E+00 | 2.5E-01 | 9.0E-01 | 1.8E-01 |
| Butyltins | | | | • | | | | |
| Tributyltin | 0.481 | 0.248 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Pesticides | | | | L | • | | | |
| BHC (delta) | 0.00154 | 0.00130 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| BHC (gamma) Lindane | 0.00968 | 0.00106 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Chlordane (alpha) | 0.000815 | 0.000776 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Chlordane (gamma) | 0.0303 | 0.0204 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Endosulfan I | 0.00269 | 0.000809 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Endosulfan II | 0.000495 | 0.000447 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Endosulfan Sulfate | 0.000953 | 0.000889 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Endrin Aldehyde | 0.00387 | 0.00344 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Endrin Ketone | 0.00309 | 0.00288 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Endrin | 0.00504 | 0.00389 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Heptachlor | 0.000464 | 0.000415 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Methoxychlor | 0.00105 | 0.00104 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| HPAHs | | | | | | | | |
| Benz(a)anthracene | 1.75 | 2.45 | *** | *** | 9.7E-02 | 1.9E-02 | 1.4E-01 | 2.7E-02 |
| Benzo(a)pyrene | 1.76 | 2.39 | *** | *** | 9.8E-02 | 2.0E-02 | 1.3E-01 | 2.7E-02 |
| Benzo(b)fluoranthene | 1.95 | 2.21 | *** | *** | 1.1E-01 | 2.2E-02 | 1.2E-01 | 2.5E-02 |
| Benzo(g,h,i)perylene | 1.07 | 1.05 | *** | *** | 5.9E-02 | 1.2E-02 | 5.8E-02 | 1.2E-02 |
| Benzo(k)fluoranthene | 0.676 | 0.764 | *** | *** | 3.8E-02 | 7.5E-03 | 4.2E-02 | 8.5E-03 |
| Benzofluoranthenes, Total | 7.76 | 10.6 | *** | *** | 4.3E-01 | 8.6E-02 | 5.9E-01 | 1.2E-01 |
| Chrysene | 2.07 | 4.59 | *** | *** | 1.2E-01 | 2.3E-02 | 2.6E-01 | 5.1E-02 |
| Dibenz(a,h)anthracene | 0.244 | 0.238 | *** | *** | 1.4E-02 | 2.7E-03 | 1.3E-02 | 2.6E-03 |
| Fluoranthene | 4.64 | 10.8 | *** | *** | 2.6E-01 | 5.2E-02 | 6.0E-01 | 1.2E-01 |

 Table 3-19b

 Hazard Quotients for Soil Invertebrates – Sandblast Area AOPC

| | EP | Cs ^d | SL | .Vs | | Hazard (| Quotients ^c | |
|-----------------------------|--|--|---|--|---|--|---|---|
| CPEC | Surface Soil (0-1 ft bgs) (mg/kg dw) | Shallow Soil (0-3 ft bgs) (mg/kg dw) | Soil Invertebrate Low SLV (mg/kg dw) ^a | Soil Invertebrate High SLV (mg/kg dw) ^a | Surface Soil (0- 1 ft bgs) Low SLV HQ | Surface Soil (0- 1 ft bgs) High SLV HQ | Shallow Soil (0- 3 ft bgs) Low SLV HQ | Shallow Soil (0-3 ft bgs) High SLV HQ |
| Indeno(1,2,3-cd)pyrene | 1.13 | 1.16 | *** | *** | 6.3E-02 | 1.3E-02 | 6.4E-02 | 1.3E-02 |
| Pyrene | 4.78 | 11.7 | *** | *** | 2.7E-01 | 5.3E-02 | 6.5E-01 | 1.3E-01 |
| Total HPAHs | 18.6 | 40.8 | 18 | 90 | 1.0E+00 | 2.1E-01 | 2.3E+00 | 4.5E-01 |
| SVOC | SVOC | | | | | | | |
| Bis(2-ethylhexyl) Phthalate | 98.0 | 77.1 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Butyl Benzyl Phthalate | 0.0171 | 0.0152 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Carbazole | 0.144 | 0.112 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Dibenzofuran | 0.0424 | 0.122 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Di-n-butyl Phthalate | 0.0804 | 0.0543 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Di-n-octyl Phthalate | 0.0251 | 0.0225 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| VOCs | | | | | | | | |
| 1,2,4-Trimethylbenzene | 0.000230 | 4.62 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| 4-Isopropyltoluene | 0.00133 | 0.0201 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| n-Propylbenzene | 0.000150 | 0.0355 | NA | NA | No SLV | No SLV | No SLV | No SLV |
| Notes: | | | | Metals HI | 2.8E+01 | 2.6E+01 | 2.3E+01 | 2.1E+01 |

a) Sources listed on Table 3-16.

b) The background concentration (95% UPL) is higher than the risk-based

SLV and replaced the SLV in this table.

c) Those chemicals that do not have established SLVs will be evaluated qualitatively.

d) The EPC is the 95% UCL, if calculated (see Table 1-2), or maximum detected concentration (if 95% UCL not calculated).

Bold indicates hazard quotient greater than 1.0.

*** = To be evaluated as Total HPAHs, per EcoSSL; however, HQs for individual HPAHs were calculated using the Total HPAH SLV to determine the primary risk drivers.

| % = percent | HPAH = high molecular weight polycyclic aromatic hydrocarbons |
|---|---|
| AOPC = area of potential concern | HQ = hazard quotient |
| BHC = benzene hexachloride | mg/kg = milligrams per kilogram |
| CPEC = chemical of potential ecological concern | NA = not available |
| dw = dry weight | SLV = screening level value |
| EcoSSL = ecological soil screening level | SVOC = semivolatile organic compound |
| EPC = exposure point concentration | VOC = volatile organic compound |
| ft bgs = feet below ground surface | UCL = upper confidence limit |
| HI = (cumulative) hazard index | UPL = upper prediction limit |
| | |

Table 3-19c Hazard Quotients for Soil Invertebrates – Pistol Range AOPC

| | EPC | SLV | | Hazard Quotients | |
|------------|--|---|--|------------------|--|
| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Soil Invertebrate Low SLV (mg/kg dw) ^a | Soil Invertebrate High SLV (mg/kg dw) ^a | 1.5 ft bgs) | Surface Soil (0· 1.5 ft bgs) High SLV HQ |
| Inorganics | | | | | |
| Lead | 365 | 1,700 | 8,500 | 2.1E-01 | 4.3E-02 |

Notes:

a) Sources listed on Table 3-16.

Bold indicates hazard quotient greater than 1.0.

% = percent

AOPC = area of potential concern

CPEC = chemical of potential ecological concern

dw = dry weight

EPC = exposure point concentration

ft bgs = feet below ground surface

HQ = hazard quotient

mg/kg = milligrams per kilogram

SLV = screening level value

UCL = upper confidence limit

 Table 3-19d

 Hazard Quotients for Soil Invertebrates – Bulb Slope AOPC

| | EPC | SI | V Hazard Quotients | | |
|------------|--|---|--|-----------|--|
| CPEC | 95% UCL in Surface Soil (0-1 ft bgs) (mg/kg dw) | Soil Invertebrate Low SLV (mg/kg dw) ^a | Soil Invertebrate High SLV (mg/kg dw) ^a | 1 ft bgs) | Surface Soil (0- 1 ft bgs) High SLV HQ |
| Inorganics | | | | | |
| Lead | 307 | 1,700 | 8,500 | 1.8E-01 | 3.6E-02 |
| Mercury | 0.720 | 0.1 | 0.5 | 7.2E+00 | 1.4E+00 |
| | | | HI | 7.4E+00 | 1.5E+00 |

Notes:

a) Sources listed on Table 3-16.

Bold indicates hazard quotient greater than 1.0.

% = percent

AOPC = area of potential concern

CPEC = chemical of potential ecological concern

dw = dry weight

EPC = exposure point concentration

ft bgs = feet below ground surface

HI = (cumulative) hazard index

HQ = hazard quotient

mg/kg = milligrams per kilogram

SLV = screening level value

UCL = upper confidence limit

| CEC | Shrew LOAEL RBC | Robin LOAEL RBC | Plant High SLV | Invertebrate High SLV | Site-Specific Background 95% UPL | Lowest of RBCs and SLVs, or Bkg ¹ |
|--------------------|--------------------|--------------------|-------------------|--------------------------|--|--|
| Landfill AOPC | <u> </u> | | 0 | 8 | | |
| Chromium | 177 | 218 | < bkg | < bkg | 28.1 | 28.1 |
| Copper | 78 | 114 | | | 56.7 | 78 |
| Lead | 149 | 54 | | | 25.5 | 54 |
| Mercury | 0.19 | 0.21 | 1.50 | 0.50 | 0.066 | 0.19 |
| Nickel | < bkg | 140 | 190 | | 26.5 | 26.5 |
| Chlordane (tech) | 0.98 | 1.38 | 1.12 | | | 0.98 |
| Total HPAHs | 5.92 | | 6.0 | 90 | | 5.92 |
| Sandblast Area AOP | PC | | | | | |
| Antimony | 2.84 | | | | 0.176 | 2.84 |
| Chromium | 177 | 218 | < bkg | < bkg | 28.1 | 28.1 |
| Lead | 157 | 56 | | | 25.5 | 56 |
| Nickel | < bkg | 138 | 190 | | 26.5 | 26.5 |
| DEHP | | 61 | | | | 61 |
| Total HPAHs | 5.92 | | 6.0 | | | 5.92 |
| Bulb Slope AOPC | | | | | | |
| Mercury | | | | 0.50 | 0.066 | 0.50 |
| Pistol Range AOPC | | | | | | |
| Lead | 151 | 78 | | | 25.5 | 78 |

 Table 3-20

 Summary of Site-Specific LOAEL- and High SLV-based Risk-Based Concentrations

Notes

Blue = Site-Specific Reference Area Background 95% UPL

All units for LOAEL-based RBCs and High SLVs expressed as mg/kg

¹ The lowest value of the LOAEL RBCs and High SLVs is shown, unless the lowest risk-based value is less than background. In which case, the site-specific background 95% UPL is shown (in blue).

AOPC = area of potential concernmg/kg = milligrams per kilogramBkg = backgroundPAH = polycyclic aromatic hydrocarbonsCEC = chemical of ecological concernRBC = risk-based concentrationDEHP = bis(2-ethylhexyl) phthalateSLV = screening level valueHPAH = high molecular weight PAHUPL = upper prediction limitLOAEL = lowest observable adverse effect level

Table 3-21Summary of Upland OU CECs for Further Evaluation in the FS

| | | Terrestrial | Soil | American | Vagrant |
|----------------|-------|-----------------------------|-------------------|----------------------------|---------------------------|
| AOPC | Media | Plants | Invertebrates | Robin | Shrew |
| Landfill | Soil | Mercury, Nickel, Chlordane, | Mercury and HPAHs | Chromium, Copper, Lead, | Chromium, Copper, Lead, |
| | | HPAHs | | Mercury, Nickel, Chlordane | Mercury, Nickel, |
| | | | | | Chlordane, HPAHs |
| Sandblast Area | Soil | Nickel and HPAHs | None | Chromium, Lead, Nickel | Antimony, Chromium, Lead, |
| | | | | | Nickel, HPAHs |
| Pistol Range | Soil | None | None | Lead | None |
| Bulb Slope | Soil | None | None | None | None |

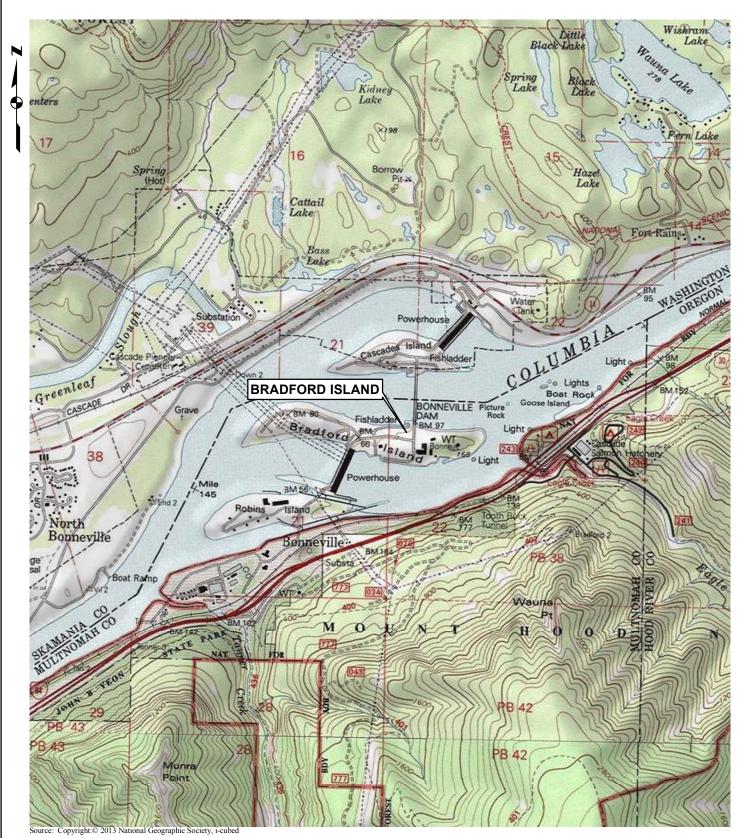
Notes:

No CECs were identified for the Canada goose, American kestrel, or American mink (Upland OU evaluation only).

No CECs were identified for the Combined AOPCs exposure unit.

AOPC = area of potential concern CEC = chemical of ecological concern FS = feasibility study HPAH = high molecular weight PAH OU = operable unit PAH = polycyclic aromatic hydrocarbons

FIGURES



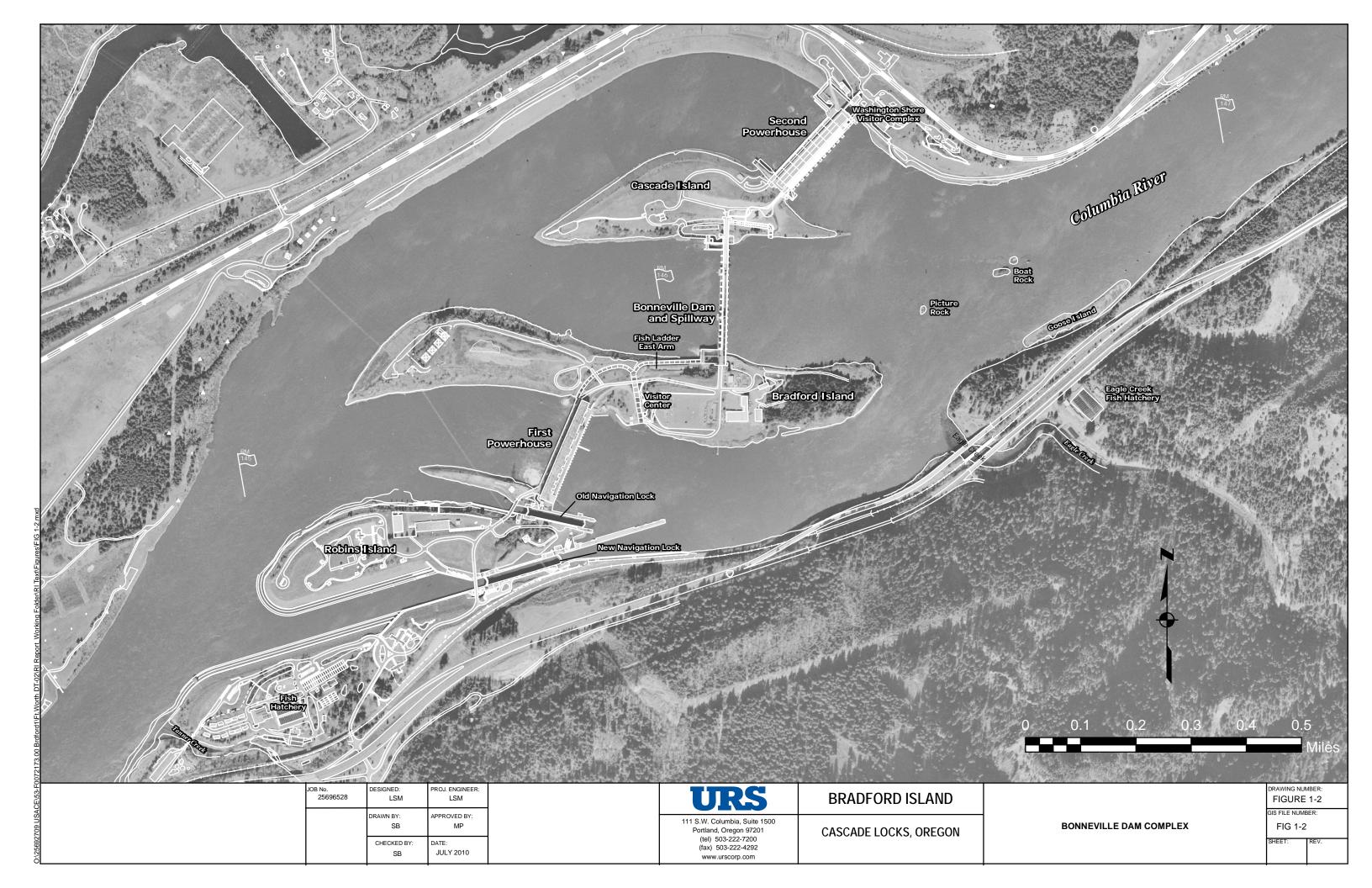
VICINITY MAP

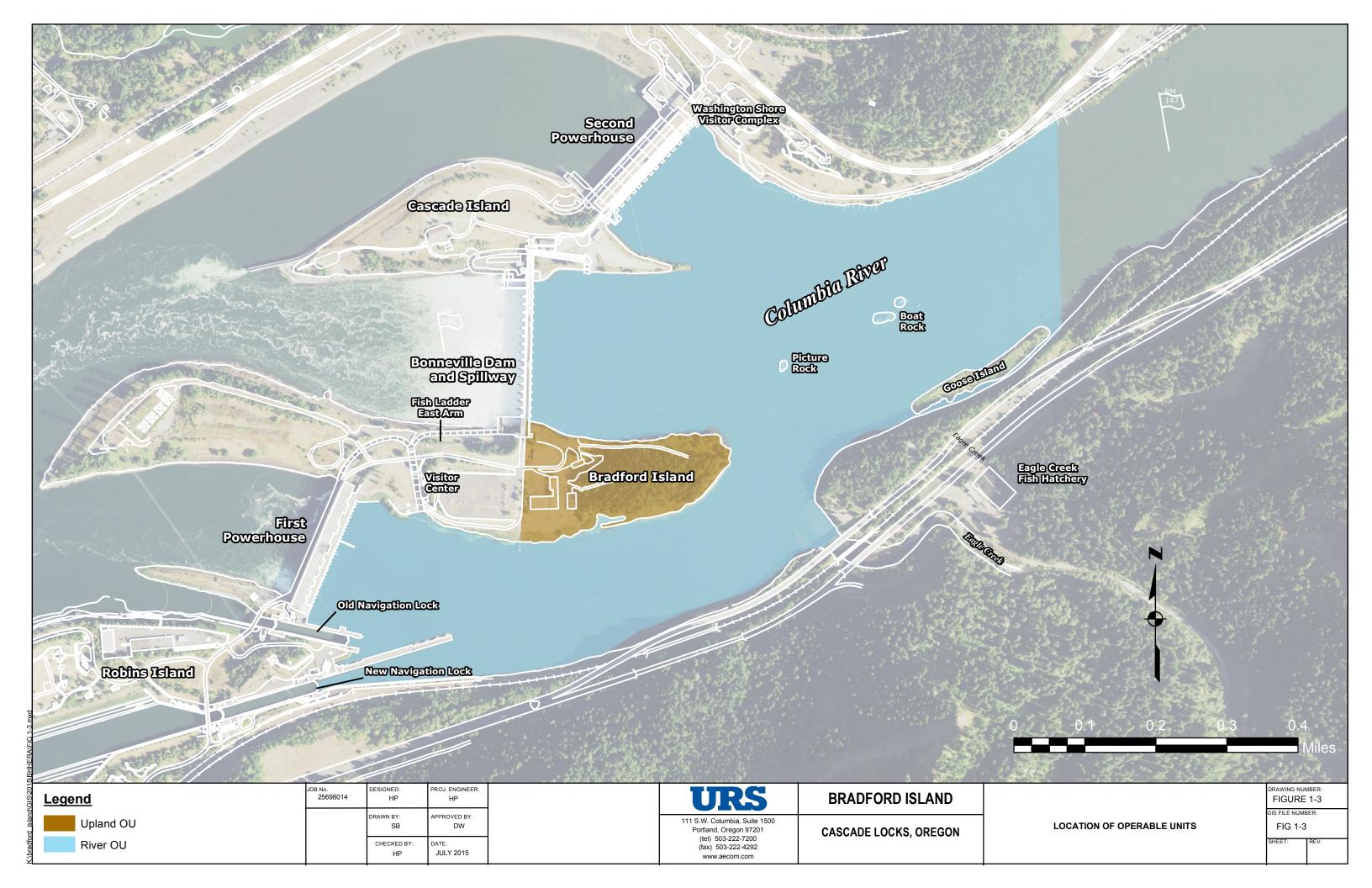
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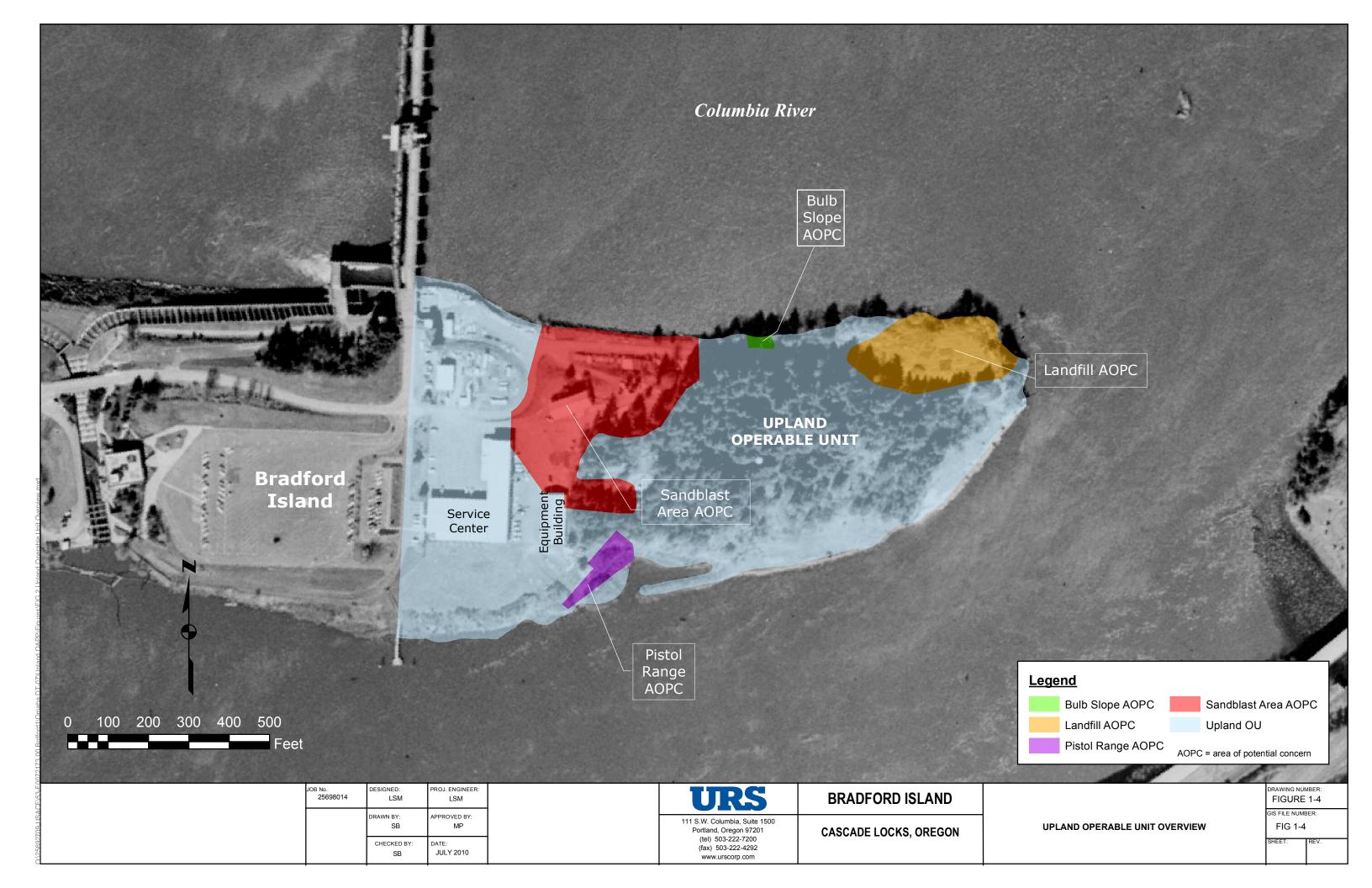
BRADFORD ISLAND CASCADE LOCKS, OREGON

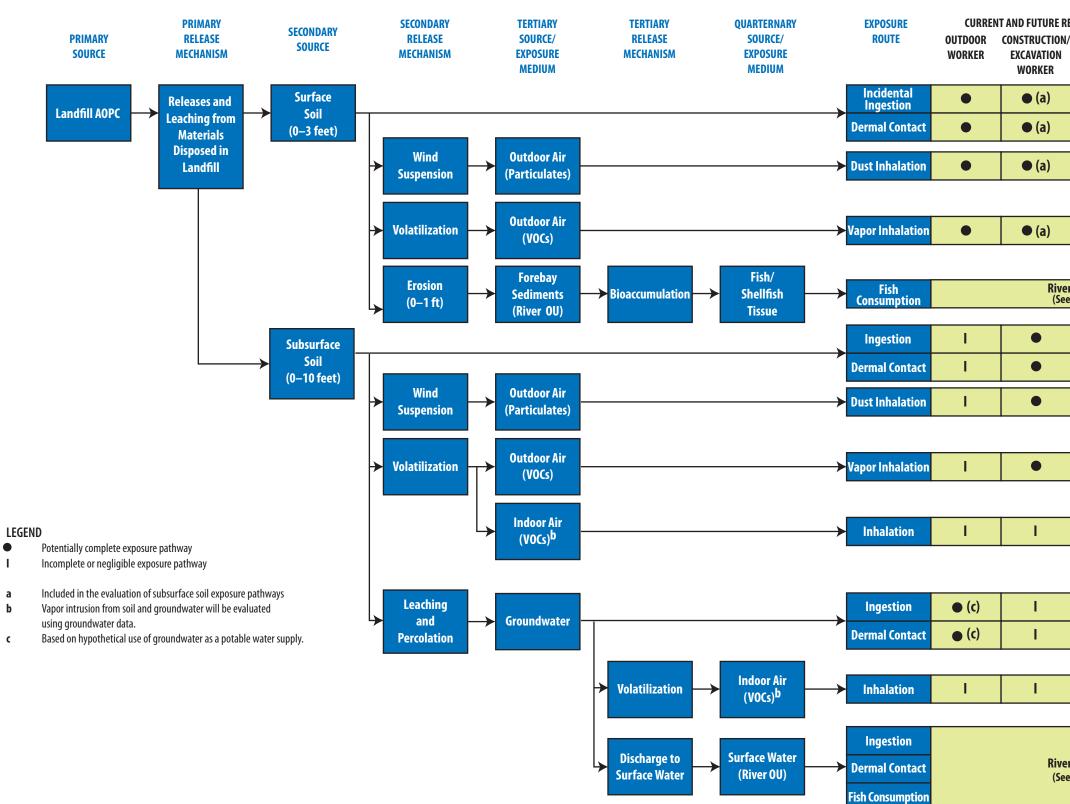
FIGURE 1-1





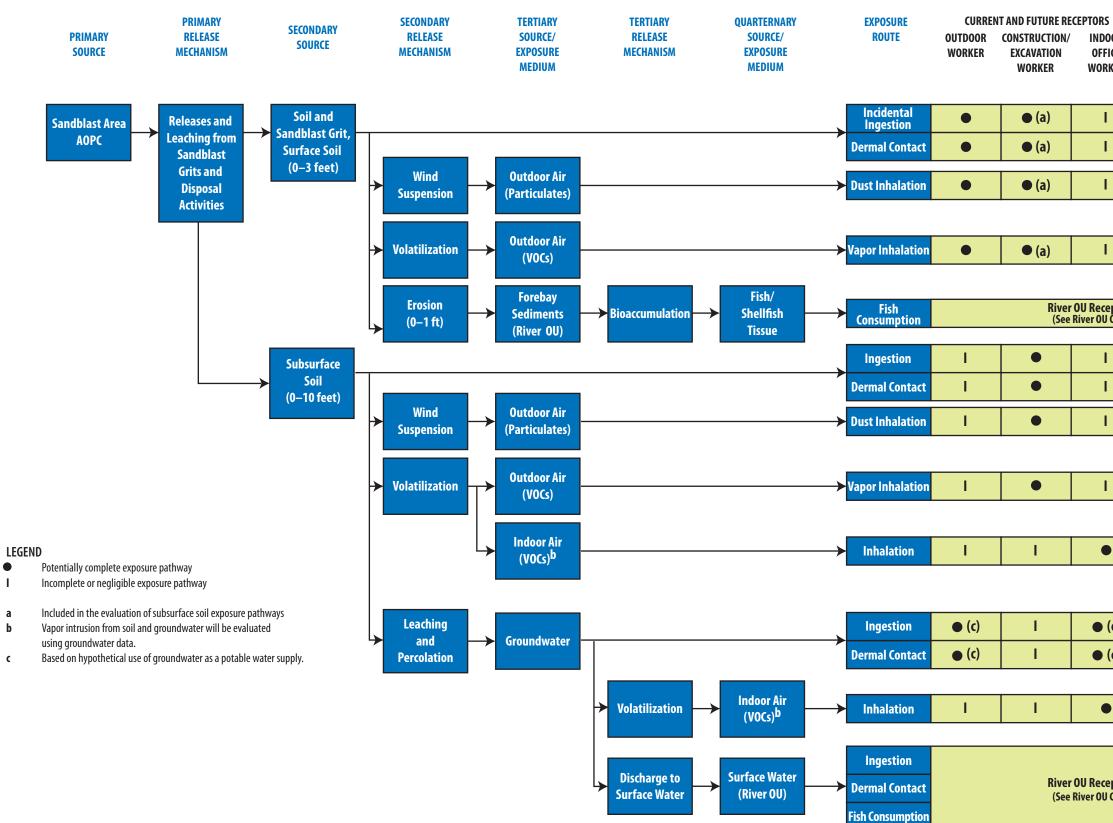






| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | |
|---------------------|-------------------|------------------------|---|-----------------------|---|
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| | CHECKED BY: HP | DATE: JULY 2010 | (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | | L |

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| | | LANDFILL AOPC | DRAWING NUMBER: |
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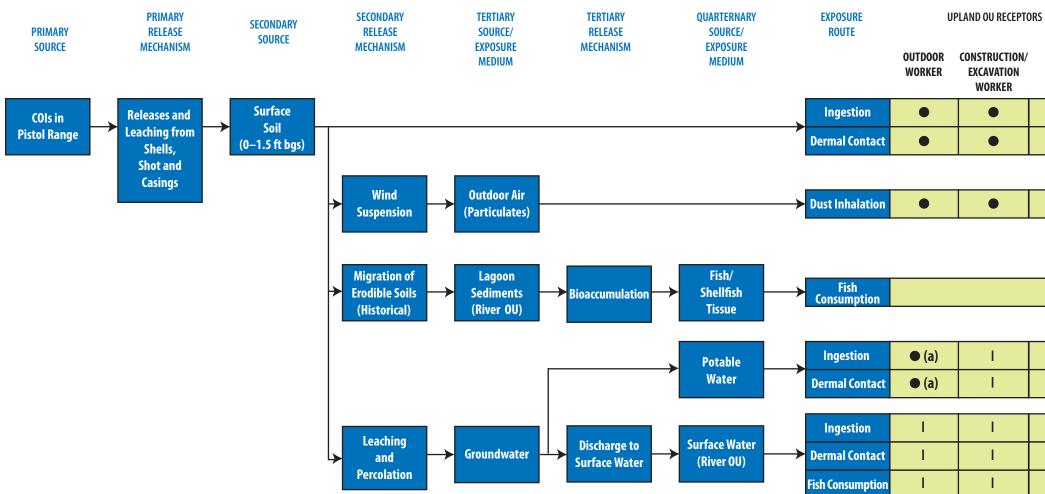
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| | SAN | NDBLAST AREA AOPC | DRAWING NUMBER: FIGURE 2-2 |
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| ŀ | IUMAN HEALT | H CONCEPTUAL EXPOSURE MODEL | FIG 2-2 |
| | | | SHEET: REV. |
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HYPOTHETICAL FUTURE



LEGEND

- Potentially complete exposure pathway
- Incomplete or negligible exposure pathway
- а Based on hypothetical use of groundwater as a potable water supply
- b Based on use of downstream river water as a potable water supply

| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | PISTOL RANGE AOPC | DRAWING NUMBER: FIGURE 2-3 |
|---------------------|-------------------|------------------------|---|-----------------------|--|-------------------------------|
| | DRAWN BY: MS | APPROVED BY: MP | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 | CASCADE LOCKS, OREGON | HUMAN HEALTH CONCEPTUAL EXPOSURE MODEL | GIS FILE NUMBER: FIG 2-3 |
| | CHECKED BY: HP | DATE: JULY 2010 | (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | | | SHEET: REV. |

RIVER OU RECEPTORS

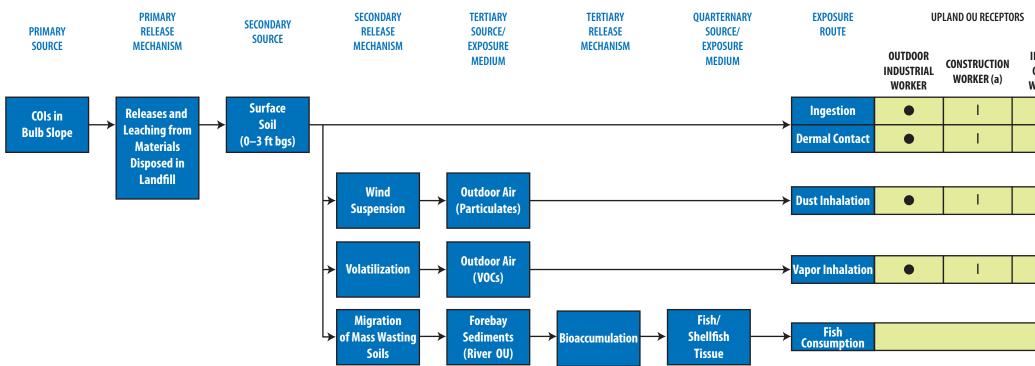
| INDOOR OFFICE WORKER | SUBSISTENCE FISHER | RECREATIONAL FISHER/WADER | DOWNSTREAM Potable Water User | FISHING PLATFORM USER (ADULT & CHILD) |
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HYPOTHETICAL FUTURE

SCENARIO

River OU Receptors (See River OU CEM)

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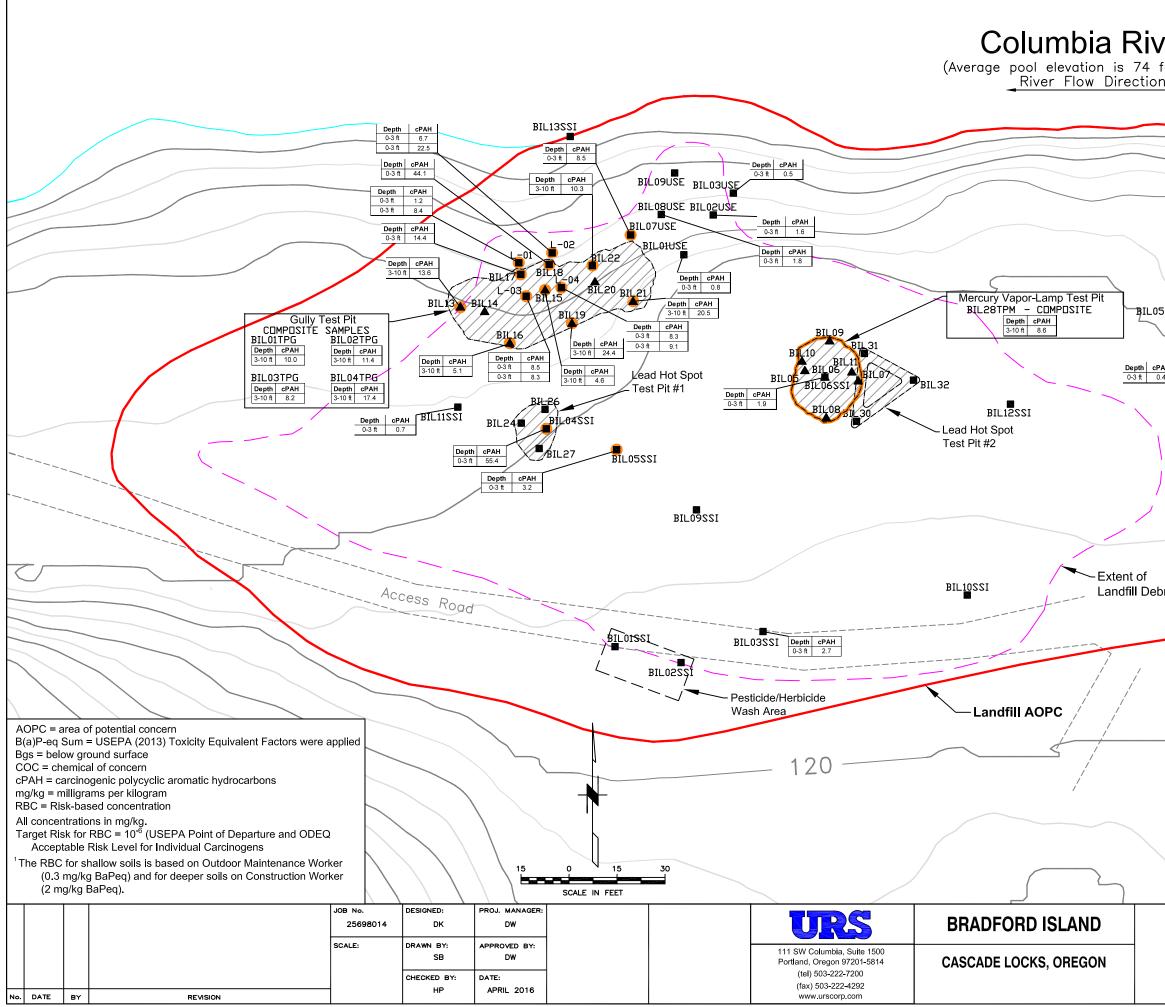
LEGEND

- Potentially complete exposure pathway
- I Incomplete or negligible exposure pathway
- (a) No construction or excavation activities are feasible or likely at the Bulb Slope AOPC

| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | | IRS | BRADFORD ISLAND | | DRAWING NUMBER: FIGURE 2-4 |
|---------------------|-------------------|------------------------|----------|---|-----------------------|--|-------------------------------|
| | DRAWN BY: MS | APPROVED BY: MP | 111 S.W. | Columbia, Suite 1500 and, Oregon 97201 | CASCADE LOCKS, OREGON | HUMAN HEALTH CONCEPTUAL EXPOSURE MODEL | GIS FILE NUMBER: FIG 2-4 |
| | CHECKED BY: HP | DATE: JULY 2010 | (fax) |) 503-222-7200 3) 503-222-4292 ww.urscorp.com | , | | SHEET: REV. |

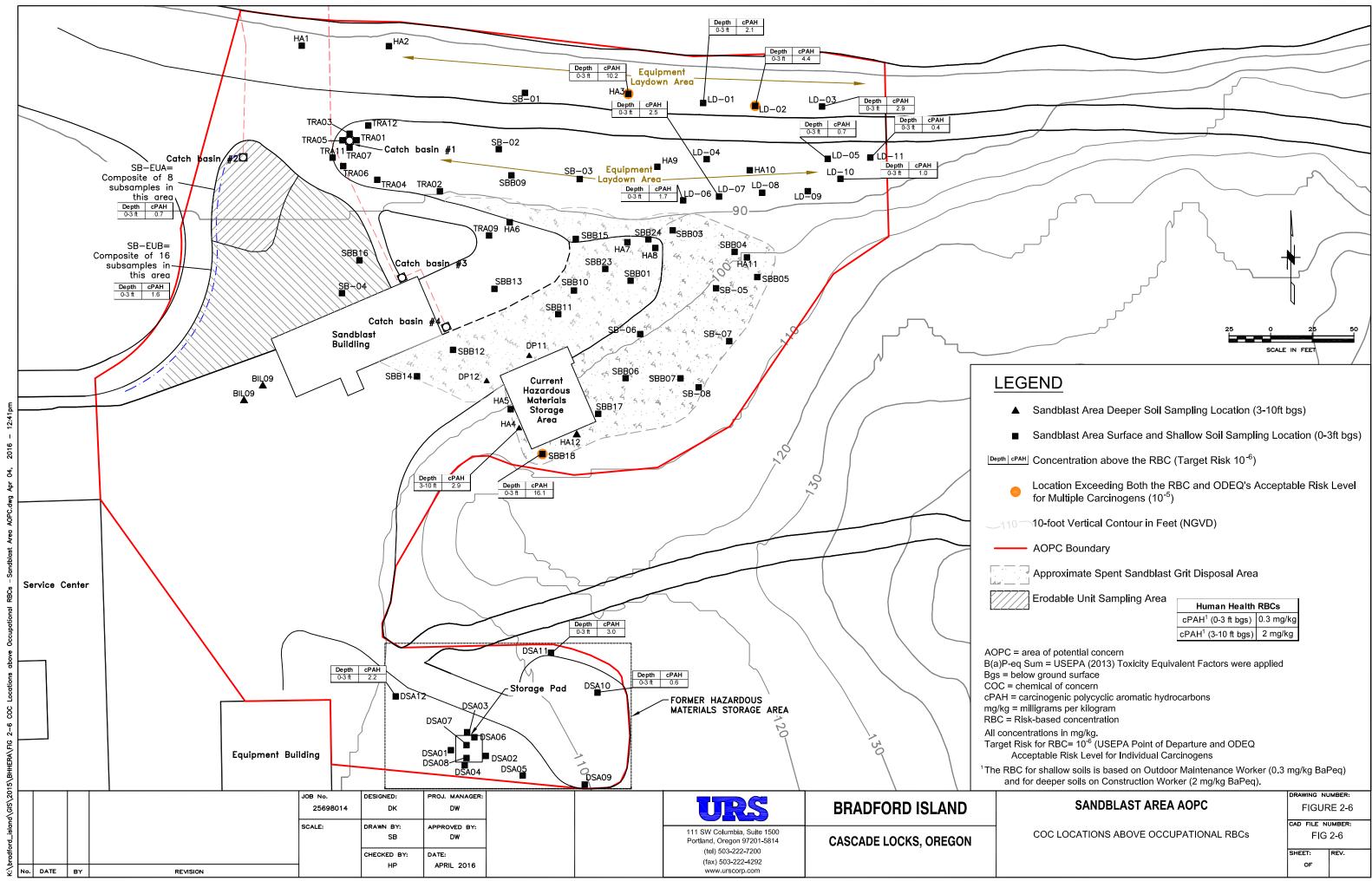
| i | RI | HYPOTHETICAL FUTURE Scenario | | |
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River OU Receptors (See River OU CEM)



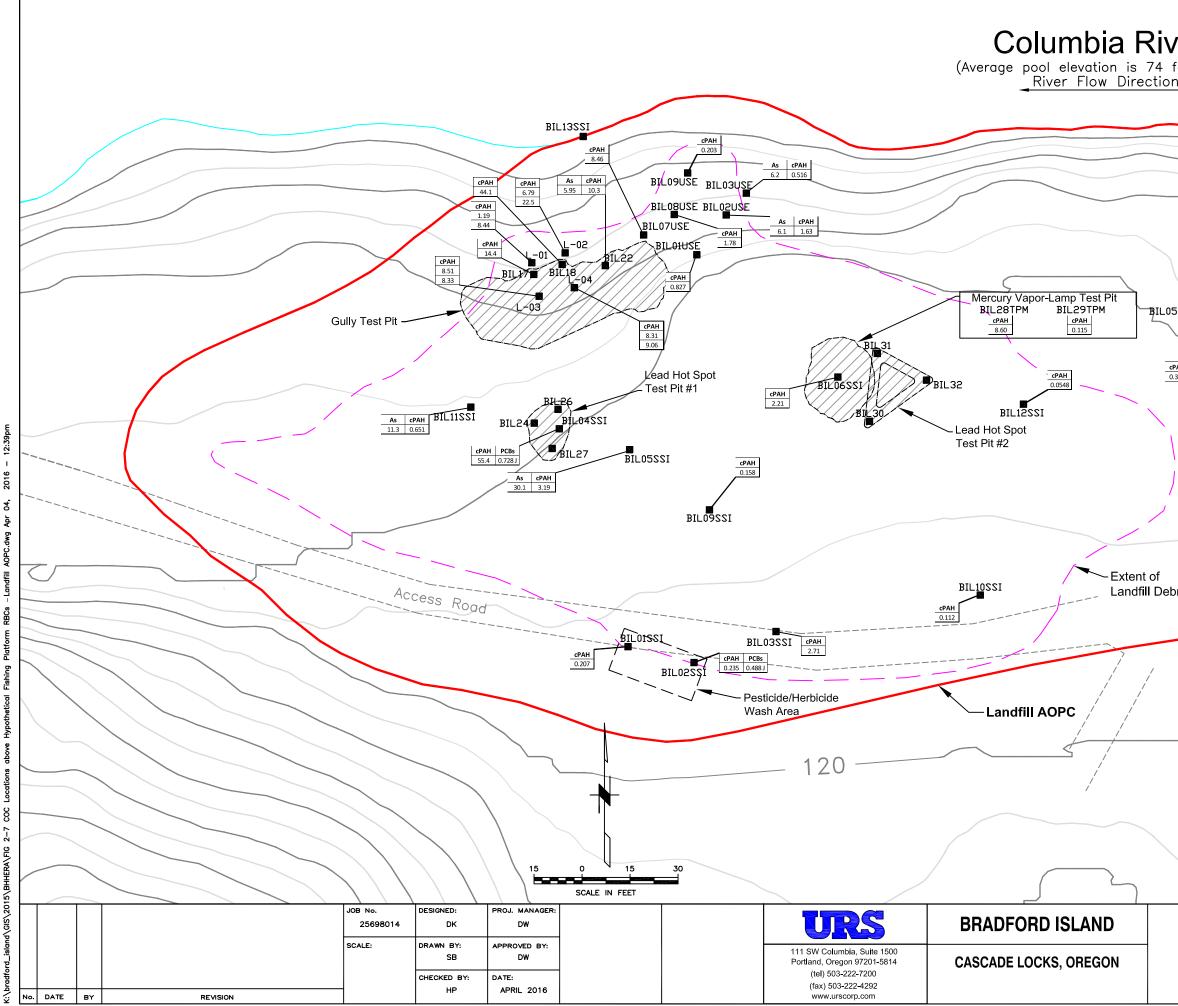
| ve fee | t NGVD) | | |
|-----------|-------------|--|--|
| | BILO4USE | $\frac{Depth}{0.3 ft} \frac{ePAH}{2.2}$ | |
| \ / [| LEGE | ND | |
| | | Landfill Surface and Shallow Soil Samplir | ng Location |
| | | Landfill Deeper Soil Sampling Location | |
| ebris | Depth cPAH | Concentration above the RBC (Target Ris | sk 10 ⁻⁶) |
| | • | Location Exceeding Both the RBC and O Acceptable Risk Level for Multiple Carcin | |
| | \square | Remediation Limits | |
| | \sim | 2- Foot vertical elevation contour in feet N | IGVD |
| | | Landfill Access Road | |
| | · | AOPC Boundary | |
| | | Approximate extent of landfill debris base geophysical reconnaissance survey using resistivity and seismic refraction | |
| | | Approximate Waterline at Mean Pool Elec (approx. 74 feet NGVD) Human He CPAH ¹ (0-3 ft b CPAH ¹ (3-10 ft b | alth RBCs gs) 0.3 mg/kg pgs) 2 mg/kg |
| | | LANDFILL AOPC | drawing number: FIGURE 2-5 |
| (| COC LOCATIO | ONS ABOVE OCCUPATIONAL RBCs | CAD FILE NUMBER: FIG 2-5 Sheet: rev. |

OF



| Human Health RBCs | | | | | |
|---------------------------------|-----------|--|--|--|--|
| cPAH ¹ (0-3 ft bgs) | 0.3 mg/kg | | | | |
| cPAH ¹ (3-10 ft bgs) | 2 mg/kg | | | | |

| DRAWING NUMBER: | | | | | |
|------------------|--|--|--|--|--|
| FIGURE 2-6 | | | | | |
| CAD FILE NUMBER: | | | | | |
| FIG 2-6 | | | | | |
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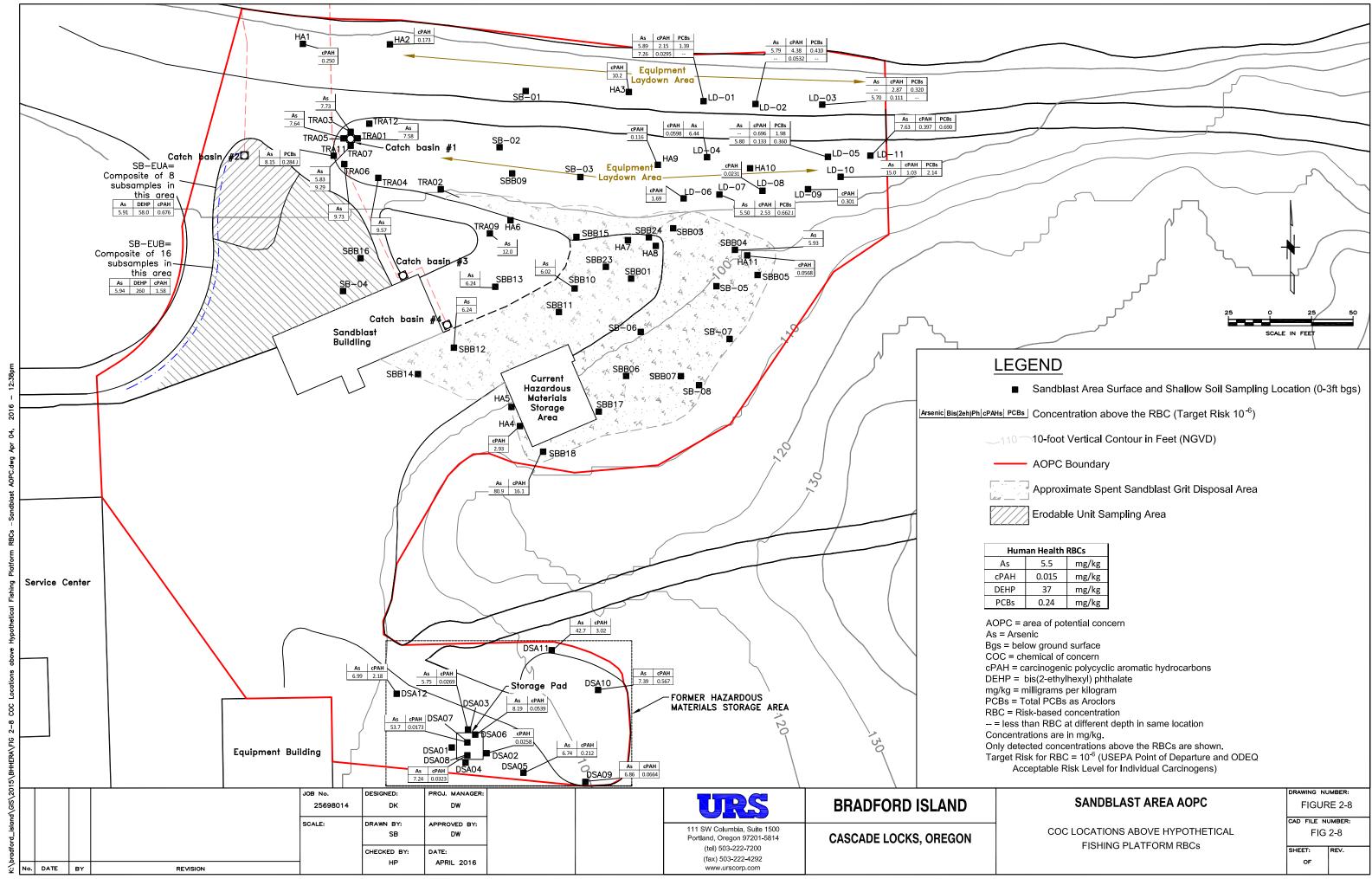


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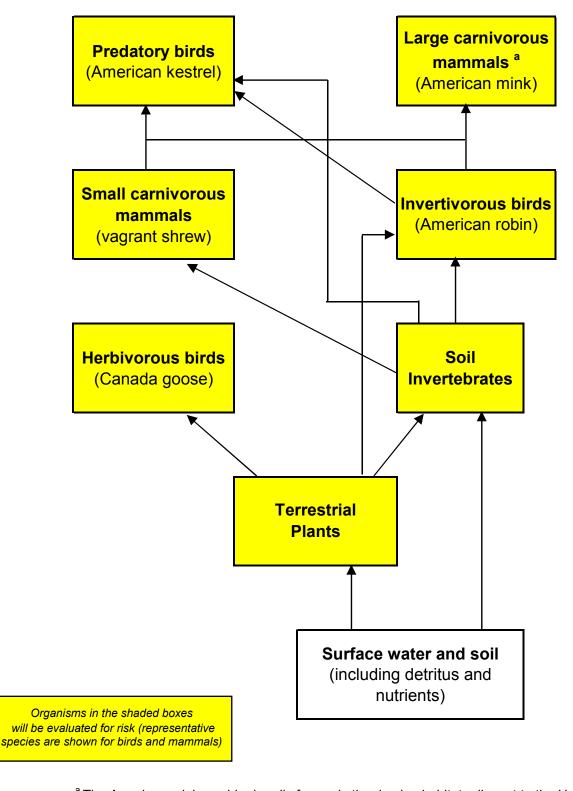
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| / e fee <u>n</u> | t NGVD) | | | | | |
|--|------------------------------|----------------------------|---------------------------|--|------------------------------------|----|
| 32U2 | PAH 0.140 BIL04USE | <u>ePAH</u> 2.20 | | | | |
| cPAH 0.392 | LEGEI | ND | | | | |
| | | | ace and Sh | nallow Soil Samplin | ig Locatio | on |
| | Arsenic cPAHs | | | ne RBC (Target Ris | | |
| | | Remediatior | | | , | |
| / | $\left \frac{2}{2} \right $ | | | on contour in feet N | IGVD | |
| | | Landfill Acce | | | | |
| | | AOPC Bour | | | | |
| bris | | Approximate | e extent of reconnaiss | landfill debris base sance survey using refraction | | ıl |
| _ | | Approximate (approx. 74 | | at Mean Pool Elev) | /ation | |
| | Hur | nan Health | RBCs | | | |
| | As | 5.5 | mg/kg | | | |
| | cPAH | 0.015 | mg/kg | | | |
| PCBs0.24mg/kgAOPC = area of potential concern As = ArsenicBgs = below ground surface COC = chemical of concern cPAH = carcinogenic polycyclic aromatic hydrocarbons mg/kg = milligrams per kilogram PCBs = Total PCBs as Aroclors RBC = Risk-based concentration Concentrations are in mg/kg.Only detected concentrations above the RBCs are shown. Target Risk for RBC = 10 ⁻⁶ (USEPA Point of Departure and ODEQ Acceptable Risk Level for Individual Carcinogens) | | | | | | |
| | | LANDFILL | AOPC | | drawing nu FIGUR | |
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| ım | iman Health RBCs | | | | | |
|----|------------------|-------|--|--|--|--|
| | 5.5 | mg/kg | | | | |
| | 0.015 | mg/kg | | | | |
| | 37 | mg/kg | | | | |
| | 0.24 | mg/kg | | | | |

| NAMING NOMBER. | | | | | | | |
|----------------------------|--|--|--|--|--|--|--|
| FIGURE 2-8 | | | | | | | |
| AD FILE NUMBER: FIG 2-8 | | | | | | | |
| HEET: REV. | | | | | | | |



^a The American mink would primarily forage in the riverine habitat adjacent to the Upland OU, but could also forage in the Upland OU when aquatic prey are scarce.

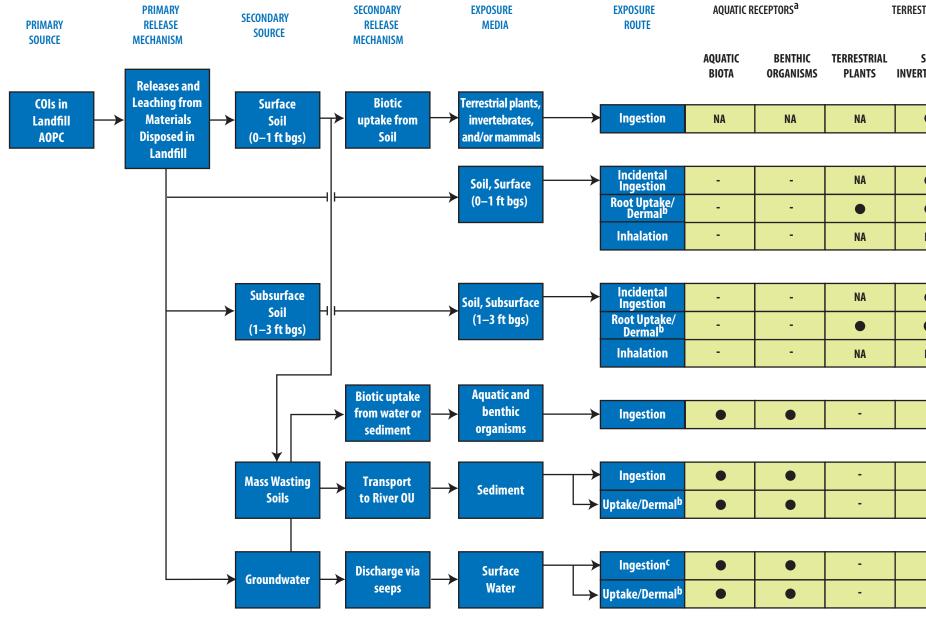


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BRADFORD ISLAND CASCADE LOCKS, OREGON

UPLAND OU FOODWEB

FIGURE 3-1



LEGEND

- Complete and potentially significant pathway
- O Complete but likely minor pathway
- Incomplete pathway
- NA Pathway not applicable to receptor

Aquatic Biota defined as aquatic plants, plankton, invertebrates, fish and aquatic-dependent wildlife. a. The Upland OU to River OU pathways evaluated at a screening level in the Final RI report (URS 2012) were not addressed, as these receptors and pathways are considered in a separate document b. The dermal exposure pathway for wildlife is expected to be minor due to the barrier offered by fur and feathers. c. Highly mobile birds and mammals could ingest water from the river, while less mobile species would likely ingest water from puddles in the Upland OU.

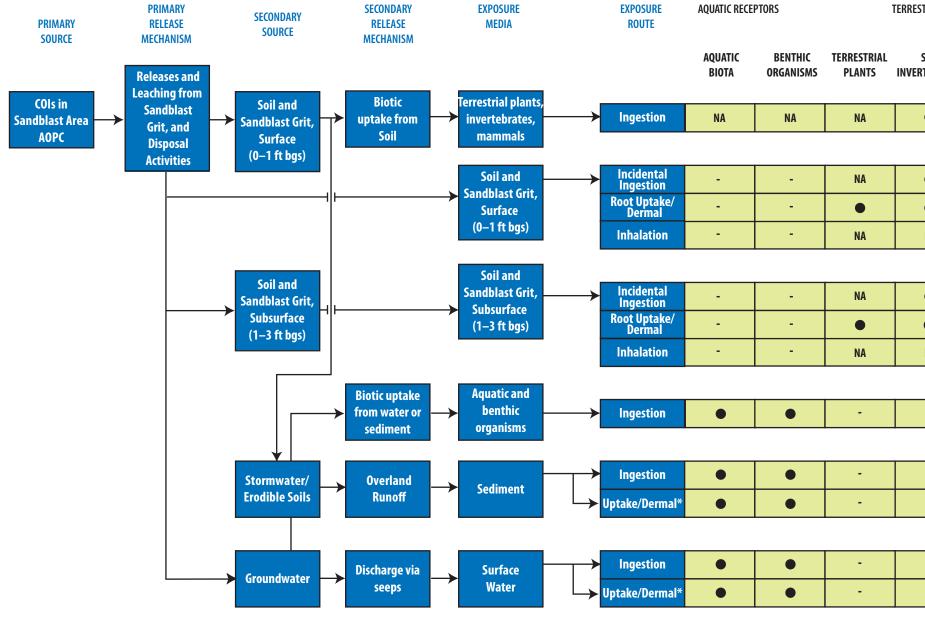
| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | LANDFILL AOPC | DRAWING NUM | 3-2 |
|---------------------|-------------------|------------------------|---|-----------------------|--------------------------------------|--------------------------|------|
| | DRAWN BY: MS | APPROVED BY: MP | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 | CASCADE LOCKS, OREGON | ECOLOGICAL CONCEPTUAL EXPOSURE MODEL | GIS FILE NUME FIG 3-2 | |
| | CHECKED BY: HP | DATE: JULY 2010 | (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | | | SHEET: | REV. |

TERRESTRIAL RECEPTORS

| SOIL TEBRATES | BIRDS | MAMMALS | |
|------------------|-------------|-------------|--|
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| NA | 0 | 0 | |
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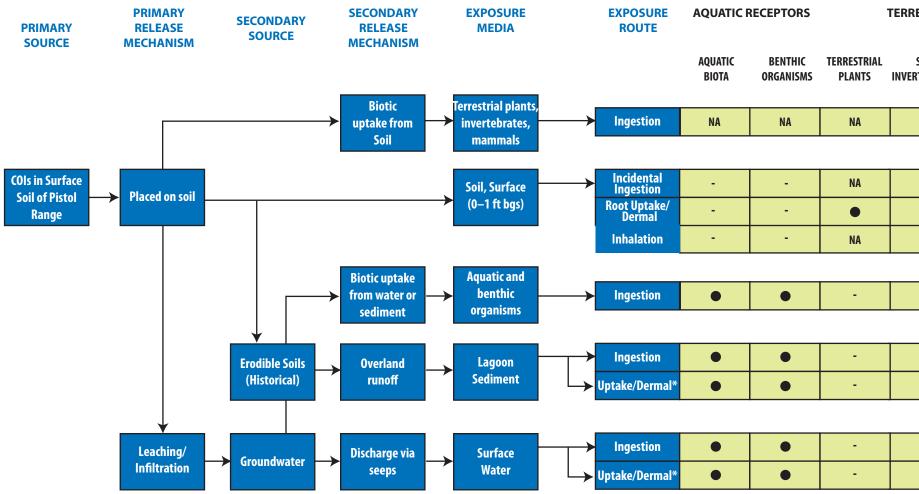
LEGEND

- Complete and potentially significant pathway
- O Complete but likely minor pathway
- Incomplete pathway
- NA Pathway not applicable to receptor

Aquatic Biota defined as aquatic plants, plankton, invertebrates, fish and aquatic-dependent wildlife. *The dermal exposure pathway for aquatic-dependent wildlife is expected to be minor due to the barrier offere

| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | |
|---------------------|-------------------|------------------------|---|-----------------------|---|
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| | CHECKED BY: HP | DATE: JULY 2010 | (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | | |

| STRIAL RECE | PTORS | | HYPOTHETICAL FUTURE Scenario | |
|-------------------|-------------------------------|-----------------------------|---|--|
| SOIL RTEBRATES | BIRDS | MAMMALS | FISHING PLATFORM USER (ADULT & CHILD) | |
| • | • | • | • | |
| | | | | |
| • | • | • | | |
| • | 0 | 0 | • | |
| NA | 0 | 0 | • | |
| | | | | |
| • | • | • | l. | |
| 0 | 0 | 0 | l I | |
| NA | 0 | 0 | l I | |
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| red by fur a | ind feathers. | | | |
| | DRAWING NUMBER: FIGURE 3-3 | | | |
| ECOL | OGICAL CON | GIS FILE NUMBER: FIG 3-3 | | |
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LEGEND

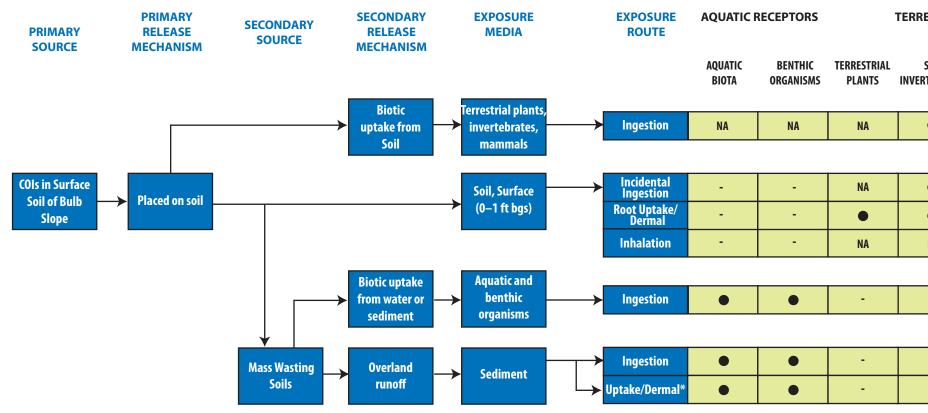
- Complete and potentially significant pathway
- O Complete but likely minor pathway
- Incomplete pathway
- NA Pathway not applicable to receptor

Aquatic Biota defined as aquatic plants, plankton, invertebrates, fish and aquatic-dependent wildlife. *The dermal exposure pathway for aquatic-dependent wildlife is expected to be minor due to the barrier offered by fur and feathers.

| JOB No. 25698014 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | PISTOL RANGE AOPC | DRAWING NU | 3-4 |
|-------------------------|-------------------|------------------------|---|-----------------|--------------------------------------|--------------|------|
| | DRAWN BY: MS | APPROVED BY: MP | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 | | ECOLOGICAL CONCEPTUAL EXPOSURE MODEL | GIS FILE NUM | |
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TERRESTRIAL RECEPTORS

| SOIL RTEBRATES | BIRDS | MAMMALS |
|-------------------|-------|---------|
| • | • | • |
| | | |
| • | • | • |
| • | 0 | 0 |
| NA | 0 | 0 |
| | | |
| - | - | - |
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LEGEND

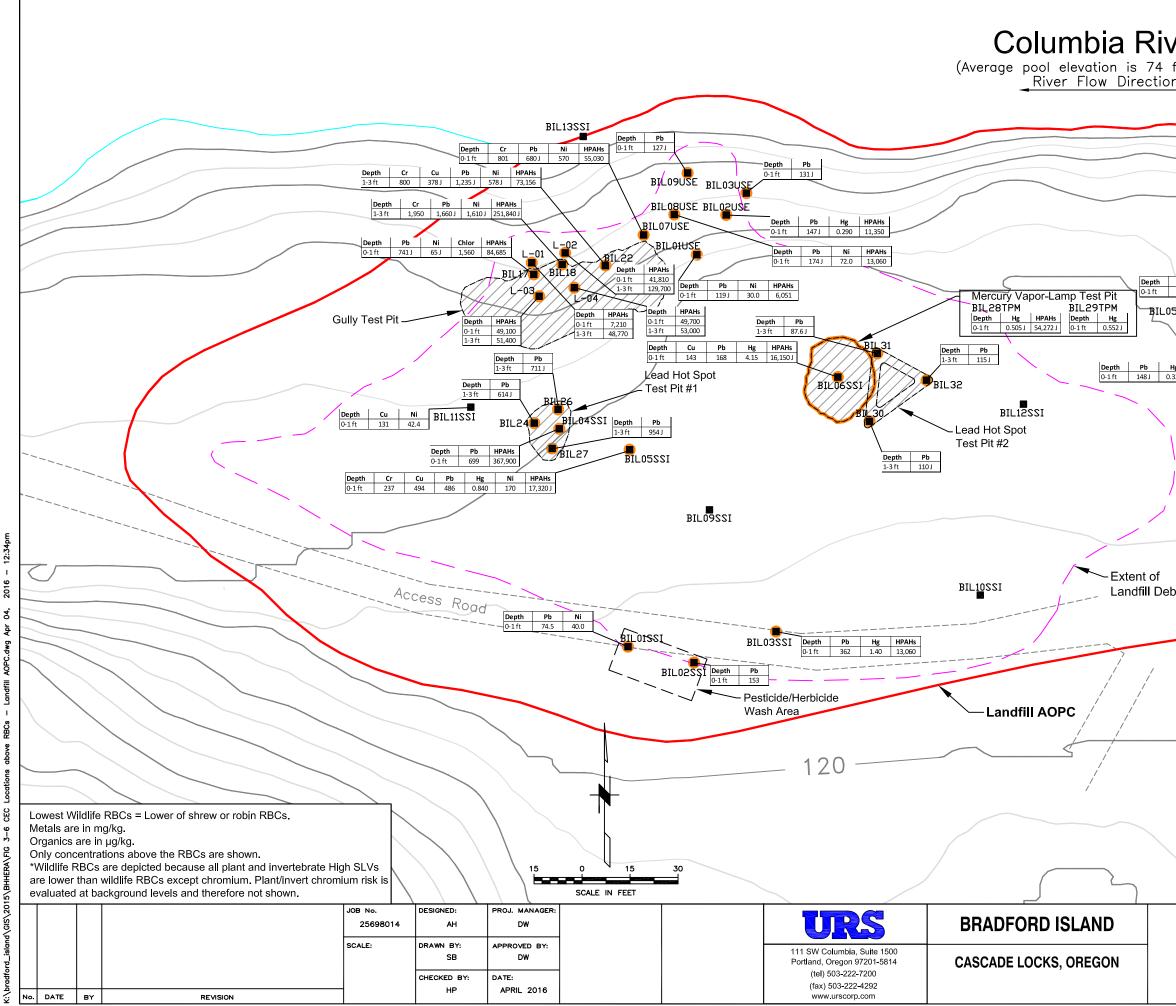
- Complete and potentially significant pathway
- O Complete but likely minor pathway
- Incomplete pathway
- NA Pathway not applicable to receptor

Aquatic Biota defined as aquatic plants, plankton, invertebrates, fish and aquatic-dependent wildlife. *The dermal exposure pathway for aquatic-dependent wildlife is expected to be minor due to the barrier offered by fur and feathers.

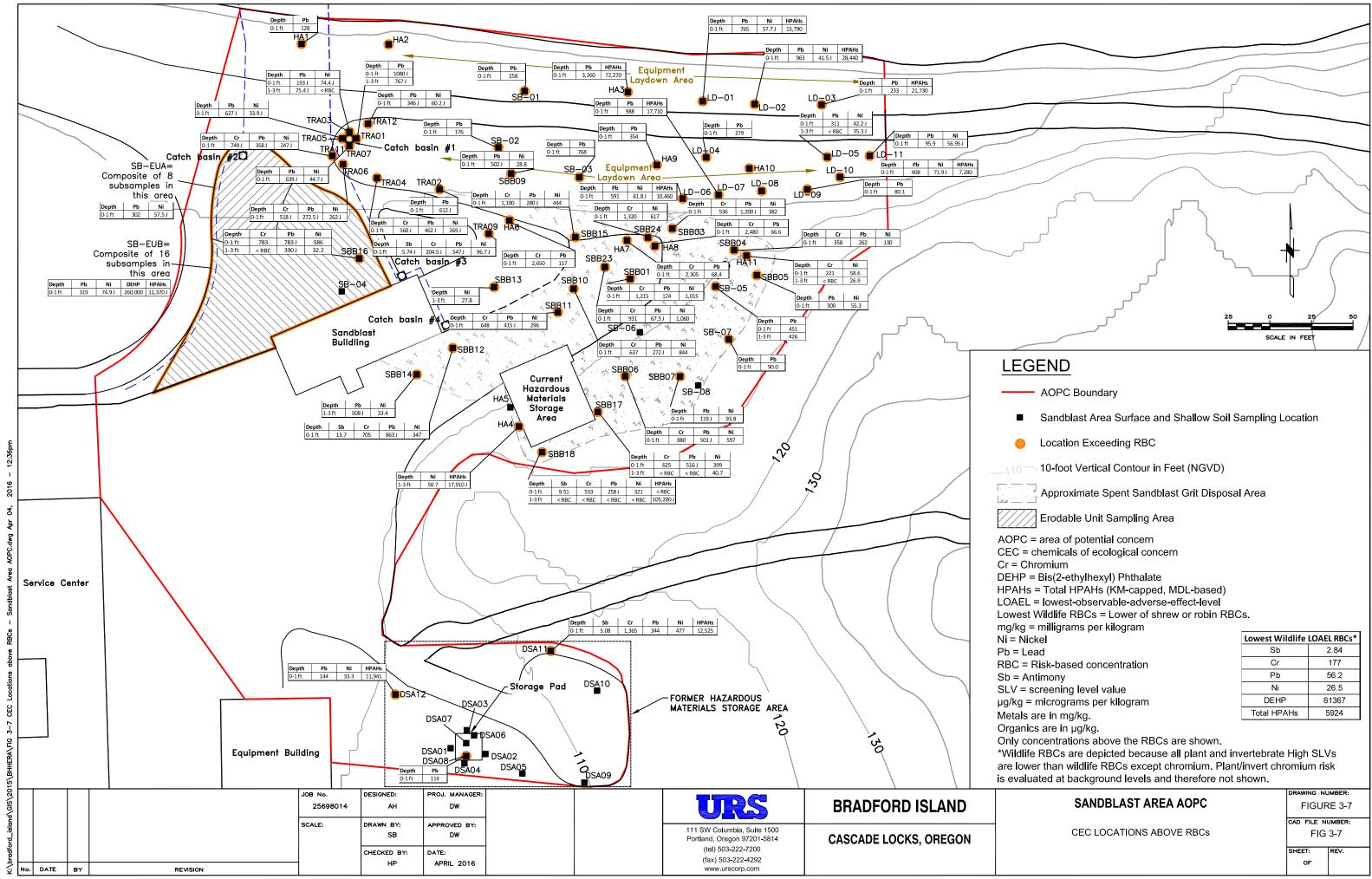
| JOB No. 25698014 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | BULB SLOPE AOPC | DRAWING NUMBER: FIGURE 3-5 |
|---------------------|-------------------|------------------------|---|-----------------------|--------------------------------------|-------------------------------|
| | DRAWN BY: MS | APPROVED BY: MP | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 (tel) 503-222-7200 | CASCADE LOCKS, OREGON | ECOLOGICAL CONCEPTUAL EXPOSURE MODEL | GIS FILE NUMBER: FIG 3-5 |
| | CHECKED BY: HP | DATE: JULY 2010 | (tax) 503-222-4292 (fax) 503-222-4292 www.urscorp.com | | | SHEET: REV. |

TERRESTRIAL RECEPTORS

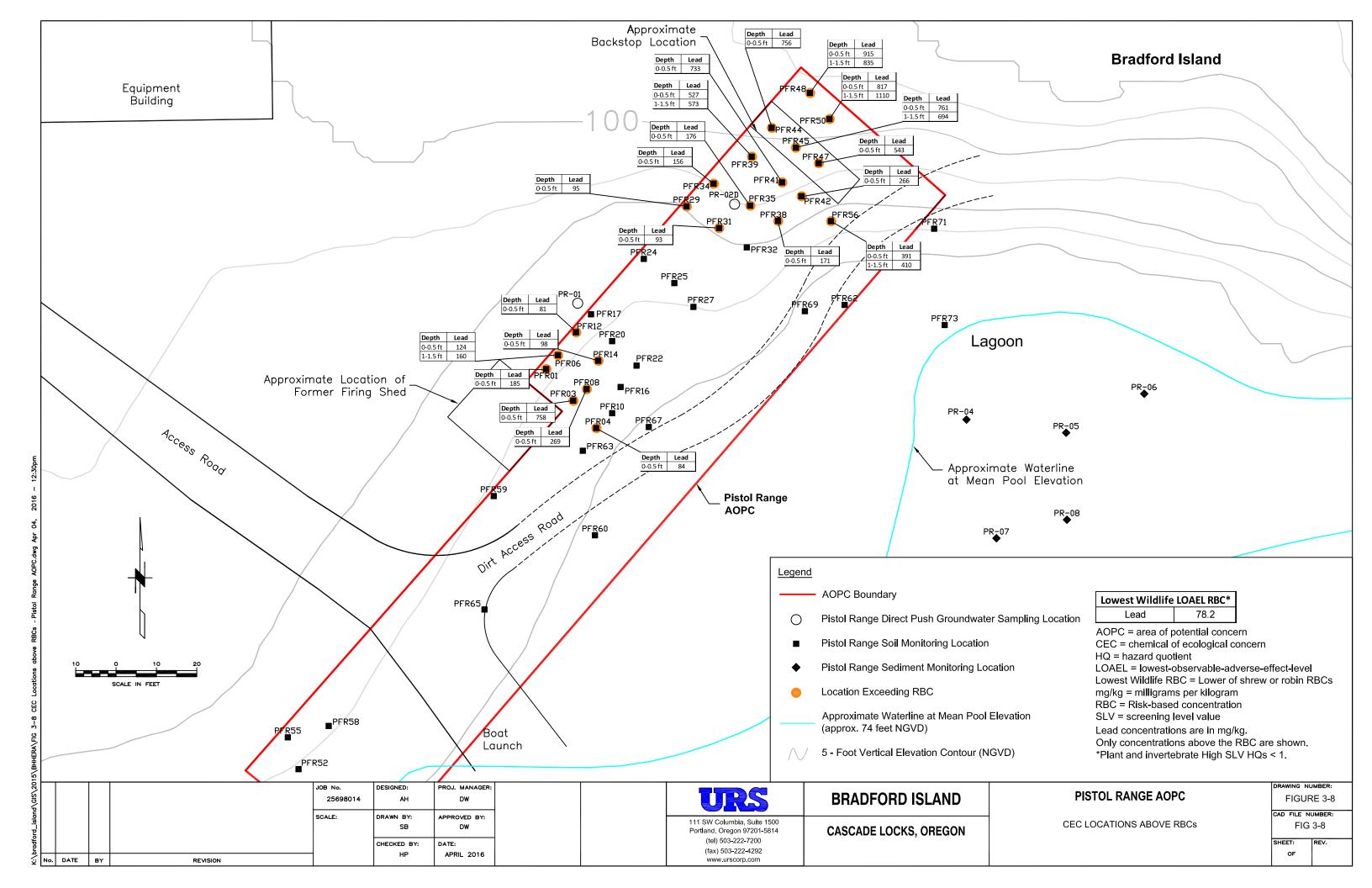
| SOIL RTEBRATES | BIRDS | MAMMALS | |
|-------------------|-------|---------|--|
| • | • | • | |
| | | | |
| • | • | • | |
| • | 0 | 0 | |
| NA | 0 | 0 | |
| | | | |
| - | - | - | |
| | | | |
| - | - | - | |
| - | - | - | |

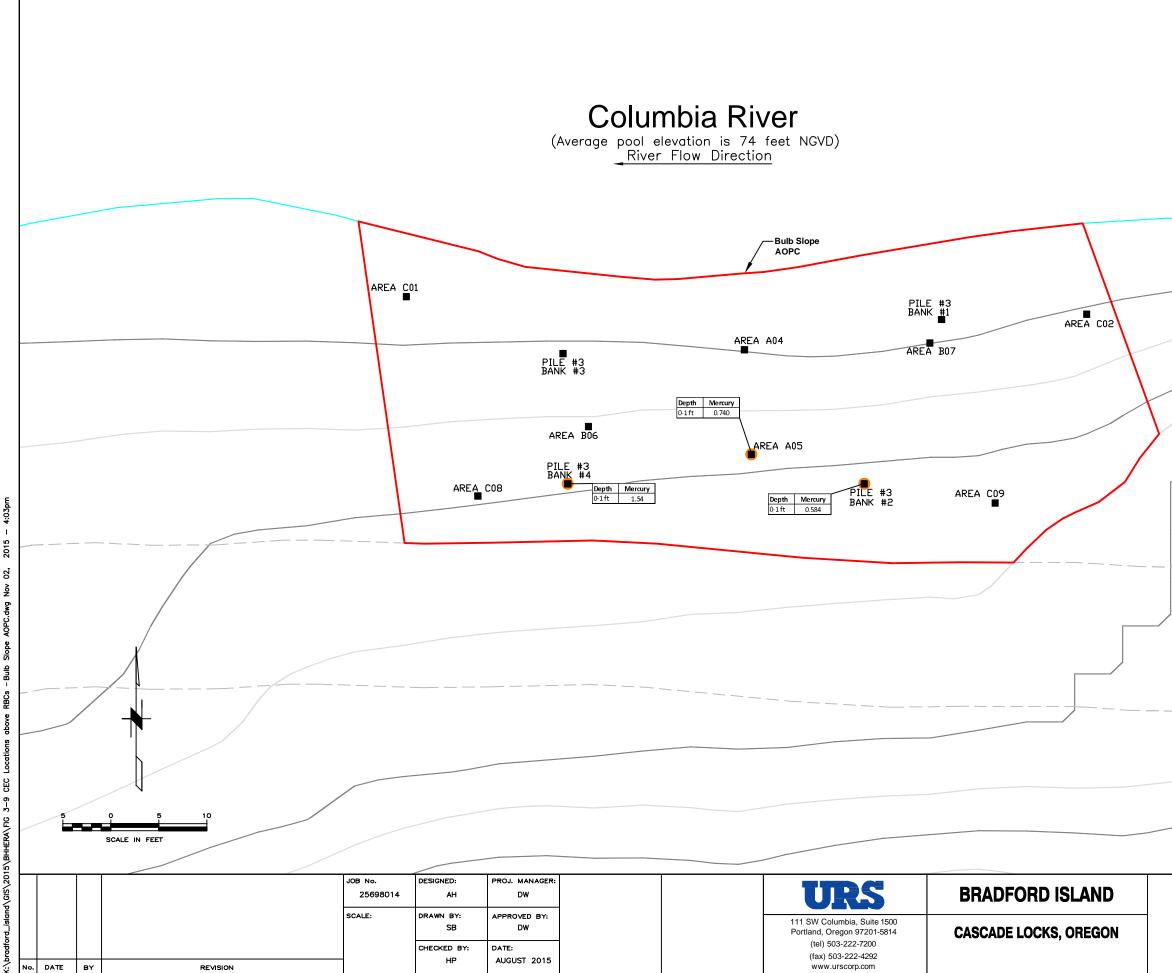


| Ve fee | t NGVD) | | | | |
|--|--------------------------------|--|------------|--------------|-----------------------|
| Hg 0.230 05US 1 Hg 0.3320 | | epth HPAHs 1ft 14,390 | | | |
| , [| LEGEN | <u>D</u> | | | |
| \ | • | Landfill Surface and Shallow S | Soil Moni | toring | Location |
| / | • | Location Exceeding RBC | | | |
| | \square | Remediation Limits | | | |
| | \sim | 2- Foot vertical elevation conte | our in fee | et NG∨ | ′D |
| | <u> </u> | Landfill Access Road | | | |
| bris | | AOPC Boundary | | | |
| _ | | Approximate extent of landfill geophysical reconnaissance s resistivity and seismic refraction | urvey us | | |
| | | Approximate Waterline at Mea (approx. 74 feet NGVD) | an Pool E | levatio | on |
| | | of potential concern | Lowest W | /ildlife | LOAEL RBCs* |
| | CEC = chemic Chlor = Chlord | als of ecological concern ane (technical) | Cr | | 177 |
| | Cr = Chromium | . , | Cu | | 78.5 |
| | Cu = Copper | | Pb Hg | | 54.0 0.192 |
| | Hg = Mercury HPAHs = Total | HPAHs (KM-capped, MDL-based) | - | | 26.5 |
| | | st-observable-adverse-effect-level | Chlorda | ane | 982 |
| | mg/kg = milligr | ams per kilogram | Total HP | | 5918 |
| | SLV = screenir | sed concentration ng level value rams per kilogram | | | |
| | L | ANDFILL AOPC | | FIG | G NUMBER: GURE 3-6 |
| | CEC LC | OCATIONS ABOVE RBCs | | F | e number: FIG 3-6 |
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| DRAWING NU | JMBER: |
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| | <u>LEGEN</u> | D | | |
|---|--|---|---------------------|--------|
| | • | Bulb Slope Soil Sampling Location | | |
| | • | Location Exceeding RBC | | |
| | \sim | 2- Foot vertical elevation contour in feet NGVD | | |
| | | Landfill Access Road | | |
| | | AOPC Boundary | | |
| ļ | | Approximate Waterline at Mean Poo Elevation (approx. 74 feet NGVD) | I | |
| | Invertebr Mercury | ate High SLV* 0.500 | | |
| | CEC = ch HQ = haz LOAEL = mg/kg = r RBC = Ri | area of potential concern emical of ecological concern ard quotient lowest-observable-adverse-effect-lev nilligrams per kilogram sk-based concentration reening level value | el | |
| | Only con *Wildlife | concentrations are in mg/kg. centrations above the RBC are showr and plant RBCs were not derived bec L/High SLV HQs < 1. | | |
| | BU | LB SLOPE AOPC | drawing ni FIGUF | |
| | CEC LO | CATIONS ABOVE RBCs | cad file n FIG | UMBER: |
| | | | SHEET: OF | REV. |

APPENDIX A BHHRA Calculation Spreadsheets

Appendix A BHHRA Calculation Spreadsheets

| Table A-1.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Outdoor Maintenance Worker RME, Landfill AOPC |
|-------------|---|
| Table A-1.2 | Derivation of Inhalation Factors for Inhalation Exposures, Outdoor Maintenance Worker RME, Landfill AOPC |
| Table A-1.3 | Risk and Hazard Estimates: Inhalation, Outdoor Maintenance Worker RME, Landfill AOPC |
| Table A-1.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Outdoor Maintenance Worker RME, Landfill AOPC |
| | (See Table 2-9. Outdoor Maintenance Worker: RME Summary - Landfill AOPC) |
| Table A-2.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Outdoor Maintenance Worker CTE, Landfill AOPC |
| Table A-2.2 | Derivation of Inhalation Factors for Inhalation Exposures, Outdoor Maintenance Worker CTE, Landfill AOPC |
| Table A-2.3 | Risk and Hazard Estimates: Inhalation, Outdoor Maintenance Worker, Landfill AOPC |
| Table A-2.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Outdoor Maintenance Worker CTE, Landfill AOPC |
| | (See Table 2-10. Outdoor Maintenance Worker: CTE Summary - Landfill AOPC) |
| Table A-3.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Construction Worker RME, Landfill AOPC |
| Table A-3.2 | Derivation of Inhalation Factors for Inhalation Exposures, Construction Worker RME, Landfill AOPC |
| Table A-3.3 | Risk and Hazard Estimates: Inhalation, Construction Worker, Landfill AOPC |
| Table A-3.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Construction Worker RME, Landfill AOPC |
| | (See Table 2-11. Construction Worker: RME Summary - Landfill AOPC) |
| Table A-4.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Construction Worker CTE, Landfill AOPC |



| Table A-4.2 | Derivation of Inhalation Factors for Inhalation Exposures, Construction Worker CTE, Landfill AOPC |
|--------------|---|
| Table A-4.3 | Risk and Hazard Estimates: Inhalation, Construction Worker, Landfill AOPC |
| Table A-4.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Construction Worker CTE, Landfill AOPC |
| | (See Table 2-12. Construction Worker: CTE Summary - Landfill AOPC) |
| Table A-5.NI | Nursing Infant: Hypothetical Fishing Platform User: RME Summary - Landfill AOPC |
| Table A-5.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: RME, Landfill AOPC |
| Table A-5.2 | Derivation of Inhalation Factors for Inhalation Exposures, Hypothetical Fishing Platform User: RME, Landfill AOPC |
| Table A-5.3 | Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Landfill AOPC |
| Table A-5.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: RME, Landfill AOPC |
| | (See Table 2-13. Hypothetical Fishing Platform User: RME Summary - Landfill AOPC) |
| Table A-6.NI | Nursing Infant: Hypothetical Fishing Platform User: CTE Summary - Landfill AOPC |
| Table A-6.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: CTE, Landfill AOPC |
| Table A-6.2 | Derivation of Inhalation Factors for Inhalation Exposures, Hypothetical Fishing Platform User: CTE, Landfill AOPC |
| Table A-6.3 | Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Landfill AOPC |
| Table A-6.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: CTE, Landfill AOPC |
| | (See Table 2-14. Hypothetical Fishing Platform User: CTE Summary - Landfill AOPC) |
| | |



| Table A-7.1 | Risk and Hazard Estimates - Ingestion of Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC |
|-------------|--|
| Table A-7.2 | Risk and Hazard Estimates - Inhalation of Volatile Compounds from Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC |
| Table A-7.3 | Estimation of Kp, Tau(event), B, and t* for Organic Compounds: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC |
| Table A-7.4 | Calculation of Dose Absorbed Per Unit Area Per Event: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC |
| Table A-7.5 | Risk and Hazard Estimates: Dermal Contact with Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC |
| | (See Table 2-15. Excavation/Trench Worker, Exposure to Groundwater: RME Summary – Landfill AOPC) |
| Table A-8.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Outdoor Maintenance Worker RME, Sandblast Area AOPC |
| Table A-8.2 | Derivation of Inhalation Factors for Inhalation Exposures, Outdoor Maintenance Worker RME, Sandblast Area AOPC |
| Table A-8.3 | Risk and Hazard Estimates: Inhalation, Outdoor Maintenance Worker, Sandblast Area AOPC |
| Table A-8.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Outdoor Maintenance Worker RME, Sandblast Area AOPC |
| | (See Table 2-16. Outdoor Maintenance Worker: RME Summary - Sandblast Area AOPC) |
| Table A-9.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Outdoor Maintenance Worker CTE, Sandblast Area AOPC |
| Table A-9.2 | Derivation of Inhalation Factors for Inhalation Exposures, Outdoor Maintenance Worker CTE, Sandblast Area AOPC |
| Table A-9.3 | Risk and Hazard Estimates: Inhalation, Outdoor Maintenance Worker, Sandblast Area AOPC |
| Table A-9.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Outdoor Maintenance Worker CTE, Sandblast Area AOPC |
| | (See Table 2-17. Outdoor Maintenance Worker: CTE Summary - Sandblast Area AOPC |



| Table A-10.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Construction Worker RME, Sandblast Area AOPC |
|---------------|--|
| Table A-10.2 | Derivation of Inhalation Factors for Inhalation Exposures, Construction Worker RME, Sandblast Area AOPC |
| Table A-10.3 | Risk and Hazard Estimates: Inhalation, Construction Worker, Sandblast Area AOPC |
| Table A-10.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Construction Worker RME, Sandblast Area AOPC |
| | (See Table 2-18. Construction Worker: RME Summary - Sandblast Area AOPC) |
| Table A-11.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Construction Worker CTE, Sandblast Area AOPC |
| Table A-11.2 | Derivation of Inhalation Factors for Inhalation Exposures, Construction Worker CTE, Sandblast Area AOPC |
| Table A-11.3 | Risk and Hazard Estimates: Inhalation, Construction Worker, Sandblast Area AOPC |
| Table A-11.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Construction Worker CTE, Sandblast Area AOPC |
| | (See Table 2-19. Construction Worker: CTE Summary - Sandblast Area AOPC) |
| Table A-12.NI | Nursing Infant: Hypothetical Fishing Platform User: RME Summary - Sandblast Area AOPC |
| Table A-12.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: RME, Sandblast Area AOPC |
| Table A-12.2 | Derivation of Inhalation Factors for Inhalation Exposures, Hypothetical Fishing Platform User: RME, Sandblast Area AOPC |
| Table A-12.3 | Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Sandblast Area AOPC |
| Table A-12.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: RME, Sandblast Area AOPC |
| | (See Table 2-20. Hypothetical Fishing Platform User: RME Summary - Sandblast Area AOPC) |



| Table A-13.NI | Nursing Infant: Hypothetical Fishing Platform User: CTE Summary - Sandblast Area AOPC |
|---------------|---|
| Table A-13.1 | Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: CTE, Sandblast Area AOPC |
| Table A-13.2 | Derivation of Inhalation Factors for Inhalation Exposures, Hypothetical Fishing Platform User: CTE, Sandblast Area AOPC |
| Table A-13.3 | Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Sandblast Area AOPC |
| Table A-13.4 | Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: CTE, Sandblast Area AOPC |
| | (See Table 2-21. Hypothetical Fishing Platform User: CTE Summary - Sandblast Area AOPC) |
| Table A-14.1 | Risk and Hazard Estimates - Ingestion of Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC |
| Table A-14.2 | Risk and Hazard Estimates - Inhalation of Volatile Compounds from Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC |
| Table A-14.3 | Estimation of Kp, Tau(event), B, and t* for Organic Compounds: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC |
| Table A-14.4 | Calculation of Dose Absorbed Per Unit Area Per Event: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC |
| Table A-14.5 | Risk and Hazard Estimates: Dermal Contact with Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC |
| | (See Table 2-22. Excavation/Trench Worker, Exposure to Groundwater: RME Summary – Sandblast Area AOPC) |
| | See Table 2-23. Indoor Worker, Exposure to Indoor Air: RME Summary – Sandblast Area AOPC See Table 2-24. Indoor Worker, Exposure to Indoor Air: CTE Summary – Sandblast |
| | See Fable 2 27. Indoor Worker, Exposure to indoor An. CTE Summary – Sandolast |

ABBREVIATIONS

% = percent --- = data not available or not calculated μ g= microgram ABS_d = Dermal Soil Absorption Fraction ADAF = age-dependent adjustment factor

Area AOPC



Afw = Soil-to-Skin Adherence Fraction, worker

AOPC = area of potential concern

Atc = Averaging Time, carcinogens

 $AT_{nc,a}$ = Averaging Time, noncarcinogens, adult

 $BW_a = Body Weight, adult$

 $C_{air} = COPC$ Concentration in air

CF = Conversion Factor

 $CF_i = Unit$ conversion factor

 $cm^2 = square centimeter$

 $cm^3 = cubic centimeter$

COPC = chemical of potential concern

 $C_{soil} = COPC$ Concentration in Soil

 $D_A = Apparent Diffusivity$

DEHP = bis(2-ethylhexyl)phthalate

D_i = Diffusivity in air

D_w = Diffusivity in water

 EC_w = Fraction of EV in Contact with Soil, worker

 $ED_w = Exposure Duration, worker$

 $EF_w = Exposure Frequency, worker$

 $ET_w = Exposure Time, worker$

EV = event

 $EV_w = Event$ Frequency, worker

FIw = Fraction Contaminated Soil Ingested, worker

 F_{OC} = Default organic-carbon content

g = gram

H' = Henry's law constant

HQ = hazard quotient

IRAF = Infant Risk Adjustment Factor

IRS_w = Soil Ingestion Rate, worker

IUR = Inhalation Unit-Risk Factor

 K_d = Soil-water partition coefficient for organics: Kd= KOC× FOC

kg = kilogram

KM - capped = Kaplan–Meier-based with Efron's bias correction, capped

K_{OC} = Soil-organic carbon partition coefficient

 K_p = permeability coefficient from water

 $m^3 = cubic meter$

MCi = maximum concentration inhaled

n = Total soil porosity (expressed as L_{porespace}/L_{soil})

NRP = not reliably predicted

PAH = polycyclic aromatic hydrocarbon



PCB = polychlorinated biphenyl

PEF = Particulate Emission Factor (non-VOCs), default

Q/C = Inverse of the mean concentration at the center of a 0.5-acre square source in Los Angeles

 θ_a = Air-filled soil porosity (expressed as L_{air}/L_{soil})

 θ_w = Water-filled soil porosity (expressed as L_{water}/L_{soil})

 $\rho_{\beta} = \text{Dry soil bulk density}$

RfC = Inhalation Reference Concentration

 $RfD_d = Oral Reference Dose Adjusted for GI Absorption$

 $RfD_0 = Oral Reference Dose$

RME = reasonable maximum exposure

RSL = regional screening level

s = second

 $SA_w = Exposed Body Surface Area, worker$

 SF_d = Oral Slope Factor Adjusted for GI Absorption

 $SF_O = Oral Slope Factor$

SVOC = semivolatile organic compound

 $t^* = time it takes to reach steady state$

 T_{worker} = Exposure interval, worker

VF = volatilization factor

VOC = volatile organic chemical

yr = year

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 Table A-1.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Outdoor Maintenance Worker RME, Landfill AOPC

| Definition | Variable | Value | | Equations | | | | |
|---|----------------------------|-------------------|---------------------------|-------------------|---|---|--|---------------|
| Averaging Time, carcinogens | AT _c | 70 | yrs | Cancer Risk: | | | | |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 25 | yrs | | ſ | | $\rightarrow FI \rightarrow CF$ | ו |
| Body Weight | BW_a | 80 | kg | $Risk = SF_{a}$ | $\times \{C_{noil} \times$ | $\frac{IRS_{w} \times EF_{w} \times EI}{BW_{a} \times AT_{c} \times 36}$ | $\mathcal{P}_{W} \times I^{\prime} I_{W} \times C I_{O}$ | } |
| Conversion Factor | CFo | 1E-06 | kg/mg | 0 | sou | $BW_a \times AT_c \times 30$ | 55day∕ year) | |
| COPC Concentration in Soil | C _{soil} | chemical-specific | mg/kg | | C | | - | |
| Exposure Duration, worker | ED_w | 25 | yrs | Noncancer Hazard: | | | | |
| Exposure Frequency, worker | EF_{w} | 225 | days/yr | | 1 | | | |
| Fraction Contaminated Soil Ingested, worker | FI_{w} | 1.0 | unitless | Hazard=- | $\frac{1}{1} \times \left\{ c \right\}$ | $C_{soil} \times \left \frac{IRS_w \times EF_v}{BW_a \times AT} \right $ | $_{v} \times ED_{w} \times FI_{w}$ | $\times CF_o$ |
| Soil Ingestion Rate, worker | IRS_w | 100 mg/day | | R | $2fD_{a}$ | $ BW_a \times AT$ | $T_{nca} \times 365 day/$ | year |
| Oral Reference Dose | RfD ₀ | chemical-specific | mg/kg-day | | | | ne,u - | (ב י |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Noncancer | |
| | | C _{soil} | SFo | Risk | % of | RfDo | Hazard | % of |
| Analyte | | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | |
| Arsenic | | 1.05E+01 | 1.50E+00 | 2.60E-06 | 7% | 3.00E-04 | 1.62E-02 | 98% |
| Chromium (III) | | 5.94E+02 | No Toxicity Value | | | 1.50E+00 | 3.05E-04 | 2% |
| Lead | | 5.11E+02 | | | | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]anthracene | | 9.96E+00 | 7.30E-01 | 2.00E-06 | 5% | No Toxicity Value | | |
| Benzo[a]pyrene | | 1.12E+01 | 7.30E+00 | 2.25E-05 | 57% | No Toxicity Value | | |
| Benzo[b]fluoranthene | | 1.52E+01 | 7.30E-01 | 3.05E-06 | 8% | No Toxicity Value | | |
| Dibenz[a,h]anthracene | | 2.06E+00 | 7.30E+00 | 4.14E-06 | 10% | No Toxicity Value | | |
| Indeno[1,2,3-cd]pyrene | | 6.80E+00 | 7.30E-01 | 1.37E-06 | 3% | No Toxicity Value | | |
| Benzofluoranthenes, total | | 1.88E+01 | 7.30E-01 | 3.77E-06 | 10% | No Toxicity Value | | |
| | | | | Cancer Risk | | | Hazard Index | |
| | | | Pathway Sum | 3.9E-05 | | | 1.6E-02 | |

 Table A-1.2

 Derivation of Inhalation Factors for Inhalation Exposures, Outdoor Maintenance Worker RME, Landfill AOPC

| Description | Variable | Value | | Units | VF | Derivation | | | | | | | | |
|---|---------------------|----------------------|--|----------------------|----------------------|---|-----------------------------|---|------------|---|----------------|---------------------------|--------------------|--------------------|
| Unit conversion factor | CFi | 1.00E-0 | $04 \text{ m}^2/\text{cm}^2$ | | | | (| | | | | | | |
| Apparent Diffusivity | D _A | derive | ed cm ² /s | | | VFs = (0) / (0) | (3 . 14) | $\frac{1}{2} \times D_A \times T_{[receptor]} ^{1/2} \times C.$ | F | | | | | |
| Diffusivity in air | Di | chemical-specifi | ic cm ² /s | | | | , , , | $(2 \times \rho_b \times D_A)$ | i i | | | | | |
| Diffusivity in water | D_w | chemical-specifi | ic cm ² /s | | | | | | | | | | | |
| Default organic-carbon content | F _{OC} | - | 01 g/g | | | | | | | | | | | |
| Henry's law constant | H' | chemical-specifi | ic dimensio | onless | | | | | | | | | | |
| Soil-water partition coefficient | K _d | chemical-specifi | ic cm ³ /g | | | | | | | | | | | |
| for organics: Kd=K _{OC} ×F _{OC} | | | | | base | d on: | | | | | | | | |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specifi | ic cm ³ /g | | | [(10 / 3 | 10 / 3 |). 21 | | | | | | |
| Total soil porosity | n | 0.4 | 43 L _{porespace} /I | L _{soil} | | $DA = \frac{\left[\left(\Theta \frac{10}{a} / 3 D_{i} H\right) + H\right]}{\rho_{R} K d}$ | $'+\Theta_w^{10} , ^{3}D_w$ | <u>]/ n ²</u>] | | | | | | |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.360E+0 | | | | $\rho_{B} Kd$ | $+\Theta_w +\Theta_a I$ | H ' | | | | | | |
| Inverse of the mean concentration at the center | Q/C | | 44 g/m ² -s pe | er kg/m ³ | | | | | | | | | | |
| of a 0.5-acre square source in Los Angeles | | | U 1 | 0 | | | | | | | | | | |
| Dry soil bulk density | ρ_b | 1. | $.5 \text{ g/cm}^3$ | | | | | | | | | | | |
| Air-filled soil porosity | θ_{a} | 0.2 | 28 L _{air} /L _{soil} | | | | | | | | | | | |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0.1 | 15 L _{water} /L _{soil} | 1 | | | | | | | | | | |
| Exposure interval, 25-yr occupational worker | T _{Worker} | 7.88E+0 | 08 seconds | | | | | | | | | | | |
| Volatilization Factor for soil | VFs | derive | ed m ³ /kg | | | | | | | | | | | |
| | | | | | | | | | | | | | VFs (VOCs) | PEF |
| | | | Di | | | $\mathbf{D}_{\mathbf{w}}$ | | H' | | K _{oc} | K _d | $\mathbf{D}_{\mathbf{A}}$ | Occupational | (non-VOCs) |
| Analyte | | (cm ² /s) | | Reference | (cm ² /s) | Reference | (dimensionless) | Reference | (cm^3/g) | Reference | (cm^3/g) | (cm^2/s) | m ³ /kg | m ³ /kg |
| Inorganic Constituents | | | | | | | | | | | | | | |
| Arsenic | | 0 | 0 | | 0 0 | | 31.61052561 | Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 29 | 0 | non-VOC | 1.360E+09 |
| Chromium (III) | | 0 | 0 | | 0 0 | | 0 | 0 | 0 | 0 | 1800000 | 0 | non-VOC | 1.360E+09 |
| Lead | | 0 | 0 | | 0 0 | | 1.001885999 | Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 900 | 0 | non-VOC | 1.360E+09 |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | | |
| Benzo[a]anthracene | | 0.0508647 | RSLs ("V | WATER9"; USEPA 2012) | 5.943E-06 RSL | s ("WATER9"; USEPA 2012) | 0.00049072 | Experimental, EPISuite v4.10 (USEPA 2011a) | 176900 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 1769 | 7.523E-10 | non-VOC | 1.360E+09 |
| Benzo[a]pyrene | | 0.0475831 | RSLs ("V | WATER9"; USEPA 2012) | 5.56E-06 RSL | s ("WATER9"; USEPA 2012) | 1.86882E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 587400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 5874 | 1.396E-11 | non-VOC | 1.360E+09 |
| Benzo[b]fluoranthene | | 0.0475831 | RSLs ("V | WATER9"; USEPA 2012) | 5.56E-06 RSL | s ("WATER9"; USEPA 2012) | 2.68669E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 599400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 5994 | 1.704E-11 | non-VOC | 1.360E+09 |
| Dibenz[a,h]anthracene | | 0.0445672 | RSLs ("V | WATER9"; USEPA 2012) | 5.207E-06 RSL | s ("WATER9"; USEPA 2012) | 5.76596E-06 | Experimental, EPISuite v4.10 (USEPA 2011a) | 1912000 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 19120 | 2.457E-12 | non-VOC | 1.360E+09 |
| Indeno[1,2,3-cd]pyrene | | 0.0447842 | | , , , | | s ("WATER9"; USEPA 2012) | 0.0000656 | EPISuite v4.10 (USEPA 2011a) | 3470000 | | 34700 | 5.359E-12 | non-VOC | 1.360E+09 |
| Benzofluoranthenes, total | | 0.0475831 | RSLs ("V | WATER9"; USEPA 2012) | 5.56E-06 RSL | s ("WATER9"; USEPA 2012) | 2.68669E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 599400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 5994 | 1.704E-11 | non-VOC | 1.360E+09 |

| Description | Variable | Val | ue | | Equations | | | | | |
|---|----------------------------|------------------------------|------------------------|-------------------|--|----------------------------------|-----------------|--|-------------------------|-------|
| Averaging Time, carcinogens | AT _c | | 70 yrs | Cancer Risk: | | (| | | | |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | | 25 yrs | | $Risk = IUR \times$ | $C_{air} \times $ | | $\frac{EF_{w} \times ED_{w} \times ET}{days / year \times 24}$ | <u>w</u> | |
| COPC Concentration in air | \mathbf{C}_{air} | chemical-speci | fic µg/m ³ | | | (AT_c) | × 365 | days / year $\times 24$ | hours / day) | |
| COPC Concentration in soil | C _{soil} | chemical-speci | fic mg/kg | | | | | | | |
| Exposure Duration, worker | ED_w | | 25 yrs | Noncancer Hazard: | 1 | (| | $FF \times FD$ | $\times FT$ | |
| Exposure Frequency, worker | EF_w | | 25 days/yr | | Hazard = $\frac{1}{Df}$ | $-\times C_{air} \times -$ | 17 | $\frac{EF_{w} \times ED_{w}}{2 \times 365 \ days \ / \ yea}$ | $\sim 21 W$ | |
| Exposure Time, worker | ET_{w} | 8. | 00 hours/day | | ĸjt | | н <i>пс</i> , и | , × 565 aays / yea | $r \times 24$ nours / a | aay j |
| Inhalation Unit-Risk Factor | IUR | chemical-speci | fic $(\mu g/m^3)^{-1}$ | | | | | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+ | 09 m ³ /kg | | | | | | | |
| Inhalation Reference Concentration | RfC | chemical-speci | fic µg/m ³ | where: | $C_{air} = \frac{C_{soil}}{VF \ or \ P}$ | $\times 1000 \frac{\mu g}{1000}$ | - | | | |
| Volatilization Factor for VOCs | VF | chemical-speci | fic m ³ /kg | | VF or P | PEF mg | | | | |
| | | | VF (VOCs) or | | | Cancer | | | Hazard | |
| | | $\mathbf{C}_{\mathbf{soil}}$ | PEF (non-VOCs) | C _{air} | IUR | Risk | % of | RfC | Quotient | % of |
| Analyte | | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) |) Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | | | |
| Arsenic | | 1.05E+01 | 1.360E+09 | 7.72E-06 | 4.30E-03 | 1.46E-09 | 57% | 1.50E-02 | 6.35E-05 | 100% |
| Chromium (III) | | 5.94E+02 | 1.360E+09 | 4.37E-04 | No Toxicity Value | | | No Toxicity Value | | |
| Lead | | 5.11E+02 | 1.360E+09 | 3.76E-04 | | | | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | |
| Benzo[a]anthracene | | 9.96E+00 | 1.360E+09 | 7.32E-06 | 1.10E-04 | 5.91E-11 | 2% | No Toxicity Value | | |
| Benzo[a]pyrene | | 1.12E+01 | 1.360E+09 | 8.24E-06 | 1.10E-03 | 6.65E-10 | 26% | No Toxicity Value | | |
| Benzo[b]fluoranthene | | 1.52E+01 | 1.360E+09 | 1.12E-05 | 1.10E-04 | 9.02E-11 | 4% | No Toxicity Value | | |
| Dibenz[a,h]anthracene | | 2.06E+00 | 1.360E+09 | 1.51E-06 | 1.20E-03 | 1.33E-10 | 5% | No Toxicity Value | | |
| Indeno[1,2,3-cd]pyrene | | 6.80E+00 | 1.360E+09 | 5.00E-06 | 1.10E-04 | 4.04E-11 | 2% | No Toxicity Value | | |
| Benzofluoranthenes, total | | 1.88E+01 | 1.360E+09 | 1.38E-05 | 1.10E-04 | 1.11E-10 | 4% | No Toxicity Value | | |
| | | | | | | Cancer Risk | | • | Hazard Index | |
| | | | | | Pathway Sums: | 2.6E-09 | | | 6.3E-05 | |

 Table A-1.3

 Risk and Hazard Estimates: Inhalation, Outdoor Maintenance Worker RME, Landfill AOPC

 Table A-1.4

 Risk and Hazard Estimates: Dermal Contact with Soil, Outdoor Maintenance Worker RME, Landfill AOPC

| Description | Variable | Valu | 9 | Equations: | | | | | |
|--|----------------------------|-------------------|---------------------------|---------------------------|---|---|--|---|---|
| Dermal Soil Absorption Fraction | ABS_d | chemical-specifi | c unitless | Cancer Risk: | | | | | |
| Soil-to-Skin Adherence Fraction, worker | AF_w | 0.12 | 2 mg/cm ² -day | | $\int \int \Delta F$ | $\times ARS \times FF$ | $F \times FD \times FV$ | EC XSA XCE | |
| Averaging Time, carcinogens | AT _c | 70 |) yrs | Risk= | $SF_d \times \{C_{soil} \times \frac{m_w}{m}$ | | $_{W}$ \sim LD_{W} \sim LV_{W} | $\frac{\langle EC_{w} \times SA_{w} \times CF_{d}}{\forall \ year}$ | > |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 2: | 5 yrs | | u sou | BW_a | $\times AT_c \times 365 day$ | / year | |
| Body Weight, adult | BW_a | 80 |) kg | | , | | | , | |
| Conversion Factor | CF_d | 1E-00 | ó kg/mg | Noncancer Hazard: | | | | | |
| COPC Concentration in Soil | C _{soil} | chemical-specifi | c mg/kg | | 1 | | $2S \vee FE \vee FD$ | $\vee FV \vee FC \vee S$ | $(\sqrt{CE}]$ |
| Fraction of EV in Contact with Soil, worker | EC_w | | unitless | Hazar | $d = \frac{1}{C_{soil}} \times \{C_{soil}\}$ | $\left \frac{AT_{w} \times AD}{M}\right $ | $S_d \times ET_w \times ED$ | $\frac{0_{w} \times EV_{w} \times EC_{w} \times SA}{\times 365 day / year}$ | $\frac{\mathbf{h}_{w} \times \mathbf{C} \mathbf{r}_{d}}{ }$ |
| Exposure Duration, worker | ED_w | 2: | 5 yrs | | RfD_d | | $BW_a \times AT_{nc,a}$ | ×365 day/ year | ļ |
| Exposure Frequency, worker | EF_{w} | 22: | 5 days/yr | | × × | - | | | _, |
| Event Frequency, worker | EV_w | | events/day | | | | | | |
| Oral Reference Dose Adjusted for GI Absorption | RfD_d | chemical-specifi | | | | | | | |
| Exposed Body Surface Area, worker | SAs | 352 | 7 cm^2 | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specifi | c (mg/kg-day) | 1 | | | | | |
| | | | | | Cancer | | | Hazard | |
| | | C _{soil} | ABS _d | SFd | Risk | % of | R f D _d | Quotient | % of |
| Analyte | | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | | |
| Arsenic | | 1.05E+01 | 3.00E-02 | 1.50E+00 | 3.30E-07 | 2% | 3.00E-04 | 2.05E-03 | 67% |
| Chromium (III) | | 5.94E+02 | 1.00E-02 | | | | 1.95E-02 | 9.93E-04 | 33% |
| Lead | | 5.11E+02 | 1.00E-02 | | | | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | |
| Benzo[a]anthracene | | 9.96E+00 | 1.30E-01 | 7.30E-01 | 1.10E-06 | 5% | | | |
| Benzo[a]pyrene | | 1.12E+01 | 1.30E-01 | 7.30E+00 | 1.24E-05 | 60% | | | |
| Benzo[b]fluoranthene | | 1.52E+01 | 1.30E-01 | 7.30E-01 | 1.68E-06 | 8% | | | |
| Dibenz[a,h]anthracene | | 2.06E+00 | 1.30E-01 | 7.30E+00 | 2.28E-06 | 11% | | | |
| Indeno[1,2,3-cd]pyrene | | 6.80E+00 | 1.30E-01 | 7.30E-01 | 7.52E-07 | 4% | | | |
| Benzofluoranthenes, total | | 1.88E+01 | 1.30E-01 | 7.30E-01 | 2.08E-06 | 10% | | | |
| | | | | D.I. C | Cancer Risk | | - | Hazard Index | |
| | | | | Pathway Sums: | 2.1E-05 | | | 3.0E-03 | |

 Table A-2.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Outdoor Maintenance Worker CTE, Landfill AOPC

| Definition | Variable | Value | | Equations | | | | |
|---|-----------------------------|-------------------|---------------------------|------------------------|------------------------------|---|---|--------------------------------|
| Averaging Time, carcinogens | AT_{c} | 70 | yrs | Cancer Risk: | | | | |
| Averaging Time, noncarcinogens, adult | $AT_{nc,a}$ | 6 | yrs | | [[| - IRS $\vee FF \vee FI$ | $D \rightarrow FI \rightarrow CF$ | ן (ך |
| Body Weight | BW_a | 80 | kg | $Risk = SF_a$ | $\times \{C_{soil} \times $ | $\frac{IRS_{w} \times EF_{w} \times EI}{BW_{a} \times AT_{c} \times 3}$ | $\mathcal{D}_{W} \wedge \Gamma I_{W} \wedge C \Gamma_{O}$ | . } |
| Conversion Factor | CFo | 1E-06 | kg/mg | 0 | 5011 | $BW_a \times AT_c \times 3$ | 65day∕ year) | |
| COPC Concentration in Soil | C _{soil} | chemical-specific | mg/kg | | | | | _, |
| Exposure Duration, worker | ED_w | 6 | yrs | Noncancer Hazard: | | | | |
| Exposure Frequency, worker | EF_{w} | 225 | days/yr | | 1 (| | | $\sim CE$ |
| Fraction Contaminated Soil Ingested, worker | FI_{w} | | unitless | Hazard = - | $\frac{1}{-} \times \{C$ | $\left \frac{IRS_{w} \times EF}{BW_{a} \times AT} \right $ | $_{W} \times ED_{W} \times FI_{W}$ | $\left \times CF_{o} \right $ |
| Soil Ingestion Rate, worker | $\mathrm{IRS}_{\mathrm{w}}$ | | mg/day | R | $2fD_o$ | $BW_a \times AT$ | $T_{nc,a} \times 365 day$ | / year |
| Oral Reference Dose | RfD ₀ | chemical-specific | | | C | L | | 1) |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Noncancer | |
| | | C _{soil} | SFo | Risk | % of | RfD _o | Hazard | % of |
| Analyte | | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | |
| Arsenic | | 1.05E+01 | 1.50E+00 | 3.12E-07 | 7% | 3.00E-04 | 8.09E-03 | 98% |
| Chromium (III) | | 5.94E+02 | No Toxicity Value | | | 1.50E+00 | 1.53E-04 | 2% |
| Lead | | 5.11E+02 | | | | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]anthracene | | 9.96E+00 | 7.30E-01 | 2.40E-07 | 5% | No Toxicity Value | | |
| Benzo[a]pyrene | | 1.12E+01 | 7.30E+00 | 2.70E-06 | 57% | No Toxicity Value | | |
| Benzo[b]fluoranthene | | 1.52E+01 | 7.30E-01 | 3.66E-07 | 8% | No Toxicity Value | | |
| Dibenz[a,h]anthracene | | 2.06E+00 | 7.30E+00 | 4.97E-07 | 10% | No Toxicity Value | | |
| Indeno[1,2,3-cd]pyrene | | 6.80E+00 | 7.30E-01 | 1.64E-07 | 3% | No Toxicity Value | | |
| Benzofluoranthenes, total | | 1.91E+01 | 7.30E-01 | 4.60E-07 | 10% | No Toxicity Value | | |
| | | | Pathway Sum | Cancer Risk 4.7E-06 | | | Hazard Index 8.2E-03 | |

 Table A-2.2

 Derivation of Inhalation Factors for Inhalation Exposures, Outdoor Maintenance Worker CTE, Landfill AOPC

| Description | Variable | Value | Units | VFL | erivation | | | | | | | | |
|---|---------------------|----------------------|---|----------------------|---|----------------------------|--|----------------------|---|---------------------------|---------------------------|--------------|----------|
| Unit conversion factor | CFi | 1.00E-04 | $4 \text{ m}^2/\text{cm}^2$ | | | (| | | | | | | |
| Apparent Diffusivity | D _A | derive | d cm ² /s | | VEs = (0 / C) | (3.14) | $\frac{\times D_A \times T_{[receptor]}}{(2 \times \rho_b \times D_A)} \times 0$ | CF | | | | | |
| Diffusivity in air | Di | chemical-specifi | c cm ² /s | | VIS = (Q / C) |) ^ | $(2 \times \rho_b \times D_A)$ | cr _i | | | | | |
| Diffusivity in water | Dw | chemical-specifi | $c cm^2/s$ | | | | | | | | | | |
| Default organic-carbon content | F _{oc} | - | 1 g/g | | | | | | | | | | |
| Henry's law constant | H' | chemical-specifi | | | | | | | | | | | |
| Soil-water partition coefficient | K _d | chemical-specifi | c cm ³ /g | | | | | | | | | | |
| for organics: $Kd = K_{OC} \times F_{OC}$ | u | Ĩ | - | based | l on: | | | | | | | | |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specifi | c cm ³ /g | | [(a 10 / 3 a ar | -2 10 (3 -) | 2] | | | | | | |
| Total soil porosity | n | 0.4 | 3 L _{porespace} /L _{soil} | | $DA = \frac{\left[\left(\Theta_{a}^{10}/3 D_{i} H\right) + P_{a} K d\right]}{\rho_{R} K d} + \frac{1}{2}$ | $+\Theta_w^{10}/5D_w^{-1}$ | <u>n - </u>] | | | | | | |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.320E+0 | 9 m ³ /kg | | $\rho_{B} Kd +$ | $\Theta_w + \Theta_a H'$ | | | | | | | |
| Inverse of the mean concentration at the center of a 0.5-acre square source in Los Angeles | Q/C | 73.4 | 4 g/m ² -s per kg/m ³ | | | | | | | | | | |
| Dry soil bulk density | ρ_{b} | 1.: | 5 g/cm ³ | | | | | | | | | | |
| Air-filled soil porosity | θ_{a} | 0.2 | 8 L _{air} /L _{soil} | | | | | | | | | | |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0.1 | 5 L _{water} /L _{soil} | | | | | | | | | | |
| Exposure interval, 6-yr occupational worker | T _{Worker} | 1.89E+0 | 8 seconds | | | | | | | | | | |
| Volatilization Factor for soil | VFs | derive | d m ³ /kg | | | | | | | | | | |
| | | | | | | | | | | | | VFs (VOCs) | |
| | | | $\mathbf{D}_{\mathbf{i}}$ | | $\mathbf{D}_{\mathbf{w}}$ | | H' | | K _{OC} | $\mathbf{K}_{\mathbf{d}}$ | $\mathbf{D}_{\mathbf{A}}$ | Occupational | ```` |
| Analyte | | (cm ² /s) | Reference | (cm ² /s) | Reference | (dimensionless) | Reference | (cm ³ /g) | Reference | (cm ³ /g) | (cm ² /s) | m³/kg | m³/kg |
| Inorganic Constituents | | | | | | | | | | | | | |
| Arsenic | | 0 | 0 | 0 0 | | 31.61052561 Bo | ond Method estimate, EPISuite v4.10 (USEPA, 2011a) | 13.22 EPI | Suite v4.10 (USEPA, 2011a); MCI-estimated | 29 | 0 | non-VOC | 1.36E+09 |
| Chromium (III) | | 0 | 0 | 0 0 | | 0 0 | | 0 0 | | 1800000 | 0 | non-VOC | 1.36E+09 |
| Lead | | 0 | 0 | 0 0 | | 1.001885999 Bo | ond Method estimate, EPISuite v4.10 (USEPA, 2011a) | 13.22 EPI | Suite v4.10 (USEPA, 2011a); MCI-estimated | 900 | 0 | non-VOC | 1.36E+09 |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | |
| Benzo[a]anthracene | | 0.0508647 | RSLs ("WATER9"; USEPA, 201 | 12) 5.94E-06 RSL | ("WATER9"; USEPA, 2012) | 0.00049072 Ex | perimental, EPISuite v4.10 (USEPA, 2011a) | 176900 EPI | Suite v4.10 (USEPA, 2011a); MCI-estimated | 1769 | 7.5E-10 | non-VOC | 1.36E+09 |
| Benzo[a]pyrene | | 0.0475831 | RSLs ("WATER9"; USEPA, 201 | 12) 5.56E-06 RSL | ("WATER9"; USEPA, 2012) | 1.86882E-05 Ex | xperimental, EPISuite v4.10 (USEPA, 2011a) | 587400 EPI | Suite v4.10 (USEPA, 2011a); MCI-estimated | 5874 | 1.4E-11 | non-VOC | 1.36E+09 |
| Benzo[b]fluoranthene | | 0.0475831 | RSLs ("WATER9"; USEPA, 201 | 12) 5.56E-06 RSL | ("WATER9"; USEPA, 2012) | 2.68669E-05 Ex | perimental, EPISuite v4.10 (USEPA, 2011a) | 599400 EPI | Suite v4.10 (USEPA, 2011a); MCI-estimated | 5994 | 1.7E-11 | non-VOC | 1.36E+09 |
| Dibenz[a,h]anthracene | | 0.0445672 | RSLs ("WATER9"; USEPA, 201 | / | ("WATER9"; USEPA, 2012) | | perimental, EPISuite v4.10 (USEPA, 2011a) | | Suite v4.10 (USEPA, 2011a); MCI-estimated | 19120 | 2.5E-12 | non-VOC | 1.36E+09 |
| Indeno[1,2,3-cd]pyrene | | 0.0447842 | RSLs ("WATER9"; USEPA, 201 | / | ("WATER9"; USEPA, 2012) | | PISuite v4.10 (USEPA, 2011a) | | Suite v4.10 (USEPA, 2011a) | 34700 | 5.4E-12 | non-VOC | 1.36E+09 |
| Benzofluoranthenes, total | | 0.0475831 | RSLs ("WATER9"; USEPA, 201 | 12) 5.56E-06 RSL | ("WATER9"; USEPA, 2012) | 2.68669E-05 Ex | perimental, EPISuite v4.10 (USEPA, 2011a) | 599400 EPI | Suite v4.10 (USEPA, 2011a); MCI-estimated | 5994 | 1.7E-11 | non-VOC | 1.36E+09 |

| Description | Variable | Valu | ue | | Equations | | | | | |
|--|----------------------------|-------------------|------------------------|-----------------------------|---|--|-------------------|---|-----------------|-------|
| Averaging Time, carcinogens | AT _c | | 70 yrs | Cancer Risk: | • | (| F | |) | |
| Averaging Time, noncarcinogens, adult | $AT_{nc,a}$ | | 6 yrs | | $Risk = IUR \times$ | $C_{air} \times \boxed{-}$ | <u> </u> | $\frac{F_{w} \times ED_{w} \times ET_{v}}{lays / year \times 24}$ | w | |
| COPC Concentration in air | C _{air} | chemical-specif | fic μg/m ³ | | | $(AT_c$ | × 365 a | lays / year × 24 | hours / day) | |
| COPC Concentration in soil | C _{soil} | chemical-specif | ïc mg/kg | | | | | | | |
| Exposure Duration, worker | ED_{w} | | 6 yrs | Noncancer Hazard: | 1 | (| | $EF \times ED$ | < ET | |
| Exposure Frequency, worker | EF_{w} | | 25 days/yr | | Hazard = $\frac{1}{Pf}$ | $\frac{1}{2} \times C_{air} \times \left \frac{1}{4} \right $ | T | $\frac{EF_{w} \times ED_{w}}{\times 365 \ days \ / \ year}$ | $\sim 21 W$ | day |
| Exposure Time, worker | ET_{w} | | 00 hours/day | | ŊĊ | | \mathbf{I} nc,w | × 505 aays / year | × 24 nours / | aay j |
| Inhalation Unit-Risk Factor | IUR | chemical-specif | $ric (\mu g/m^3)^{-1}$ | | | | | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+0 | 09 m ³ /kg | | | | | | | |
| Inhalation Reference Concentration | RfC | chemical-specif | fic µg/m ³ | where: | $C_{air} = \frac{C_{soil}}{VF \text{ or } F}$ | $\times 1000 \frac{\mu g}{1000}$ | | | | |
| Volatilization Factor for VOCs | VF | chemical-specif | fic m ³ /kg | | VF or F | PEF mg | | | | |
| | | | VF (VOCs) or | | | Cancer | | | Hazard | |
| | | C _{soil} | PEF (non-VOCs) | $\mathbf{C}_{\mathbf{air}}$ | IUR | Risk | % of | RfC | Quotient | % of |
| Analyte | | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | | | |
| Arsenic | | 1.05E+01 | 1.360E+09 | 7.72E-06 | 4.30E-03 | 3.51E-10 | 57% | 1.50E-02 | 6.35E-05 | 100% |
| Chromium (III) | | 5.94E+02 | 1.360E+09 | 4.37E-04 | No Toxicity Value | | | No Toxicity Value | | |
| Lead | | 5.11E+02 | 1.360E+09 | 3.76E-04 | | | | | | |
| Polycyclic Aromatic Hydrocarbons (PA | Hs) | | | | | | | | | |
| Benzo[a]anthracene | | 9.96E+00 | 1.360E+09 | 7.32E-06 | 1.10E-04 | 1.42E-11 | 2% | No Toxicity Value | | |
| Benzo[a]pyrene | | 1.12E+01 | 1.360E+09 | 8.24E-06 | 1.10E-03 | 1.60E-10 | 26% | No Toxicity Value | | |
| Benzo[b]fluoranthene | | 1.52E+01 | 1.360E+09 | 1.12E-05 | 1.10E-04 | 2.17E-11 | 4% | No Toxicity Value | | |
| Dibenz[a,h]anthracene | | 2.06E+00 | 1.360E+09 | 1.51E-06 | 1.20E-03 | 3.20E-11 | 5% | No Toxicity Value | | |
| Indeno[1,2,3-cd]pyrene | | 6.80E+00 | 1.360E+09 | 5.00E-06 | 1.10E-04 | 9.69E-12 | 2% | No Toxicity Value | | |
| Benzofluoranthenes, total | | 1.91E+01 | 1.360E+09 | 1.40E-05 | 1.10E-04 | 2.72E-11 | 4% | No Toxicity Value | | |
| | | | | | | Cancer Risk | | | Hazard Index | |
| | | | | | Pathway Sums: | 6.2E-10 | | | 6.3E-05 | |

 Table A-2.3

 Risk and Hazard Estimates: Inhalation, Outdoor Maintenance Worker CTE, Landfill AOPC

 Table A-2.4

 Risk and Hazard Estimates: Dermal Contact with Soil, Outdoor Maintenance Worker CTE, Landfill AOPC

| Description | Variable | Value | 2 | Equations: | | | | | |
|--|----------------------------|---------------------|---------------------------|---------------------------|---|-----------------------------------|------------------------------|--|--|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specific | unitless | Cancer Risk: | | | | | |
| Soil-to-Skin Adherence Fraction, worker | AF_w | 0.02 | 2 mg/cm ² -day | | $\begin{bmatrix} & & & \\ & & & & \\ & & & & & \\ & & & & $ | | $\nabla FD \nabla FV $ | $(FC \vee SA \vee CF)$ | 1 |
| Averaging Time, carcinogens | AT _c | 70 |) yrs | Risk=S | $SF_d \times \{C_{uvil} \times \frac{\pi r_w}{2}\}$ | $\wedge ADS_d \wedge ET_1$ | $W^{ALD}_W^{ALV}_W^{ALV}$ | $\langle EC_{w} \times SA_{w} \times CF_{d} $ | } |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 6 | ō yrs | | a sou | BW_a | $\times AT_c \times 365 day$ | / year | |
| Body Weight, adult | BW_a | 80 |) kg | | · - | | | _/ | |
| Conversion Factor | CF _d | 1E-06 | 6 kg/mg | Noncancer Hazard: | | | | | |
| COPC Concentration in Soil | C _{soil} | chemical-specific | c mg/kg | | 1 [| | | $\nabla EV \nabla EC \nabla C$ | |
| Fraction of EV in Contact with Soil, worker | EC_w | 1 | unitless | Hazard | $l = \frac{1}{1} \times \{C_{uuil} \times$ | $\frac{A\Gamma_{W} \times AD}{M}$ | $S_d \times EF_w \times ED$ | $\frac{EV_{w} \times EV_{w} \times EC_{w} \times SA}{\times 365 day / year}$ | $\frac{\mathbf{I}_{w} \times \mathbf{C} \mathbf{\Gamma}_{d}}{ }$ |
| Exposure Duration, worker | ED_w | 6 | 5 yrs | | RfD_d sould | | $BW_a \times AT_{nc,a}$ | ×365 day/ year | |
| Exposure Frequency, worker | EF_{w} | 225 | 5 days/yr | | C | - | | | _) |
| Event Frequency, worker | EV_{w} | 1 | events/day | | | | | | |
| Oral Reference Dose Adjusted for GI Absorption | RfD_d | chemical-specific | c mg/kg-day | | | | | | |
| Exposed Body Surface Area, worker | SA_w | 3300 |) cm^2 | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF _d | chemical-specific | c (mg/kg-day) | 1 | | | | | |
| | | | | | Cancer | | | Hazard | |
| | | \mathbf{C}_{soil} | ABS _d | SFd | Risk | % of | RfD _d | Ouotient | % of |
| Analyte | | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | | |
| Arsenic | | 1.05E+01 | 3.00E-02 | 1.50E+00 | 1.24E-08 | 2% | 3.00E-04 | 3.20E-04 | 67% |
| Chromium (III) | | 5.94E+02 | 1.00E-02 | | | | 1.95E-02 | 1.55E-04 | 33% |
| Lead | | 5.11E+02 | 1.00E-02 | | | | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | 01112102 | 11002 02 | | | | | | |
| Benzo[a]anthracene | | 9.96E+00 | 1.30E-01 | 7.30E-01 | 4.12E-08 | 5% | | | |
| Benzo[a]pyrene | | 1.12E+01 | 1.30E-01 | 7.30E+00 | 4.63E-07 | 60% | | | |
| Benzo[b]fluoranthene | | 1.52E+01 | 1.30E-01 | 7.30E-01 | 6.29E-08 | 8% | | | |
| Dibenz[a,h]anthracene | | 2.06E+00 | 1.30E-01 | 7.30E+00 | 8.52E-08 | 11% | | | |
| Indeno[1,2,3-cd]pyrene | | 6.80E+00 | 1.30E-01 | 7.30E-01 | 2.81E-08 | 4% | | | |
| Benzofluoranthenes, total | | 1.91E+01 | 1.30E-01 | 7.30E-01 | 7.90E-08 | 10% | | | |
| | | | | Pathway Sums: | <u>Cancer Risk</u> 7.7E-07 | | · | Hazard Index 4.8E-04 | |

 Table A-3.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Construction Worker RME, Landfill AOPC

| Definition | Variable | Value | | Equations | | | | |
|--|---------------------|-------------------|---------------------------|-----------------------|-----------------------|---|-----------------------------------|---------------|
| Averaging Time, carcinogens | AT _c | 70 | yrs | Cancer Risk: | | | | |
| Averaging Time, noncarcinogens, construction worker | AT _{nc,cw} | 1 | yrs | Г | (| | | 1 |
| Body Weight | BW_{cw} | | | $Risk = SF_o \times$ | $C_{soil} \times -$ | $\frac{RS_{cw} \times EF_{cw} \times ED}{BW_{cw} \times AT_{c} \times 3}$ | $V_{cw} \times FI_{cw} \times CF$ | <u>o</u> |
| Conversion Factor | CFo | | | L | (| $\bigcup_{cw} \times AI_c \times 505 \ uuy \ / \ yeur$ | | |
| COPC Concentration in Soil | C _{soil} | chemical-specific | mg/kg | | | | | |
| Exposure Duration, construction worker | ED_{cw} | 1 | yrs | Noncancer Hazard: | | | | |
| Exposure Frequency, construction worker | EF_{cw} | 250 | days/yr | | Г | | | |
| Fraction Contaminated Soil Ingested, construction worker | FI _{cw} | 1.0 | unitless | Hazard $=\frac{1}{1}$ | $-\times C_{soil} $ | $\times \left(\frac{IRS_{cw} \times EF_{cw}}{BW_{cw} \times AT_{m}}\right)$ | $\times ED_{cw} \times FI_{cw}$ | $\times CF_o$ |
| Soil Ingestion Rate, worker | IRS _{cw} | 330 | mg/day | RfD | 0 | $(BW_{cw} \times AT_{no})$ | $c_{,cw} \times 365 day / g$ | year J |
| Oral Reference Dose | RfD ₀ | chemical-specific | mg/kg-day | | | | | |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Noncancer | |
| | | C _{soil} | SFo | Risk | % of | RfDo | Hazard | % of |
| Analyte | | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Volatile Organic Compounds (VOCs) | | | | | | | | |
| Tetrachloroethene | | 1.27E+02 | 2.10E-03 | 1.08E-08 | < 1% | 6.00E-03 | 5.98E-02 | 100% |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]pyrene | | 1.08E+01 | 7.30E+00 | 3.18E-06 | 100% | No Toxicity Value | | |
| | | | | Cancer Risk | | | Hazard Index | |
| | | | Pathway Sum | : 3.2E-06 | | | 6.0E-02 | |

 Table A-3.2

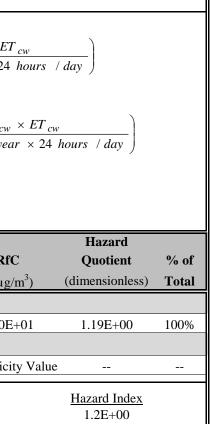
 Derivation of Inhalation Factors for Inhalation Exposures, Construction Worker RME, Landfill AOPC

| Description | Variable | Value Units | VF Derivation | | | | | | | | |
|---|------------------|---|---|----------------------------------|---|--------------------------|--|----------------------|------------|--------------------|--------------------|
| Unit conversion factor | CF _i | $1.00E-04 m^2/cm^2$ | | (| \mathbf{N} / 2 | | | | | | |
| Apparent Diffusivity | D_A | derived cm ² /s | VFs = (O / C) |) $\times \frac{(3.14)}{(3.14)}$ | $\frac{D_A \times T_{[receptor]}}{(2 \times \rho_{+} \times D_{+})} \times 0$ | CF : | | | | | |
| Diffusivity in air | Di | chemical-specific cm ² /s | | (| $(2 \times \rho_b \times D_A)$ | 1 | | | | | |
| Diffusivity in water | D_w | chemical-specific cm ² /s | | | | | | | | | |
| Default organic-carbon content | F _{OC} | 0.01 g/g | | | | | | | | | |
| Henry's law constant | H' | chemical-specific dimensionless | | | | | | | | | |
| Soil-water partition coefficient | K _d | chemical-specific cm ³ /g | | | | | | | | | |
| for organics: Kd= K_{OC} × F_{OC} | | | based on: | | | | | | | | |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specific cm ³ /g | $\left[\left(\Theta^{10} \right)^{3} D \right] H$ | $(+ \Theta^{10}/3D)$ | (n^2) | | | | | | |
| Total soil porosity | n | 0.43 L _{porespace} /L _{soil} | $DA = \frac{\left[\left(\Theta_{a}^{10} / 3 D_{i} H\right)\right]}{\rho_{B} K d}$ | $+0_w D_w$ | | | | | | | |
| Particulate Emission Factor (non-VOCs), default | PEF | $1.000E+06 \text{ m}^3/\text{kg}$ | $\rho_{B} \kappa a$ | $+ \Theta_w + \Theta_a \Pi$ | | | | | | | |
| Inverse of the mean concentration at the center | Q/C | 73.44 g/m ² -s per kg/m ³ | | | | | | | | | |
| of a 0.5-acre square source | | | | | | | | | | | |
| Dry soil bulk density | ρ_b | 1.5 g/cm^3 | | | | | | | | | |
| Air-filled soil porosity | θ_{a} | $0.28 L_{air}/L_{soil}$ | | | | | | | | | |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0.15 L _{water} /L _{soil} | | | | | | | | | |
| Exposure interval, 1-yr construction worker | T_{cw} | 3.15E+07 seconds | | | | | | | | | |
| Volatilization Factor for soil | VFs | derived m ³ /kg | | | | | | | | | |
| | | | | | | | | | D | VFs (VOCs) | PEF |
| | | Di | D _w | | H' | 2 | K _{oc} | K _d | DA | Construction | `` ´ |
| Analyte | | (cm ² /s) Reference | (cm ² /s) Reference | (dimensionless) | Reference | (cm^3/g) | Reference | (cm ³ /g) | (cm^2/s) | m ³ /kg | m ³ /kg |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | |
| Tetrachloroethene | | 0.0504664 RSLs ("WATER9"; USEPA, 2012 | 2) 9.46E-06 RSLs ("WATER9"; USEPA, 2012) | 0.7238 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 94.94 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 0.9494 | 0.001597 | 609.6009273 | use VF |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | |
| Benzo[a]pyrene | | 0.0475831 RSLs ("WATER9"; USEPA, 2012 | 2) 5.56E-06 RSLs ("WATER9"; USEPA, 2012) | 1.87E-05 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 587400 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 5874 | 1.4E-11 | non-VOC | 1.000E+06 |

| Description | Variable | Val | ue | | Equations | | | |
|---|---------------------|-------------------|------------------------|-----------------------------|----------------------------|--|---------------------|---|
| Averaging Time, carcinogens | AT _c | , | 70 yrs | Cancer Risk: | | | | |
| Averaging Time, noncarcinogens, construction worker | AT _{nc,cw} | | 1 yrs | | | (E | D × | $EF \times ET$ |
| COPC Concentration in air | C _{air} | chemical-specif | fic µg/m ³ | Risk = | $IUR \times C_{ai}$ | $r \times \left(\frac{E}{AT_c \times 365}\right)$ | $\frac{davs}{davs}$ | $\frac{21}{cw} \times 21$ vear $\times 24$ |
| COPC Concentration in soil | C _{soil} | chemical-specif | fic mg/kg | | | $(\prod_{c} \times 505)$ | uuys 1 | <i>year</i> ~ 21 |
| Exposure Duration, construction worker | ED_{cw} | | 1 yrs | Noncancer Hazard: | | | | |
| Exposure Frequency, construction worker | EF_{cw} | 2: | 50 days/yr | | 1 | (| FD | $\vee FF$ |
| Exposure Time, construction worker | ET_{cw} | | 8 hours/day | Hazara | $l = \frac{1}{PfC} \times$ | $C_{air} \times \left(\frac{1}{AT_{nc,cr}}\right)$ | <u> </u> | $\frac{c_W}{days} / \frac{c_W}{vaa}$ |
| Inhalation Unit-Risk Factor | IUR | chemical-specif | fic $(\mu g/m^3)^{-1}$ | | ŊĊ | (AI nc, c) | _N × 303 | uuys / yeu |
| Particulate emission factor (non-VOCs) | PEF | 1.00E+ | 06 m ³ /kg | | | | | |
| Inhalation Reference Concentration | RfC | chemical-specif | fic µg/m ³ | where: | C | C _{soil} | $\rho \mu g$ | |
| Volatilization Factor for VOCs | VF | chemical-specif | fic m ³ /kg | | $C_{air} = \frac{1}{V}$ | $\frac{C_{soil}}{VF \text{ or } PEF} \times 100$ | $\frac{10}{mg}$ | |
| | | | VF (VOCs) or | | | Cancer | | |
| | | C _{soil} | PEF (non-VOCs) | $\mathbf{C}_{\mathbf{air}}$ | IUR | Risk | % of | RfC |
| Analyte | | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | (µg/r |
| Volatile Organic Compounds (VOCs) | | 1 | | | | | | |
| Tetrachloroethene | | 1.27E+02 | 6.10E+02 | 2.08E+02 | 2.60E-07 | 1.77E-07 | 82% | 4.00E+ |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]pyrene | | 1.08E+01 | 1.00E+06 | 1.08E-02 | 1.10E-03 | 3.87E-08 | 18% | No Toxicity |
| | | | | Pa | thway Sums | <u>Cancer Risk</u> : 2.2E-07 | | - |

 Table A-3.3

 Risk and Hazard Estimates: Inhalation, Construction Worker, Landfill AOPC



| Description | Variable | Val | ue | Equations: | | | | | |
|--|---------------------|-------------------|-------------------------------|----------------------------|--|-------------------------|--|--|---------|
| Dermal Soil Absorption Fraction | ABS_d | chemical-specif | fic unitless | Cancer Risk: | | | | | |
| Soil-to-Skin Adherence Fraction, construction worker | AF_{cw} | 0 | .3 mg/cm ² -day | | $(AF_{cw} \times ABS)$ | $_{d} \times EF_{cw} >$ | $\times ED_{cw} \times EV_{cw} \times$ | $EC_{cw} \times SA_{cw} \times CF_d$ |)] |
| Averaging Time, carcinogens | AT_{c} | , | 70 yrs | $Risk = SF_d \times$ | $C_{soil} \times \left(\frac{AF_{cw} \times ABS}{S}\right)$ | $BW_{cw} \times$ | $AT_c \times 365 \ day /$ | year | J |
| Averaging Time, noncarcinogens, construction worker | AT _{nc,cw} | | 1 yrs | - | | | | | - |
| Body Weight, adult | BW_{cw} | 5 | 80 kg | | | | | | |
| Conversion Factor | CF_d | 1E-0 | 06 kg/mg | | | | | | |
| COPC Concentration in Soil | C _{soil} | chemical-specif | fic mg/kg | Noncancer Hazard: | | | | | |
| Fraction of EV in Contact with Soil, construction worker | EC_{cw} | | 1 unitless | | $\int AF_{CW} \times AF_{CW}$ | ABS $_d \times EF$ | $C_{cw} \times ED_{cw} \times EV_{c}$ | $w \times EC_{cw} \times SA_{cw} \times C$ | $[F_d]$ |
| Exposure Duration, construction worker | ED_{cw} | | 1 yrs | $Hazard = \frac{1}{RfD_d}$ | $-\times C_{soil} \times \left(\frac{AF_{cw} \times AF_{cw} \times AF_$ | BW _{cw} | $\times AT_{nc,cw} \times 365$ | day / year | |
| Exposure Frequency, construction worker | EF_{cw} | 23 | 50 days/yr | | L | | | | L . |
| Event Frequency, construction worker | EV_{cw} | | 1 events/day | | | | | | |
| Oral Reference Dose Adjusted for GI Absorption | RfD _d | chemical-specif | fic mg/kg-day | | | | | | |
| Exposed Body Surface Area, construction worker | SA_{cw} | 352 | 27 cm^2 | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specif | fic (mg/kg-day) ⁻¹ | | | | | | |
| | | | | | Cancer | | | Hazard | |
| | | C _{soil} | ABS _d | $\mathbf{SF}_{\mathbf{d}}$ | Risk | % of | RfD _d | Quotient | % of |
| Analyte | | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Volatile Organic Compounds (VOCs) | | 1.00E+00 | | | | | | | |
| Tetrachloroethene | | 1.27E+02 | 0.00E+00 | 2.10E-03 | 0.00E+00 | < 1% | 6.00E-03 | 0.00E+00 | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | |
| Benzo[a]pyrene | | 1.08E+01 | 1.30E-01 | 7.30E+00 | 1.33E-06 | 100% | | | |
| | | | | - | Cancer Risk | | • | Hazard Index | |
| | | | | Pathway Sums: | 1.3E-06 | | | <0.1 | |

Table A-3.4 Risk and Hazard Estimates: Dermal Contact with Soil, Construction Worker RME, Landfill AOPC

 Table A-4.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Construction Worker CTE, Landfill AOPC

| Variable | Value | | Equations | | | | |
|---|---|---|--|---|--|--|---|
| AT _c | 70 | yrs | Cancer Risk: | | | | |
| AT _{nc,cw} | 1 | yrs | | Г | | | 1 |
| BW_{cw} | 80 | kg | $Risk = SF_o \times C_{soil} \times $ | | $\frac{IKS_{cw} \times EF_{cw} \times ED}{DW}$ | $\frac{cw}{cw} \times FI_{cw} \times CF_{c}$ | <u>></u> |
| CFo | 1E-06 | kg/mg | | L | $(BW_{cw} \times AI_c \times 3)$ | 65 day / year | 刀 |
| C _{soil} | chemical-specific | mg/kg | | | | | |
| ED_{cw} | 1 | 1 yrs N | | : | | | |
| EF_{cw} | 250 | days/yr | | . Г | | | |
| FI _{cw} | 1.0 | unitless | Hazard $=$ $-$ | $\frac{1}{C} \times C_{1}$ | soil × $\frac{IKS_{cw} \times EF_{cw}}{DW}$ | $\times ED_{cw} \times FI_{cw} >$ | $\frac{\langle CF_0}{\langle CF_0}$ |
| uction worker FI_{cw} 1.0 unitlessHazard $=$ IRS _{cw} 100 mg/day | | | $(BW_{cw} \times AT_{no})$ | $_{r,cw}$ × 365 day / y | vear [] | | |
| RfD ₀ | chemical-specific | mg/kg-day | | | | | |
| SF_{O} | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | Cancer | | | Noncancer | |
| | C _{soil} | SFo | Risk | % of | RfDo | Hazard | % of |
| | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| | | | | | | | |
| | 1.27E+02 | 2.10E-03 | 1.63E-09 | < 1% | 6.00E-03 | 1.81E-02 | 100% |
| | | | | | | | |
| | 1.08E+01 | 7.30E+00 | 4.82E-07 | 100% | No Toxicity Value | | |
| | | Pathway Sum | Cancer Risk 4.8E-07 | | | Hazard Index 1.8E-02 | |
| | $\begin{array}{c} AT_c\\ AT_{nc,cw}\\ BW_{cw}\\ CF_o\\ C_{soil}\\ ED_{cw}\\ EF_{cw}\\ FI_{cw}\\ IRS_{cw}\\ RfD_0 \end{array}$ | $\begin{array}{cccc} AT_{c} & 70 \\ AT_{nc,cw} & 1 \\ BW_{cw} & 80 \\ CF_{o} & 1E-06 \\ C_{soil} & chemical-specific \\ ED_{cw} & 1 \\ EF_{cw} & 250 \\ FI_{cw} & 1.0 \\ IRS_{cw} & 100 \\ RfD_{O} & chemical-specific \\ SF_{O} & chemical-specific \\ SF_{O} & chemical-specific \\ \hline \\ $ | $\begin{array}{ccccc} AT_{c} & 70 \ yrs \\ AT_{nc,cw} & 1 \ yrs \\ BW_{cw} & 80 \ kg \\ CF_{o} & 1E-06 \ kg/mg \\ C_{soil} & chemical-specific \ mg/kg \\ ED_{cw} & 1 \ yrs \\ EF_{cw} & 250 \ days/yr \\ FI_{cw} & 1.0 \ unitless \\ IRS_{cw} & 100 \ mg/day \\ RfD_{O} & chemical-specific \ mg/kg-day \\ SF_{O} & chemical-specific \ mg/kg-day \\ SF_{O} & chemical-specific \ mg/kg-day \\ \end{array}$ | $\begin{array}{c cccc} AT_{c} & 70 \ yrs & Cancer Risk: \\ AT_{nc,cw} & 1 \ yrs \\ BW_{cw} & 80 \ kg & Risk = SF_{o} \times \\ CF_{o} & 1E-06 \ kg/mg \\ C_{soil} & chemical-specific \ mg/kg \\ ED_{cw} & 1 \ yrs & Noncancer Hazard \\ EF_{cw} & 250 \ days/yr \\ FI_{cw} & 1.0 \ unitless & Hazard & = \\ RfD_{o} & chemical-specific \ mg/kg-day \\ RfD_{o} & chemical-specific \ mg/kg-day \\ SF_{o} & chemical-specific \ (mg/kg-day)^{-1} & Cancer \\ \hline & C_{soil} & SF_{o} & Risk \\ (mg/kg) & (mg/kg-day)^{-1} & (dimensionless) \\ \hline & & 1.27E+02 & 2.10E-03 & 1.63E-09 \\ \hline & & 1.08E+01 & 7.30E+00 & 4.82E-07 \\ \hline \end{array}$ | $\begin{array}{c cccc} AT_{c} & 70 \ yrs & Cancer Risk: \\ AT_{nc,cw} & 1 \ yrs \\ BW_{cw} & 80 \ kg & Risk = SF_{o} \times \left[C_{soil} \times CF_{o} & 1E-06 \ kg/mg \\ C_{soil} & chemical-specific \ mg/kg \\ ED_{cw} & 1 \ yrs & Noncancer Hazard: \\ EF_{cw} & 250 \ days/yr \\ FI_{cw} & 1.0 \ unitless & Hazard & = \frac{1}{R/D_{o}} \times \left[C_{soil} \times 100 \ mg/day \\ RfD_{O} & chemical-specific \ mg/kg-day \\ SF_{O} & chemical-specific \ (mg/kg-day)^{-1} & (dimensionless) \\ \hline \end{array} \right] \times \left[C_{soil} \times Cancer \\ C_{soil} & SF_{o} & Risk & \% & of \\ (mg/kg) & (mg/kg-day)^{-1} & (dimensionless) \\ \hline \end{array} \right] + \frac{1.27E+02}{1.08E+01} & 2.10E-03 & 1.63E-09 & <1\% \\ \hline \end{array}$ | $\begin{array}{c cccc} AT_{c} & 70 \ yrs & Cancer Risk: \\ AT_{nc.cw} & 1 \ yrs \\ BW_{cw} & 80 \ kg & Risk = SF_{o} \times \left[C_{soil} \times \left(\frac{IRS_{cw} \times EF_{cw} \times ED}{BW_{cw} \times AT_{c} \times 3}\right)\right] \\ CF_{o} & 1E-06 \ kg/mg & Risk = SF_{o} \times \left[C_{soil} \times \left(\frac{IRS_{cw} \times EF_{cw} \times ED}{BW_{cw} \times AT_{c} \times 3}\right)\right] \\ C_{soil} & chemical-specific \ mg/kg & Risk = SF_{o} \times \left[C_{soil} \times \left(\frac{IRS_{cw} \times EF_{cw} \times EF_{cw} \times EF_{cw} \times AT_{c} \times 3}{BW_{cw} \times AT_{c} \times 3}\right)\right] \\ FI_{cw} & 1 \ yrs & Noncancer \ Hazard & = \frac{1}{RfD_{o}} \times \left[C_{soil} \times \left(\frac{IRS_{cw} \times EF_{cw} \times EF_{cw} \times AT_{c} \times 3}{BW_{cw} \times AT_{c} \times 3}\right)\right] \\ RS_{cw} & 100 \ mg/day & RfD_{o} \ chemical-specific \ mg/kg-day \ SF_{o} & chemical-specific \ (mg/kg-day)^{-1} \ (dimensionless) & Total \ (mg/kg-day) & (mg/kg-day) \ \hline \\ \hline \\ \hline \\ 1.27E+02 & 2.10E-03 \ 1.63E-09 \ < 1\% \ 6.00E-03 \ \hline \\ 1.08E+01 & 7.30E+00 \ 4.82E-07 \ 100\% \ No \ Toxicity \ Value \ \hline \\ \hline$ | $\begin{array}{c cccc} AT_{c} & 70 \ yrs & Cancer Risk: \\ AT_{nc,cw} & 1 \ yrs & \\ BW_{cw} & 80 \ kg & Risk = SF_{o} \times \left[C_{soil} \times \left(\frac{IRS_{cw} \times EF_{cw} \times ED_{cw} \times FI_{cw} \times CF_{c}}{BW_{cw} \times AT_{c} \times 365 \ day \ / \ year} \right) \\ CF_{o} & IE-06 \ kg/mg & \\ CF_{o} & IE-06 \ kg/mg & \\ Cs_{oil} & chemical-specific \ mg/kg & \\ ED_{cw} & 1 \ yrs & Noncancer Hazard: \\ EF_{cw} & 250 \ days/yr & \\ FI_{cw} & 1.0 \ unitless & Hazard \ = \frac{1}{RfD_{o}} \times \left[C_{soil} \times \left(\frac{IRS_{cw} \times EF_{cw} \times ED_{cw} \times FI_{cw} \times FI_{cw}$ |

 Table A-4.2

 Derivation of Inhalation Factors for Inhalation Exposures, Construction Worker CTE, Landfill AOPC

| Description | Variable | Value | Units | VFD | Derivation | | | | | | | | |
|---|------------------|-----------------------------------|---------------------------------------|--------------|---|----------------------------|---|------------|--|----------------------|------------|--------------------|--------------------|
| Unit conversion factor | CFi | 1.00E-04 m ² /c | cm ² | | | 6 | | | | | | | |
| Apparent Diffusivity | D _A | derived cm ² | /s | | VFs = (O / C) | × <u>(3.14</u> × | $\frac{D_A \times T_{[receptor]}}{2 \times \rho + \times D_A} \times 0$ | CF , | | | | | |
| Diffusivity in air | Di | chemical-specific cm ² | /s | | (2) | (| $2 \times \rho_b \times D_A$) | 1 | | | | | |
| Diffusivity in water | D_w | chemical-specific cm ² | /s | | | | | | | | | | |
| Default organic-carbon content | F _{oc} | 0.01 g/g | | | | | | | | | | | |
| Henry's law constant | H' | chemical-specific dim | ensionless | | | | | | | | | | |
| Soil-water partition coefficient | K _d | chemical-specific cm ³ | /g | | | | | | | | | | |
| for organics: Kd=K _{OC} ×F _{OC} | | | | based | l on: | | | | | | | | |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specific cm ³ | /g | | $\Theta^{10}/{}^3D$ | $'+ \Theta \frac{10}{3} D$ | (n^{2}) | | | | | | |
| Total soil porosity | n | 0.43 L _{por} | espace/L _{soil} | | $DA = \frac{\left[\left(\Theta \frac{10}{a} / 3 D_{i} H \right) \right]}{\rho_{i} K d_{i} + 1}$ | $\Theta + \Theta H$ | | | | | | | |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.000E+06 m ³ /l | kg | | r B | - w - a | | | | | | | |
| Inverse of the mean concentration at the center | Q/C | 73.44 g/m | ² -s per kg/m ³ | | | | | | | | | | |
| of a 0.5-acre square source | | C | | | | | | | | | | | |
| Dry soil bulk density | ρ_b | 1.5 g/cr | n ³ | | | | | | | | | | |
| Air-filled soil porosity | θ_{a} | 0.28 L _{air} / | L _{soil} | | | | | | | | | | |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0.15 L _{wat} | er/L _{soil} | | | | | | | | | | |
| Exposure interval, 0.5-yr construction worker | T _{cw} | 1.58E+07 seco | onds | | | | | | | | | | |
| Volatilization Factor for soil | VFs | derived m ³ /l | kg | | | | | | | | | | |
| | | | - | | | | | | | | | VFs (VOCs) (a) | PEF |
| | | 2 | D _i | 2 | D _w | | Н' | 2 | K _{oc} | K _d | | Construction | (non-VOCs) |
| Analyte | | (cm ² /s) | Reference | (cm^2/s) | Reference | (dimensionless) | Reference | (cm^3/g) | Reference | (cm ³ /g) | (cm^2/s) | m ³ /kg | m ³ /kg |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | | | |
| Tetrachloroethene | | 0.0504664 RSI | Ls ("WATER9"; USEPA, 2012) | 9.46E-06 RSL | ("WATER9"; USEPA, 2012) | 0.724 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 94.94 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 0 | 0.00805 | | use VF |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | |
| Benzo[a]pyrene | | 0.0475831 RSI | Ls ("WATER9"; USEPA, 2012) | 5.56E-06 RSL | ("WATER9"; USEPA, 2012) | 1.87E-05 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 587400 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 0 | 8.2E-07 | non-VOC | 1.000E+06 |
| | | 0.0475831 RSI | Ls ("WATER9"; USEPA, 2012) | 5.56E-06 RSL | s ("WATER9"; USEPA, 2012) | 1.87E-05 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 587400 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 0 | 8.2E-07 | non-VOC | _ |

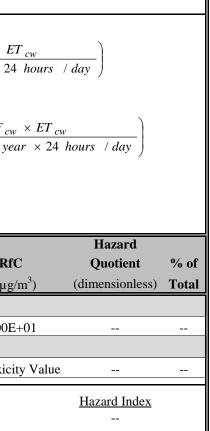
Notes

(a) The VF for this receptor pathway is not calculated.

| Description | Variable | Val | ue | | Equations | | | |
|---|-------------------|-------------------|------------------------|-------------------|----------------------------|---|-----------------|---|
| Averaging Time, carcinogens | AT_{c} | , | 70 yrs | Cancer Risk: | | | | |
| Averaging Time, noncarcinogens, construction worker | $AT_{nc,cw}$ | 0 | 0.5 yrs | | | (| ED × | $EF \rightarrow E7$ |
| COPC Concentration in air | C _{air} | chemical-specif | fic µg/m ³ | Risk = | = $IUR \times C_a$ | $_{iir} \times \left(\frac{1}{AT_{c} \times 36}\right)$ | $\frac{22}{cw}$ | $\frac{1}{\sqrt{vear}} \times 24$ |
| COPC Concentration in soil | C _{soil} | chemical-specif | fic mg/kg | | | (11 c) | e aays | , , , |
| Exposure Duration, construction worker | ED_{cw} | 0 |).5 yrs | Noncancer Hazard: | | | | |
| Exposure Frequency, construction worker | EF_{cw} | 2: | 50 days/yr | | 1 | (| FΓ | $\rightarrow FF$ |
| Exposure Time, construction worker | ET_{cw} | | 8 hours/day | Hazard | $l = \frac{1}{RfC} \times$ | $\propto C_{air} \times \left(\frac{1}{AT_{nc}}\right)$ | × 36 | $\frac{\sigma_{cw} \wedge Er_{cw}}{5 days / yac}$ |
| Inhalation Unit-Risk Factor | IUR | chemical-specif | fic $(\mu g/m^3)^{-1}$ | | ŊĊ | (Πnc) | $c_W \wedge 50$ | s uuys 7 yeu |
| Particulate Emission Factor (non-VOCs) | PEF | 1.00E+0 | 06 m ³ /kg | | | | | |
| Inhalation Reference Concentration | RfC | chemical-specif | fic µg/m ³ | where: | C | $\frac{C_{soil}}{VF \text{ or } PEF} \times 10$ | $\rho \mu g$ | |
| Volatilization Factor for VOCs | VF | chemical-specif | fic m ³ /kg | | $C_{air} = -V$ | $\overline{F} \text{ or } PEF \times 10$ | $\frac{00}{mg}$ | |
| | | | VF (VOCs) or | | | Cancer | | |
| | | C _{soil} | PEF (non-VOCs) | C _{air} | IUR | Risk | % of | RfC |
| Analyte | | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | (µg/1 |
| Volatile Organic Compounds (VOCs) | | 1 | | | | | | |
| Tetrachloroethene | | 1.27E+02 | | | 2.60E-07 | | | 4.00E- |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]pyrene | | 1.08E+01 | 1.00E+06 | 1.08E-02 | 1.10E-03 | 1.94E-08 | 100% | No Toxicit |
| | | | | | thway Sums | Cancer Risk | | |

 Table A-4.3

 Risk and Hazard Estimates: Inhalation, Construction Worker CTE, Landfill AOPC



Value Description Variable Equations: Dermal Soil Absorption Fraction ABS_d chemical-specific unitless Cancer Risk: $Risk = SF_{d} \times \left[C_{soil} \times \left(\frac{AF_{cw} \times ABS_{d} \times EF_{cw} \times ED_{cw} \times EV_{cw} \times EC_{cw} \times SA_{cw} \times CF_{d}}{BW_{cw} \times AT_{c} \times 365 \ day \ / \ year} \right) \right]$ Soil-to-Skin Adherence Fraction, construction worker AF_{cw} 0.1 mg/cm^2 -day Averaging Time, carcinogens AT_c 70 yrs Averaging Time, noncarcinogens, construction worker AT_{nc.cw} 1 yrs Body Weight, adult BW_{cw} 80 kg Conversion Factor CF_d 1E-06 kg/mg COPC Concentration in Soil chemical-specific mg/kg C_{soil} Noncancer Hazard: Fraction of EV in Contact with Soil, construction worker EC_{cw} 1 unitless $Hazard = \frac{1}{RfD_d} \times \left[C_{soil} \times \left(\frac{AF_{cw} \times ABS_d \times EF_{cw} \times ED_{cw} \times EV_{cw} \times EC_{cw} \times SA_{cw} \times CF_d}{BW_{cw} \times AT_{nc,cw} \times 365 \ day / year} \right) \right]$ Exposure Duration, construction worker ED_{cw} 1 yrs Exposure Frequency, construction worker EF_{cw} 250 days/yr Event Frequency, construction worker EV_{cw} 1 events/day Oral Reference Dose Adjusted for GI Absorption chemical-specific mg/kg-day RfD_d 3300 cm^2 Exposed Body Surface Area, construction worker SA_{cw} chemical-specific (mg/kg-day)⁻¹ Oral Slope Factor Adjusted for GI Absorption SF_d Cancer Hazard C_{soil} ABS_d SF_d Risk % of **RfD**_d Quotient % of Analyte (dimensionless) (mg/kg-day) (dimensionless) Total (mg/kg) (unitless) (mg/kg-day)⁻¹ Total Volatile Organic Compounds (VOCs) Tetrachloroethene 1.27E+02 0.00E+002.10E-03 0.00E+00< 1%6.00E-03 0.00E+00--Polycyclic Aromatic Hydrocarbons (PAHs) Benzo[a]pyrene 1.08E+01 1.30E-01 7.30E+00 2.07E-07 100% ----___ Cancer Risk Hazard Index

Pathway Sums:

2.1E-07

 Table A-4.4

 Risk and Hazard Estimates: Dermal Contact with Soil, Construction Worker CTE, Landfill AOPC

0.00E+00

| | | Table A-5. | NI | | | | |
|--|--------------------|-------------------|--------------------|------------------------------------|---------------------|--|--|
| Nursing Infant: Hypothetics | al Fishing Plat | tform User: Cur | rent Scenario : RI | ME Summary - Landf | ill AOPC | | |
| Definition | Variable | Value | Source | Equation | ons | | |
| Infant Risk Adjustment Factor | IRAF | Chemical Specific | ODEQ 2010 | | | | |
| Carcinogenic IRAFc | | | | | | | |
| Total PCB | IRAFc_pcb | 0.6 | ODEQ 2010 | Infant Cancer Risk = Mother Risk x | | | |
| Noncancer IRAFnc | | | | | | | |
| Total PCB | IRAFnc_pcb | 4 | ODEQ 2010 | Infant Noncancer Hazard | = Child HQ x IRAFnc | | |
| | | | | | | | |
| | | other | Infa | ant | | | |
| | C _{soilt} | Cancer Risk | Hazard Quotient | Cancer Risk | Hazard Quotient | | |
| Analyte | (mg/kg) | (adult) | (child) | | | | |
| Total PCBs as Congeners (KM-based, capped) | 4E-01 | 1.7E-06 | 3.6E-01 | 1.0E-06 | 1.5 | | |

| Definition | Variable | Valu | 1e | Equations | Val | ue | |
|--|---|---|--|---|---|--|--|
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 | yrs | Cancer Risk: | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 | yrs | Nonmutagens: | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 | yrs |] | | $(IRS_a \times I$ | ED_{a} |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 | yrs | $Risk = SF_o \times \left C_{soil} \times \right $ | $EF_r \times FI \times C$ | $CF_0 \times \left(\frac{u}{DUU} \right)$ | $\frac{a}{a + a + b}$ |
| Averaging Time, carcinogens | AT _c | 70 | yrs | | | $\left\langle BW_a \times AT_c \times \right\rangle$ | $\frac{365}{year}$ E |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 20 | yrs | Mutagens: | | , | |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | 6 | yrs | $Risk = SF_0 \times C_{soil} >$ | $\langle EE_{*} \times FI \times$ | CF_{α} | |
| Body Weight | BWa | 80 | kg | | | | |
| Body Weight | BW _c | 15 | kg | $\times \left(-\frac{1}{2} \right)$ | $RS_c \times ED_{0-2}$ | $\frac{2 \times ADAF_{0-2}}{65 days/year} + \frac{IR}{BW_0}$ | $RS_c \times ED_{2-6} \times ED_{2-6}$ |
| Conversion Factor | CFo | 1E-06 | • | | | | |
| COPC Concentration in soil | C _{soil} | chemical-specific | | $+\frac{IR}{IR}$ | $S_a \times ED_{6-16}$ | $\frac{\times ADAF_{6-16}}{5 \text{ days/year}} + \frac{IRS_a}{BW_a}$ | $\times ED_{16-26} \times$ |
| Exposure Duration, child 0-2 | ED ₀₋₂ | - | yrs | ' BW | $T_a \times AT_c \times 36$ | 5 days/year ' BW _a | $\times AT_c \times 365$ |
| Exposure Duration, child 2-6 | ED ₀₋₂ ED ₂₋₆ | | yrs | Vinyl Chloride: | | | ר איי איי |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 | yrs | Vinyl Chloride: $Risk = SF_0 \times \begin{cases} C_{sol} \\ \\ \end{cases}$ | _ | $EF_r \times \left(\frac{ED_c \times IR}{BW_c}\right)$ | $\frac{S_c}{RW_a} + \frac{ED_a \times H}{RW_a}$ |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 | yrs | $Risk = SF_0 \times \left\{ C_{sol} \right\}$ | $_{il} \times FI \times CF_0$ | $X = \frac{1}{12}$ | days |
| Exposure Duration, adult | ED_a | 20 | yrs | | | $AT_c \times 3$ | $\frac{165}{year}$ |
| Exposure Duration, child | ED _c | 6 | yrs | | | | |
| Exposure Frequency, resident | EF_r | 365 | days/yr | Noncancer Hazard: | | | |
| Fraction Contaminated Soil Ingested | FI | 1.0 | unitless | 1 Г | | | |
| - | | | | | $I I R S_{a}$ | $\times ED_{*} \times EE_{*} \times EI \times$ | (F_{a}) |
| Soil Ingestion Rate (adult) | IRS _a | | mg/day | $Hazard = \frac{1}{RfD} \times C$ | $_{soil} \times \left(\frac{IRS_c}{BW}\right)$ | $\frac{\times ED_c \times EF_r \times FI \times}{\times AT} \times \frac{365 day}{2}$ | $\left(\frac{CF_0}{Poar}\right)$ |
| Soil Ingestion Rate (child) | IRS _c | 200 | mg/day | $Hazard = \frac{1}{RfD_o} \times \left[C\right]$ | $soil \times \left(\frac{IRS_c}{BW_c}\right)$ | $\frac{\times ED_c \times EF_r \times FI \times}{\times AT_{nc,c} \times 365 day/y}$ | $\left[\frac{CF_o}{vear}\right]$ |
| Soil Ingestion Rate (child) Oral Reference Dose | IRS _c RfD _O | 200 chemical-specific | mg/day (mg/kg-day) | $Hazard = \frac{1}{RfD_o} \times \left[C\right]$ | $_{soil} \times \left(\frac{IRS_c}{BW_c}\right)$ | $\frac{\times ED_c \times EF_r \times FI \times}{\times AT_{nc,c} \times 365 \ day/y}$ | $\left[\frac{CF_o}{vear}\right]$ |
| Soil Ingestion Rate (child) | IRS _c | 200 | mg/day (mg/kg-day) | $Hazard = \frac{1}{RfD_o} \times \left[C\right]$ | $soil \times \left(\frac{IRS_c}{BW_c}\right)$ | $\frac{\times ED_c \times EF_r \times FI \times}{\times AT_{nc,c} \times 365 \ day/y}$ | $\left(\frac{CF_o}{vear}\right)$ |
| Soil Ingestion Rate (child) Oral Reference Dose | IRS _c RfD _O | 200 chemical-specific chemical-specific | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ | Cancer | (2112) | | Noncance |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor | IRS _c RfD ₀ SF ₀ | 200 chemical-specific chemical-specific C _{soil} | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o | Cancer Risk | % of | RfD _o | Noncance Hazard (Chi |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor | IRS _c RfD _O | 200 chemical-specific chemical-specific | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ | Cancer | (2112) | | Noncance |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor | IRS _c RfD ₀ SF ₀ | 200 chemical-specific chemical-specific C _{soil} | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o | Cancer Risk | % of | RfD _o | Noncance Hazard (Chi |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte | IRS _c RfD ₀ SF ₀ | 200 chemical-specific chemical-specific C _{soil} | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o | Cancer Risk | % of | RfD _o | Noncance Hazard (Chi |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents | IRS _c RfD _O SF _O Mutagen? | 200 chemical-specific chemical-specific C _{soil} (mg/kg) | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ | Cancer Risk (dimensionless) | % of Total | RfD _o (mg/kg-day) | Noncancer Hazard (Chi (dimensionle |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic | IRS _c RfD ₀ SF ₀ Mutagen? | 200 chemical-specific chemical-specific C _{soil} (mg/kg) 1.05E+01 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ 1.50E+00 | Cancer Risk (dimensionless) | % of Total | RfD _o (mg/kg-day) 3.00E-04 | Noncance Hazard (Chi (dimensionle 2.80E-01 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) | IRS _c RfD ₀ SF ₀ Mutagen? 0.00E+00 0.00E+00 | 200 chemical-specific chemical-specific (mg/kg) 1.05E+01 5.94E+02 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ 1.50E+00 No Toxicity Value | Cancer Risk (dimensionless) 1.42E-05 | % of Total 2% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 | Noncance Hazard (Chi (dimensionle 2.80E-01 5.28E-03 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) Lead | IRS _c RfD ₀ SF ₀ Mutagen? 0.00E+00 0.00E+00 0.00E+00 | 200 chemical-specific chemical-specific C _{soil} (mg/kg) 1.05E+01 5.94E+02 0.00E+00 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF_{o} $(mg/kg-day)^{-1}$ $1.50E+00$ No Toxicity Value 8.50E-03 | Cancer Risk (dimensionless) 1.42E-05 0.00E+00 | % of Total 2% < 1% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 No Toxicity Value | Noncancer Hazard (Chi (dimensionle 2.80E-01 5.28E-03 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) Lead Mercury | IRS _c RfD ₀ SF ₀ Mutagen? 0.00E+00 0.00E+00 0.00E+00 | 200 chemical-specific chemical-specific C _{soil} (mg/kg) 1.05E+01 5.94E+02 0.00E+00 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF_{o} $(mg/kg-day)^{-1}$ $1.50E+00$ No Toxicity Value 8.50E-03 | Cancer Risk (dimensionless) 1.42E-05 0.00E+00 | % of Total 2% < 1% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 No Toxicity Value | Noncancer Hazard (Chi (dimensionle 2.80E-01 5.28E-03 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) Lead Mercury Polychlorinated Biphenyls (PCBs) - Mixtures | IRS _c RfD ₀ SF ₀ Mutagen? 0.00E+00 0.00E+00 0.00E+00 0.00E+00 | 200 chemical-specific chemical-specific Csoil (mg/kg) 1.05E+01 5.94E+02 0.00E+00 1.45E+00 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ 1.50E+00 No Toxicity Value 8.50E-03 No Toxicity Value | Cancer Risk (dimensionless) 1.42E-05 0.00E+00 | % of <u>Total</u> 2% < 1% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 No Toxicity Value 1.60E-04 | Noncance Hazard (Chi (dimensionle 2.80E-01 5.28E-03 1.21E-01 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) Lead Mercury Polychlorinated Biphenyls (PCBs) - Mixtures Total PCBs | IRS _c RfD ₀ SF ₀ Mutagen? 0.00E+00 0.00E+00 0.00E+00 0.00E+00 | 200 chemical-specific chemical-specific Csoil (mg/kg) 1.05E+01 5.94E+02 0.00E+00 1.45E+00 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ 1.50E+00 No Toxicity Value 8.50E-03 No Toxicity Value | Cancer Risk (dimensionless) 1.42E-05 0.00E+00 | % of <u>Total</u> 2% < 1% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 No Toxicity Value 1.60E-04 | Noncance Hazard (Chi (dimensionle 2.80E-01 5.28E-03 1.21E-01 2.73E-01 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) Lead Mercury Polychlorinated Biphenyls (PCBs) - Mixtures Total PCBs Volatile Organic Compounds (VOCs) | IRS _c RfD ₀ SF ₀ Mutagen? 0.00E+00 0.00E+00 0.00E+00 0.00E+00 | 200 chemical-specific chemical-specific Csoil (mg/kg) 1.05E+01 5.94E+02 0.00E+00 1.45E+00 1.45E+00 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ 1.50E+00 No Toxicity Value 8.50E-03 No Toxicity Value 2.00E+00 | Cancer Risk (dimensionless) 1.42E-05 0.00E+00 1.23E-06 | % of Total 2% < 1% < 1% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 No Toxicity Value 1.60E-04 2.00E-05 | Noncance Hazard (Chi (dimensionle 2.80E-01 5.28E-03 1.21E-01 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) Lead Mercury Polychlorinated Biphenyls (PCBs) - Mixtures Total PCBs Volatile Organic Compounds (VOCs) Naphthalene | IRS _c RfD ₀ SF ₀ Mutagen? 0.00E+00 0.00E+00 0.00E+00 0.00E+00 | 200 chemical-specific chemical-specific Csoil (mg/kg) 1.05E+01 5.94E+02 0.00E+00 1.45E+00 1.45E+00 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ 1.50E+00 No Toxicity Value 8.50E-03 No Toxicity Value 2.00E+00 | Cancer Risk (dimensionless) 1.42E-05 0.00E+00 1.23E-06 | % of Total 2% < 1% < 1% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 No Toxicity Value 1.60E-04 2.00E-05 | Noncance Hazard (Chi (dimensionle 2.80E-01 5.28E-03 1.21E-01 2.73E-01 2.55E-03 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) Lead Mercury Polychlorinated Biphenyls (PCBs) - Mixtures Total PCBs Volatile Organic Compounds (VOCs) Naphthalene Pesticides | IRSc RfDo SFo Mutagen? 0.00E+00 | 200 chemical-specific chemical-specific chemical-specific chemical-specific (mg/kg) 1.05E+01 5.94E+02 0.00E+00 1.45E+00 1.45E+00 3.83E+00 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ 1.50E+00 No Toxicity Value 8.50E-03 No Toxicity Value 2.00E+00 No Toxicity Value | Cancer Risk (dimensionless) 1.42E-05 0.00E+00 1.23E-06 | % of <u>Total</u> <u>2%</u> <u></u> < 1% <u></u> < 1% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 No Toxicity Value 1.60E-04 2.00E-05 2.00E-02 | Noncance Hazard (Chi (dimensionle 2.80E-01 5.28E-03 1.21E-01 2.73E-01 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) Lead Mercury Polychlorinated Biphenyls (PCBs) - Mixtures Total PCBs Volatile Organic Compounds (VOCs) Naphthalene Pesticides MCPP | IRSc RfDo SFo Mutagen? 0.00E+00 | 200 chemical-specific chemical-specific chemical-specific chemical-specific (mg/kg) 1.05E+01 5.94E+02 0.00E+00 1.45E+00 1.45E+00 3.83E+00 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ 1.50E+00 No Toxicity Value 8.50E-03 No Toxicity Value 2.00E+00 No Toxicity Value | Cancer Risk (dimensionless) 1.42E-05 0.00E+00 1.23E-06 | % of <u>Total</u> <u>2%</u> <u></u> < 1% <u></u> < 1% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 No Toxicity Value 1.60E-04 2.00E-05 2.00E-02 | Noncance Hazard (Chi (dimensionle 2.80E-01 5.28E-03 1.21E-01 2.73E-01 2.55E-03 |
| Soil Ingestion Rate (child) Oral Reference Dose Oral Slope Factor Analyte Inorganic Constituents Arsenic Chromium (III) Lead Mercury Polychlorinated Biphenyls (PCBs) - Mixtures Total PCBs Volatile Organic Compounds (VOCs) Naphthalene Pesticides MCPP Polycyclic Aromatic Hydrocarbons (PAHs) | IRSc RfDo SFo Mutagen? 0.00E+00 0.00E+00 | 200 chemical-specific chemical-specific chemical-specific chemical-specific chemical-specific (mg/kg) 1.05E+01 1.05E+01 1.45E+00 1.45E+00 3.83E+00 1.12E+01 | mg/day (mg/kg-day) (mg/kg-day) ⁻¹ SF _o (mg/kg-day) ⁻¹ 1.50E+00 No Toxicity Value 8.50E-03 No Toxicity Value 2.00E+00 No Toxicity Value No Toxicity Value | Cancer Risk (dimensionless) 1.42E-05 0.00E+00 1.23E-06 | % of Total 2% <1% <1% | RfD _o (mg/kg-day) 3.00E-04 1.50E+00 No Toxicity Value 1.60E-04 2.00E-05 2.00E-02 1.00E-03 | Noncancer Hazard (Chi (dimensionle 2.80E-01 5.28E-03 1.21E-01 2.73E-01 2.55E-03 1.49E-01 |

Table A-5.1 Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: Current Scenario RME, Landfill AOPC

$$= \frac{IRS_c \times ED_c}{BW_c \times AT_c \times 365 \frac{day}{year}} \bigg) \bigg]$$

$$= \frac{5 \times ADAF_{2-6}}{65 \ days/year} \times ADAF_{16-26}} \bigg) \bigg]$$

$$= \frac{IRS_a}{55 \ days/year} + \frac{IRS_c}{BW_c} \bigg] \bigg\}$$

$$= \frac{IRS_a}{V_a} + \frac{IRS_c}{BW_c} \bigg] \bigg\}$$

$$= \frac{101 \ 34\%}{1.13E-02}$$

$$= \frac{101 \ 33\%}{2.56E-02}$$

$$= \frac{101 \ 33\%}{2.39E-04}$$

$$= \frac{101 \ 18\%}{1.40E-02}$$

| Definition | Variable | Val | ue | Equations | Valu | ıe | |
|--|-----------------------|-------------------|---------------------------|---|---|--|---------------------------------|
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 | yrs | Cancer Risk: | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 | yrs | Nonmutagens: | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 | yrs |] | | $(IRS_a \times$ | ED_{a} |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 | yrs | $Risk = SF_o \times C_{soil}$ | \times EF _r \times FI \times C | $F_0 \times \left(\frac{u}{1 + u} \right)$ | $\frac{a}{day} + \frac{b}{day}$ |
| Averaging Time, carcinogens | AT _c | 70 | yrs | | | $\left\langle BW_a \times AT_c \times \right\rangle$ | $\frac{365}{year}$ |
| Averaging Time, noncarcinogens, adult | $AT_{nc,a}$ | 20 | yrs | Mutagens: | | , | |
| Averaging Time, noncarcinogens, child | AT _{nc.c} | 6 | yrs | $Risk = SF_0 \times \left[C_{soil}\right]$ | $\times FF \times FI \times I$ | C F . | |
| Body Weight | BWa | 80 | kg | | | 010 | |
| Body Weight | BWc | 15 | kg | ×(- | $IRS_c \times ED_{0-2}$ | $\frac{11}{55 \text{ days/year}} + \frac{11}{BW}$ | $RS_c \times ED_{2-6}$ |
| Conversion Factor | CF _o | | kg/mg | | | | |
| COPC Concentration in soil | C _{soil} | chemical-specific | | + | $RS_a \times ED_{6-16}$ | $\frac{\times ADAF_{6-16}}{5 days/year} + \frac{IRS_a}{BW_a}$ | $L \times ED_{16-26} \times$ |
| Exposure Duration, child 0-2 | ED ₀₋₂ | - | yrs | B | $W_a \times AT_c \times 365$ | 5 days/year ' BW _a | $_{t} \times AT_{c} \times 365$ |
| Exposure Duration, child 2-6 | ED_{2-6} | | yrs | Vinyl Chloride: | | | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | | yrs | Vinyl Chloride: $Risk = SF_0 \times \begin{cases} C_s \end{cases}$ | | $EF_r \times \left(\frac{ED_c \times IR}{BW_c}\right)$ | $\frac{ED_a \times T}{BW_a}$ |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 | yrs | $Risk = SF_O \times \left\{ C_s \right\}$ | $_{soil} \times FI \times CF_0$ | × | $\frac{2m_a}{davs}$ |
| Exposure Duration, adult | ED_{a} | 20 | yrs | | | $AT_c \times 3$ | $365 \frac{day 2}{year}$ |
| Exposure Duration, child | ED_{c} | 6 | yrs | | | | - |
| Exposure Frequency, resident | EF_r | 365 | days/yr | Noncancer Hazard: | | | |
| Fraction Contaminated Soil Ingested | FI | 1.0 | unitless | 1 [| (IRS | ×FD ×FF ×FI× | (F) |
| Soil Ingestion Rate (adult) | IRS _a | 100 | mg/day | $Hazard = \frac{1}{RfD_{\star}} \times$ | $C_{soil} \times \left(\frac{IRS_c}{PM_c}\right)$ | $\frac{1}{\sqrt{T}} = \frac{1}{\sqrt{2}} \frac{1}{\sqrt{T}} + \frac{1}{\sqrt{2}} \frac{1}{\sqrt{T}}$ | $\frac{CT_0}{VOGT}$ |
| Soil Ingestion Rate (child) | IRS _c | | mg/day | $K_J D_0$ | (DW_c) | $\operatorname{AI}_{nc,c} \times \operatorname{SOS} \operatorname{uuy}_{1}$ | yeur /] |
| Oral Reference Dose | RfD ₀ | chemical-specific | | | | | |
| Oral Slope Factor | SFo | chemical-specific | $(mg/kg-day)^{-1}$ | | | | |
| | | | | Cancer | | | Noncance |
| | | C _{soil} | SFo | Risk | % of | RfD _o | Hazard (Ch |
| Analyte | Mutagen? | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionle |
| Benzo[g,h,i]perylene | 0.00E+00 | 5.82E+00 | No Toxicity Value | | | 3.00E-02 | 2.59E-03 |
| Benzo[k]fluoranthene | М | 2.63E+01 | 7.30E-02 | 1.31E-05 | 1% | No Toxicity Value | |
| Benzofluoranthenes, Total | М | 1.88E+01 | 7.30E-01 | 9.34E-05 | 10% | No Toxicity Value | |
| Dibenz[a,h]anthracene | М | 2.06E+00 | 7.30E+00 | 1.02E-04 | 11% | No Toxicity Value | |
| Indeno[1,2,3-cd]pyrene | М | 6.80E+00 | 7.30E-01 | 3.38E-05 | 4% | No Toxicity Value | |
| | | | | Cancer Risk | | | Hazard Index (|
| | | | Pathway Sum | 9.4E-04 | | Pathway Sum: | 0.83 |

Table A-5.1 Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: Current Scenario RME, Landfill AOPC

$$-\frac{IRS_{c} \times ED_{c}}{BW_{c} \times AT_{c} \times 365 \frac{day}{year}})$$

$$= \frac{3 \times ADAF_{2-6}}{55 \ days/year} \times ADAF_{16-26}} = \frac{1RS_{a}}{5 \ days/year} + \frac{1RS_{c}}{BW_{c}} = \frac{1}{5}$$

$$= \frac{1RS_{a}}{\sqrt{a}} + \frac{1RS_{c}}{BW_{c}} + \frac{1}{BW_{c}} = \frac{1}{5}$$

$$= \frac{1}{5} = \frac{$$

 Table A-5.2

 Derivation of Inhalation Factors for Inhalation Exposures, Hypothetical Fishing Platform User: Current Scenario RME, Landfill AOPC

| Description | Variable | Value | Units VF Derivation |
|---|-----------------------|--|--|
| Unit conversion factor | CFi | $1.00E-04 \text{ m}^2/\text{cm}^2$ | |
| Apparent Diffusivity | D_A | derived cm ² /s | $VFs = Q/C_{vol} \times \frac{(3.14 \times D_A \times T_{resident})^{1/2}}{(2 \times \rho_b \times D_A)} \times CF_i$ |
| Diffusivity in air | D_i | chemical-specific cm ² /s | $(2 \times \rho_b \times D_A) $ |
| Diffusivity in water | D_w | chemical-specific cm ² /s | |
| Default organic-carbon content | Foc | 0.006 g/g | |
| Henry's law constant | Η' | chemical-specific dimensionless | |
| Soil-water partition coefficient | K _d | chemical-specific cm3/g | |
| for organics: Kd=K _{OC} ×F _{OC} | | | based on: |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specific cm3/g | $[(a \ 10/3 n \ u) + a \ 10/3 n \) (-2]$ |
| Total soil porosity | n | 0.43 L _{porespace} /L _{soil} | $D_{A} = \frac{\left[\left(\theta_{a}^{10/3} D_{i} H' + \theta_{w}^{10/3} D_{w} \right) / n^{2} \right]}{\rho_{B} K_{d} + \theta_{w} + \theta_{a} H'}$ |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.360E+09 m ³ /kg | $\rho_B K_d + \theta_w + \theta_a H'$ |
| Inverse of the mean concentration at the center | Q/C _{vol} | 73.44 g/m ² -s per kg/m | 3 |
| of a 0.5-acre square source in Los Angeles | | | |
| Dry soil bulk density | ρ_b | 1.5 g/cm ³ | |
| Air-filled soil porosity | θ_{a} | 0.28 L _{air} /L _{soil} | |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0.15 L _{water} /L _{soil} | |
| Exposure interval, 30-yr resident | T _{Resident} | 8.20E+08 seconds | |
| Volatilization Factor for soil | VFs | derived m ³ /kg | |

| Water-filled soil porosity | Θ_{w} | | 5 L _{water} /L _{soil} | | | | | | | | | | |
|---|-----------------------|------------|---|---------------|-----------------------------|-----------------|--|------------|--|---------------------|------------|--------------------|--------------------|
| Exposure interval, 30-yr resident | T _{Resident} | | 08 seconds | | | | | | | | | | |
| Volatilization Factor for soil | VFs | derive | ed m ³ /kg | | | | | | | | | | |
| | | | | | | | | | | | | VFs (VOCs) | PEF |
| | | | \mathbf{D}_{i} | | $\mathbf{D}_{\mathbf{w}}$ | | H' | | K _{OC} | K _d | DA | Residential | (non-VOCs) |
| Analyte | | (cm^2/s) | Reference | (cm^2/s) | Reference | (dimensionless) | Reference | (cm^3/g) | Reference | (cm ³ /g | | m ³ /kg | m ³ /kg |
| Inorganic Constituents | | (**** (**) | | (111, 10) | | | | (**** , 8) | | (**** / 8 | (0000 / 0) | | |
| Arsenic | | 0 | 0 | 0 | 0 | 31.61052561 | Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 | EPISuite v4.10 (USEPA, 2011a): MCI-estimated | 20 | 0 | non-VOC | 1.360E+09 |
| Chromium (III) | | 0 | 0 | 0 | 0 | 0 | Dona Method estimate, EFISane V4.10 (USEFA 2011a) | 13.22 | 0 | 180000 | | non-VOC | 1.360E+09 |
| Lead | | 0 | 0 | 0 | 0 | 1.001885999 | Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 | EPISuite v4.10 (USEPA, 2011a): MCI-estimated | 900 | | non-VOC | 1.360E+09 |
| Mercury | | 0.0307 | Exhibit C-1. USEPA 2002e | 0 000062 | Exhibit C-1, USEPA [2002e] | 0.467116057 | RSL ("SSL"; USEPA, 2012) | 13.22 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 52 | 1.42E-05 | 32929.7937 | use VF |
| Polychlorinated Biphenyls (PCBs) - Mixtures | | 0.0307 | Exhibit C-1, USEPA 2002e | 0.0000005 | Exhibit C-1, USEPA [2002e] | 0.407110037 | KSL (SSL ; USEPA, 2012) | 13.22 | EFISuite v4.10 (USEFA, 2011a); MCI-estimated | 32 | 1.42E-03 | 52929.1931 | use vr |
| Total PCBs | | 0.040076 | RSLs ("WATER9"; USEPA 2012) | 4 6926E 06 | RSLs ("WATER9": USEPA 2012) | 0.011572806 | Experimental, EPISuite v4.10 (USEPA 2011a) | 130500 | EPISuite v4.10 (USEPA, 2011a): MCI-estimated | 0 | 0.000235 | non-VOC | 1.360E+09 |
| Volatile Organic Compounds (VOCs) | l | 0.040070 | RSLS (WATER9; USEPA 2012) | 4.0820E-00 | KSLS (WATER9; USEPA 2012) | 0.011372800 | Experimental, EPISuite V4.10 (USEPA 2011a) | 130300 | EFISuite v4.10 (USEFA, 2011a); MCI-estimated | 0 | 0.000255 | non-voc | 1.300E+09 |
| Naphthalene | 0 | 0.0604994 | RSLs ("WATER9"; USEPA 2012) | 8 277E 06 | RSLs ("WATER9"; USEPA 2012) | 0.017002055 | Experimental, EPISuite v4.10 (USEPA 2011a) | 1544 | EPISuite v4.10 (USEPA, 2011a): MCI-estimated | 0 | 0.000546 | 5316.643215 | use VF |
| Pesticides | 0 | 1.0004994 | RSLS (WATER9; USEPA 2012) | 8.377E-00 | KSLS (WATER9; USEPA 2012) | 0.017993033 | Experimental, EPISuite V4.10 (USEPA 2011a) | 1344 | EFISURE V4.10 (USEFA, 2011a); MCI-estimated | 0 | 0.000340 | 3310.043213 | use vr |
| MCPP | 0 | 0.0529988 | RSLs ("WATER9": USEPA 2012) | C 1025E 0C | RSLs ("WATER9": USEPA 2012) | 2 ((404E 09 | Experimental, EPISuite v4.10 (USEPA 2011a) | 48.51 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 4.01E-07 | non-VOC | 1.360E+09 |
| | 0 | 0.0529988 | RSLS (WATER9 ; USEPA 2012) | 0.1925E-00 | RSLS (WATER9 ; USEPA 2012) | 3.00404E-08 | Experimental, EPISuite V4.10 (USEPA 2011a) | 48.51 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 4.01E-07 | non-voc | 1.300E+09 |
| Polycyclic Aromatic Hydrocarbons (PAHs) | 0 | 0509647 | DEL - ("WATEDO": LICEDA 2012) | 5 0421E 0C | RSLs ("WATER9": USEPA 2012) | 0.00040072 | Emerimental EDISaite and 10 (USEDA 2011a) | 17(000 | EDIS-site and 10 (LISEDA 2011-); MCL antimated | 0 | 1.22E.05 | nen VOC | 1.2000.00 |
| Benzo[a]anthracene | | 0.0508647 | RSLs ("WATER9"; USEPA 2012) | | | 0.00049072 | Experimental, EPISuite v4.10 (USEPA 2011a) | 176900 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 1.33E-05 | non-VOC | 1.360E+09 |
| Benzo[a]pyrene | | 0.0475831 | RSLs ("WATER9"; USEPA 2012) | | RSLs ("WATER9"; USEPA 2012) | | Experimental, EPISuite v4.10 (USEPA 2011a) | 587400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 8.2E-07 | non-VOC | 1.360E+09 |
| Benzo[b]fluoranthene | 0 | 0.0475831 | RSLs ("WATER9"; USEPA 2012) | | RSLs ("WATER9"; USEPA 2012) | | Experimental, EPISuite v4.10 (USEPA 2011a) | 599400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 1.02E-06 | non-VOC | 1.360E+09 |
| Benzo[g,h,i]perylene | | 0 | 0 | 0 | 0 | 1.35357E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 1951000 | | 0 | 0 | non-VOC | 1.360E+09 |
| Benzo[k]fluoranthene | | 0.0475831 | RSLs ("WATER9"; USEPA 2012) | 2.000 / 12 00 | RSLs ("WATER9"; USEPA 2012) | 2.38817E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 587400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 |)og 0, | non-VOC | 1.360E+09 |
| Benzofluoranthenes, Total | | 0.0475831 | RSLs ("WATER9"; USEPA 2012) | | RSLs ("WATER9"; USEPA 2012) | 2.68669E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 599400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 1.02E-06 | non-VOC | 1.360E+09 |
| Dibenz[a,h]anthracene | | 0.0445672 | RSLs ("WATER9"; USEPA 2012) | | RSLs ("WATER9"; USEPA 2012) | 5.76596E-06 | Experimental, EPISuite v4.10 (USEPA 2011a) | 1912000 | | 0 | 4.7E-07 | non-VOC | 1.360E+09 |
| Indeno[1,2,3-cd]pyrene | 0 | 0.0447842 | RSLs ("WATER9"; USEPA 2012) | 5.2327E-06 | RSLs ("WATER9"; USEPA 2012) | 0.0000656 | EPISuite v4.10 (USEPA 2011a) | 3470000 | EPISuite v4.10 (USEPA 2011a) | 0 | 1.86E-06 | non-VOC | 1.360E+09 |

| Description | Variable | Valu | e | Equations | | | | | | | |
|---|-----------------------|-------------------|----------------------|------------------|--|---|--|---|---|-----------------|-----------------|
| Averaging Time, carcinogens | AT _c | 7 | 0 yrs | Cancer Risk: | | | | | | | |
| Averaging Time, noncarcinogens, resident | AT _{nc,r} | 2 | 6 yrs | Nonmutagens: | | / | | , | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | | 0 yrs | | $Risk = IIIR \times C$ | $ED_r \times (ED_r \times ED_r \times ED_r$ | $EF_r \times ET_r$ | _) | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | | 3 yrs | | | $AT_c \times 365$ | $\frac{day}{day} \times 24 \frac{ho}{day}$ | <u>ur</u> | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | | 3 yrs | | $Risk = IUR \times C_{0}$ $Risk = IUR \times \left[C_{0}\right]$ | | year do | iy / | | | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | | 1 yrs | Mutagens: | $Risk = IUR \times C_{0}$ | $_{vir} \times EF_r \times ET_r$ | | | | | |
| Averaging Time, carcinogens | AT_{c} | 7 | 0 yrs | | L | (EDo | $x ADAF_{0}$ | | $ED_{2} \leftarrow \times ADAF_{2}$ | | |
| COPC Concentration in air | C_{air} | chemical-specifi | $c \mu g/m^3$ | | × | $\frac{220-2}{4T \times 365 days}$ | $\frac{1}{\sqrt{2}}$ | $\overline{ours/day}^+ \overline{AT_c \times 36}$ | 5 days /vear × 2 | -6 A hours/a | lav |
| COPC Concentration in soil | C_{soil} | chemical-specifi | c mg/kg | | | $(H_c \times 305 uuys)$ | Y ADAF. | G_{α} | $\frac{55 \text{uuys}}{50 \text{uuys}} \times \frac{404 F}{50}$ | - 11041 5/1 |)] |
| Exposure Duration, child 0-2 | ED ₀₋₂ | | 2 yrs | | + | $\frac{L D_{6-16}}{4T \times 26E days}$ | $\times HDH_{6-16}$ | $\frac{h}{urs/day} + \frac{h}{AT_c \times 36}$ | $\frac{D_{16-26} \times D_{11}}{5}$ | -26 | <u></u> { |
| Exposure Duration, child 2-6 | ED ₂₋₆ | | 4 yrs | | | $AI_c \times 305 \ uuys/$ | yeur × 24 no | $arspace Arc \times 50$ | 5 uuys/yeur × 24 | noursju | <i>uy</i>)] |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 1 | 0 yrs | Vinyl Chloride: | 1 | (| FD × FF × I | <i>чт</i>)] | | | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 1 | 0 yrs | | $Risk = IUR \times C_a$ | $_{ir} \times \left\{ \frac{1}{4T \times 26E} \right\}$ | $D_r \land D_r \land T_r \land T_r$ | $\frac{5T_r}{24 hours/day} + 1$ | | | |
| Exposure Duration, child | ED_{c} | | 6 yrs | | L | $(AI_c \times 505 u)$ | uys/yeur x | 24 nours/uuy)] | | | |
| Exposure Duration, resident | ED_r | | 6 yrs | | | | | | | | |
| Exposure Frequency | EF_r | 36 | 5 days/yr | | | | | | | | |
| Exposure Time | ET_r | 2 | 4 hours/day | Noncancer Haza | rd: 1 | / FD | $) \times FF \times FT$ | , _\ | | | |
| Inhalation Unit-Risk Factor | IUR | chemical-specifi | $(\mu g/m^3)^{-1}$ | | Hazard = $\frac{1}{RfC}$ × | $C_{air} \times ($ | $\frac{dav}{dav}$ | r hour | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+0 | 9 m ³ /kg | | ĸjt | $AT_{nc,r} \times$ | $365 \frac{uuy}{vear} \times 2$ | $\left(4\frac{dour}{day}\right)$ | | | |
| Inhalation Reference Concentration | RfC | chemical-specifi | $c \mu g/m^3$ | | | N N | <i>y</i> e ea. | | | | |
| Volatilization Factor for VOCs | VF | chemical-specifi | c m [°] /kg | Where: | | | | $C_{air} = \frac{C_{soil}}{VF \text{ or } PH}$ | $\frac{\mu g}{EF} \times 1000 \frac{\mu g}{mg}$ | | |
| | | | VF (VOCs) or | | | Cancer | | | Noncancer | | Noncancer |
| | | C _{soil} | PEF (non-VOCs) | C _{air} | IUR | Risk | % of | RfC | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total | (dimensionless) |
| Inorganic Constituents | | | | | | | | | | | |
| Arsenic | 0.00E+00 | 1.05E+01 | 1.360E+09 | 7.72E-06 | 4.30E-03 | 7.40E-09 | < 1% | 1.50E-02 | 3.09E-04 | < 1% | 3.09E-04 |
| Chromium (III) | 0.00E+00 | 5.94E+02 | 1.360E+09 | 4.37E-04 | No Toxicity Value | | | No Toxicity Value | | | |
| Lead | 0.00E+00 | 0.00E+00 | 1.360E+09 | 0.00E+00 | 1.20E-05 | 0.00E+00 | < 1% | No Toxicity Value | | | |
| Mercury | 0.00E+00 | 1.45E+00 | 3.293E+04 | 4.40E-02 | No Toxicity Value | | | 3.00E-01 | 1.47E-01 | 38% | 1.47E-01 |
| Polychlorinated Biphenyls (PCBs) - Mixtur | | 1.451100 | 5.2551104 | | Tto Toxicity Value | | | 5.002 01 | 1.472.01 | 5070 | 1.472.01 |
| Total PCBs | 0.00E+00 | 4.10E-01 | 1.360E+09 | 3.01E-07 | 5.70E-04 | 6.38E-11 | < 1% | No Toxicity Value | | | |
| | 0.00E+00 | 4.10E-01 | 1.500E+09 | 5.01E-07 | J.70E-04 | 0.38E-11 | < 1% | No Toxicity value | | | |
| Volatile Organic Compounds (VOCs) | 0.007.00 | 2.025.00 | 5.2155.02 | 5 3 05 01 | 2 405 05 | 0.105.07 | 1000/ | 2.005.00 | 2 40 E 04 | 60 04 | 0.407.04 |
| Naphthalene | 0.00E+00 | 3.83E+00 | 5.317E+03 | 7.20E-01 | 3.40E-05 | 9.10E-06 | 100% | 3.00E+00 | 2.40E-01 | 62% | 2.40E-01 |
| Pesticides | | | | | | | | | | | |
| MCPP | 0.00E+00 | 1.12E+01 | 1.360E+09 | 8.24E-06 | No Toxicity Value | | | No Toxicity Value | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) |) | | | | | | | | | | |
| Benzo[a]anthracene | М | 9.96E+00 | 1.360E+09 | 7.32E-06 | 1.10E-04 | 8.29E-10 | < 1% | No Toxicity Value | | | |
| Benzo[a]pyrene | М | 1.12E+01 | 1.360E+09 | 8.24E-06 | 1.10E-03 | 9.32E-09 | < 1% | No Toxicity Value | | | |
| Benzo[b]fluoranthene | М | 1.52E+01 | 1.360E+09 | 1.12E-05 | 1.10E-04 | 1.26E-09 | < 1% | No Toxicity Value | | | |
| Benzo[g,h,i]perylene | 0.00E+00 | 5.82E+00 | 1.360E+09 | 4.28E-06 | No Toxicity Value | | | No Toxicity Value | | | |
| | | | | | Ť | | | - | | | |
| = | 111 | 2.001.01 | 1.000001000 | 1.702 00 | | | / 0 | ite i enterty varae | | | |
| Benzo[k]fluoranthene | М | 2.63E+01 | 1.360E+09 | 1.93E-05 | 1.10E-04 | 2.19E-09 | < 1% | No Toxicity Value | | | |

Table A-5.3Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Landfill AOPC

| Description | Variable | Valu | ie | Equations | | | | | | | |
|--|-----------------------|-------------------|-----------------------|------------------|--|--|---|---|--|-----------------------|---------------------|
| Averaging Time, carcinogens | AT _c | 7 | 0 yrs | Cancer Risk: | | | | | | | |
| Averaging Time, noncarcinogens, resident | $AT_{nc,r}$ | 2 | 6 yrs | Nonmutagens: | | , | | 、 、 | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 1 | 0 yrs | | $Risk - IIIR \times C$ | $ED_r \times \left(\frac{ED_r \times ED_r}{ED_r} \right)$ | $EF_r \times ET_r$ | _) | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | | 3 yrs | | $Risk = IOR \times C_0$ | $AT_{a} \times 365$ | $\frac{day}{day} \times 24 \frac{hou}{day}$ | <u>r</u>] | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | | 3 yrs | | ſ | | year | ' / | | | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | | 1 yrs | Mutagens: | $Risk = IUR \times C_{0}$ $Risk = IUR \times \left[C_{0}\right]$ | $_{nir} \times EF_r \times ET_r$ | | | | | |
| Averaging Time, carcinogens | AT _c | 7 | 0 yrs | | L | (FD _a | $\sim ADAF_{\circ}$ | | $FD_{a} \propto ADAF_{a}$ | | |
| COPC Concentration in air | C_{air} | chemical-specif | ic μg/m ³ | | × | $\frac{1}{4T} \times 365 days$ | $\frac{1}{\sqrt{2}}$ | $\frac{1}{\sqrt{T}} + \frac{1}{\sqrt{T}}$ | $ED_{2-6} \times ADAF_{2}$ 65 days/year × 24 $ED_{16-26} \times ADAF_{16}$ 5 days/year × 24 | -6 | day |
| COPC Concentration in soil | C_{soil} | chemical-specif | ic mg/kg | | | $(AI_c \land 303 uuys)$ | yeur ∧ 24 noi ∨ ADAE | $H_c \wedge S$ | $5 uuys/yeur \land 2$ | e nour s _j | 11 1 |
| Exposure Duration, child 0-2 | ED ₀₋₂ | | 2 yrs | | + | $\frac{LD_{6-16}}{4T \times 26\Gamma days}$ | $\sim ADAP_{6-16}$ | $\frac{1}{\pi a/day} + \frac{1}{4\pi \sqrt{2}}$ | $LD_{16-26} \times RDRP_{16}$ | -26 | <u></u> {} |
| Exposure Duration, child 2-6 | ED ₂₋₆ | | 4 yrs | | | $AI_c \times 505 uuys/$ | yeur x 24 nou | $I_{c} \times 30$ | 5 uuys/yeur × 24 | nours/u | (<i>ay</i>)] |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 1 | 0 yrs | Vinyl Chloride: | 1 | (| $FD \times FF \times FT$ | י א | | | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 1 | 0 yrs | | $Risk = IUR \times C_a$ | $_{ir} \times \left\{ \frac{1}{AT} \times 26E \right\}$ | $D_r \times D_r \times D_r$ | $\left\{\frac{r_r}{4 hours/day} + 1\right\}$ | | | |
| Exposure Duration, child | ED_{c} | | 6 yrs | | L | $(AI_c \times 505 u)$ | uys/yeur × 2 | + <i>nours</i> / <i>uuy</i>)] | | | |
| Exposure Duration, resident | ED_r | | 6 yrs | | | | | | | | |
| Exposure Frequency | EF_r | 36 | 5 days/yr | | | | | | | | |
| Exposure Time | ET_r | 2 | 4 hours/day | Noncancer Haza | ırd: 1 | / FI | $) \vee FF \vee FT$ | \ | | | |
| Inhalation Unit-Risk Factor | IUR | chemical-specif | ic $(\mu g/m^3)^{-1}$ | | Hazard = $\frac{1}{Dfc}$ × | $C_{air} \times \left($ | $\frac{d_{c} \wedge B_{r} \wedge B_{r}}{d_{c}}$ | hour | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+0 | 9 m ³ /kg | | Hazard = $\frac{1}{RfC}$ × | $AT_{nc,r} \times$ | $365 \frac{uuy}{vear} \times 24$ | $\frac{dav}{dav}$ | | | |
| Inhalation Reference Concentration | RfC | chemical-specif | ic $\mu g/m^3$ | | | \ \ | yeur | 5 / | | | |
| Volatilization Factor for VOCs | VF | chemical-specif | ic m ³ /kg | Where: | | | | $C_{air} = \frac{C_{soil}}{VF \text{ or } P}$ | $\frac{\mu g}{mg} \times 1000 \frac{\mu g}{mg}$ | | |
| | | | VF (VOCs) or | | | Cancer | | | Noncancer | | Noncancer |
| | | C _{soil} | PEF (non-VOCs) | C _{air} | IUR | Risk | % of | RfC | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total | (dimensionless) |
| Dibenz[a,h]anthracene | М | 2.06E+00 | 1.360E+09 | 1.51E-06 | 1.20E-03 | 1.87E-09 | < 1% | No Toxicity Value | | | |
| Indeno[1,2,3-cd]pyrene | М | 6.80E+00 | 1.360E+09 | 5.00E-06 | 1.10E-04 | 5.66E-10 | < 1% | No Toxicity Value | | | |
| | | | | | | Cancer Risk | | | Hazard Index (Child) | | Hazard Index (Adult |
| | | | | | Pathway Sum: | 9.1E-06 | | Pathway Sum: | 0.39 | | 0.39 |

Table A-5.3Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Landfill AOPC

| Description | Variable | Value | e | Equations: | | | | |
|--|----------------------------|-------------------|-----------------------------|----------------------------|--|---------------------------|---|-------------------------|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specifi | c unitless | Cancer Risk: | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 |) yrs | Nonmutagens: | $Pick - SE_{\star}$ | $\int C - \nabla F$ | $D_a \times EF \times EV \times D$ | $FC \sim ABS$ |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | ź | 3 yrs | | , u | 2 3011 | ü | u u |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | ź | 3 yrs | | × (| $AF_a \times SA_a$ | $\frac{1}{av/vear} + \frac{A}{BW_c \times W_c}$ | $AF_c \times ED_c$ |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | | l yrs | | $(BW_a \times A)$ | $G_c \times 365 da$ | ay/year ' $BW_c \times$ | $AT_c \times 365$ |
| Soil-to-Skin Adherence Fraction | AF_{a} | 0.0 | 7 mg/cm ² -day | Mutagens: | (| | | |
| Soil-to-Skin Adherence Fraction | AF_{c} | 0. | 2 mg/cm ² -day | Risk = SF | $_{d} \times \{C_{soil} \times EF_{r}\}$ | $\times EV \times Ec$ | $C \times ABS_d \times CF_d$ | |
| Averaging Time, Carcinogens | AT _c | 70 |) yrs | | | | | D 4 F |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 20 |) yrs | | $\times \left \frac{ED_{0-2}}{2} \right $ | $\times AF_{c} \times SA$ | $\frac{c}{b} \frac{ADAF_{0-2}}{days/year} + \frac{E}{B}$ | $D_{2-6} \times AF_{6}$ |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | | 5 yrs | | $BW_c \times$ | $AT_c \times 365$ | days/year l | $BW_c \times AT_c$ |
| Body Weight, adult | $\mathbf{BW}_{\mathbf{a}}$ | 80 |) kg | | $+\frac{ED_{6-16}}{ED_{6-16}}$ | $\times AF_a \times SA$ | $A_a \times ADAF_{6-16} + 5 days/year$ | $ED_{16-26} \times$ |
| Body Weight, child | BW_{c} | 1: | 5 kg | | $BW_a \times$ | $AT_c \times 365$ | 5 days/year ' | $BW_a \times$ |
| Conversion Factor | CF_d | 1E-0 | 6 kg/mg | Vinyl Chloride: | $Risk - SF. \times$ | {C × F) | $F_r \times EV \times EC \times A$ | $RS \cdot \times CF$ |
| COPC Concentration in Soil | C_{soil} | chemical-specifi | c mg/kg | | $Risk = 5I_d \wedge$ | C _{soil} ~ L | $r \wedge L V \wedge L C \wedge H$ | $D_d \wedge C_d$ |
| Fraction of EV in Contact with Soil | EC | | l unitless | | $\int_{EE} \sqrt{ED}$ | $_c \times AF_c \times S$ | $SA_{c} \perp ED_a \times AF_a$ | $\times SA_a$ |
| Exposure Duration, child 0-2 | ED ₀₋₂ | , | 2 yrs | | $\times \frac{ET_r \wedge ()}{2}$ | BW_c | $\frac{BH_c}{BW_a} + \frac{ED_a \times AT_a}{BW_a}$ $\frac{365 \frac{days}{magga}}{BW_a}$ |) |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 2 | 4 yrs | | | $AT_{-} \times$ | 365 <u>days</u> | 1 |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 |) yrs | | L | m _c ~ | year | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | |) yrs | | | | | |
| Exposure Duration, adult | ED _a | |) yrs | Noncancer Haza | rd: | | | |
| Exposure Duration, child | ED _c | | 5 yrs | II | $a_{a} = \frac{1}{\sqrt{2}}$ | | $F_c \times ABS_d \times ED_{nc,c} \times EF_r$ | ×EC×SA _c ×CF |
| Exposure Frequency, resident | EFr | | 5 days/yr | П | $IZUTU = \frac{1}{RfD_d} \times$ | $c_{soil} \times (-$ | $\frac{F_c \times ABS_d \times ED_{nc,c} \times EF_r}{BW_c \times AT_{nc,c} \times 36}$ | $5\frac{day}{vaar}$ |
| Event Frequency | EV | | l events/day | | | | | yeur |
| Reference Dose Adjusted for GI Absorption | RfD _d | chemical-specifi | | | | | | |
| Exposed Body Surface Area, adult | SA_a | | 2 cm^2 | | | | | |
| Exposed Body Surface Area, child | SA_{c} | | 3 cm^2 | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specifi | c (mg/kg-day) ⁻¹ | | | | | |
| | | | | | Cancer | | | Nonc |
| | | C _{soil} | ABS _d | $\mathbf{SF}_{\mathbf{d}}$ | Risk | % of | $\mathbf{RfD}_{\mathbf{d}}$ | Hazard |
| Analyte | Mutagen? | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimens |
| Inorganic Constituents | | | | | | | | |
| Arsenic | 0.00E+00 | 1.05E+01 | 3.00E-02 | 1.50E+00 | 1.20E-06 | < 1% | 3.00E-04 | 1.99 |
| Chromium (III) | 0.00E+00 | 5.94E+02 | 1.00E-02 | | | | 1.95E-02 | 9.64 |
| Lead | 0.00E+00 | 0.00E+00 | 1.00E-02 | 8.50E-03 | 0.00E+00 | < 1% | | |
| Mercury | 0.00E+00 | 1.45E+00 | 1.00E-02 | | | | 1.60E-04 | 2.87 |
| Polychlorinated Biphenyls (PCBs) - Mixtures | | | | | | | | |
| Total PCBs | 0.00E+00 | 4.10E-01 | 1.40E-01 | 2.00E+00 | 4.84E-07 | < 1% | 2.00E-05 | 9.08 |
| | | | | | | | | |

0.00E+00

0.00E+00

Volatile Organic Compounds (VOCs)

Naphthalene

Pesticides MCPP

3.83E+00

1.12E+01

1.30E-01

1.00E-01

--

2.00E-02

1.00E-03

Table A-5.4 Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: RME, Landfill AOPC

| Description | Variable | Valu | e | Equations: | | | | |
|---|------------------------|-------------------|-----------------------------|----------------------------|--|----------------------------|---|-----------------------------|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specifi | c unitless | Cancer Risk: | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 | 0 yrs | Nonmutagens: | Dick - SE | | $D_a \times EF \times EV \times I$ | $C \sim ADC$ |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | - | 3 yrs | | | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | ź | 3 yrs | | × (| $4F_a \times SA_a$ | $\frac{A}{W_{ear}} + \frac{A}{BW_{eX}}$ | $F_c \times ED_c$ |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | | 1 yrs | | $(BW_a \times A)$ | $T_c \times 365 da$ | $y/year' BW_c \times$ | $AT_c \times 36$ |
| Soil-to-Skin Adherence Fraction | AF_{a} | 0.0 | 7 mg/cm ² -day | Mutagens: | (| | | |
| Soil-to-Skin Adherence Fraction | AF _c | 0.2 | 2 mg/cm ² -day | Risk = SF | $C_d \times \{C_{soil} \times EF_r\}$ | $\times EV \times EC$ | $C \times ABS_d \times CF_d$ | |
| Averaging Time, Carcinogens | AT _c | 70 | 0 yrs | | | | | D 41 |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 20 | 0 yrs | | $\times \left \frac{ED_{0-2}}{2} \right $ | $\times AF_c \times SA$ | $\frac{c \times ADAF_{0-2}}{days/year} + \frac{E}{H}$ | $D_{2-6} \times AF$ |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | | б yrs | | | | | |
| Body Weight, adult | BW_a | 80 | 0 kg | | $+\frac{ED_{6-16}}{2}$ | $\times AF_a \times SA$ | $\frac{1}{6} \times ADAF_{6-16} + \frac{1}{6} days/year$ | $ED_{16-26} \times$ |
| Body Weight, child | BW_{c} | 1: | 5 kg | | $BW_a \times$ | $AT_c \times 365$ | 5 days/year | $BW_a \times$ |
| Conversion Factor | CF_d | 1E-0 | 6 kg/mg | Vinyl Chloride: | $Pick - SE \vee$ | | $F_r \times EV \times EC \times A$ | $PC \sim CE$ |
| COPC Concentration in Soil | C_{soil} | chemical-specifi | c mg/kg | | $\pi i s \kappa - s r_d \wedge$ | C _{soil} ~ EI | $r_r \wedge EV \wedge EU \wedge A$ | $D_d \wedge C_d$ |
| Fraction of EV in Contact with Soil | EC | | 1 unitless | | $\int_{EE} \sqrt{ED}$ | $P_c \times AF_c \times S$ | $SA_{c} \perp ED_{a} \times AF_{a}$ | $\times SA_a$ |
| Exposure Duration, child 0-2 | ED ₀₋₂ | , | 2 yrs | | $\times \left[\frac{Er_r \times ()}{2} \right]$ | BW_c | $-+$ BW_a | <u> </u> |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 4 | 4 yrs | | | $AT_{-} \times$ | $\frac{BA_c}{BW_a} + \frac{BD_a \times AF_a}{BW_a}$ $365 \frac{days}{BW_a}$ | I |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 | 0 yrs | | L | m _c x | year | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | | 0 yrs | | | | | |
| Exposure Duration, adult | ED _a | | 0 yrs | Noncancer Haza | ard: | | | |
| Exposure Duration, child | ED _c | | 6 yrs | U | $a_{a}a_{a}d = \frac{1}{\sqrt{2}}$ | $C \sim (AF)$ | $C_c \times ABS_d \times ED_{nc,c} \times EF_r$ | <ec×sa<sub>c×Cl</ec×sa<sub> |
| Exposure Frequency, resident | EF _r | | 5 days/yr | П | $uzuru - \frac{RfD_d}{RfD_d}$ | $c_{soil} \wedge ($ | $\frac{C_c \times ABS_d \times ED_{nc,c} \times EF_r}{BW_c \times AT_{nc,c} \times 36}$ | 5 day vear |
| Event Frequency Refererenc Dose Adjusted for GI Absorption | EV RfD _d | chemical-specifi | 1 events/day | | | | | year |
| | _ | - | 2 cm^2 | | | | | |
| Exposed Body Surface Area, adult | SA _a | | 3 cm^2 | | | | | |
| Exposed Body Surface Area, child | SA_c | | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specific | c (mg/kg-day) ⁻¹ | | | | | |
| | | | | | Cancer | | | None |
| | | C _{soil} | ABS _d | $\mathbf{SF}_{\mathbf{d}}$ | Risk | % of | $\mathbf{RfD}_{\mathbf{d}}$ | Hazaro |
| Analyte | Mutagen? | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimen |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]anthracene | М | 9.96E+00 | 1.30E-01 | 7.30E-01 | 1.65E-05 | 5% | | |
| Benzo[a]pyrene | М | 1.12E+01 | 1.30E-01 | 7.30E+00 | 1.86E-04 | 60% | | |
| Benzo[b]fluoranthene | М | 1.52E+01 | 1.30E-01 | 7.30E-01 | 2.52E-05 | 8% | | |
| Benzo[g,h,i]perylene | 0.00E+00 | 5.82E+00 | 1.30E-01 | | | | 3.00E-02 | 7.98 |
| Benzo[k]fluoranthene | М | 2.63E+01 | 1.30E-01 | 7.30E-02 | 4.36E-06 | 1% | | |
| Benzofluoranthenes, Total | М | 1.88E+01 | 1.30E-01 | 7.30E-01 | 3.12E-05 | 10% | | |
| Dibenz[a,h]anthracene | М | 2.06E+00 | 1.30E-01 | 7.30E+00 | 3.42E-05 | 11% | | |
| Indeno[1,2,3-cd]pyrene | М | 6.80E+00 | 1.30E-01 | 7.30E-01 | 1.13E-05 | 4% | | |
| - * * | | | | | | | | |

Table A-5.4 Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: RME, Landfill AOPC

Cancer Risk

Pathway Sum:

3.1E-04

Pathway Sum:

Table A-6.NI Nursing Infant: Hypothetical Fishing Platform User: CTE Summary - Landfill AOPC

| Definition | Variable | Value | Source | Equations | |
|--|--------------------|--------------------|-----------------|---------------------------|---------------------|
| Infant Risk Adjustment Factor | IRAF | Chemical Specific | ODEQ 2010 | | |
| Carcinogenic IRAFc | | | | | |
| Total PCB | IRAFc_pcb | 0.6 | ODEQ 2010 | Infant Cancer Risk = | Mother Risk x IRAFc |
| Noncancer IRAFnc | | | | | |
| Total PCB | IRAFnc_pcb | 4 | ODEQ 2010 | Infant Noncancer Hazard = | Child HQ x IRAFnc |
| | | | | | |
| | | Mo | ther | Infa | ant |
| | C _{soilt} | Cancer Risk | Hazard Quotient | Cancer Risk | Hazard Quotient |
| Analyte | (mg/kg) | (adult) | (child) | | |
| Total PCBs as Congeners (KM-based, capped) | 4E-01 | 2.3E-07 | 6.4E-02 | 1.4E-07 | 0.26 |
| | | | | | |

| Definition | Variable | Value | Equations | Value |
|--|-----------------------|---|--------------------------------|---|
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 yrs | Cancer Risk: | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 yrs | Nonmutagens: | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 yrs |] | $(IRS_a \times ED_a)$ |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 yrs | $Risk = SF_o \times C_{soi} $ | $T_{l} \times EF_r \times FI \times CF_O \times \left(\frac{IRS_a \times ED_a}{BW_a \times AT_c \times 365 \frac{day}{year}} + \frac{BT_a}{BT_a}\right)$ |
| Averaging Time, carcinogens | AT _c | 70 yrs | | $\left(BW_a \times AT_c \times 365 \frac{1}{year} \right) B$ |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 3 yrs | Mutagens: | , |
| Averaging Time, noncarcinogens, child | AT _{nc.c} | 6 yrs | $Risk - SF_{a} \times \int C$ | × FF × FI × CF |
| Body Weight | BWa | 80 kg | | $EF_r \times FI \times CF_o$ |
| Body Weight | BW _c | 15 kg | × | $\left(\frac{IRS_c \times ED_{0-2} \times ADAF_{0-2}}{BW_c \times AT_c \times 365 \ days/year} + \frac{IRS_c \times ED_{2-6} \times BW_c \times AT_c \times 365 \ days/year}{BW_c \times AT_c \times 365 \ days/year} + \frac{IRS_c \times ED_{2-6} \times BW_c \times AT_c \times 365 \ days/year}{BW_c \times AT_c \times 365 \ days/year} + \frac{IRS_c \times ED_{2-6} \times BW_c \times $ |
| Conversion Factor | CF _o | 1E-06 kg/mg | X | $BW_c \times AT_c \times 365 \ days/year \ BW_c \times AT_c \times 365$ |
| COPC Concentration in soil | C_{soil} | chemical-specific mg/kg | +- | $\frac{IRS_a \times ED_{6-16} \times ADAF_{6-16}}{BW_a \times AT_c \times 365 \ days/year} + \frac{IRS_a \times ED_{16-26} \times ADAF_{6-16}}{BW_a \times AT_c \times 365 \ days/year}$ |
| Exposure Duration, child 0-2 | ED ₀₋₂ | 2 yrs | · · | $BW_a \times AT_c \times 365 \ days/year ' BW_a \times AT_c \times 365 \ days/yea$ |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 4 yrs | Vinyl Chloride: | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 yrs | | $C_{soil} \times FI \times CF_{O} \times \left[\frac{EF_{r} \times \left(\frac{ED_{c} \times IRS_{c}}{BW_{c}} + \frac{ED_{a} \times IR}{BW_{a}}\right)}{AT_{c} \times 365 \frac{days}{vear}} \right]$ |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 yrs | $Risk = SF_0 \times \{$ | $C_{soil} \times FI \times CF_0 \times $ |
| Exposure Duration, adult | ED_a | 3 yrs | | $AT_c \times 365 \frac{1}{year}$ |
| Exposure Duration, child | ED_{c} | 6 yrs | | |
| Exposure Frequency, resident | EF_r | 152 days/yr | Noncancer Hazard: | |
| Fraction Contaminated Soil Ingested | FI | 1.0 unitless | 1 | $\begin{bmatrix} / IPS \times FD \times FE \times FI \times CE \end{bmatrix}$ |
| Soil Ingestion Rate (adult) | IRS _a | 50 mg/day | Hazard = $\frac{1}{D(D)}$ × | $\left(C_{soil} \times \left(\frac{IRS_c \times ED_c \times EF_r \times FI \times CF_o}{BW_s \times AT_{res} \times 365 day/year}\right)\right)$ |
| Soil Ingestion Rate (child) | IRS _c | 100 mg/day | Rf D _o | $[W_c \times AI_{nc,c} \times 365 \ aay/year]$ |
| Oral Reference Dose | RfD _o | chemical-specific (mg/kg-day) | | |
| Oral Slope Factor | SFo | chemical-specific (mg/kg-day) ⁻¹ | | |
| | | | Cancer | Nonca |
| | | C | Risk | % of RfD Hazard (|

Table A-6.1 Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: CTE, Landfill AOPC

| | | | | Cancer | | | Noncar |
|---|----------|-------------------|--------------------|-----------------|-------|-------------------|-----------|
| | | C _{soil} | SFo | Risk | % of | RfDo | Hazard (|
| Analyte | Mutagen? | (mg/kg) | $(mg/kg-day)^{-1}$ | (dimensionless) | Total | (mg/kg-day) | (dimensio |
| Inorganic Constituents | | | | | | | |
| Arsenic | 0.00E+00 | 1.05E+01 | 1.50E+00 | 2.35E-06 | 1% | 3.00E-04 | 5.83E- |
| Chromium (III) | 0.00E+00 | 5.94E+02 | No Toxicity Value | | | 1.50E+00 | 1.10E- |
| Lead | 0.00E+00 | 0.00E+00 | 8.50E-03 | 0.00E+00 | < 1% | No Toxicity Value | |
| Mercury | 0.00E+00 | 1.45E+00 | No Toxicity Value | | | 1.60E-04 | 2.52E- |
| Polychlorinated Biphenyls (PCBs) - Mixtures | | | | | | | |
| Total PCBs | 0.00E+00 | 4.10E-01 | 2.00E+00 | 2.04E-07 | < 1% | 2.00E-05 | 5.69E- |
| Volatile Organic Compounds (VOCs) | | | | | | | |
| Naphthalene | 0.00E+00 | 3.83E+00 | No Toxicity Value | | | 2.00E-02 | 5.32E- |
| Pesticides | | | | | | | |
| МСРР | 0.00E+00 | 1.12E+01 | No Toxicity Value | | | 1.00E-03 | 3.11E- |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | |
| Benzo[a]anthracene | М | 9.96E+00 | 7.30E-01 | 1.03E-05 | 5% | No Toxicity Value | |
| Benzo[a]pyrene | М | 1.12E+01 | 7.30E+00 | 1.16E-04 | 59% | No Toxicity Value | |
| Benzo[b]fluoranthene | М | 1.52E+01 | 7.30E-01 | 1.57E-05 | 8% | No Toxicity Value | |

$$\frac{IRS_c \times ED_c}{BW_c \times AT_c \times 365 \frac{day}{year}}\right)$$

$$\frac{\langle ADAF_{2-6}}{\langle days/year} \\ \frac{ADAF_{16-26}}{\langle days/year} \\ \frac{ADAF_{16-26}}{\langle days/year} \\ \end{pmatrix} + \frac{IRS_c}{BW_c} \right)$$

$$\frac{RS_a}{-} + \frac{IRS_c}{BW_c} \right)$$

$$\frac{RS_a}{-} + \frac{IRS_c}{BW_c} \\ + \frac{IRS$$

| Definition | Variable | Val | ue | Equations | Valı | ie | | | |
|--|-----------------------|-------------------|---------------------------|--|--|--|---|----------------------------|--------------------|
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 | yrs | Cancer Risk: | | | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 | yrs | Nonmutagens: | | | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 | yrs | 1 | | $IRS_a \times ED$ | , IRS _c | $\times ED_{c}$ | \] |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 | yrs | $Risk = SF_o \times C_{sol}$ | $E_{l} \times EF_r \times FI \times CF_c$ | $_{D} \times \left(\frac{IRS_{a} \times ED}{BW_{a} \times AT_{c} \times 36}\right)$ | $\frac{d}{dav} + \frac{dav}{dav} + \frac{dav}{dav}$ | | |
| Averaging Time, carcinogens | AT _c | 70 | yrs | | | $BW_a \times AT_c \times 36$ | $5\frac{W}{year}$ $BW_c \times AT_c$ | $\times 365 \frac{1}{yec}$ | <u>r</u> |
| Averaging Time, noncarcinogens, adult | AT _{nc.a} | 3 | yrs | Mutagens: | | X | | | / - |
| Averaging Time, noncarcinogens, child | AT _{nc.c} | 6 | yrs | $Risk - SF_{a} \times \begin{bmatrix} C \end{bmatrix}$ | $ \times FF \times FI \times CF$ | | | | |
| Body Weight | BWa | 80 | kg | $Risk = SF_O \times \left[C_{so}\right]$ | | | | | |
| Body Weight | BWc | 15 | • | × | $(IRS_c \times ED_{0-2} \times$ | $\frac{ADAF_{0-2}}{days/year} + \frac{IRS_c}{BW_c \times}$ | $\times ED_{2-6} \times ADAF_{2-6}$ | <u>.</u> | |
| Conversion Factor | CF _o | 1E-06 | - | | | | | | |
| COPC Concentration in soil | C _{soil} | chemical-specific | | <u></u> . | $IRS_a \times ED_{6-16} \times D_{6-16}$ | $\frac{ADAF_{6-16}}{avs/vear} + \frac{IRS_a \times I}{BW_a \times I}$ | $ED_{16-26} \times ADAF_{16-26}$ | 26 | |
| Exposure Duration, child 0-2 | ED ₀₋₂ | - | yrs | і | $BW_a \times AT_c \times 365 d$ | lays/year ' $BW_a \times D_b$ | АТ _с × 365 days/yea | r]] | |
| Exposure Duration, child 2-6 | ED ₂₋₆ | | yrs | Vinyl Chloride: | | | | 1) | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | | yrs | | | $\frac{EF_r \times \left(\frac{ED_c \times IRS_c}{BW_c}\right)}{AT_c \times 365}$ | $+\frac{ED_a \times IRS_a}{RW}$) IR | s.) | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | | yrs | $Risk = SF_0 \times \{$ | $C_{soil} \times FI \times CF_0 \times$ | | $\frac{BW_a}{days} + \frac{M}{BW_a}$ | $\left \frac{v}{v}\right $ | |
| Exposure Duration, adult | EDa | | yrs | (| | $AT_c \times 365$ | year by | c | |
| Exposure Duration, child | ED _c | 6 | yrs | · · · · · · · · · · · · · · · · · · · | | - | 2 | | |
| Exposure Frequency, resident | EF_{r} | 152 | days/yr | Noncancer Hazard: | | | | | |
| Fraction Contaminated Soil Ingested | FI | 1.0 | unitless | 1 | | | \] | | |
| Soil Ingestion Rate (adult) | IRS _a | 50 | mg/day | Hazard = $\frac{1}{D \in D}$ × | $C_{soil} \times \left(\frac{IKS_c \times I}{DM_c}\right)$ | ED _c × EF _r × FI × CF _c AT _{ress} × 365 day/yea | <u>)</u> | | |
| Soil Ingestion Rate (child) | IRS _c | 100 | mg/day | RfD_o | $\begin{bmatrix} BW_c \times A \end{bmatrix}$ | $T_{nc,c} \times 365 day/yea$ | r)] | | |
| Oral Reference Dose | RfD ₀ | chemical-specific | (mg/kg-day) | | | | | | |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | | |
| | | | | Cancer | | | Noncancer | | Noncancer |
| | | C _{soil} | SFo | Risk | % of | \mathbf{RfD}_{0} | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total | (dimensionless) |
| Benzo[g,h,i]perylene | 0.00E+00 | 5.82E+00 | No Toxicity Value | | | 3.00E-02 | 5.39E-04 | < 1% | 5.05E-05 |
| Benzo[k]fluoranthene | М | 2.63E+01 | 7.30E-02 | 2.72E-06 | 1% | No Toxicity Value | | | |
| Benzofluoranthenes, Total | М | 1.88E+01 | 7.30E-01 | 1.94E-05 | 10% | No Toxicity Value | | | |
| Dibenz[a,h]anthracene | М | 2.06E+00 | 7.30E+00 | 2.13E-05 | 11% | No Toxicity Value | | | |
| Indeno[1,2,3-cd]pyrene | М | 6.80E+00 | 7.30E-01 | 7.04E-06 | 4% | No Toxicity Value | | | |
| | | | | Cancer Risk | | | Hazard Index (Child) | | Hazard Index (Adul |
| | | | Pathway Sum | n: 2.0E-04 | | Pathway Sum: | 0.17 | | 0.016 |

Table A-6.1 Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: CTE, Landfill AOPC

| | | | | Cancer | | | Noncar |
|---------------------------|----------|-------------------|--------------------|-----------------|-------|-------------------|-------------|
| | | C _{soil} | SFo | Risk | % of | RfDo | Hazard (|
| Analyte | Mutagen? | (mg/kg) | $(mg/kg-day)^{-1}$ | (dimensionless) | Total | (mg/kg-day) | (dimensio |
| Benzo[g,h,i]perylene | 0.00E+00 | 5.82E+00 | No Toxicity Value | | | 3.00E-02 | 5.39E- |
| Benzo[k]fluoranthene | М | 2.63E+01 | 7.30E-02 | 2.72E-06 | 1% | No Toxicity Value | |
| Benzofluoranthenes, Total | М | 1.88E+01 | 7.30E-01 | 1.94E-05 | 10% | No Toxicity Value | |
| Dibenz[a,h]anthracene | М | 2.06E+00 | 7.30E+00 | 2.13E-05 | 11% | No Toxicity Value | |
| Indeno[1,2,3-cd]pyrene | М | 6.80E+00 | 7.30E-01 | 7.04E-06 | 4% | No Toxicity Value | |
| | | | | | | | |
| | | | | Cancer Risk | | | Hazard Inde |
| | | | Pathway Sum: | 2.0E-04 | | Pathway Sum: | 0.17 |

 Table A-6.2

 Derivation of Inhalation Factors for Inhalation Exposures, Hypothetical Fishing Platform User: CTE, Landfill AOPC

| escription | Variable | Value | Units VF Derivation | |
|---|-----------------------|--|---------------------|--|
| Unit conversion factor | CFi | $1.00E-04 \text{ m}^2/\text{cm}^2$ | | |
| Apparent Diffusivity | D _A | derived cm ² /s | VFs | $= Q/C_{vol} \times \frac{(3.14 \times D_A \times T_{resident})^{1/2}}{(2 \times \rho_b \times D_A)} \times CF_i$ |
| Diffusivity in air | D_i | chemical-specific cm ² /s | | $(2 \times \rho_b \times D_A) $ |
| Diffusivity in water | D_w | chemical-specific cm ² /s | | |
| Default organic-carbon content | F _{OC} | 0.006 g/g | | |
| Henry's law constant | H' | chemical-specific dimensionless | | |
| Soil-water partition coefficient | K _d | chemical-specific cm ³ /g | | |
| for organics: Kd=K _{OC} ×F _{OC} | | | based on: | |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specific cm3/g | | [(a, 10/3 p, yy + a, 10/3 p,)/(2)] |
| Total soil porosity | n | 0.43 L _{porespace} /L _{soil} | D | $=\frac{\left[\left(\theta_{a}^{10/3}D_{i}H'+\theta_{w}^{10/3}D_{w}\right)/n^{2}\right]}{\rho_{B}K_{d}+\theta_{w}+\theta_{a}H'}$ |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.360E+09 m ³ /kg | - | $ \rho_B K_d + \theta_w + \theta_a H' $ |
| Inverse of the mean concentration at the center of a 0.5-acre square source in Los Angeles | Q/C _{vol} | 73.44 g/m ² -s per kg/r | 1 ³ | |
| Dry soil bulk density | ρ_b | 1.5 g/cm ³ | | |
| Air-filled soil porosity | θ_{a} | 0.28 L _{air} /L _{soil} | | |
| Water-filled soil porosity | $\theta_{\rm w}$ | $0.15 L_{water}/L_{soil}$ | | |
| Exposure interval, 30-yr resident | T _{Resident} | 2.80E+08 seconds | | |
| Volatilization Factor for soil | VFs | derived m ³ /kg | | |

| Volatilization Factor for soil | VFs | deriv | ed m ³ /kg | | | | | | | | | | |
|---|------|-------|-----------------------------|---------------|---------------------------|-----------------|---|----------------------|--|----------------|---------------------------|--------------------|--------------------|
| | | | | | | | | | | | | VFs (VOCs) |) PEF |
| | | | Di | | D _w | | Н' | | K _{OC} | K _d | \mathbf{D}_{A} | Residential | (non-VOCs) |
| Analyte | (c | n²/s) | Reference | (cm^2/s) | Reference | (dimensionless) | Reference | (cm ³ /g) | Reference | (cm^3/g) |) (cm^{2}/s) | m ³ /kg | m ³ /kg |
| Inorganic Constituents | | | | | | | | | | | | | |
| Arsenic | | 0 | 0 | 0 0 | | 31.61052561 | Bond Method estimate, EPISuite v4.10 (USEPA, 2011a) | 13.22 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 29 | 0 | non-VOC | 1.360E+09 |
| Chromium (III) | | 0 | 0 | 0 0 | | 0 | 0 | 0 | 0 | 180000 | 0 0 | non-VOC | 1.360E+09 |
| Lead | | 0 | 0 | 0 0 | | 1.001885999 | Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 900 | 0 | non-VOC | 1.360E+09 |
| Mercury | 0. | 0307 | Exhibit C-1, USEPA 2002e | 0.0000063 Exh | ibit C-1, USEPA 2002e | 0.467116057 | RSL ("SSL"; USEPA 2012) | 13.22 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 52 | 1.42E-05 | 19242.4825 | use VF |
| Polychlorinated Biphenyls (PCBs) - Mixtures | | | | | | | | | | | | | |
| Total PCBs | 0.0 | 40076 | RSLs ("WATER9"; USEPA 2012) | 4.6826E-06 RS | Ls ("WATER9"; USEPA 2012) | 0.011572806 | Experimental, EPISuite v4.10 (USEPA 2011a) | 130500 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 0.000235 | non-VOC | 1.360E+09 |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | | | |
| Naphthalene | 0.0 | 04994 | RSLs ("WATER9"; USEPA 2012) | 8.377E-06 RS | Ls ("WATER9"; USEPA 2012) | 0.017993055 | Experimental, EPISuite v4.10 (USEPA 2011a) | 1544 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 0.000546 | 3106.7736 | use VF |
| Pesticides | | | | | | | | | | | | | |
| MCPP | 0.03 | 29988 | RSLs ("WATER9"; USEPA 2012) | 6.1925E-06 RS | Ls ("WATER9"; USEPA 2012) | 3.66404E-08 | Experimental, EPISuite v4.10 (USEPA 2011a) | 48.51 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 4.01E-07 | non-VOC | 1.360E+09 |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | |
| Benzo[a]anthracene | 0.03 | 08647 | RSLs ("WATER9"; USEPA 2012) | 5.9431E-06 RS | Ls ("WATER9"; USEPA 2012) | 0.00049072 | Experimental, EPISuite v4.10 (USEPA 2011a) | 176900 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 1.33E-05 | non-VOC | 1.360E+09 |
| Benzo[a]pyrene | 0.04 | 75831 | RSLs ("WATER9"; USEPA 2012) | 5.5597E-06 RS | Ls ("WATER9"; USEPA 2012) | 1.86882E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 587400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 8.2E-07 | non-VOC | 1.360E+09 |
| Benzo[b]fluoranthene | 0.04 | 75831 | RSLs ("WATER9"; USEPA 2012) | 5.5597E-06 RS | Ls ("WATER9"; USEPA 2012) | 2.68669E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 599400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 1.02E-06 | non-VOC | 1.360E+09 |
| Benzo[g,h,i]perylene | | 0 | 0 | 0 0 | | 1.35357E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 1951000 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 0 | non-VOC | 1.360E+09 |
| Benzo[k]fluoranthene | 0.04 | 75831 | RSLs ("WATER9"; USEPA 2012) | 5.5597E-06 RS | Ls ("WATER9"; USEPA 2012) | 2.38817E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 587400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 9.48E-07 | non-VOC | 1.360E+09 |
| Benzofluoranthenes, Total | 0.04 | 75831 | RSLs ("WATER9"; USEPA 2012) | 5.5597E-06 RS | Ls ("WATER9"; USEPA 2012) | 2.68669E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 599400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 1.02E-06 | non-VOC | 1.360E+09 |
| Dibenz[a,h]anthracene | 0.04 | 45672 | RSLs ("WATER9"; USEPA 2012) | 5.2073E-06 RS | Ls ("WATER9"; USEPA 2012) | 5.76596E-06 | Experimental, EPISuite v4.10 (USEPA 2011a) | 1912000 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0 | 4.7E-07 | non-VOC | 1.360E+09 |
| Indeno[1,2,3-cd]pyrene | 0.04 | 47842 | RSLs ("WATER9"; USEPA 2012) | 5.2327E-06 RS | Ls ("WATER9"; USEPA 2012) | 0.0000656 | EPISuite v4.10 (USEPA 2011a) | 3470000 | EPISuite v4.10 (USEPA 2011a) | 0 | 1.86E-06 | non-VOC | 1.360E+09 |
| | | | | | | | | | | | | | |

Description Variable Value Equations AT_{c} 70 yrs Cancer Risk: Averaging Time, carcinogens Averaging Time, noncarcinogens, resident AT_{nc.r} 9 yrs Nonmutagens:
$$\begin{split} Risk &= IUR \times C_{air} \times \left(\frac{ED_r \times EF_r \times ET_r}{AT_c \times 365 \frac{day}{year} \times 24 \frac{hour}{day}} \right) \\ Risk &= IUR \times \left[C_{air} \times EF_r \times ET_r \right. \\ &\quad \times \left\{ \frac{ED_{0-2} \times ADAF_{0-2}}{AT_c \times 365 \ days/year \times 24 \ hours/day} + \frac{ED_{2-6} \times ADAF_{2-6}}{AT_c \times 365 \ days/year \times 24 \ hours/day} + \frac{ED_{16-26} \times ADAF_{16-26}}{AT_c \times 365 \ days/year \times 24 \ hours/day} + \frac{ED_{16-26} \times ADAF_{16-26}}{AT_c \times 365 \ days/year \times 24 \ hours/day} \right\} \end{split}$$
Age-dependent Adjustment Factor, 0-2 $ADAF_{0-2}$ 10 yrs Age-dependent Adjustment Factor, 2-6 $ADAF_{2-6}$ 3 yrs Age-dependent Adjustment Factor, 6-16 $ADAF_{6-16}$ 3 yrs Age-dependent Adjustment Factor, 16-26 Mutagens: ADAF₁₆₋₂₆ 1 yrs Averaging Time, carcinogens 70 yrs AT_{c} COPC Concentration in air C_{air} chemical-specific $\mu g/m^3$ COPC Concentration in soil C_{soil} chemical-specific mg/kg Exposure Duration, child 0-2 ED_{0-2} 2 yrs ED_{2-6} Exposure Duration, child 2-6 4 yrs Vinyl Chloride: $Risk = IUR \times \left[C_{air} \times \left\{ \frac{ED_r \times EF_r \times ET_r}{AT_r \times 365 \ days/year \times 24 \ hours/day} + 1 \right\} \right]$ ED₆₋₁₆ Exposure Duration, child 6-16 10 yrs Exposure Duration, adult 16-26 ED₁₆₋₂₆ 10 yrs Exposure Duration, child ED_c 6 yrs Exposure Duration, resident ED_r 9 yrs **Exposure Frequency** EF_r 152 days/yr Exposure Time ET_r 24 hours/day Noncancer Hazard: $Hazard = \frac{1}{RfC} \times C_{air} \times \left(\frac{ED_c \times EF_r \times ET_r}{AT_{nc,r} \times 365 \frac{day}{vear} \times 24 \frac{hour}{day}}\right)$ chemical-specific $(\mu g/m^3)^{-1}$ Inhalation Unit-Risk Factor IUR $1.360E+09 \text{ m}^3/\text{kg}$ Particulate emission factor (non-VOCs) PEF chemical-specific µg/m³ Inhalation Reference Concentration RfC $C_{air} = \frac{C}{VF}$ Volatilization Factor for VOCs VF chemical-specific m³/kg Where: VF (VOCs) or Cancer C_{soil} Cair RfC PEF (non-VOCs) IUR Risk % of $(\mu g/m^3)^{-1}$ Analyte Mutagen? (mg/kg) (m^3/kg) $(\mu g/m^3)$ (dimensionless) Total $(\mu g/m^3)$ **Inorganic Constituents** 7.72E-06 1.07E-09 1.50E-02 Arsenic 0.00E+001.05E+011.360E+09 4.30E-03 <1% Chromium (III) 0.00E+00 5.94E+02 1.360E+09 4.37E-04 No Toxicity Value No Toxicity Va -----Lead 0.00E+00 0.00E+001.360E+09 0.00E+00 1.20E-05 0.00E+00 < 1% No Toxicity Val Mercurv 0.00E+00 1.45E+00 1.924E+04 7.54E-02 No Toxicity Value 3.00E-01 -----**Polychlorinated Biphenyls (PCBs) - Mixtures Total PCBs** 0.00E+00 4.10E-01 3.01E-07 5.70E-04 9.20E-12 < 1% 1.360E+09 No Toxicity Va Volatile Organic Compounds (VOCs) Naphthalene 0.00E+00 3.83E+00 3.107E+03 1.23E+00 3.40E-05 2.24E-06 100% 3.00E+00 Pesticides 0.00E+00 MCPP 1.12E+01 1.360E+09 8.24E-06 No Toxicity Value ------No Toxicity Val **Polycyclic Aromatic Hydrocarbons (PAHs)** Benzo[a]anthracene Μ 9.96E+00 1.360E+09 7.32E-06 1.10E-04 3.45E-10 < 1% No Toxicity Val Μ Benzo[a]pyrene 1.12E+01 1.360E+09 8.24E-06 1.10E-03 3.88E-09 < 1% No Toxicity Va Benzo[b]fluoranthene Μ 1.52E+01 1.360E+09 1.12E-05 1.10E-04 5.27E-10 < 1% No Toxicity Va Benzo[g,h,i]perylene 0.00E+00 5.82E+00 1.360E+09 4.28E-06 No Toxicity Value No Toxicity Va -----Benzo[k]fluoranthene Μ 2.63E+01 1.360E+09 1.93E-05 1.10E-04 9.11E-10 < 1% No Toxicity Val М Benzofluoranthenes, Total 1.88E+01 1.360E+09 1.38E-05 1.10E-04 6.51E-10 < 1% No Toxicity Va

Table A-6.3 Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Landfill AOPC

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| Description | Variable | Valu | ie | Equations | | | | | | | |
|--|-----------------------|-------------------|-----------------------|------------------|--|---|--|--|---|---------------|---------------------|
| Averaging Time, carcinogens | AT _c | 7 | '0 yrs | Cancer Risk: | | | | | | | |
| Averaging Time, noncarcinogens, resident | AT _{nc,r} | | 9 yrs | Nonmutagens: | | | | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 1 | 0 yrs | | $Dick = IIID \times C$ | $\sim \int \frac{ED_r \times ED_r}{ED_r}$ | $EF_r \times ET_r$ | | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | | 3 yrs | | $\pi i s \kappa = I 0 \pi \wedge C_{ai}$ | $\frac{1}{AT \times 365}$ | $\frac{day}{x 24}$ hour | | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | | 3 yrs | | 1 | $\begin{pmatrix} m_c \times 505 \end{pmatrix}$ | year ^{~ 2 °} day / | | | | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | | 1 yrs | Mutagens: | $Risk = IUR \times C_{ai}$ $Risk = IUR \times \left[C_{ai}\right]$ | $T_r \times EF_r \times ET_r$ | | | | | |
| Averaging Time, carcinogens | AT_{c} | 7 | 0 yrs | | L | FD _o | $\times ADAF_{a}$ | FD | $a \land X A D A F_{a} \land$ | | |
| COPC Concentration in air | C_{air} | chemical-specif | ic μg/m ³ | | × { | $\frac{LD_{0-2}}{AT}$ | $\frac{2}{10000} \times \frac{2}{100000000000000000000000000000000000$ | $\frac{ED}{day} + \frac{ED}{AT_c \times 365 \ da}$ | $\frac{2-6}{2-6} \times \frac{10}{2-6}$ | a / day | |
| COPC Concentration in soil | C_{soil} | chemical-specif | ic mg/kg | | (| $AI_c \land 303 uuys / FD$ | $\sim ADAF$ | $H_c \sim 303 \text{uu}$ | \sqrt{D} | 3/ <i>uuy</i> | |
| Exposure Duration, child 0-2 | ED ₀₋₂ | | 2 yrs | | + - | $\frac{\nu}{4T} \times 26E dame / 2$ | $\sim \pi \nu \pi r_{6-16}$ | $\frac{ED_{16-}}{AT_c \times 365 \ day}$ | $26 \land \Pi \cup \Pi^{1} 16 - 26$ | <u></u> } | |
| Exposure Duration, child 2-6 | ED ₂₋₆ | | 4 yrs | | | $a_{1c} \times 305 \ aays/j$ | yeur × 24 nours/a | uy AI _c × 305 aay | syyeur × 24 nours | s/aay)] | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 1 | 0 yrs | Vinyl Chloride: | : ſ | (F | $7D \times FF \times FT$ |)] | | | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 1 | 0 yrs | | $Risk = IUR \times C_{ai}$ | $r \times \left\{ \frac{1}{AT} \times 26E \right\}$ | $\frac{D_r \times D_r}{2} \times \frac{D_r}{2} \times \frac{D_r}{2}$ | $\frac{1}{1}$ | | | |
| Exposure Duration, child | ED_{c} | | 6 yrs | | L | $(AI_c \times 505 u)$ | uys/yeur x 24 nou | irs/uuy)] | | | |
| Exposure Duration, resident | ED_r | | 9 yrs | | | | | | | | |
| Exposure Frequency | EF_r | 15 | 2 days/yr | | | | | | | | |
| Exposure Time | ET_r | 2 | 4 hours/day | Noncancer Haza | ard: | / 50 | $\nabla F F \nabla F T$ | \ | | | |
| Inhalation Unit-Risk Factor | IUR | chemical-specif | ic $(\mu g/m^3)^{-1}$ | | $Hazard = \frac{1}{RfC} \times C$ | $C_{air} \times \left($ | $\frac{dav}{dav}$ how | .) | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+0 | 9 m ³ /kg | | ĸju | $AT_{nc,r} \times 1$ | $365 \frac{duy}{vear} \times 24 \frac{h0u}{day}$ | -] | | | |
| Inhalation Reference Concentration | RfC | chemical-specif | ic µg/m ³ | | | Υ. | year aay | / | | | |
| Volatilization Factor for VOCs | VF | chemical-specif | ic m [°] /kg | Where: | | | | $C_{air} = \frac{C_{soil}}{VF \text{ or } P}$ | $\frac{\mu g}{ma} \times 1000 \frac{\mu g}{ma}$ | | |
| | | | VF (VOCs) or | | | Cancer | | | Noncancer | | Noncancer |
| | | C _{soil} | PEF (non-VOCs) | C _{air} | IUR | Risk | % of | RfC | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total | (dimensionless) |
| Dibenz[a,h]anthracene | М | 2.06E+00 | 1.360E+09 | 1.51E-06 | 1.20E-03 | 7.79E-10 | < 1% | No Toxicity Value | | | |
| Indeno[1,2,3-cd]pyrene | М | 6.80E+00 | 1.360E+09 | 5.00E-06 | 1.10E-04 | 2.36E-10 | < 1% | No Toxicity Value | | | |
| | | | | | | Cancer Risk | | | Hazard Index (Child) | | Hazard Index (Adult |
| | | | | | Pathway Sum: | 2.3E-06 | | Pathway Sum: | 0.28 | | 0.28 |

Table A-6.3Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Landfill AOPC

 Table A-6.4

 Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: CTE, Landfill AOPC

| Description | Variable | Value | 2 | Equations | : | | | | | |
|--|-------------------------|-------------------|-------------------------|---------------------------|---|--|--|---|------------------|-----------------------------|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specific | unitless | Cancer Risk: | | | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 |) yrs | Nonmutagens | Dial - SE V | | | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 | yrs | | ü | 2 3011 | u | $EC \times ABS_d \times CF_d$ | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 | 3 yrs | | ×(| $4F_a \times SA_a$ | | $\frac{AF_c \times ED_c \times SA_c}{\langle AT_c \times 365 \ dav/\gamma}$ |) | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 | yrs | | $(BW_a \times AT)$ | $T_c \times 365 da$ | ay/year ' BW _c > | $\times AT_c \times 365 day/y$ | vear]] | |
| Soil-to-Skin Adherence Fraction | AF_a | 0.01 | mg/cm ² -day | Mutagens | | | | | | |
| Soil-to-Skin Adherence Fraction | AF_{c} | 0.04 | mg/cm ² -day | Risk = SF | $d_d \times \left\{ C_{soil} \times EF_r \right\}$ | $\times EV \times EC$ | $C \times ABS_d \times CF_d$ | | | |
| Averaging Time, Carcinogens | AT_{c} | 70 |) yrs | | (| | | | VADAE | |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 3 | s yrs | | $\times \left \frac{ED_{0-2} \times D_{0-2}}{D} \right $ | $AF_{c} \times SA_{c}$ | $\frac{1}{2} \times ADAF_{0-2} + \frac{L}{2}$ | $\frac{D_{2-6} \times AF_c \times SA_c}{BW_c \times AT_c \times 365}$ | X ADAF | 2-6 |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | e | 5 yrs | | | | | | | |
| Body Weight, adult | BW_a | 80 |) kg | | $+\frac{ED_{6-16}}{}$ | $\langle AF_a \times SA$ | $a \times ADAF_{6-16} +$ | $\frac{ED_{16-26} \times AF_a \times$ | $SA_a \times AL$ | DAF_{16-26} |
| Body Weight, child | BW_c | 15 | 5 kg | | $BW_a \times$ | $AT_c \times 365$ | days/year | $BW_a \times AT_c \times 3$ | 65 days | /year]) |
| Conversion Factor | CF_d | | 5 kg/mg | Vinyl Chloride | $Risk = SF_{a} \times$ | $\{C_{coil} \times E\}$ | $F_r \times EV \times EC \times A$ | $ABS_{d} \times CF_{d}$ | | |
| COPC Concentration in Soil | C _{soil} | chemical-specific | 0 0 | | | C-SULL II | , . <u></u> <u>_</u> <u>_</u> <u>_</u> <u>_</u> | - <i>a a</i> | | |
| Fraction of EV in Contact with Soil | EC | | unitless | | $EE_{r} \times (\frac{ED}{r})$ | $c \times AF_c \times S$ | $\frac{SA_c}{A_c} + \frac{ED_a \times AF_a}{ED_a \times AF_a}$ | $(\times SA_a)$ | .]) | |
| Exposure Duration, child 0-2 | ED ₀₋₂ | | 2 yrs | | × | BW_c | BWa | $\left(\frac{AF_c \times SA_a}{BW_c}\right) + \frac{AF_c \times SB_c}{BW_c}$ | A_c | |
| Exposure Duration, child 2-6 | ED ₂₋₆ | | yrs | | | $AT_c \times$ | $365 \frac{days}{vaar}$ | BW_c | | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | |) yrs | | L | Ũ | yeur | | J) | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | |) yrs | NY YY | | | | | | |
| Exposure Duration, adult | ED _a | | yrs | Noncancer Haza | ra: | | | | | |
| Exposure Duration, child | ED _c | | 5 yrs | Н | $nzard = \frac{1}{m} \times [$ | $C_{1} \times \left(\frac{AF_0}{2}\right)$ | $\times ABS_d \times ED_{nc,c} \times EF_r$ | $\times EV \times EC \times SA_c \times CF_d$ | | |
| Exposure Frequency, resident | EF _r EV | | 2 days/yr | | RfD _d | | $BW_c \times AT_{nc,c} \times$ | $\left(\frac{XEV \times EC \times SA_c \times CF_d}{365 \frac{day}{year}}\right)$ | | |
| Event Frequency Reference Dose Adjusted for GI Absorption | E v RfD _d | chemical-specific | events/day | | | | | - | | |
| Exposed Body Surface Area, adult | SA _a | | 2 cm^2 | | | | | | | |
| Exposed Body Surface Area, child | SA _c | 2373 | - | | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF _d | chemical-specific | | | | | | | | |
| Of al Slope Pactor Adjusted for Of Absorption | 31 _d | enemical-specific | (iiig/kg-uuy) | | <i>a</i> | | | | | |
| | | C _{soil} | ABS _d | SF_d | Cancer Risk | % of | RfD _d | Noncancer | % of | Noncancer Hazard (Adult) |
| | 34 4 9 | | - | - | | | - | Hazard (Child) | | ``´´ |
| Analyte | Mutagen? | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total | (dimensionless) |
| Inorganic Constituents | | | | | | | | | | |
| Arsenic | 0.00E+00 | 1.05E+01 | 3.00E-02 | 1.50E+00 | 6.79E-08 | < 1% | 3.00E-04 | 1.66E-03 | 12% | 1.98E-04 |
| Chromium (III) | 0.00E+00 | 5.94E+02 | 1.00E-02 | | | | 1.95E-02 | 8.03E-04 | 6% | 9.56E-05 |
| Lead | 0.00E+00 | 0.00E+00 | 1.00E-02 | 8.50E-03 | 0.00E+00 | < 1% | | | | |
| Mercury | 0.00E+00 | 1.45E+00 | 1.00E-02 | | | | 1.60E-04 | 2.39E-04 | 2% | 2.85E-05 |
| Polychlorinated Biphenyls (PCBs) - Mixtures | | | | | | | | | | |
| Total PCBs | 0.00E+00 | 4.10E-01 | 1.40E-01 | 2.00E+00 | 2.75E-08 | < 1% | 2.00E-05 | 7.56E-03 | 57% | 9.01E-04 |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | |
| Naphthalene | 0.00E+00 | 3.83E+00 | 1.30E-01 | | | | 2.00E-02 | 6.56E-05 | < 1% | 7.82E-06 |
| Pesticides | | | | | | | | | | |
| МСРР | 0.00E+00 | 1.12E+01 | 1.00E-01 | | | | 1.00E-03 | 2.95E-03 | 22% | 3.52E-04 |

 Table A-6.4

 Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: CTE, Landfill AOPC

| Description | Variable | Value | | Equations: | | | | | | | | |
|--|-----------------------|----------------------|--------------------------|---|---|------------------------------------|---|---|------------------|--------------------|--|--|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specific un | itless | Cancer Risk: | | | | | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 yrs | 5 | Nonmutagens: | Dials - SE V | | N EE N EU N | $EC \times ABS_d \times CF_d$ | | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 yrs | 5 | | | | | | | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 yrs | 5 | $\times \left(\frac{AF_a \times SA_a}{BW_c \times AT_c \times 365 \ day/year} + \frac{AF_c \times ED_c \times SA_c}{BW_c \times AT_c \times 365 \ day/year} \right)$ | | | | | | | | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 yrs | 5 | | $(BW_a \times AT)$ | $T_c \times 365 dag$ | $y/year' BW_c \times$ | $AT_c \times 365 \ day/y$ | rear /] | | | |
| Soil-to-Skin Adherence Fraction | AF_a | 0.01 mg | g/cm ² -day | Mutagens: | (| | | | | | | |
| Soil-to-Skin Adherence Fraction | AF_{c} | 0.04 m | g/cm ² -day | $Risk = SF_d$ | $\times \{C_{soil} \times EF_r\}$ | \times EV \times EC | $\times ABS_d \times CF_d$ | | | | | |
| Averaging Time, Carcinogens | AT_{c} | 70 yr: | 5 | | (| | | | | , | | |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 3 yrs | 5 | | $\times \left \frac{ED_{0-2} \times D_{0-2}}{D_{0-2}} \right $ | $AF_c \times SA_c$ | $\frac{\times ADAF_{0-2}}{1} + \frac{EI}{1}$ | $\frac{D_{2-6} \times AF_c \times SA_c}{BW_c \times AT_c \times 365 \ a}$ | X ADAF | 2-6 | | |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | 6 yrs | 5 | | | | | | | | | |
| Body Weight, adult | BW_a | 80 kg | | | $+\frac{ED_{6-16}}{2}$ | $\langle AF_a \times SA_a \rangle$ | $\times ADAF_{6-16} + 1$ | $\frac{ED_{16-26} \times AF_a \times S}{BW_a \times AT_c \times 3}$ | $A_a \times A_b$ | DAF_{16-26} | | |
| Body Weight, child | BW_c | 15 kg | | | $BW_a \times$ | $AT_c \times 365$ | days/year | $BW_a \times AT_c \times 3$ | 65 days | /year]) | | |
| Conversion Factor | CF_d | 1E-06 kg | /mg | Vinyl Chloride: | Risk - SF. X | \$C × FF | $X \times EV \times EC \times A$ | $RS \cdot \times CF$ | | | | |
| COPC Concentration in Soil | C _{soil} | chemical-specific ma | g/kg | | $max = 3I_d \land$ | Usoil ~ Er | | | | | | |
| Fraction of EV in Contact with Soil | EC | 1 un | itless | | $\begin{bmatrix} EE \\ \hline \end{bmatrix} \begin{pmatrix} ED \end{pmatrix}$ | $_c \times AF_c \times S$ | $A_{c} \perp ED_{a} \times AF_{a}$ | $\times SA_a$ | . 1) | | | |
| Exposure Duration, child 0-2 | ED_{0-2} | 2 yrs | 5 | | \times | BW_c | $+$ BW_a | $\frac{1}{1} + \frac{AF_c \times S}{1}$ | A_c | | | |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 4 yrs | 5 | | ^ | $AT_{-} \times 3$ | 365 <u>days</u> | $\frac{(X - SA_a)}{BW_c} + \frac{AF_c \times S}{BW_c}$ | | | | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 yrs | 5 | | L | | year year | | IJ | | | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 yrs | | | | | | | | | | |
| Exposure Duration, adult | ED_a | 3 yrs | | Noncancer Hazar | d: | | | | | | | |
| Exposure Duration, child | ED _c | 6 yr: | | IJa | -1 | $c (AF_c)$ | <abs<sub>d×ED_{nc,c}×EF_r×</abs<sub> | $EV \times EC \times SA_c \times CF_d$ | | | | |
| Exposure Frequency, resident | EF_r | 152 da | | на | $ara = \frac{1}{RfD_d} \times \left[\frac{1}{RfD_d} \right]$ | L _{soil} × (| $BW_c \times AT_{nc,c} \times 3$ | $\frac{EV \times EC \times SA_c \times CF_d}{B65 \frac{day}{vaar}}$ | | | | |
| Event Frequency | EV | | ents/day | | L. | , | | yeur , 1 | | | | |
| Reference Dose Adjusted for GI Absorption | RfD _d | chemical-specific mg | | | | | | | | | | |
| Exposed Body Surface Area, adult | SA_a | 6032 cn | | | | | | | | | | |
| Exposed Body Surface Area, child | SA_c | 2373 cm | 1 | | | | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specific (m | ng/kg-day) ⁻¹ | | | | | | | | | |
| | | | | | Cancer | | | Noncancer | | Noncancer | | |
| | | C _{soil} | ABS _d | SF_d | Risk | % of | $\mathbf{RfD}_{\mathbf{d}}$ | Hazard (Child) | % of | Hazard (Adult) | | |
| Analyte | Mutagen? | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total | (dimensionless) | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | |
| Benzo[a]anthracene | М | 9.96E+00 | 1.30E-01 | 7.30E-01 | 1.31E-06 | 5% | | | | | | |
| Benzo[a]pyrene | М | 1.12E+01 | 1.30E-01 | 7.30E+00 | 1.47E-05 | 60% | | | | | | |
| Benzo[b]fluoranthene | М | 1.52E+01 | 1.30E-01 | 7.30E-01 | 2.00E-06 | 8% | | | | | | |
| Benzo[g,h,i]perylene | 0.00E+00 | 5.82E+00 | 1.30E-01 | | | | 3.00E-02 | 6.65E-05 | < 1% | 7.92E-06 | | |
| Benzo[k]fluoranthene | М | 2.63E+01 | 1.30E-01 | 7.30E-02 | 3.45E-07 | 1% | | | | | | |
| Benzofluoranthenes, Total | М | 1.88E+01 | 1.30E-01 | 7.30E-01 | 2.47E-06 | 10% | | | | | | |
| Dibenz[a,h]anthracene | М | 2.06E+00 | 1.30E-01 | 7.30E+00 | 2.71E-06 | 11% | | | | | | |
| Indeno[1,2,3-cd]pyrene | М | 6.80E+00 | 1.30E-01 | 7.30E-01 | 8.93E-07 | 4% | | | | | | |
| | | | | | Cancer Risk | | ŀ | Hazard Index (Child) | | Hazard Index (Adul | | |
| | | | | Pathway Sum: | 2.5E-05 | | Pathway Sum: | 0.013 | | 0.0016 | | |
| | | | | ,, | | | , <i></i> , , | | | | | |

 Table A-7.1

 Risk and Hazard Estimates - Ingestion of Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC

| Definitions | Variable | Value | | Equations | | | | |
|--|-----------------------------|--------------------|---------------------------|------------------------|------------------------------|---|--------------------------------------|-------|
| Averaging Time, carcinogens | AT _c | 70 | yrs | Carcinogenic: | | | | |
| Averaging Time, noncarcinogens, worker | AT _{nc,w} | 1 | yrs | Г | | | | |
| Body Weight, adult | \mathbf{BW}_{a} | 80 | kg | $Risk = SF_{a}$ | $\times C_{water} \times$ | $\frac{IRW_{w} \times EF_{w} \times ED_{w}}{BW_{a} \times AT_{c} \times 365}$ | $\times FI_{w} \times CF_{o}$ | |
| Conversion Factor, ingestion | CFo | 1.0E-03 | mg/µg | 0 | water | $BW_a \times AT_c \times 365$ | day/year) | |
| COPC Concentration in Water | Cwater | chemical-specific | µg/L | | | | | |
| Exposure Duration, worker | ED_{w} | 1 | yrs | | | | | |
| Exposure Frequency, worker | EF_{w} | 9 | days/yr | Noncarcinogenic: | | | | |
| Fraction Contaminated Water Ingested, worker | FI_{w} | | unitless | | Г | (| | ν |
| Water Ingestion Rate, worker | $\mathrm{IRW}_{\mathrm{w}}$ | | L/day | Hazard = - | $\frac{1}{C} \times C_{max}$ | $_{ver} \times \left(\frac{IRW_{w} \times EF_{w} \times BW_{a} \times AT_{nc,w}}{BW_{a} \times AT_{nc,w}}\right)$ | $ED_{w} \times FI_{w} \times CF_{o}$ | |
| Oral Reference Dose | RfD _O | chemical-specific | | R | $2fD_o \mid w^{uu}$ | $BW_a \times AT_{nc,w}$ | × 365 day / year | |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Noncancer | |
| | | C _{water} | SFo | Risk | % of | RfDo | Hazard | % of |
| Analyte | | $(\mu g/L)$ | $(mg/kg-day)^{-1}$ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | |
| Antimony | | 2.23E-01 | No Toxicity Value | | | 4.00E-04 | 3.44E-06 | < 1% |
| Iron | | 2.91E+04 | No Toxicity Value | | | 7.00E-01 | 2.56E-04 | 55% |
| Mercury | | 7.69E-02 | No Toxicity Value | | | 1.60E-04 | 2.96E-06 | < 1% |
| Thallium | | 2.38E-01 | No Toxicity Value | | | 1.00E-05 | 1.47E-04 | 31% |
| Zinc | | 2.07E+03 | No Toxicity Value | | | 3.00E-01 | 4.25E-05 | 9% |
| Phthalates | | | | | | | | |
| di-n-octyl phthalate | | 4.61E+00 | No Toxicity Value | | | 1.00E-02 | 2.84E-06 | < 1% |
| Volatile Organic Compounds (VOCs) | | | | | | | | |
| Acetone | | 1.54E+01 | No Toxicity Value | | | 9.00E-01 | 1.05E-07 | < 1% |
| Chloroform | | 3.70E+00 | 3.10E-02 | 1.01E-11 | 14% | 1.00E-02 | 2.28E-06 | <1% |
| Isopropylbenzene | | 4.60E+00 | No Toxicity Value | | | 1.00E-01 | 2.84E-07 | <1% |
| Tetrachloroethene | | 8.78E+00 | 2.10E-03 | 1.62E-12 | 2% | 6.00E-03 | 9.02E-06 | 2% |
| Vinyl Chloride | | 9.55E-01 | 7.20E-01 | 6.06E-11 | 84% | 3.00E-03 | 1.96E-06 | < 1% |
| n-propylbenzene | | 2.00E+00 | No Toxicity Value | | | 1.00E-01 | 1.23E-07 | < 1% |
| 1,2,4-trimethylbenzene | | 5.20E+00 | No Toxicity Value | | | No Toxicity Value | | |
| | | | Pathway Sum: | Cancer Risk 7.2E-11 | | 1 | Hazard Index 4.7E-04 | |

| Definition | Variable | e Valu | e | Risk and Ha | azard Equations | | | | | |
|--|----------------------------|--------------------|---------------------------|---------------|---------------------------------|---|-------------------------|-------------------------|-------------------------|-------|
| Averaging Time, carcinogens | AT _c | 7 | 0 yrs | Cancer Risk | | | | | | |
| Averaging Time, noncarcinogens, worker | $AT_{nc,w}$ | | 1 yrs | | | Г | (| | $\sim FT$ |) |
| Body Weight, adult | BW_a | 8 | 0 kg | | Risk = URF | $\times C \times$ | | | | _ |
| COPC Concentration in air | C _{air} | chemical-specifi | c μg/m ³ | | | ur | $\left(AT \right)_{c}$ | \times 365 days | / year $\times 24$ | 4) |
| COPC Concentration in water | C _{water} | chemical-specifi | c μg/L | Noncancer I | Hazard: | _ | | | | _ |
| Exposure Duration, worker | ED_{w} | | 1 yrs | | 1 | [(| Fl | $F \times FD \times FT$ |) | |
| Exposure Frequency, worker | EF_{w} | | 9 days/yr | 1 | $Hazard = \frac{1}{RfC} \times$ | $ C_{air} \times $ | | | | |
| Exposure Time, worker | ET_{w} | 4.0 | 0 hours | | RfC | $\begin{bmatrix} AI_{nc,w} \end{bmatrix}$ | ×365 <i>d</i> | ays/ year×24ho | ours/day | |
| Air Inhalation Rate, worker | IRA_w | 2 | 0 m ³ /workday | 7 | | | | | | |
| Inhalation Reference Concentration | RfC | chemical-specifi | c μg/m ³ | | | | | | | |
| Inhalation Unit-Risk Factor | URF | chemical-specifi | $c (\mu g/m^3)^{-1}$ | where: | $C_{air} = C_{water} \times V$ | Æ | | | | |
| Volatilization Factor | VF | chemical-specifi | c L/m ³ | | $C_{air} = C_{water} \times V$ | Γ | | | | |
| | | | | | | Cancer | | | Hazard | |
| | | C _{water} | VF | C_{air} | URF _i | Risk | % of | RfC | Quotient | % of |
| Analyte | | $(\mu g/L)$ | (L/m^3) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | | | |
| Phthalates | | | | | | | | | | |
| di-n-octyl phthalate | | 4.61E+00 | non-VOC | | No Toxicity Value | | | No Toxicity Value | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | |
| Acetone | | 1.54E+01 | 4.12E+00 | 6.34E+01 | No Toxicity Value | | | 3.10E+04 | 8.41E-06 | < 1% |
| Chloroform | | 3.70E+00 | 9.35E+00 | 3.46E+01 | 2.30E-05 | 4.67E-08 | 94% | 9.80E+01 | 1.45E-03 | 4% |
| Isopropylbenzene | | 4.60E+00 | 9.55E+00 | 4.39E+01 | No Toxicity Value | | | 4.00E+02 | 4.51E-04 | 1% |
| Tetrachloroethene | | 8.78E+00 | 7.53E+00 | 6.61E+01 | 2.60E-07 | 1.01E-09 | 2% | 4.00E+01 | 6.79E-03 | 20% |
| Vinyl Chloride | | 9.55E-01 | 8.31E+00 | 7.93E+00 | 4.40E-06 | 2.05E-09 | 4% | 1.00E+02 | 3.26E-04 | < 1% |
| n-propylbenzene | | 2.00E+00 | 8.46E+00 | 1.69E+01 | No Toxicity Value | | | 1.00E+03 | 6.96E-05 | < 1% |
| 1,2,4-trimethylbenzene | | 5.20E+00 | 8.36E+00 | 4.35E+01 | No Toxicity Value | | | 7.00E+00 | 2.55E-02 | 74% |
| | | | | | Pathway Sums: | Cancer Risk 5.0E-08 | | | Hazard Index 3.5E-02 | |
| | | | | | | | | | | |

 Table A-7.2

 Risk and Hazard Estimates - Inhalation of Volatile Compounds from Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC

 Table A-7.3

 Estimation of Kp, Tau(event), B, and t* for Organic Compounds: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC

| Definition | Variable | Value | | Equations | | | | | |
|--|-----------------|--------------------------------|---------------------|--|--------------------------------|--|----------|-----------------|----------|
| Permeability coefficient from water | K _p | Calculated (Equation 1) | cm/hr | ¹⁾ $V = 10^{(-1)}$ | $2.8 + 0.66 Log K_{OW}$ | -0.0056 MW) | | | |
| Octanol:water partition coefficient | Kow | Chemical-specific | dimensionless | $\mathbf{\Lambda}_{p} = 10^{\circ}$ | | | | | |
| Molecular weight | MW | Chemical-specific | g/mole | 2) $(1)^{2}$ | 2 | | | | |
| Lag time per event | τ_{event} | Calculated (Equation 2) | hr/event | 2) $\tau_{event} = \frac{(l_{sc})^2}{6D_{sc}}$ | _ | | | | |
| Thickness of the strateum corneum | l _{sc} | 0.001 | cm | $6D_{sc}$ | | | | | |
| Effective diffusion coefficient, through the stratum | 30 | | | | | | | | |
| corneum | D _{sc} | Calculated (Equation 2) | cm ² /hr | where: D | $= l \times 10^{(-1)}$ | $2.8 - 0.0056 \ MW$) | | | |
| Relative contribution of permeability coefficients in | B | Calculated (Equation 2) | | | v _{sc} ··· 1 0 | | | | |
| stratum corneum and viable epidermis | 2 | Calculated (Equation 5) | unionorioneos | | | | | | |
| Time it takes to reach steady state | t* | Calculated (Equation 4) | hr | 3) [| 1/11/ | | | | |
| Correlation coefficient | b | Calculated (Equation 4) | | $B = K_p \frac{\sqrt{2}}{2}$ | | (as an approximation | on) | | |
| Correlation coefficient | с | Calculated (Equation 4) | dimensionless | P | 2.6 | | , | | |
| | | | | ⁴⁾ If $B \le 0$. | $6 \cdot t^* = 2/$ | 1τ | | | |
| | | | | | | | | | |
| | | | | | | $e_{vent} \left(b - \sqrt{b^2} \right)$ | 2) | | |
| | | | | If $B > 0$. | $6: t^* = 6\tau$ | $event \left(b - \sqrt{b^2} \right)$ | $-c^{2}$ | | |
| | | | | | | | | | |
| | | | | where: $c = \frac{1+3}{3}$ | $3B + 3B^2$ | | | | |
| | | | | 3: | (1 + B) | | | | |
| | | | | 2 ~ | $(1 + R)^2$ | | | | |
| | | | | $b = \frac{2 \times 1}{2}$ | $\frac{(1+B)^2}{c} - c$ | | | | |
| | | | | | π | | | | |
| | | MW | LogKow | Kp ^a | τ _{event} | В | с | b | t* |
| Analyte | | (g/mole) | (dimensionless) | (cm/hr) | (hr/event) | (dimensionless) | | (dimensionless) | (hr) |
| Inorganic Constituents | | (8, | () | (0.02.00.) | (, | (| (| (| () |
| Phthalates | | | | | | | | | |
| di-n-octyl phthalate | | 390.57 | 8.10E+00 | NRP | 1.62E+01 | | | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | |
| Acetone | | 58.08 | -2.40E-01 | 5.20E-04 | 2.22E-01 | 1.53E-03 | 3.34E-01 | 3.04E-01 | 5.34E-01 |
| Chloroform | | 119.38 | 1.97E+00 | 6.79E-03 | 4.90E-01 | 2.85E-02 | 3.53E-01 | 3.21E-01 | 1.18E+00 |
| Isopropylbenzene | | 120.2 | 3.66E+00 | 8.76E-02 | 4.95E-01 | 3.69E-01 | 6.13E-01 | 5.81E-01 | 1.19E+00 |
| Tetrachloroethene | | 165.83 | 3.40E+00 | 3.28E-02 | 8.92E-01 | 1.62E-01 | 4.49E-01 | 4.11E-01 | 2.14E+00 |
| Vinyl Chloride | | 62.5 | 1.62E+00 | 8.30E-03 | 2.35E-01 | 2.52E-02 | 3.50E-01 | 3.19E-01 | 5.65E-01 |
| n-propylbenzene | | 120.2 | 3.69E+00 | 9.17E-02 | 4.95E-01 | 3.87E-01 | 6.27E-01 | 5.97E-01 | 1.19E+00 |
| 1,2,4-trimethylbenzene | | 120.2 | 3.63E+00 | 8.37E-02 | 4.95E-01 | 3.53E-01 | 5.99E-01 | 5.66E-01 | 1.19E+00 |
| | | | | | | | | | |
| ^a NRP = Not reliably predicted; the compound's chem | ical propert | ies fall outside the Effection | ve Prediction Doma | ain for Kp (Equations | 3.9 and 3.10; USE | EPA, 2004a). | | | |
| | | | | | | | | | |
| = No value | | | | | | | | | |

 Table A-7.4

 Calculation of Dose Absorbed Per Unit Area Per Event: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC

| Definition | Variable | Value | | Equations | | | |
|--|--------------------------------|----------------------|---------------------------------|------------------------------------|---|---|----------------------------|
| Relative contribution of permeability coefficients | В | Chemical-specific | dimensionless | Organics: | | | |
| in stratum corneum and viable epidermis | | | | If $t_{event} \le t^*$, then: | | | |
| Concentration of chemical in water | C _{water} | Measured | μg/L | DA = 2 EA | $\times K \times (C)$ | $CE = \left(\frac{6\tau_{event}}{t_e} \right)$ | event |
| Conversion Factor | CF _d | 1.0E-06 | $(mg L) / (\mu g cm^3)$ | $DA_{event} = 2 PA_{W}$ | $X \times \mathbf{K}_p \times (\mathbf{C}_{water} \wedge \mathbf{C}_{water})$ | CF_d) × $\left(\sqrt{\frac{6\tau_{event} t_e}{\pi}}\right)$ | J |
| Dose absorbed per unit area per event | DA _{event} | Calculated | mg/cm ² -event | If $t_{event} > t^*$, then: | | Г Г | |
| Fraction Absorbed, worker | FAw | Chemical-specific | dimensionless | $DA = EA \times$ | $V \to (C \to CE)$ | tevent 12 7 | $(1+3B+3B^2)$ |
| Permeability coefficient from water | K _p | Chemical-specific | cm/hr | $DA_{event} = \Gamma A_{W} \times$ | $\mathbf{K}_p \times (\mathbf{C}_{water} \times \mathbf{C}\mathbf{F}_a)$ | $(t_{event}) \times \left(\frac{t_{event}}{1+B} + 2\tau_{event} \right)$ | $\frac{(1+B)^2}{(1+B)^2}$ |
| Lag time per event | τ _{event} | Chemical-specific | | | | | |
| Duration of event | t _{event} | 1 | hr/event | Inorganics: | | | |
| Time it takes to reach steady state | t* | Chemical-specific | | 0 | | | |
| The ft takes to reach steady state | L. | Chemieu speenie | | $DA_{event} = K_p \times$ | $(C_{water} \times CF_d) \times t_e$ | vent | |
| | | | | | | | |
| | $\mathbf{FA}_{\mathbf{W}}^{a}$ | $\mathbf{K_{p}}^{b}$ | C _{water} ^c | $\mathbf{\tau_{event}}^{d}$ | t* ^d | \mathbf{B}^{d} | DA _{event} |
| Analyte | (dimensionless) | (cm/hr) | $(\mu g/L)$ | (hr/event) | (<i>hr</i>) | (dimensionless) | $(mg/cm^2 - event)$ |
| Inorganic Constituents | | | | | | | |
| Antimony | | 1.00E-03 | 2.23E-01 | 5.26E-01 | 1.26E+00 | 4.30E-03 | 4.46E-10 |
| Iron | | 1.00E-03 | 2.91E+04 | 2.16E-01 | 5.19E-01 | 2.87E-03 | 5.82E-05 |
| Mercury | | 1.00E-03 | 7.69E-02 | 1.40E+00 | 3.35E+00 | 5.45E-03 | 1.54E-10 |
| Thallium | | 1.00E-03 | 2.38E-01 | 1.47E+00 | 3.52E+00 | 5.50E-03 | 4.76E-10 |
| Zinc | | 6.00E-04 | 2.07E+03 | 2.51E-01 | 6.02E-01 | 1.89E-03 | 2.48E-06 |
| Phthalates | | | | | | | |
| di-n-octyl phthalate | 1 | NRP | 4.61E+00 | 1.62E+01 | | | |
| Volatile Organic Compounds (VOCs) | | | | | | | |
| Acetone | 1 | 5.20E-04 | 1.54E+01 | 2.22E-01 | 5.34E-01 | 1.53E-03 | 1.96E-08 |
| Chloroform | 1 | 6.79E-03 | 3.70E+00 | 4.90E-01 | 1.18E+00 | 2.85E-02 | 7.42E-08 |
| Isopropylbenzene | 1 | 8.76E-02 | 4.60E+00 | 4.95E-01 | 1.19E+00 | 3.69E-01 | 1.12E-06 |
| Tetrachloroethene | 1 | 3.28E-02 | 8.78E+00 | 8.92E-01 | 2.14E+00 | 1.62E-01 | 1.06E-06 |
| Vinyl Chloride | 1 | 8.30E-03 | 9.55E-01 | 2.35E-01 | 5.65E-01 | 2.52E-02 | 1.93E-08 |
| n-propylbenzene | 1 | 9.17E-02 | 2.00E+00 | 4.95E-01 | 1.19E+00 | 3.87E-01 | 5.11E-07 |
| 1,2,4-trimethylbenzene | 1 | 8.37E-02 | 5.20E+00 | 4.95E-01 | 1.19E+00 | 3.53E-01 | 1.22E-06 |

FAr for organic chemicals is from Exhibit B-3, USEPA (2004. RAGS Part E. EPA/540/R/99/005); for chemicals not listed in Exhibit B-3, a default value of 1.0 was used.

K_p for inorganics is from USEPA (2004. RAGS Part E. EPA/540/R/99/005).

 K_p for organics is from Table A-9.3

NRP = Not Reliably Predicted (see Table A-9.3).

from Table A-9.3.

from Table A-9.3.

 Table A-7.5

 Risk and Hazard Estimates: Dermal Contact with Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Landfill AOPC

| Definition | Variable | Value | | Equations | | | | |
|--|----------------------------|---------------------|---------------------------|------------------------|---------------------------|---|---|-------|
| Averaging Time, carcinogens | AT_{c} | 70 | yrs | Carcinogenic: | | | | |
| Averaging Time, noncarcinogens, worker | AT _{nc,w} | 1 | yrs | | | $(EF_w \times ED_w \times I)$ | $EV_{w} \times SA_{w}$ | |
| Body Weight, adult | BW_a | 80 | kg | KISK = SF | $d \times DA_{even}$ | $_{t} \times \left(\frac{EF_{w} \times ED_{w} \times B}{BW_{a} \times AT_{c} \times 36}\right)$ | 5 day / year | |
| COPC Absorbed Dose per Event | DA _{event} | chemical-specific | mg/cm ² -event | | L | | | |
| Exposure Duration, worker | ED_{w} | 1 | yrs | Noncarcinogenic: | 1 | | | , T |
| Exposure Frequency, worker | EF_{w} | 9 | days/yr | Hazard = | $\frac{1}{D(D)} \times D$ | $PA_{event} \times \left(\frac{EF_w \times EI}{BW_a \times AT_{not}}\right)$ | $\frac{D_{w} \times EV_{w} \times SA_{w}}{2}$ | |
| Event Frequency, worker | EV_{w} | 2 | events/day | | K_{fD}_{d} | $(BW_a \times AI_n)$ | $_{x,w} \times 365 \ aay / year$ | ·] |
| Oral Reference Dose Adjusted for GI Absorption | RfD_d | chemical-specific | (mg/kg-day) | | | | | |
| Exposed Body Surface Area, worker | SAa | 5700 | cm ² | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF _d | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Hazard | |
| | | DA _{event} | SFD | Risk | % of | RfD _D | Quotient | % of |
| Analyte | | $(mg/cm^2$ -event) | $(mg/kg-day)^{-1}$ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | |
| Antimony | | 4.46E-10 | | | | 6.00E-05 | 2.61E-05 | 2% |
| Iron | | 5.82E-05 | | | | 7.00E-01 | 2.92E-04 | 23% |
| Mercury | | 1.54E-10 | | | | 1.60E-04 | 3.38E-06 | < 1% |
| Thallium | | 4.76E-10 | | | | 1.00E-05 | 1.67E-04 | 13% |
| Zinc | | 2.48E-06 | | | | 3.00E-01 | 2.91E-05 | 2% |
| Phthalates | | | | | | | | |
| di-n-octyl phthalate | | | | | | 1.00E-02 | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | |
| Acetone | | 1.96E-08 | | | | 9.00E-01 | 7.64E-08 | <1% |
| Chloroform | | 7.42E-08 | 3.10E-02 | 1.15E-10 | 12% | 1.00E-02 | 2.61E-05 | 2% |
| Isopropylbenzene | | 1.12E-06 | | | | 1.00E-01 | 3.95E-05 | 3% |
| Tetrachloroethene | | 1.06E-06 | 2.10E-03 | 1.12E-10 | 12% | 6.00E-03 | 6.22E-04 | 50% |
| Vinyl Chloride | | 1.93E-08 | 7.20E-01 | 6.97E-10 | 75% | 3.00E-03 | 2.26E-05 | 2% |
| n-propylbenzene | | 5.11E-07 | | | | 1.00E-01 | 1.80E-05 | 1% |
| 1,2,4-trimethylbenzene | | 1.22E-06 | | | | | | |
| | | | Pathway Sums | Cancer Risk 9.2E-10 | | | Hazard Index 1.2E-03 | |

Table A-8.1

| Risk and Hazard Estimates: Incidental Ingestion of Soil, Outdoor Maintenance Worker RME, Sandblast Area AOPC |
|--|
|--|

| Definition | Variable | Value | | Equations | | | | |
|---|----------------------------|-------------------|---------------------------|-------------------|----------------------------|--|---|---------------|
| Averaging Time, carcinogens | AT_{c} | 70 | yrs | Cancer Risk: | | | | |
| Averaging Time, noncarcinogens, adult | $AT_{nc,a}$ | 25 | yrs | | ſ | | $D \times FI \times CF$ | J) |
| Body Weight | BW_a | 80 | kg | $Risk = SF_a$ | $\times \{C_{soil} \times$ | $\frac{IRS_{w} \times EF_{w} \times EI}{BW_{a} \times AT_{c} \times 30}$ | $\mathcal{D}_{W} \wedge \Gamma I_{W} \wedge C \Gamma_{O}$ | } |
| Conversion Factor | CFo | 1E-06 | kg/mg | 0 | 301 | $BW_a \times AT_c \times 30$ | 65day/ year) | |
| COPC Concentration in Soil | C _{soil} | chemical-specific | mg/kg | | | _ | | _, |
| Exposure Duration, worker | ED_w | 25 | yrs | Noncancer Hazard: | | | | |
| Exposure Frequency, worker | EF_{w} | 225 | days/yr | | 1 | | | |
| Fraction Contaminated Soil Ingested, worker | FI_{w} | 1.0 | unitless | Hazard=- | $\frac{1}{2} \times C$ | $\left \frac{IRS_{w} \times EF}{BW_{a} \times AT} \right $ | $_{W} \times ED_{W} \times FI_{W}$ | $\times CF_o$ |
| Soil Ingestion Rate, worker | IRS_w | | mg/day | R | $2fD_o$ | $ BW_a \times AT$ | $T_{nc,a} \times 365 day/$ | 'year |
| Oral Reference Dose | RfD ₀ | chemical-specific | mg/kg-day | | , (| | | (L |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Noncancer | |
| | | C _{soil} | SFo | Risk | % of | RfDo | Hazard | % of |
| Analyte | | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | |
| Arsenic | | 9.71E+00 | 1.50E+00 | 2.40E-06 | 27% | 3.00E-04 | 1.50E-02 | 80% |
| Chromium (III) | | 5.79E+02 | No Toxicity Value | | | 1.50E+00 | 2.97E-04 | 2% |
| Lead | | 0.00E+00 | 8.50E-03 | 0.00E+00 | < 1% | No Toxicity Value | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]pyrene | | 2.39E+00 | 7.30E+00 | 4.80E-06 | 53% | No Toxicity Value | | |
| Benzofluoranthenes Total | | 6.36E+00 | 7.30E-01 | 1.28E-06 | 14% | No Toxicity Value | | |
| Dibenz[a,h]anthracene | | 2.38E-01 | 7.30E+00 | 4.78E-07 | 5% | No Toxicity Value | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | |
| Tetrachloroethene | | 2.73E+01 | 2.10E-03 | 1.58E-08 | < 1% | 6.00E-03 | 3.51E-03 | 19% |
| | | | | Cancer Risk | | 1 | Hazard Index | |
| | | | Pathway Sum | : 9.0E-06 | | | 1.9E-02 | |

Table A-8.2

Derivation of Inhalation Factors for Inhalation Exposures, Outdoor Maintenance Worker RME, Sandblast Area AOPC

| Description: | Variable | Value | Units | VF Deriv | ation | | | | | | | | |
|---|---------------------|----------------------|---|----------------------|---|-------------------------|--|------------|--|----------------------|----------------|--------------------|--------------------|
| Unit conversion factor | CFi | 1.00E-04 | m^2/cm^2 | | | (| | | | | | | |
| Apparent Diffusivity | D_A | derived | cm ² /s | I | VFs = (0 / 0) | (3.1) | $\frac{4 \times D_A \times T_{[receptor]}}{(2 \times \rho_b \times D_A)} \times 0$ | CF . | | | | | |
| Diffusivity in air | Di | chemical-specific | $c cm^2/s$ | , | | | $(2 \times \rho_b \times D_A)$ | j, | | | | | ſ |
| Diffusivity in water | D_w | chemical-specific | c cm ² /s | | | | | | | | | | ſ |
| Default organic-carbon content | F _{OC} | 0.01 | g/g | | | | | | | | | | I |
| Henry's law constant | H' | chemical-specific | | | | | | | | | | | I |
| Soil-water partition coefficient for organics: Kd=K _{OC} ×F _{OC} | K _d | chemical-specific | $c \text{ cm}^3/\text{g}$ | based on | : | | | | | | | | |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specific | $c \text{ cm}^3/\text{g}$ | | F (10 / 2 | 10 / 2 |) 1 | | | | | | |
| Total soil porosity | n | - | Lporespace/Lsoil | L | $DA = \frac{\left[\left(\Theta_{a}^{10}\right)^{3} D_{i} H\right]}{\rho_{B} K d + 1}$ | $'+\Theta_w^{10}/^3D_w$ | $)/n^2$ | | | | | | |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.360E+09 | m ³ /kg | | $\rho_B Kd$ + | $\Theta_w + \Theta_a h$ | H ' | | | | | | ſ |
| Inverse of the mean concentration at the center of a 0.5-acre square source in Los Angeles | Q/C | 73.44 | g/m ² -s per kg/m ³ | | | | | | | | | | |
| Dry soil bulk density | ρ_b | 1.5 | g/cm ³ | | | | | | | | | | |
| Air-filled soil porosity | θ_{a} | 0.28 | B L _{air} /L _{soil} | | | | | | | | | | I |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0.15 | Lwater/Lsoil | | | | | | | | | | ļ |
| Exposure interval, 25-yr occupational worker | T _{Worker} | 7.88E+08 | seconds | | | | | | | | | | I |
| Volatilization Factor for soil | VFs | derived | l m ³ /kg | | | | | | | | | | |
| | | | | | | | | | | | | VFs (VOCs) | PEF |
| | | | \mathbf{D}_{i} | | $\mathbf{D}_{\mathbf{w}}$ | | Η' | | K _{OC} | K _d | D _A | Occupational | (non-VOCs) |
| Analyte | | (cm ² /s) | Reference | (cm ² /s) | Reference | (dimensionless) | Reference | (cm^3/g) | Reference | (cm ³ /g) | (cm^2/s) | m ³ /kg | m ³ /kg |
| Inorganic Constituents | | | | | | | | | | | | | |
| Arsenic | | 0 | 0 | 0 0 | | 31.61052561 | Bond Method estimate, EPISuite v4.10 (USEPA, 2011a) | 13.22 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 29 | 0 | non-VOC | 1.360E+09 |
| Chromium (III) | | 0 | 0 | 0 0 | | 0 | 0 | 0 | 0 | 1800000 | 0 | non-VOC | 1.360E+09 |
| Lead | | 0 | 0 | 0 0 | | 1.001885999 | Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 900 | 0 | non-VOC | 1.360E+09 |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | |
| Benzo[a]pyrene | | 0.0475831 | RSLs ("WATER9"; USEPA 2012 |) 5.6E-06 RSLs ("V | WATER9"; USEPA 2012) | 1.86882E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 587400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 5874 | 1.4E-11 | non-VOC | 1.360E+09 |
| Benzofluoranthenes Total | | 0.0475831 | RSLs ("WATER9"; USEPA 2012 |) 5.6E-06 RSLs ("V | WATER9"; USEPA 2012) | 2.68669E-05 | Experimental, EPISuite v4.10 (USEPA 2011a) | 599400 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 5994 | 1.7E-11 | non-VOC | 1.360E+09 |
| Dibenz[a,h]anthracene | | 0.0445672 | RSLs ("WATER9"; USEPA 2012 |) 5.2E-06 RSLs ("V | WATER9"; USEPA 2012) | 5.76596E-06 | Experimental, EPISuite v4.10 (USEPA 2011a) | 1912000 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 19120 | 2.46E-12 | non-VOC | 1.360E+09 |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | | | |
| Tetrachloroethene | | 0.0504664 | RSLs ("WATER9"; USEPA 2012 |) 9.5E-06 RSLs ("V | WATER9"; USEPA 2012) | 0.723811518 | Experimental, EPISuite v4.10 (USEPA 2011a) | 94.94 | EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0.9494 | 0.001597 | 3048.004637 | use VF |

| | | | ates: Inhalation, | Outdoor Mainte | hance worker, | Sanublast Al | ta AU | | | | |
|--|--------------------|-------------------|-----------------------|-------------------|--|---------------------------------|---|--|-------------------------|-------|--|
| Description | Variable | Val | ue | | Equations | | | | | | |
| Averaging Time, carcinogens | AT _c | | 70 yrs | Cancer Risk: | | (| E | |) | | |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | | 25 yrs | | $Risk = IUR \times$ | $C_{air} \times $ | $\left(\frac{EF_{w} \times ED_{w} \times ET_{w}}{AT_{c} \times 365 \ days \ / \ year \ \times 24 \ hours \ / \ day}\right)$ | | | | |
| COPC Concentration in air | C_{air} | chemical-specif | ïc μg/m ³ | | | $(AT_c$ | × 365 a | days / year $\times 24$ | hours / day) | | |
| COPC Concentration in soil | C _{soil} | chemical-specif | ïc mg/kg | | | | | | | | |
| Exposure Duration, worker | ED_w | 2 | 25 yrs | Noncancer Hazard: | 1 | (| | $FF \rightarrow FD$ | $\sim FT$ |) | |
| Exposure Frequency, worker | EF_w | 22 | 25 days/yr | | Hazard = $\frac{1}{DC}$ | $-\times C_{air} \times $ | T | $\frac{EF_{w} \times ED_{w}}{\times 365 \ days \ / \ yea}$ | | | |
| Exposure Time, worker | ET_w | 8.0 | 00 hours/day | | RfC | | $I_{nc,w}$ | × 365 aays / yea | $r \times 24$ nours / | aay) | |
| Inhalation Unit-Risk Factor | IUR | chemical-specif | ic $(\mu g/m^3)^{-1}$ | | | | | | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+0 | 09 m ³ /kg | | | | | | | | |
| Inhalation Reference Concentration | RfC | chemical-specif | ïc μg/m ³ | where: | $C_{air} = \frac{C_{soil}}{VF \ or \ P}$ | $ \times 1000 \ \mu g$ | | | | | |
| Volatilization Factor for VOCs | VF | chemical-specif | ïc m ³ /kg | | $C_{air} = \frac{1}{VF} \text{ or } P$ | \overline{EF} \overline{mg} | | | | | |
| - | | | VF (VOCs) or | | | Cancer | | | Hazard | | |
| | | C _{soil} | PEF (non-VOCs) | C _{air} | IUR | Risk | % of | RfC | Quotient | % of | |
| Analyte | | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total | |
| Inorganic Constituents | | | | | | | | | | | |
| Arsenic | | 9.71E+00 | 1.360E+09 | 7.14E-06 | 4.30E-03 | 1.35E-09 | < 1% | 1.50E-02 | 5.87E-05 | < 1% | |
| Chromium (III) | | 5.79E+02 | 1.360E+09 | 4.26E-04 | No Toxicity Value | | | No Toxicity Value | | | |
| Lead | | 0.00E+00 | 1.360E+09 | 0.00E+00 | 1.20E-05 | 0.00E+00 | < 1% | No Toxicity Value | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs | s) | | | | | | | | | | |
| Benzo[a]pyrene | | 2.39E+00 | 1.360E+09 | 1.76E-06 | 1.10E-03 | 1.42E-10 | <1% | No Toxicity Value | | | |
| Benzofluoranthenes Total | | 6.36E+00 | 1.360E+09 | 4.68E-06 | 1.10E-04 | 3.78E-11 | < 1% | No Toxicity Value | | | |
| Dibenz[a,h]anthracene | | 2.38E-01 | 1.360E+09 | 1.75E-07 | 1.20E-03 | 1.54E-11 | < 1% | No Toxicity Value | | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | |
| Tetrachloroethene | | 2.73E+01 | 3.048E+03 | 8.96E+00 | 2.60E-07 | 1.71E-07 | 99% | 4.00E+01 | 4.60E-02 | 100% | |
| | | | | | Pathway Sums: | Cancer Risk 1.7E-07 | | | Hazard Index 4.6E-02 | | |

Risk and Hazard Estimates: Inhalation, Outdoor Maintenance Worker, Sandblast Area AOPC

Table A-8.3

Table A-8.4

| Description | Variable | Valu | , | Equations: | | , | | | |
|--|----------------------------|-------------------|--------------------------|---------------------------|---|------------------------------|---|---|--|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specifi | c unitless | Cancer Risk: | | | | | |
| Soil-to-Skin Adherence Fraction, worker | AF_w | 0.12 | 2 mg/cm^2 -day | | $\begin{bmatrix} & & & \\ & & & & \\ & & & & & \\ & & & & $ | × ∆RS × FE | $\nabla \times FD \times FV \times$ | $EC \times SA \times CE$ | |
| Averaging Time, carcinogens | AT_{c} | 70 | 0 yrs | Risk=S | $ F_d \times \{C_{soil} \times \frac{m_w}{m}\}$ | $\times IDO_d \times LI$ | $_{W} \wedge LD_{W} \wedge LV_{W} \wedge$ | $\frac{EC_{w} \times SA_{w} \times CF_{d}}{year}$ | } |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 2: | 5 yrs | | u sou | BW_a | ×AT _c ×365day | year | ļ |
| Body Weight, adult | BW_a | 8 | 0 kg | | x – | | | | , , |
| Conversion Factor | CF _d | 1E-0 | 6 kg/mg | Noncancer Hazard: | | | | | |
| COPC Concentration in Soil | C _{soil} | chemical-specifi | c mg/kg | | 1 | [AE VAD | | $\nabla FV \times FC \times S$ | |
| Fraction of EV in Contact with Soil, worker | EC_w | | 1 unitless | Hazara | $l = \frac{1}{2} \times \{C_{soil} \times$ | $\frac{AT_{w} \times AD}{2}$ | $S_d \times ET_w \times ED_v$ | $\frac{1}{2} \times EV_{w} \times EC_{w} \times Sa}{\times 365 day / year}$ | $\frac{\mathbf{A}_{w} \times \mathbf{C} \mathbf{\Gamma}_{d}}{ }$ |
| Exposure Duration, worker | ED_w | 2: | 5 yrs | | RfD_d ³⁰⁰ | | $BW_a \times AT_{nc,a}$ | ×365 day/ year | |
| Exposure Frequency, worker | EF_{w} | 22: | 5 days/yr | | (| - | | | |
| Event Frequency, worker | EV_w | | 1 events/day | | | | | | |
| Oral Reference Dose Adjusted for GI Absorption | RfD _d | chemical-specifi | c mg/kg-day | | | | | | |
| Exposed Body Surface Area, worker | SA_w | 352 | 7 cm^2 | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specifi | c (mg/kg-day) | 1 | | | | | |
| | | | | | Cancer | | | Hazard | |
| | | C _{soil} | ABS _d | SF_d | Risk | % of | $\mathbf{RfD}_{\mathbf{d}}$ | Quotient | % of |
| Analyte | | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | | |
| Arsenic | | 9.71E+00 | 3.00E-02 | 1.50E+00 | 3.05E-07 | 8% | 3.00E-04 | 1.90E-03 | 66% |
| Chromium (III) | | 5.79E+02 | 1.00E-02 | | | | 1.95E-02 | 9.68E-04 | 34% |
| Lead | | 0.00E+00 | 1.00E-02 | 8.50E-03 | 0.00E+00 | < 1% | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | |
| Benzo[a]pyrene | | 2.39E+00 | 1.30E-01 | 7.30E+00 | 2.64E-06 | 68% | | | |
| Benzofluoranthenes Total | | 6.36E+00 | 1.30E-01 | 7.30E-01 | 7.03E-07 | 18% | | | |
| Dibenz[a,h]anthracene | | 2.38E-01 | 1.30E-01 | 7.30E+00 | 2.63E-07 | 7% | | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | |
| Tetrachloroethene | | 2.73E+01 | 0.00E+00 | 2.10E-03 | 0.00E+00 | < 1% | 6.00E-03 | 0.00E+00 | < 1% |
| | | | | Pathway Sums: | Cancer Risk 3.9E-06 | | | Hazard Index 2.9E-03 | |

Risk and Hazard Estimates: Dermal Contact with Soil, Outdoor Maintenance Worker RME, Sandblast Area AOPC

 Table A-9.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Outdoor Maintenance Worker CTE, Sandblast Area AOPC

| Definition | Variable | Value | | Equations | | | | |
|---|----------------------------|-------------------|---------------------------|-------------------------------|------------------------------|---|------------------------------------|-------------------------|
| Averaging Time, carcinogens | AT_{c} | 70 | yrs | Cancer Risk: | | | | |
| Averaging Time, noncarcinogens, adult | $AT_{nc,a}$ | 6 | yrs | | ſ | $\begin{bmatrix} IRS & FE & F \end{bmatrix}$ | $D \vee FI \vee CF$ | ר |
| Body Weight | BW_a | 80 | kg | $Risk = SF_{a}$ | $\times \{C_{soil} \times$ | $\frac{IRS_{w} \times EF_{w} \times E}{BW_{a} \times AT_{c} \times 3}$ | $D_W \wedge I I_W \wedge C I_o$ | . } |
| Conversion Factor | CF_o | 1E-06 | kg/mg | 0 | 5011 | $BW_a \times AT_c \times 3$ | 65day/ year) | |
| COPC Concentration in Soil | C _{soil} | chemical-specific | mg/kg | | C C | - | | _, |
| Exposure Duration, worker | ED_w | 6 | yrs | Noncancer Hazard: | | | | |
| Exposure Frequency, worker | EF_{w} | 225 | days/yr | | 1 [| | | |
| Fraction Contaminated Soil Ingested, worker | FI_{w} | | unitless | Hazard = - | $\frac{1}{-} \times \{ c \}$ | $\sum_{v,v} \times \left \frac{IKS_w \times EF_v}{V} \right $ | $_{v} \times ED_{w} \times FI_{w}$ | $\frac{\times CF_o}{ }$ |
| Soil Ingestion Rate, worker | IRS_w | 50 | mg/day | R | fD_o | $C_{soil} \times \left \frac{IRS_w \times EF_v}{BW_a \times AT} \right $ | $T_{nc,a} \times 365 day/$ | year |
| Oral Reference Dose | RfD _O | chemical-specific | mg/kg-day | | | | | (L |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Noncancer | |
| | | C _{soil} | SFo | Risk | % of | \mathbf{RfD}_{0} | Hazard | % of |
| Analyte | | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | |
| Arsenic | | 9.71E+00 | 1.50E+00 | 2.89E-07 | 26% | 3.00E-04 | 7.48E-03 | 80% |
| Chromium (III) | | 5.79E+02 | No Toxicity Value | ; | | 1.50E+00 | 1.49E-04 | 2% |
| Lead | | 4.18E+02 | | | | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]pyrene | | 2.39E+00 | 7.30E+00 | 5.76E-07 | 52% | No Toxicity Value | | |
| Benzofluoranthenes Total | | 7.84E+00 | 7.30E-01 | 1.89E-07 | 17% | No Toxicity Value | | |
| Dibenz[a,h]anthracene | | 2.38E-01 | 7.30E+00 | 5.74E-08 | 5% | No Toxicity Value | | |
| Tetrachloroethene | | 2.73E+01 | 2.10E-03 | 1.89E-09 | <1% | 6.00E-03 | 1.75E-03 | 19% |
| | | | Pathway Sum: | <u>Cancer Risk</u> 1.1E-06 | | | Hazard Index 9.4E-03 | |

 Table A-9.2

 Derivation of Inhalation Factors for Inhalation Exposures, Outdoor Maintenance Worker CTE, Sandblast Area AOPC

| Description | Variable | Value | | Units | VFI | Derivation | | | | | | | | |
|---|---------------------|-----------------|---|-------------------|--------------|---|---------------------------|--|--------------------|-------------------------------------|----------------|------------|----------------------------|--------------------|
| Unit conversion factor | CFi | 1.00E-0 | $04 \text{ m}^2/\text{cm}^2$ | | | | (| | | | | | | |
| Apparent Diffusivity | D_A | derive | ed cm ² /s | | | VFs = (O / O) | $(3) \times (3)$ | $\frac{14 \times D_A \times T_{[receptor]})^{1/2}}{(2 \times \rho_A \times D_A)} \times C$ | CF . | | | | | |
| Diffusivity in air | D_i | chemical-specif | ic cm ² /s | | | | , , | $(2 \times \rho_b \times D_A)$ | i | | | | | |
| Diffusivity in water | D_w | chemical-specif | fic cm ² /s | | | | | | | | | | | |
| Default organic-carbon content | Foc | 0.0 | 01 g/g | | | | | | | | | | | |
| Henry's law constant | H' | chemical-specif | ic dimensionless | | | | | | | | | | | |
| Soil-water partition coefficient | K _d | chemical-specif | ic cm ³ /g | | | | | | | | | | | |
| for organics: $Kd=K_{OC} \times F_{OC}$ | | | | | base | d on: | | | | | | | | |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specif | ic cm ³ /g | | | $\int (0 10 / 3 p) u$ | 10/3 D |) $(2]$ | | | | | | |
| Total soil porosity | n | | 43 $L_{\text{porespace}}/L_{\text{soil}}$ | | | $DA = \frac{\left[\left(\Theta \stackrel{10}{a} \stackrel{/}{}^{3} D_{i} H\right) + \frac{1}{\rho_{B} K d}\right]}{\rho_{B} K d}$ | $+\Theta_{w}^{i} + D_{w}$ | $()/n^{-1}$ | | | | | | |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.360E+0 | 09 m ³ /kg | | | $\rho_{B} Kd$ + | $\Theta_w + \Theta_a$ | Н ' | | | | | | |
| Inverse of the mean concentration at the center of a 0.5-acre square source in Los Angeles | Q/C | 73.4 | 44 g/m ² -s per kg/m | n ³ | | | | | | | | | | |
| Dry soil bulk density | ρ_{b} | 1. | $.5 \text{ g/cm}^3$ | | | | | | | | | | | |
| Air-filled soil porosity | θ_{a} | 0.2 | 28 L _{air} /L _{soil} | | | | | | | | | | | |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0.1 | 15 L _{water} /L _{soil} | | | | | | | | | | | |
| Exposure interval, 6-yr occupational worker | T _{Worker} | 1.89E+0 | 08 seconds | | | | | | | | | | | |
| Volatilization Factor for soil | VFs | derive | ed m ³ /kg | | | | | | | | | | | |
| | | | Di | | | D _w | | H' | | K _{OC} | K _d | DA | VFs (VOCs) Occupational | PEF (non-VOCs) |
| Analyte | | (cm^2/s) | | eference | (cm^2/s) | Reference | (dimensionless) | | (cm^{3}/g) | Reference | (cm^3/g) | (cm^2/s) | m ³ /kg | m ³ /kg |
| Inorganic Constituents | | (011170) | | | (011170) | | (| | (611178) | | (em /g) | (011175) | iii / iig | |
| Arsenic | | 0 | 0 | | 0 0 | | 31.6 | Bond Method estimate, EPISuite v4.10 (USEPA, 2011a) | 13.22 EPISuite v | 74.10 (USEPA, 2011a); MCI-estimated | 29 | 0 | non-VOC | 1.360E+09 |
| Chromium (III) | | 0 | 0 | | 0 0 | | 0 | 0 | 0 0 | | 1800000 | 0 | non-VOC | 1.360E+09 |
| Lead | | 0 | 0 | | 0 0 | | 1.002 | Bond Method estimate, EPISuite v4.10 (USEPA, 2011a) | 13.22 EPISuite v | 4.10 (USEPA, 2011a); MCI-estimated | 900 | 0 | non-VOC | 1.360E+09 |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | | |
| Benzo[a]pyrene | | 0.0475831 | RSLs ("WATE | R9"; USEPA, 2012) | 5.56E-06 RSL | s ("WATER9"; USEPA, 2012) | 0.0000187 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 587400 EPISuite v | 4.10 (USEPA, 2011a); MCI-estimated | 0 | 8.2E-07 | non-VOC | 1.360E+09 |
| Benzofluoranthenes Total | | 0.0475831 | RSLs ("WATE | R9"; USEPA, 2012) | 5.56E-06 RSL | s ("WATER9"; USEPA, 2012) | 0.00002687 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 599400 EPISuite v | v4.10 (USEPA, 2011a); MCI-estimated | 0 | 1.02E-06 | non-VOC | 1.360E+09 |
| Dibenz[a,h]anthracene | | 0.0445672 | RSLs ("WATE | R9"; USEPA, 2012) | 5.21E-06 RSL | s ("WATER9"; USEPA, 2012) | 0.00000577 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 1912000 EPISuite v | v4.10 (USEPA, 2011a); MCI-estimated | 0 | 4.7E-07 | non-VOC | 1.360E+09 |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | | | | |
| Tetrachloroethene | | 0.0504664 | RSLs ("WATE | R9"; USEPA, 2012) | 9.46E-06 RSL | s ("WATER9"; USEPA, 2012) | 0.724 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 94.94 EPISuite | v4.10 (USEPA, 2011a); MCI-estimated | 0 | 0.008045 | | use VF |

| Description | Variable | Valu | le | | Equations | | | | | |
|--|----------------------------|-------------------|------------------------------------|------------------|---|---|----------------|-------------------------------------|---------------------|-------|
| Averaging Time, carcinogens | AT _c | 7 | 70 yrs | Cancer Risk: | | (| EI | |) | |
| Averaging Time, noncarcinogens, adult | $AT_{nc,a}$ | | 6 yrs | | $Risk = IUR \times C$ | $C_{air} \times $ —— | EF | $F_{W} \times ED_{W} \times EI_{V}$ | <u>v</u> | |
| COPC Concentration in air | Cair | chemical-specif | ic μg/m ³ | | | $(AT_c \times$ | 365 d | ays / year × 24 | hours / day) | |
| COPC Concentration in soil | C _{soil} | chemical-specif | ic mg/kg | | | | | | | |
| Exposure Duration, worker | ED_w | | 6 yrs No | | 1 | (| | | $\sim FT$ |) |
| Exposure Frequency, worker | EF_{w} | | 225 days/yr | | $Hazard = \frac{1}{RfC}$ | $- \times C_{air} \times \left \frac{1}{\sqrt{2}} \right $ | 7 | $EF_W \wedge ED_W$ | | / 1 |
| Exposure Time, worker | ET_w | 8.0 | 8.00 hours/day | | RfC | (AI) | $nc, w \times$ | < 365 aays / year | $\times 24$ nours / | aay j |
| Inhalation Unit-Risk Factor | IUR | chemical-specif | emical-specific $(\mu g/m^3)^{-1}$ | | | | | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+0 | 1.360E+09 m ³ /kg | | | | | | | |
| Inhalation Reference Concentration | RfC | chemical-specif | nical-specific µg/m ³ | | $C_{air} = \frac{C_{soil}}{VF \ or \ PI}$ | $-\times 1000 \frac{\mu g}{1000}$ | | | | |
| Volatilization Factor for VOCs | VF | chemical-specif | emical-specific m ³ /kg | | $C_{air} = VF \text{ or } PI$ | EF mg | | | | |
| | | | VF (VOCs) or | | | Cancer | | | Hazard | |
| | | C _{soil} | PEF (non-VOCs) | C _{air} | IUR | Risk | % of | RfC | Quotient | % of |
| Analyte | | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | | | |
| Arsenic | | 9.71E+00 | 1.360E+09 | 7.14E-06 | 4.30E-03 | 3.24E-10 | 87% | 1.50E-02 | 5.87E-05 | 100% |
| Chromium (III) | | 5.79E+02 | 1.360E+09 | 4.26E-04 | No Toxicity Value | | | No Toxicity Value | | |
| Lead | | 4.18E+02 | 1.360E+09 | 3.07E-04 | | | | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs | s) | | | | | | | | | |
| Benzo[a]pyrene | | 2.39E+00 | 1.360E+09 | 1.76E-06 | 1.10E-03 | 3.40E-11 | 9% | No Toxicity Value | | |
| Benzofluoranthenes Total | | 7.84E+00 | 1.360E+09 | 5.77E-06 | 1.10E-04 | 1.12E-11 | 3% | No Toxicity Value | | |
| Dibenz[a,h]anthracene | | 2.38E-01 | 1.360E+09 | 1.75E-07 | 1.20E-03 | 3.70E-12 | < 1% | No Toxicity Value | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | |
| Tetrachloroethene | | 2.73E+01 | | | 2.60E-07 | | | 4.00E+01 | | |
| | | | | | | Cancer Risk | | 1 | Hazard Index | |
| | | | | | Pathway Sums: | | | | 5.9E-05 | |

 Table A-9.3

 Risk and Hazard Estimates: Inhalation, Outdoor Maintenance Worker, Sandblast Area AOPC

Description Variable Value Equations: Dermal Soil Absorption Fraction ABS chemical-specific unitless Cancer Risk: 0.02 mg/cm^2 -day Soil-to-Skin Adherence Fraction, worker AF_{w} $Risk = SF_d \times \left\{ C_{soil} \times \left| \frac{AF_w \times ABS_d \times EF_w \times ED_w \times EV_w \times EC_w \times SA_w \times CF_d}{BW_a \times AT_c \times 365 day / year} \right| \right\}$ Averaging Time, carcinogens AT_{c} 70 yrs Averaging Time, noncarcinogens, adult AT_{nc a} 6 yrs Body Weight, adult BW_a 80 kg **Conversion Factor** CF_d 1E-06 kg/mg Noncancer Hazard: COPC Concentration in Soil C_{soil} chemical-specific mg/kg $Hazard = \frac{1}{RfD_d} \times \left\{ C_{soil} \times \left[\frac{AF_w \times ABS_d \times EF_w \times ED_w \times EV_w \times EC_w \times SA_w \times CF_d}{BW_a \times AT_{nc,a} \times 365 \, day / \, year} \right] \right\}$ Fraction of EV in Contact with Soil, worker EC_w 1 unitless Exposure Duration, worker ED. 6 yrs Exposure Frequency, worker EF_w 225 days/yr Event Frequency, worker EV_w 1 events/day Oral Reference Dose Adjusted for GI Absorption RfD₄ chemical-specific mg/kg-day 3300 cm^2 SAw Exposed Body Surface Area, worker chemical-specific (mg/kg-day)⁻¹ Oral Slope Factor Adjusted for GI Absorption SF_{d} Cancer Hazard Csoil ABS_d SFd Risk % of RfD_d Ouotient % of Analyte Total (mg/kg-day) Total (mg/kg) (unitless) (mg/kg-day)⁻¹ (dimensionless) (dimensionless) Inorganic Constituents Arsenic 9.71E+00 3.00E-02 1.50E+001.14E-08 3.00E-04 2.96E-04 66% 7% Chromium (III) 5.79E+02 1.00E-02 1.95E-02 1.51E-04 34% ___ --Lead 4.18E+02 1.00E-02 ---------Polycyclic Aromatic Hydrocarbons (PAHs) Benzo[a]pyrene 2.39E+00 1.30E-01 7.30E+00 9.89E-08 65% ---___ Benzofluoranthenes Total 7.84E+00 1.30E-01 7.30E-01 3.24E-08 21% ___ ___ 2.38E-01 1.30E-01 7.30E+00 9.85E-09 Dibenz[a,h]anthracene 6% --___ --Volatile Organic Compounds (VOCs) Tetrachloroethene 2.73E+01 0.00E+00 2.10E-03 0.00E+00 < 1% 6.00E-03 0.00E+00< 1% Cancer Risk Hazard Index Pathway Sums: 1.5E-07 4.5E-04

 Table A-9.4

 Risk and Hazard Estimates: Dermal Contact with Soil, Outdoor Maintenance Worker CTE, Sandblast Area AOPC

| Definition | Variable | Value | | Equations | | | | | | |
|--|---------------------|---|---------------------------|------------------------|-------------------|---|---|--------------|--|--|
| Averaging Time, carcinogens | AT _c | 70 | yrs | Cancer Risk: | | | | | | |
| Averaging Time, noncarcinogens, construction worker | AT _{nc,cw} | 1 | yrs | | Г | $ \frac{IRS_{cw} \times EF_{cw} \times ED_{cw} \times FI_{cw} \times C}{BW_{cw} \times AT_{c} \times 365 \ day \ / \ year } $ | | | | |
| Body Weight | BW_{cw} | 80 | kg | $Risk = SF_o \times$ | $C_{soil} \times$ | $\frac{IKS_{CW} \times EF_{CW} \times ED}{PW}$ | $\frac{c_W}{c_W} \times \frac{FT}{c_W} \times CF_c$ | <u>^ </u> | | |
| Conversion Factor | CFo | 1E-06 kg/mg chemical-specific mg/kg | | | <u> </u> | $\bigcup_{cw} M_{cw} \times AI_{c} \times S$ | sos ady / year | <u>ار</u> | | |
| COPC Concentration in Soil | C _{soil} | | | | | | | | | |
| Exposure Duration, construction worker | ED_{cw} | _{cw} 1 yrs | | Noncancer Hazard | | | | | | |
| Exposure Frequency, construction worker | EF_{cw} 250 da | | days/yr | 1 | Г | | | $(CE)^{7}$ | | |
| Fraction Contaminated Soil Ingested, construction worker | FI _{cw} | 1.0 | unitless | Hazard = $\frac{1}{n}$ | $- \times C_s $ | $oil \times \left(\frac{IRS_{cw} \times EF_{cw}}{BW_{cw} \times AT_{nc}}\right)$ | $\frac{\times ED_{cw} \times FI_{cw} \times}{265}$ | | | |
| Soil Ingestion Rate, worker | IRS _{cw} | 330 mg/day chemical-specific mg/kg-day | | RjL | 0 | $(BW_{cw} \times AI_{nc})$ | $_{r,cw}$ × 365 day / y | ear) | | |
| Oral Reference Dose | RfD ₀ | | | | | | | | | |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | | | |
| | | | | Cancer | | | Noncancer | | | |
| | | C _{soil} | SFo | Risk | % of | RfD _o | Hazard | % of | | |
| Analyte | | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | |
| Tetrachloroethene | | 9.90E+01 | 2.10E-03 | 8.39E-09 | < 1% | 6.00E-03 | 4.66E-02 | 100% | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | |
| Benzo[a]pyrene | | 3.98E+00 | 7.30E+00 | 1.17E-06 | 99% | No Toxicity Value | | | | |
| | | | | Cancer Risk | | | Hazard Index | | | |
| | | | Pathway Sum: | 1.2E-06 | | | 4.7E-02 | | | |

Risk and Hazard Estimates: Incidental Ingestion of Soil, Construction Worker RME, Sandblast Area AOPC

| Description: | Variable | Value | Units | VF | Derivation | | | | | | | | |
|---|------------------|----------------------|---|----------------------|--|--------------------------------|--|----------------------|--|----------------------|------------|--------------------|--------------------|
| Unit conversion factor | CFi | 1.00E-04 | m ² /cm ² | | | 6 | \mathbf{D} \mathbf{T} \mathbf{V} / 2 | | | | | | |
| Apparent Diffusivity | D_A | derived | cm ² /s | | VFs = (Q / C) | $\times \frac{(3.14)}{(3.14)}$ | $\frac{\times D}{(2 \times \rho + \times D)} \times \frac{D}{(2 \times \rho + \times D)} \times C$ | CF i | | | | | |
| Diffusivity in air | D_i | chemical-specific | cm ² /s | | | | $(2 \times \rho_b \times D_A)$ | | | | | | |
| Diffusivity in water | D_w | chemical-specific | cm ² /s | | | | | | | | | | |
| Default organic-carbon content | F _{OC} | 0.01 | g/g | | | | | | | | | | |
| Henry's law constant | H' | chemical-specific | dimensionless | | | | | | | | | | |
| Soil-water partition coefficient | K _d | chemical-specific | cm ³ /g | | | | | | | | | | |
| for organics: $Kd=K_{OC} \times F_{OC}$ | | | | base | ed on: | | | | | | | | |
| Soil-organic carbon partition coefficient | K _{oc} | chemical-specific | cm ³ /g | | $\left[\left(\Theta^{10} \right)^{3} D H^{1} \right]$ | $\Theta^{10/3}D$ | (n^{2}) | | | | | | |
| Total soil porosity | n | 0.43 | L _{porespace} /L _{soil} | | $DA = \frac{\left[\left(\Theta_{a}^{10}/3 D_{i} H^{-1}\right)\right]}{\rho_{B} K d} + \frac{1}{2}$ | | <u>), n]</u> | | | | | | |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.000E+06 | | | $\rho_B \kappa a +$ | $y_w + 0_a I$ | 1 | | | | | | |
| Inverse of the mean concentration at the center | Q/C | 73.44 | g/m^2 -s per kg/m ³ | | | | | | | | | | |
| of a 0.5-acre square source | - | | | | | | | | | | | | |
| Dry soil bulk density | ρ_b | 1.5 | g/cm ³ | | | | | | | | | | |
| Air-filled soil porosity | θ_{a} | 0.28 | L _{air} /L _{soil} | | | | | | | | | | |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0.15 | Lwater/Lsoil | | | | | | | | | | |
| Exposure interval, 1-yr construction worker | T_{cw} | 3.15E+07 | seconds | | | | | | | | | | |
| Volatilization Factor for soil | VFs | derived | m ³ /kg | | | | | | | | | | |
| | | | | | _ | | | | | | | VFs (VOCs) | PEF |
| | | 2 | D _i | 2 | $\mathbf{D}_{\mathbf{w}}$ | | Η' | 2 | K _{oc} | K _d | DA | Construction | (non-VOCs) |
| Analyte | | (cm ² /s) | Reference | (cm ² /s) | Reference | (dimensionless) | Reference | (cm ³ /g) | Reference | (cm ³ /g) | (cm^2/s) | m ³ /kg | m ³ /kg |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | | | |
| Tetrachloroethene | | 0.0504664 | RSLs ("WATER9"; USEPA, 2012) | 9.46E-06 RSI | Ls ("WATER9"; USEPA, 2012) | 0.724 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 94.94 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 0.9494 | 0.0016 | 609.6009273 | use VF |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | |
| Benzo[a]pyrene | | 0.0475831 | RSLs ("WATER9"; USEPA, 2012) | 5.56E-06 RSI | Ls ("WATER9"; USEPA, 2012) | 1.87E-05 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 587400 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 0 | 8.2E-07 | non-VOC | 1.000E+06 |

| Risk and Hazard Estimates | Inhalation, Construction Worker, Sandblast Area A | OPC |
|---------------------------|---|-----|
|---------------------------|---|-----|

| Description | Variable | Val | ue | | Equations | | | | | | |
|---|---------------------|----------------|------------------------------|---|--|---|------------------|--|-----------------|-------|--|
| Averaging Time, carcinogens | AT _c | | 70 yrs | Cancer Risk: | 1 | | | | | | |
| Averaging Time, noncarcinogens, construction worker | AT _{nc,cw} | | 1 yrs | | | (| | EE VET | | | |
| COPC Concentration in air | Cair | chemical-speci | fic µg/m ³ | Risk = | $= IUR \times C_{air} \times \left(\frac{ED_{cw} \times EF_{cw} \times ET_{cw}}{AT_{c} \times 365 \ days \ / \ year \ \times 24 \ hours \ / \ day} \right)$ | | | | | | |
| COPC Concentration in soil | C _{soil} | chemical-speci | fic mg/kg | $(AI_c \times 365 \ adys \ / \ year \ \times 24)$ | | | | year ~ 24 nours 7 | uuy) | | |
| Exposure Duration, construction worker | ED_{cw} | | 1 yrs | Noncancer Hazard: | | | | | | | |
| Exposure Frequency, construction worker | EF_{cw} | w 250 days/yr | | | 1 | (| ED | | | | |
| Exposure Time, construction worker | ET_{cw} | | 8 hours/day | | $=\frac{1}{PfC}\times$ | $C_{air} \times \frac{1}{AT}$ | ED 0 | $\frac{1}{cw} \times EF_{cw} \times ET_{cw}}{days / year \times 24 h}$ | ours / day | | |
| Inhalation Unit-Risk Factor | IUR | chemical-speci | fic $(\mu g/m^3)^{-1}$ | | ŊĊ | (AI nc, c | w × 505 | aays / year × 24 n | ours (aay) | | |
| Particulate emission factor (non-VOCs) | PEF | 1.00E+ | 06 m ³ /kg | | | | | | | | |
| Inhalation Reference Concentration | RfC | chemical-speci | hemical-specific $\mu g/m^3$ | | C | C _{soil} | μg | | | | |
| Volatilization Factor for VOCs | VF | chemical-speci | fic m ³ /kg | | $C_{air} = \frac{1}{V_{air}}$ | $\frac{C_{soil}}{F \text{ or } PEF} \times 100$ | $\frac{100}{mg}$ | | | | |
| | | | VF (VOCs) or | | | Cancer | | | Hazard | | |
| | | C_{soil} | PEF (non-VOCs) | C _{air} | IUR | Risk | % of | RfC | Quotient | % of | |
| Analyte | | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | |
| Tetrachloroethene | | 9.90E+01 | 6.10E+02 | 1.62E+02 | 2.60E-07 | 1.38E-07 | 91% | 4.00E+01 | 9.27E-01 | 100% | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | |
| Benzo[a]pyrene | | 3.98E+00 | 1.00E+06 | 3.98E-03 | 1.10E-03 | 1.43E-08 | 9% | No Toxicity Value | | | |
| | | | | | | Cancer Risk | | • | Hazard Index | | |
| | | | | Pa | thway Sums: | 1.5E-07 | | | 9.3E-01 | | |

| Description | Variable | Valu | ue | Equations | | | | | |
|--|-----------------------------|--|----------------------------|-----------------------------------|-------------------------------|--------------------------------|-----------------------------------|---|-------|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specif | ïc unitless | Cancer Risk: | | | | | |
| Soil-to-Skin Adherence Fraction, construction worker | AF_{cw} | 0 | .3 mg/cm ² -day | | $(AF_{cw} \times ABS_{d})$ | $_{l} \times EF_{cw} \times .$ | $ED_{cw} \times EV_{cw} \times E$ | $EC_{cw} \times SA_{cw} \times CF_d$ |] |
| Averaging Time, carcinogens | AT_{c} | - | 70 yrs | $Risk = SF_d \times C_d$ | soil × | $BW_{cw} \times I$ | $AT_c \times 365 \ day / y$ | $\left(\frac{BC_{cw} \times SA_{cw} \times CF_d}{ear}\right)$ | |
| Averaging Time, noncarcinogens, construction worker | AT _{nc,cw} | | 1 yrs | - | | | | | - |
| Body Weight, adult | BW_{cw} | 8 | 30 kg | | | | | | |
| Conversion Factor | CF_d | 1E-0 | 06 kg/mg | | | | | | |
| COPC Concentration in Soil | C _{soil} | 1 unitless 1 yrs 250 days/yr 1 events/day | | Noncancer Hazard: | | | | | |
| Fraction of EV in Contact with Soil, construction worker | EC_{cw} | | | 1 | $\int (AF \times A$ | $BS \to EF$ | $\times ED \dots \times EV$ | $\times EC = \times SA = \times CI$ | ₹,)] |
| Exposure Duration, construction worker | ED_{cw} | | | $Hazard = \frac{1}{RfD_d} \times$ | $ C_{soil} \times $ | BW_{cw} | $< AT_{nc} \sim 365 c$ | lay / year | |
| Exposure Frequency, construction worker | $\mathrm{EF}_{\mathrm{cw}}$ | | | 5 u | L | C W | nc ,cw | 5 5 | 7] |
| Event Frequency, construction worker | EV_{cw} | | | | | | | | |
| Oral Reference Dose Adjusted for GI Absorption | RfD_d | chemical-specif | ïc mg/kg-day | | | | | | |
| Exposed Body Surface Area, construction worker | SA_{cw} | 352 | 27 cm^2 | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | $\mathbf{SF}_{\mathbf{d}}$ | chemical-specif | ic (mg/kg-day) | I | | | | | |
| | | | | | Cancer | | | Hazard | |
| | | C _{soil} | ABS _d | $\mathbf{SF}_{\mathbf{d}}$ | Risk | % of | RfD _d | Quotient | % of |
| Analyte | | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Volatile Organic Compounds (VOCs) | | | | | | | | | |
| Tetrachloroethene | | 9.90E+01 | 0.00E+00 | 2.10E-03 | 0.00E+00 | < 1% | 6.00E-03 | 0.00E+00 | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | |
| Benzo[a]pyrene | | 3.98E+00 | 1.30E-01 | 7.30E+00 | 4.89E-07 | 100% | | | |
| | | | | Pathway Sums: | <u>Cancer Risk</u> 4.9E-07 | | | Hazard Index 0.00E+00 | |

Risk and Hazard Estimates: Dermal Contact with Soil, Construction Worker RME, Sandblast Area AOPC

 Table A-11.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Construction Worker CTE, Sandblast Area AOPC

| Definition | Variable | Value | | Equations | | | | |
|--|---------------------|------------------------------|---------------------------|---------------------------|---|---|---|---------------|
| Averaging Time, carcinogens | AT _c | 70 | yrs | Cancer Risk: | | | | |
| Averaging Time, noncarcinogens, construction worker | AT _{nc,cw} | 1 | yrs | Г | (| | | 17 |
| Body Weight | BW _{cw} | 80 | kg | $Risk = SF_o \times$ | $C_{soil} \times \left[-\frac{1}{2} \right]$ | $\frac{RS_{cw} \times EF_{cw} \times EL}{BW_{cw} \times AT_{c} \times 3}$ | <u>o</u> | |
| Conversion Factor | CFo | 1E-06 | kg/mg | | (| $BW_{cw} \times AT_c \times C$ | 365 day / year | 刀 |
| COPC Concentration in Soil | C _{soil} | chemical-specific | mg/kg | | | | | |
| Exposure Duration, construction worker | ED_{cw} | 1 | yrs | Noncancer Hazard: | | | | |
| Exposure Frequency, construction worker | EF_{cw} | 250 | days/yr | | Г | | | |
| Fraction Contaminated Soil Ingested, construction worker | r FI _{cw} | | unitless | Hazard = $\frac{1}{D(D)}$ | $-\times C_{soil} $ | $\frac{1}{2} \times \left(\frac{1RS_{cw} \times EF_{cw}}{BW_{cw} \times AT_{m}}\right)$ | $\frac{\times ED_{cw} \times FI_{cw}}{265 l}$ | $\times CF_o$ |
| Soil Ingestion Rate, worker | IRS _{cw} | 100 | mg/day | RfD _o | | $(BW_{cw} \times AT_n)$ | $c_{,cw} \times 365 day / 2$ | year J |
| Oral Reference Dose | RfD ₀ | chemical-specific | mg/kg-day | | | | | |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Noncancer | |
| | | $\mathbf{C}_{\mathbf{soil}}$ | SFo | Risk | % of | RfD _o | Hazard | % of |
| Analyte | | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Volatile Organic Compounds (VOCs) | | | | | | | | |
| Tetrachloroethene | | 9.90E+01 | 2.10E-03 | 1.27E-09 | < 1% | 6.00E-03 | 1.41E-02 | 100% |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]pyrene | | 3.98E+00 | 7.30E+00 | 1.78E-07 | 99% | No Toxicity Value | | |
| | | | | Cancer Risk | | | Hazard Index | |
| | | | Pathway Sum | : 1.8E-07 | | | 1.4E-02 | |

 Table A-11.2

 Derivation of Inhalation Factors for Inhalation Exposures, Construction Worker CTE, Sandblast Area AOPC

| Description | Variable | Value | Units | VFD | erivation | | | | | | | | <u> </u> |
|---|------------------|--|---------------------|----------------------|---|--|---|----------------------|--|----------------------|------------|--------------------|--------------------|
| Unit conversion factor | CFi | $1.00E-04 \text{ m}^2/\text{cm}^2$ | | | (- | | $\lambda l \neq 2$ | | | | | | |
| Apparent Diffusivity | D_A | derived cm ² /s | | | $VFs = (Q / C) \times \frac{(3.)}{2}$ | $\frac{14 \times D_A \times T_{[rec}}{\sqrt{1-1}}$ | $\frac{eptor}{2} \times CF_{i}$ | | | | | | |
| Diffusivity in air | D_i | chemical-specific cm ² /s | | | | $(2 \times \rho_b \times L)$ | | | | | | | |
| Diffusivity in water | D_w | chemical-specific cm ² /s | | | | | | | | | | | |
| Default organic-carbon content | F _{OC} | 0.01 g/g | | | | | | | | | | | |
| Henry's law constant | H' | chemical-specific dimensionles | s | | | | | | | | | | ł |
| Soil-water partition coefficient | K _d | chemical-specific cm ³ /g | | | | | | | | | | | ł |
| for organics: Kd=K _{OC} ×F _{OC} | | | | based | on: | | | | | | | | ł |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specific cm ³ /g | | | $\int O \frac{10}{3} D H$ | $(10^{10}/3^{3})$ | , ² | | | | | | ł |
| Total soil porosity | n | 0.43 L _{porespace} /L _{soil} | | | $DA = \frac{\left[\left(\Theta \right _{a}^{10} / 3 D_{i} H \right) \right]_{i}}{\rho_{R} K d} + \frac{1}{2} \left[\left(\Theta \right)_{i}^{10} + \left(\Theta \right)_{i}^{10} \right]_{i}}$ | $+ 0_w D_w J$ | <u></u> | | | | | | ł |
| Particulate Emission Factor (non-VOCs), default | PEF | $1.000E+06 \text{ m}^3/\text{kg}$ | | | $\rho_B \kappa a +$ | $\Theta_w + \Theta_a H$ | | | | | | | ł |
| Inverse of the mean concentration at the center | Q/C | 73.44 g/m ² -s per kg | $/m^3$ | | | | | | | | | | ł |
| of a 0.5-acre square source | | | | | | | | | | | | | ł |
| Dry soil bulk density | $\rho_{\rm b}$ | 1.5 g/cm^3 | | | | | | | | | | | ł |
| Air-filled soil porosity | θ_{a} | 0.28 Lair/Lsoil | | | | | | | | | | | ł |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0.15 L _{water} /L _{soil} | | | | | | | | | | | |
| Exposure interval, 0.5-yr construction worker | T_{cw} | 1.58E+07 seconds | | | | | | | | | | | |
| Volatilization Factor for soil | VFs | derived m3/kg | | | | | | | | | | | |
| | | | | | | | | | | | _ | VFs (VOCs) | |
| | | D _i | | | $\mathbf{D}_{\mathbf{w}}$ | | Н' | | K _{OC} | K _d | DA | Construction | · , |
| Analyte | | (cm^2/s) | Reference | (cm ² /s) | Reference | (dimensionless) | Reference | (cm ³ /g) | Reference | (cm ³ /g) | (cm^2/s) | m ³ /kg | m ³ /kg |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | | | |
| Tetrachloroethene | | 0.0504664 RSLs ("WAT | ER9"; USEPA, 2012) | 9.46E-06 RSL | ("WATER9"; USEPA, 2012) | 0.723811518 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 94.94 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 0 | 0.008045 | | use VF |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | |
| Benzo[a]pyrene | | 0.0475831 RSLs ("WAT | 'ER9"; USEPA, 2012) | 5.56E-06 RSL | ("WATER9"; USEPA, 2012) | 1.86882E-05 | Experimental, EPISuite v4.10 (USEPA, 2011a) | 587400 | EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 0 | 8.2E-07 | non-VOC | 1.000E+06 |

| | Table A-11.3 | |
|-----------------------------------|--------------------------|-----------------------------|
| Risk and Hazard Estimates: | Inhalation, Construction | Worker, Sandblast Area AOPC |

| Description | Variable | Val | ue | | Equations | | | | | | | | |
|---|---------------------|-------------------|-------------------------------|--|--|---|---|---|-----------------|-------|--|--|--|
| Averaging Time, carcinogens | AT _c | | 70 yrs | Cancer Risk: | | | | | | | | | |
| Averaging Time, noncarcinogens, construction worker | AT _{nc,cw} | (| 0.5 yrs | | | (E | D × | EF × ET | | | | | |
| COPC Concentration in air | C_{air} | chemical-speci | fic µg/m ³ | Risk = | $= IUR \times C_{air} \times \left(\frac{ED_{cw} \times EF_{cw} \times ET_{cw}}{AT_{c} \times 365 \ days \ / \ year \ \times 24 \ hours \ / \ day} \right)$ | | | | | | | | |
| COPC Concentration in soil | C _{soil} | chemical-speci | fic mg/kg | $(AI_c \times 505 \text{ uuys } 7 \text{ ye})$ | | | | <i>year 21 nours ,</i> | | | | | |
| Exposure Duration, construction worker | ED_{cw} | (| 0.5 yrs | Noncancer Hazard: | | | | | | | | | |
| Exposure Frequency, construction worker | EF_{cw} | 2 | 250 days/yr | | 1 | (| FD | $\vee FF \vee FT$ |) | | | | |
| Exposure Time, construction worker | ET_{cw} | | 8 hours/day | | $=\frac{1}{RfC}\times$ | $C_{air} \times \boxed{{AT}}$ | $\frac{c_W}{c_W} \sim Er_{c_W} \sim Er_{c_W}$ | $\frac{v \times EF_{cw} \times ET_{cw}}{days / year \times 24 \ hours / day}$ | | | | | |
| Inhalation Unit-Risk Factor | IUR | chemical-speci | fic $(\mu g/m^3)^{-1}$ | | ŊĊ | (III nc,cv | _w × 505 | uuys / yeur × 24 n | ours (uuy) | | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.00E+ | 06 m ³ /kg | | | | | | | | | | |
| Inhalation Reference Concentration | RfC | chemical-speci | chemical-specific $\mu g/m^3$ | | C | C _{soil} | μg | | | | | | |
| Volatilization Factor for VOCs | VF | chemical-speci | fic m ³ /kg | | $C_{air} = -V$ | $\frac{C_{soil}}{F \text{ or } PEF} \times 100$ | $\frac{10}{mg}$ | | | | | | |
| | | | VF (VOCs) or | | | Cancer | | | Hazard | | | | |
| | | C _{soil} | PEF (non-VOCs) | C_{air} | IUR | Risk | % of | RfC | Quotient | % of | | | |
| Analyte | | (mg/kg) | (m ³ /kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total | | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | | | |
| Tetrachloroethene | | 9.90E+01 | | | 2.60E-07 | | | 4.00E+01 | | | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | | | |
| Benzo[a]pyrene | | 3.98E+00 | 1.00E+06 | 3.98E-03 | 1.10E-03 | 7.14E-09 | 100% | No Toxicity Value | | | | | |
| | | | | | | Cancer Risk | | | Hazard Index | | | | |
| | | | | Pa | thway Sums: | 7.1E-09 | | | | | | | |

Table A-11.4 Risk and Hazard Estimates: Dermal Contact with Soil, Construction Worker CTE, Sandblast Area AOPC

| Description | Variable | Valu | ue | Equations | | | | | | | |
|--|---------------------|--|----------------------------|------------------------------|---|--|---|---|-------|--|--|
| Dermal Soil Absorption Fraction | ABS_d | chemical-specif | ic unitless | Cancer Risk: | | | | | | | |
| Soil-to-Skin Adherence Fraction, construction worker | | 0 | .1 mg/cm ² -day | | $AF_{cw} \times ABS$ | $_d \times EF_{cw} \times$ | 1 | | | | |
| Averaging Time, carcinogens | AT_{c} | - | 70 yrs | $RISK = SF_d \times 0$ | c _{soil} × | $\frac{d \times EF_{cw} \times ED_{cw} \times EV_{cw} \times EC_{cw} \times SA_{cw} \times CF_{d}}{BW_{cw} \times AT_{c} \times 365 \ day / year}$ | | | | | |
| Averaging Time, noncarcinogens, construction worker | AT _{nc,cw} | | 1 yrs | - | | | | | - | | |
| Body Weight, adult | BW_{cw} | 5 | 30 kg | | | | | | | | |
| Conversion Factor | CF_d | 1E-0 | 06 kg/mg | | | | | | | | |
| COPC Concentration in Soil | C _{soil} | chemical-specific mg/kg 1 unitless 1 yrs | | Noncancer Hazard: | | | | | | | |
| Fraction of EV in Contact with Soil, construction worker | EC_{cw} | | | 1 | | $ABS \cdot \vee FF$ | $EF_{cw} \times ED_{cw} \times EV_{cw} \times EC_{cw} \times SA_{cw} \times CF_{d}$ $ = \left \frac{1}{2} 1$ | | | | |
| Exposure Duration, construction worker | ED_{cw} | | | $Hazard = \frac{1}{RfD_{d}}$ | $\times C_{soil} \times \frac{m_{cw} \times r}{r}$ | BW_{min} | $\times AT_{m} \times 365$ | $\frac{1}{2} \times LC_{cw} \times SM_{cw} \times CL$ | | | |
| Exposure Frequency, construction worker | EF_{cw} | 25 | 50 days/yr | -5- u | | = · · <i>cw</i> | nc,cw coc | | 7] | | |
| Event Frequency, construction worker | EV_{cw} | 1 events/day | | | | | | | | | |
| Oral Reference Dose Adjusted for GI Absorption | RfD_d | chemical-specific mg/kg-day | | | | | | | | | |
| Exposed Body Surface Area, construction worker | SA_{cw} | 3300 cm^2 | | | | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specific (mg/kg-day) ⁻¹ | | 1 | | | | | | | |
| | | | | | Cancer | | Hazard | | | | |
| | | C _{soil} | ABS _d | SF _d | Risk | % of | $\mathbf{RfD}_{\mathbf{d}}$ | Quotient | % of | | |
| Analyte | | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | | |
| Tetrachloroethene | | 9.90E+01 | 0.00E+00 | 2.10E-03 | 0.00E+00 | < 1% | 6.00E-03 | 0.00E+00 | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | | |
| Benzo[a]pyrene | | 3.98E+00 | 1.30E-01 | 7.30E+00 | 7.62E-08 | 100% | | | | | |
| | | | | • | Cancer Risk | | • | Hazard Index | | | |
| | | | | Pathway Sums: | 7.6E-08 | | | 0.00E+00 | | | |

Table A-12.NI Nursing Infant: Hypothetical Fishing Platform User: RME Summary - Sandblast Area AOPC

| Definition | Variable | Value | Source | Equations | | |
|--|--------------------|-------------------|-----------------|---|-----|--|
| Infant Risk Adjustment Factor | IRAF | Chemical Specific | ODEQ 2010 | | | |
| Carcinogenic IRAFc | | | | | | |
| Total PCB | IRAFc_pcb | 0.6 | ODEQ 2010 | Infant Cancer Risk = Mother Risk x IRAF | c | |
| Noncancer IRAFnc | | | | | | |
| Total PCB | IRAFnc_pcb | 4 | ODEQ 2010 | Infant Noncancer Hazard = Child HQ x IRAFnc | | |
| | | Mother | | Infant | | |
| | C _{soilt} | Cancer Risk | Hazard Quotient | Cancer Risk Hazard Quotic | ent | |
| Analyte | (mg/kg) | (adult) | (child) | | | |
| Total PCBs as Congeners (KM-based, capped) | 4E-01 | 1.8E-06 | 3.7E-01 | 1.1E-06 1.5 | | |
| | | | | | | |

Definition Variable Equations Value Value Cancer Risk: Age-dependent Adjustment Factor, 0-2 ADAF₀₋₂ 10 yrs Age-dependent Adjustment Factor, 2-6 ADAF2-6 Nonmutagens: 3 yrs Age-dependent Adjustment Factor, 6-16 ADAF₆₋₁₆ 3 yrs $Risk = SF_o \times \left| C_{soil} \times EF_r \times FI \times CF_o \times \left(\frac{IRS_a \times ED_a}{BW_a \times AT_c \times 365 \frac{day}{vear}} + \frac{IRS_c \times ED_c}{BW_c \times AT_c \times 365 \frac{day}{vear}} \right) \right|$ Age-dependent Adjustment Factor, 16-26 ADAF₁₆₋₂₆ 1 yrs Averaging Time, carcinogens 70 yrs AT_c Averaging Time, noncarcinogens, adult AT_{nc a} 20 yrs Mutagens: $Risk = SF_0 \times C_{soil} \times EF_r \times FI \times CF_0$ Averaging Time, noncarcinogens, child AT_{nc.c} 6 yrs Body Weight BW. 80 kg $\times \left(\frac{IRS_c \times ED_{0-2} \times ADAF_{0-2}}{BW_c \times AT_c \times 365 \text{ days/year}} + \frac{IRS_c \times ED_{2-6} \times ADAF_{2-6}}{BW_c \times AT_c \times 365 \text{ days/year}}\right)$ Body Weight BW_c 15 kg Conversion Factor CF_{o} 1E-06 kg/mg $+\frac{IRS_a \times ED_{6-16} \times ADAF_{6-16}}{BW_a \times AT_c \times 365 \ days/year} + \frac{IRS_a \times ED_{16-26} \times ADAF_{16-26}}{BW_a \times AT_c \times 365 \ days/year}\right)$ COPC Concentration in soil C_{soil} chemical-specific mg/kg Exposure Duration, child 0-2 ED_{0-2} 2 yrs Exposure Duration, child 2-6 Vinyl Chloride: ED_{2-6} 4 yrs $Risk = SF_{0} \times \left\{ C_{soil} \times FI \times CF_{0} \times \left| \frac{EF_{r} \times \left(\frac{ED_{c} \times IKS_{c}}{BW_{c}} + \frac{ED_{a} \times IKS_{a}}{BW_{a}}\right)}{AT_{c} \times 365 \frac{days}{yaar}} + \frac{IRS_{c}}{BW_{c}} \right| \right\}$ Exposure Duration, child 6-16 ED₆₋₁₆ 10 yrs Exposure Duration, adult 16-26 ED₁₆₋₂₆ 10 yrs Exposure Duration, adult ED_{a} 20 yrs Exposure Duration, child ED_{c} 6 yrs Exposure Frequency, resident EF. 365 days/yr Noncancer Hazard: Fraction Contaminated Soil Ingested 1.0 unitless FI $Hazard = \frac{1}{RfD_{c}} \times \left[C_{soil} \times \left(\frac{IRS_{c} \times ED_{c} \times EF_{r} \times FI \times CF_{o}}{BW_{c} \times AT_{roc} \times 365 \text{ day/year}} \right) \right]$ Soil Ingestion Rate (adult) IRS_a 100 mg/day Soil Ingestion Rate (child) IRS_c 200 mg/dav Oral Reference Dose **R**fD₀ chemical-specific (mg/kg-day) chemical-specific (mg/kg-day)⁻¹ Oral Slope Factor SFo Cancer Noncancer Noncancer C_{soil} SF_o Risk % of RfD_o Hazard (Child) % of Hazard (Adult) (dimensionless) (dimensionless) Analyte Mutagen? (mg/kg) $(mg/kg-day)^{-1}$ Total (mg/kg-day) (dimensionless) Total Inorganic Constituents 9.71E+00 1.31E-05 Arsenic 0.00E+00 1.50E+006% 3.00E-04 2.59E-01 30% 2.43E-02 Chromium (III) 0.00E+00 5.79E+02 No Toxicity Value ---1.50E+00 5.15E-03 < 1% 4.83E-04 ---Lead 0.00E+00 4.18E+02 No Toxicity Value --------------Nickel 2.00E-02 1.67E-01 1.57E-02 0.00E+00 2.51E+02 No Toxicity Value 19% ------Polychlorinated Biphenyls (PCBs) - Mixtures Total PCBs 0.00E+00 4.19E-01 2.00E+00 1.26E-06 < 1% 2.00E-05 2.79E-01 32% 2.62E-02 Volatile Organic Compounds (VOCs) Tetrachloroethene 0.00E+00 2.63E+01 2.10E-03 8.28E-08 < 1%6.00E-03 5.84E-02 7% 5.48E-03 Trichloroethene 0.00E+00 1.78E+00 4.60E-02 1.23E-07 < 1% 5.00E-04 4.75E-02 5% 4.45E-03

1.40E-02

1.62E-06

< 1%

2.00E-02

 Table A-12.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: RME, Sandblast Area AOPC

4.82E-03

5.14E-02

6%

DEHP

Semivolatile Organic Compounds (SVOCs)

0.00E+00

7.71E+01

Age-dependent Adjustment Factor, 0-2 ADAF₀₋₂ Cancer Risk: 10 yrs Age-dependent Adjustment Factor, 2-6 ADAF2-6 Nonmutagens: 3 yrs Age-dependent Adjustment Factor, 6-16 ADAF₆₋₁₆ 3 yrs $Risk = SF_o \times \left| C_{soil} \times EF_r \times FI \times CF_o \times \left(\frac{IRS_a \times ED_a}{BW_a \times AT_c \times 365 \frac{day}{vear}} + \frac{IRS_c \times ED_c}{BW_c \times AT_c \times 365 \frac{day}{vear}} \right) \right|$ Age-dependent Adjustment Factor, 16-26 ADAF₁₆₋₂₆ 1 yrs 70 yrs Averaging Time, carcinogens AT_c Averaging Time, noncarcinogens, adult AT_{nc.a} 20 yrs Mutagens: $Risk = SF_0 \times C_{soil} \times EF_r \times FI \times CF_0$ Averaging Time, noncarcinogens, child AT_{nc.c} 6 yrs Body Weight BW. 80 kg $\times \left(\frac{IRS_c \times ED_{0-2} \times ADAF_{0-2}}{BW_c \times AT_c \times 365 \text{ days/year}} + \frac{IRS_c \times ED_{2-6} \times ADAF_{2-6}}{BW_c \times AT_c \times 365 \text{ days/year}}\right)$ Body Weight BW_c 15 kg Conversion Factor CF_{o} 1E-06 kg/mg $+\frac{IRS_a \times ED_{6-16} \times ADAF_{6-16}}{BW_a \times AT_c \times 365 \ days/year} + \frac{IRS_a \times ED_{16-26} \times ADAF_{16-26}}{BW_a \times AT_c \times 365 \ days/year}\right)^{-1}$ COPC Concentration in soil C_{soil} chemical-specific mg/kg Exposure Duration, child 0-2 ED_{0-2} 2 yrs Exposure Duration, child 2-6 Vinyl Chloride: ED_{2-6} 4 yrs $Risk = SF_{O} \times \left\{ C_{soil} \times FI \times CF_{O} \times \left| \frac{EF_{r} \times \left(\frac{ED_{c} \times IRS_{c}}{BW_{c}} + \frac{ED_{a} \times IRS_{a}}{BW_{a}}\right)}{AT_{c} \times 365 \frac{days}{days}} + \frac{IRS_{c}}{BW_{c}} \right| \right\}$ Exposure Duration, child 6-16 ED₆₋₁₆ 10 yrs Exposure Duration, adult 16-26 ED₁₆₋₂₆ 10 yrs Exposure Duration, adult ED_{a} 20 yrs Exposure Duration, child ED_{c} 6 yrs Exposure Frequency, resident EF. 365 days/yr Noncancer Hazard: Fraction Contaminated Soil Ingested 1.0 unitless FI $Hazard = \frac{1}{RfD_{c}} \times \left[C_{soil} \times \left(\frac{IRS_{c} \times ED_{c} \times EF_{r} \times FI \times CF_{o}}{BW_{c} \times AT_{rec} \times 365 \text{ dav/year}} \right) \right]$ Soil Ingestion Rate (adult) IRS_a 100 mg/day Soil Ingestion Rate (child) IRS_c 200 mg/dav **R**fD_O Oral Reference Dose chemical-specific (mg/kg-day) chemical-specific (mg/kg-day)⁻¹ Oral Slope Factor SFo Cancer Noncancer Noncancer C_{soil} SF_o Risk % of RfD_o Hazard (Child) % of Hazard (Adult) (dimensionless) (dimensionless) Analyte Mutagen? (mg/kg) $(mg/kg-day)^{-1}$ Total (mg/kg-day) (dimensionless) Total **Polycyclic Aromatic Hydrocarbons (PAHs)** Benzo[a]anthracene Μ 2.45E+007.30E-01 1.22E-05 6% No Toxicity Value ---------Μ 2.39E+00 1.19E-04 57% Benzo[a]pyrene 7.30E+00 No Toxicity Value ---------Benzo[b]fluoranthene Μ 2.21E+00 7.30E-01 1.10E-05 5% No Toxicity Value ---------Benzo[g,h,i]perylene 0.00E+00 1.05E+00No Toxicity Value ------3.00E-02 4.67E-04 < 1%4.38E-05 Benzofluoranthenes, Total Μ 6.36E+00 7.30E-01 3.16E-05 15% No Toxicity Value --------Μ 2.38E-01 7.30E+00 1.18E-05 6% No Toxicity Value Dibenz[a,h]anthracene ---------Μ 7.30E-01 Indeno[1,2,3-cd]pyrene 1.16E+00 5.77E-06 3% No Toxicity Value ---------Cancer Risk Hazard Index (Child) Hazard Index (Adult) Pathway Sum: 2.1E-04 Pathway Sum: 0.87 0.081

 Table A-12.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: RME, Sandblast Area AOPC

Equations

Value

Definition

Variable

Value

 Table A-12.2

 Derivation of Inhalation Factors for Inhalation Exposures, Hypothetical Fishing Platform User: RME, Sandblast Area AOPC

1.10412E-05 Experimental, EPISuite v4.10 (USEPA 2011a)

0.00049072 Experimental, EPISuite v4.10 (USEPA 2011a)

1.86882E-05 Experimental, EPISuite v4.10 (USEPA 2011a)

2.68669E-05 Experimental, EPISuite v4.10 (USEPA 2011a)

1.35357E-05 Experimental, EPISuite v4.10 (USEPA 2011a)

2.68669E-05 Experimental, EPISuite v4.10 (USEPA 2011a)

5.76596E-06 Experimental, EPISuite v4.10 (USEPA 2011a)

0.0000656 EPISuite v4.10 (USEPA 2011a)

119600 EPISuite v4.10

176900EPISuite v4.10587400EPISuite v4.10

599400 EPISuite v4.10

1951000 EPISuite v4.10

599400 EPISuite v4.10

1912000 EPISuite v4.10

3470000 EPISuite v4.10

| Description | Variable | Value | Units | | VF Derivation | | | | |
|---|-----------------------|--|--|----------------------|--|--------------------------------|---|------------|--------------------|
| Unit conversion factor | CFi | 1.00E-0 | $14 \text{ m}^2/\text{cm}^2$ | | | | | | |
| Apparent Diffusivity | D _A | deriv | ed cm ² /s | | $VFs = Q/C_{vol} \times \frac{(3.14 \times C_{vol})}{(2.14 \times C_{vol})}$ | $(D_A \times T_{resident})^1$ | /2 | | |
| Diffusivity in air | Di | chemical-specif | ic cm ² /s | | $VFS = Q/C_{vol} \times \frac{1}{(1-1)^{1/2}}$ | $2 \times \rho_b \times D_A$) | | | |
| Diffusivity in water | D _w | chemical-specif | - | | | | | | |
| Default organic-carbon content | F _{OC} | - | 06 g/g | | | | | | |
| Henry's law constant | H' | | ic dimensionless | | | | | | |
| Soil-water partition coefficient | K _d | chemical-specif | - | | | | | | |
| for organics: Kd= K_{OC} × F_{OC} | u | 1 | C . | i | based on: | | | | |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specif | ic cm ³ /g | | $[(-10)^3)$ | - 10/2 -) - 21 | | | |
| Total soil porosity | n | 0.43 L _{porespace} /L _{soil} | | | $D_A = \frac{\left[\left(\theta_a^{10/3} D_i H' + \rho_B K_d + \theta\right)\right]}{\rho_B K_d + \theta}$ | $\theta_w^{10/3} D_w)/n^2]$ | | | |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.360E+0 | | | $\rho_B K_d + \theta$ | $\theta_w + \theta_a H'$ | | | |
| Inverse of the mean concentration at the center | Q/C _{vol} | | 14 g/m ² -s per kg/m ³ | | | | | | |
| of a 0.5-acre square source in Los Angeles | Q/ UV01 | 75. | | | | | | | |
| Dry soil bulk density | ρ_{b} | 1 | .5 g/cm ³ | | | | | | |
| Air-filled soil porosity | θ_{a} | 0.2 | 28 L _{air} /L _{soil} | | | | | | |
| Water-filled soil porosity | $\theta_{\rm w}$ | 0. | 5 L _{water} /L _{soil} | | | | | | |
| Exposure interval, 30-yr resident | T _{Resident} | 8.20E+0 | 08 seconds | | | | | | |
| Volatilization Factor for soil | VFs | derive | ed m ³ /kg | | | | | | |
| | | | \mathbf{D}_{i} | | $\mathbf{D}_{\mathbf{w}}$ | | Н' | | K |
| Analyte | | (cm ² /s) | Reference | (cm ² /s) | Reference | (dimensionless) | Reference | (cm^3/g) | |
| Inorganic Constituents | | | | | | | | | |
| Arsenic | | 0 | 0 | 0 |) | 31.61052561 | Bond Method estimate, EPISuite v4.10 (USEPA, 2011a) | 13.22 | EPISuite v4.10 (US |
| Chromium (III) | | 0 | 0 | 0 |) | 0 | 0 | 0 | 0 |
| Lead | | 0 | 0 | 0 |) | 1.001885999 | Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 | EPISuite v4.10 (US |
| Nickel | | 0 | 0 | 0 |) | 1.001885999 | Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 | EPISuite v4.10 (US |
| Polychlorinated Biphenyls (PCBs) - Mixtures | | | | | | | | | |
| Total PCBs | | 0.040076 | RSLs ("WATER9"; USEPA 2012) | 4.7E-06 | RSLs ("WATER9"; USEPA 2012) | 0.011572806 | Experimental, EPISuite v4.10 (USEPA 2011a) | 130500 | EPISuite v4.10 (US |
| Volatile Organic Compounds (VOCs) | | | | | | | | | |
| Tetrachloroethene | | 0.0504664 | RSLs ("WATER9"; USEPA 2012) | | RSLs ("WATER9"; USEPA 2012) | 0.723811518 | Experimental, EPISuite v4.10 (USEPA 2011a) | 94.94 | EPISuite v4.10 (US |
| Trichloroethene | | 0.0686618 | RSLs ("WATER9"; USEPA 2012) | 1E-05 | RSLs ("WATER9"; USEPA 2012) | 0.402799065 | Experimental, EPISuite v4.10 (USEPA 2011a) | 60.7 | EPISuite v4.10 (US |

4.2E-06 RSLs ("WATER9"; USEPA 2012)

5.9E-06 RSLs ("WATER9"; USEPA 2012)

5.6E-06 RSLs ("WATER9"; USEPA 2012)

5.6E-06 RSLs ("WATER9"; USEPA 2012)

5.6E-06 RSLs ("WATER9"; USEPA 2012)

5.2E-06 RSLs ("WATER9"; USEPA 2012)

5.2E-06 RSLs ("WATER9"; USEPA 2012)

0 0

Semivolatile Organic Compounds (SVOCs)

Polycyclic Aromatic Hydrocarbons (PAHs)

0.0173403

0.0508647

0.0475831

0.0475831

0

0.0475831

0.0445672

0.0447842

RSLs ("WATER9"; USEPA 2012)

DEHP

Benzo[a]anthracene

Benzo[b]fluoranthene Benzo[g,h,i]perylene Benzofluoranthenes, Total

Dibenz[a,h]anthracene

Indeno[1,2,3-cd]pyrene

Benzo[a]pyrene

| | | | | DEE |
|-------------------------------|---------------------------|---------------------------|--------------------|--------------------|
| | | _ | VFs (VOCs) | PEF |
| K _{oc} | $\mathbf{K}_{\mathbf{d}}$ | $\mathbf{D}_{\mathbf{A}}$ | Residential | (non-VOCs) |
| Reference | (cm^3/g) | (cm^2/s) | m ³ /kg | m ³ /kg |
| | | | | |
| (USEPA, 2011a); MCI-estimated | 29 | 0 | non-VOC | 1.360E+09 |
| | 1800000 | 0 | non-VOC | 1.360E+09 |
| (USEPA 2011a); MCI-estimated | 900 | 0 | non-VOC | 1.360E+09 |
| (USEPA 2011a); MCI-estimated | 65 | 0 | non-VOC | 1.360E+09 |
| | | | | |
| (USEPA 2011a); MCI-estimated | 0 | 0.00024 | non-VOC | 1.360E+09 |
| | | | | |
| (USEPA 2011a); MCI-estimated | 0 | 0.00805 | 1384.893971 | use VF |
| (USEPA 2011a); MCI-estimated | 0.3642 | 0.00266 | 2410.681404 | use VF |
| | | | | |
| (USEPA 2011a); MCI-estimated | 0 | 3.7E-07 | non-VOC | 1.360E+09 |
| | | | | |
| (USEPA 2011a); MCI-estimated | 0 | 1.3E-05 | non-VOC | 1.360E+09 |
| (USEPA 2011a); MCI-estimated | 0 | 8.2E-07 | non-VOC | 1.360E+09 |
| (USEPA 2011a); MCI-estimated | 0 | 1E-06 | non-VOC | 1.360E+09 |
| (USEPA 2011a); MCI-estimated | 0 | 0 | non-VOC | 1.360E+09 |
| (USEPA 2011a); MCI-estimated | 0 | 1E-06 | non-VOC | 1.360E+09 |
| (USEPA 2011a); MCI-estimated | 0 | 4.7E-07 | non-VOC | 1.360E+09 |
| (USEPA 2011a) | 0 | 1.9E-06 | non-VOC | 1.360E+09 |

 Table A-12.3

 Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Sandblast Area AOPC

| Description | Variable | Value | e | Equations | | | | | | | |
|--|-----------------------|-------------------|---------------------------|-----------------------------|---|---|---|---|---|--------------|-----------------|
| Averaging Time, carcinogens | AT _c | 71 | 0 yrs | Cancer Risk: | | | | | | | |
| Averaging Time, noncarcinogens, resident | AT _{nc,r} | 20 | 6 yrs | Nonmutagens: | | | | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 | 0 yrs | | $Pisk - IIIR \times C$ | $ED_r \times E$ | $F_r \times ET_r$ | | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | | 3 yrs | | $\pi i s \kappa = I 0 \pi \wedge C_{c}$ | $\frac{1}{4T \times 365} dc$ | $\frac{1}{24}$ hour | | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | | 3 yrs | | ſ | $\left(\prod_{c} \times 303 \text{ ye} \right)$ | $ar^{24} day$ | | | | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | | 1 yrs | Mutagens: | $Risk = IUR \times C_a$ | $_{uir} \times EF_r \times ET_r$ | | $\frac{ED_{2}}{AT_c \times 365 \ days} + \frac{ED_{16-26}}{AT_c \times 365 \ days}$ | | | |
| Averaging Time, carcinogens | AT _c | 70 | 0 yrs | | L | (ED_{n-1}) | ADAFo o | EDa | $\langle \times ADAF_{2} \rangle$ | | |
| COPC Concentration in air | C _{air} | chemical-specific | c μg/m ³ | | × | $\frac{1}{4T \times 365 days/v}$ | par × 24 hours | $\frac{-2}{4\pi}$ + $\frac{-2}{4\pi}$ × 365 days | s/vear × 24 hours | Iday | |
| COPC Concentration in soil | C _{soil} | chemical-specifie | c mg/kg | | | $(\Pi_c \land 303 uuys/y)$ | | $H_c \wedge 505 uuys$ | $\sim ADAF$ |)1 | |
| Exposure Duration, child 0-2 | ED ₀₋₂ | <i>.</i> | 2 yrs | | + | 4T x 26T days (200 | $\frac{ADAP_{6-16}}{2}$ | $\frac{LD_{16-26}}{4\pi} + \frac{LD_{16-26}}{4\pi}$ | $5 \times ADAP_{16-26}$ | <u> </u> | |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 4 | 4 yrs | | | $AI_c \times 365 \ aays/ye$ | ar × 24 nours/c | $ay AI_c \times 365 aays,$ | /year × 24 nours/ | aay)] | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 | 0 yrs | Vinyl Chloride | ». Г | (FD | $\times FF \times FT$ |)] | | | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 | 0 yrs | | $Risk = IUR \times C_a$ | $_{ir} \times \left\{ \frac{ED}{AT_c \times 365 day} \right\}$ | $r \wedge L_r \wedge L_r$ | $\frac{1}{1}$ | | | |
| Exposure Duration, child | ED_{c} | (| 6 yrs | | L | $(AI_c \times 365 aa)$ | /s/year × 24 no | urs/aay)] | | | |
| Exposure Duration, resident | ED_r | | 6 yrs | | | | | | | | |
| Exposure Frequency | EF_r | 36. | 5 days/yr | | | | | | | | |
| Exposure Time | ET_r | 24 | 4 hours/day | Noncancer Haz | ard: 1 | (ED - | $\vee FF \vee FT$ | \ | | | |
| Inhalation Unit-Risk Factor | IUR | chemical-specifi | $(\mu g/m^3)^{-1}$ | | Hazard = $\frac{1}{DCC} \times$ | $C_{air} \times \left(- \frac{ED_c}{C} \right)$ | $\frac{\Delta E_{r} \wedge E_{r}}{day}$ | <u></u>) | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+09 | $9 \text{ m}^3/\text{kg}$ | | $Hazard = \frac{1}{RfC} \times$ | $AT_{nc,r} \times 36$ | $55 \frac{uuy}{vear} \times 24 \frac{nou}{day}$ | | | | |
| Inhalation Reference Concentration | RfC | chemical-specifi | | | | | yeur uu | | | | |
| Volatilization Factor for VOCs | VF | chemical-specific | | Where: | | | | $C_{air} = \frac{C_{soil}}{VF \text{ or } PE}$ | $\frac{\mu g}{\mu} \times 1000 \frac{\mu g}{\mu}$ | | |
| | | | VF (VOCs) or | | | Cancer | | VF OF PE | Noncancer | | Noncancer |
| | | C _{soil} | PEF (non-VOCs) | $\mathbf{C}_{\mathbf{air}}$ | IUR | Risk | % of | RfC | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (m ³ /kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total | (dimensionless) |
| Inorganic Constituents | | | | | | · · · · · · | | | × * | | × / |
| Arsenic | 0.00E+00 | 9.71E+00 | 1.360E+09 | 7.14E-06 | 4.30E-03 | 6.84E-09 | < 1% | 1.50E-02 | 2.86E-04 | < 1% | 2.86E-04 |
| Chromium (III) | 0.00E+00 | 5.79E+02 | 1.360E+09 | 4.26E-04 | No Toxicity Value | | | No Toxicity Value | | | |
| Lead | 0.00E+00 | 4.18E+02 | 1.360E+09 | 3.07E-04 | | | | No Toxicity Value | | | |
| Nickel | 0.00E+00 | 2.51E+02 | 1.360E+09 | 1.85E-04 | 2.60E-04 | 1.78E-08 | < 1% | 9.00E-02 | 2.05E-03 | < 1% | 2.05E-03 |
| Polychlorinated Biphenyls (PCBs) - Mixtu | | 2.5112+02 | 1.500E+07 | 1.051-04 | 2.001-04 | 1.76L-06 | < 170 | J.00L-02 | 2.05E-05 | < 1 /0 | 2.0512-05 |
| Total PCBs | 0.00E+00 | 4.19E-01 | 1.360E+09 | 3.08E-07 | 5.70E-04 | 6.52E-11 | < 1% | No Toxicity Value | | | |
| | 0.00E+00 | 4.19E-01 | 1.300E+09 | 3.08E-07 | 3.70E-04 | 0.32E-11 | < 1% | No Toxicity value | | | |
| Volatile Organic Compounds (VOCs) | 0.005.00 | 2 (25, 01 | 1.0055.00 | 1.005.01 | A (0E 05 | 1.005.07 | C1 0/ | 4.005.01 | | 5 60/ | |
| Tetrachloroethene | 0.00E+00 | 2.63E+01 | 1.385E+03 | 1.90E+01 | 2.60E-07 | 1.83E-06 | 61% | 4.00E+01 | 4.75E-01 | 56% | 4.75E-01 |
| Trichloroethene | 0.00E+00 | 1.78E+00 | 2.411E+03 | 7.38E-01 | 4.10E-06 | 1.12E-06 | 38% | 2.00E+00 | 3.69E-01 | 44% | 3.69E-01 |
| Semivolatile Organic Compounds (SVOCs) | | | | | | | | | | | |
| DEHP | 0.00E+00 | 7.71E+01 | 1.360E+09 | 5.67E-05 | 2.40E-06 | 5.05E-11 | < 1% | No Toxicity Value | | | |
| Polycyclic Aromatic Hydrocarbons (PAHs) |) | | | | | | | | | | |
| Benzo[a]anthracene | М | 2.45E+00 | 1.360E+09 | 1.80E-06 | 1.10E-04 | 2.04E-10 | < 1% | No Toxicity Value | | | |
| Benzo[a]pyrene | М | 2.39E+00 | 1.360E+09 | 1.76E-06 | 1.10E-03 | 1.99E-09 | < 1% | No Toxicity Value | | | |
| Benzo[b]fluoranthene | М | 2.21E+00 | 1.360E+09 | 1.63E-06 | 1.10E-04 | 1.84E-10 | < 1% | No Toxicity Value | | | |
| Benzo[g,h,i]perylene | 0.00E+00 | 1.05E+00 | 1.360E+09 | 7.72E-07 | No Toxicity Value | | | No Toxicity Value | | | |
| Benzofluoranthenes, Total | М | 6.36E+00 | 1.360E+09 | 4.68E-06 | 1.10E-04 | 5.29E-10 | < 1% | No Toxicity Value | | | |
| Dibenz[a,h]anthracene | M | 2.38E-01 | 1.360E+09 | 1.75E-07 | 1.20E-03 | 2.16E-10 | < 1% | No Toxicity Value | | | |
| | M | 1.16E+00 | 1.360E+09 | 8.53E-07 | 1.10E-04 | 9.65E-11 | < 1% | No Toxicity Value | | | |
| Indeno[1 2 3-cd]nyrene | 111 | 1.100-00 | 1.5000-00 | 0.556-07 | 1.100-04 | 7.056-11 | \ 1 /0 | The reality value | | | |
| Indeno[1,2,3-cd]pyrene | | | | | | | | | | | |

| Description | Variable | Value | Equations: |
|--|----------------------------|---|--|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specific unitless | Cancer Risk: |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 yrs | Nonmutagens: $Risk = SF_d \times [C_{soil} \times ED_a \times EF \times EV \times EC \times A]$ |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 yrs | u l'obti u |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 yrs | $\times \left(\frac{AF_a \times SA_a}{BW_a \times AT_c \times 365 \ dav/vear} + \frac{AF_c \times B}{BW_c \times AT_c \times 365 \ dav/vear} + \frac{AF_c \times B}{B$ |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 yrs | $(BW_a \times AT_c \times 365 \ day/year \ BW_c \times AT_c \times$ |
| Soil-to-Skin Adherence Fraction | AF_a | 0.07 mg/cm^2 -day | Mutagens: |
| Soil-to-Skin Adherence Fraction | AF_{c} | 0.2 mg/cm^2 -day | $Risk = SF_d \times \left\{ C_{soil} \times EF_r \times EV \times EC \times ABS_d \times CF_d \right\}$ |
| Averaging Time, Carcinogens | AT _c | 70 yrs | |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 20 yrs | $\times \left \frac{ED_{0-2} \times AF_c \times SA_c \times ADAF_{0-2}}{BW_c \times AT_c \times 365 \ days/year} + \frac{ED_{2-6} \times ABAF_{0-2}}{BW_c \times ABAF_{0-2}} \right $ |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | 6 yrs | |
| Body Weight, adult | \mathbf{BW}_{a} | 80 kg | $+\frac{ED_{6-16} \times AF_a \times SA_a \times ADAF_{6-16}}{BW_a \times AT_c \times 365 \ days/year} + \frac{ED_{16-26}}{BW_a}$ |
| Body Weight, child | BW_{c} | 15 kg | $T BW_a \times AT_c \times 365 days/year BW_a$ |
| Conversion Factor | CF_d | 1E-06 kg/mg | Vinyl Chloride: $Risk = SF_d \times \{C_{soil} \times EF_r \times EV \times EC \times ABS_d \times EV \times EC \times EV \times EV$ |
| COPC Concentration in Soil | C_{soil} | chemical-specific mg/kg | $RISR = SF_d \times \{C_{soil} \times EF_r \times EV \times EC \times ABS_d \times A$ |
| Fraction of EV in Contact with Soil | EC | 1 unitless | $\begin{bmatrix} ED_c \times AF_c \times SA_c \end{bmatrix} ED_a \times AF_a \times SA_a$ |
| Exposure Duration, child 0-2 | ED ₀₋₂ | 2 yrs | $EF_r \times \left(\frac{BW_c}{BW_c} + \frac{BW_a}{BW_a} \right)$ |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 4 yrs | $\times \frac{\frac{EF_r \times \left(\frac{ED_c \times AT_c \times SA_c}{BW_c} + \frac{ED_a \times AT_a \times SA_a}{BW_a}\right)}{AT_c \times 365 \frac{days}{max}}$ |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 yrs | $H_c \times 303 \overline{year}$ |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 yrs | |
| Exposure Duration, adult | ED_{a} | 20 yrs | Noncancer Hazard: |
| Exposure Duration, child | ED_{c} | 6 yrs | - 1 Γ (AF _a ×ABS _d ×ED _{max} ×EF _a ×EV×EC× |
| Exposure Frequency, resident | $\mathrm{EF_{r}}$ | 365 days/yr | $Hazard = \frac{1}{RfD_d} \times \left[C_{soil} \times \left(\frac{AF_c \times ABS_d \times ED_{nc,c} \times EF_r \times EV \times EC \times EV}{BW_c \times AT_{nc,c} \times 365 \frac{day}{com}} \right) \right]$ |
| Event Frequency | EV | 1 events/day | y u [(Dwc/mnc,c/303 year |
| Refererenc Dose Adjusted for GI Absorption | RfD_d | chemical-specific mg/kg-day | |
| Exposed Body Surface Area, adult | SA_a | 6032 cm^2 | |
| Exposed Body Surface Area, child | SA_c | 2373 cm^2 | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specific (mg/kg-day) ⁻¹ | |
| | | | Cancer N |
| | | C _{soil} ABS _d | SF _d Risk % of RfD _d Haz |

| Table A-12.4 |
|---|
| Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: RME, Sandblast Area AO |

| | | | | | Cancer | | | N |
|---|----------|-------------------|------------------|---------------------------|-----------------|-------|-----------------------------|------|
| | | C _{soil} | ABS _d | SF_d | Risk | % of | $\mathbf{RfD}_{\mathbf{d}}$ | Haz |
| Analyte | Mutagen? | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (din |
| Inorganic Constituents | | | | | | | | |
| Arsenic | 0.00E+00 | 9.71E+00 | 3.00E-02 | 1.50E+00 | 1.11E-06 | 2% | 3.00E-04 | 1 |
| Chromium (III) | 0.00E+00 | 5.79E+02 | 1.00E-02 | | | | 1.95E-02 | 9 |
| Lead | 0.00E+00 | 4.18E+02 | 1.00E-02 | | | | | |
| Nickel | 0.00E+00 | 2.51E+02 | 1.00E-02 | | | | 8.00E-04 | 9 |
| Polychlorinated Biphenyls (PCBs) - Mixtures | | | | | | | | |
| Total PCBs | 0.00E+00 | 4.19E-01 | 1.40E-01 | 2.00E+00 | 4.95E-07 | < 1% | 2.00E-05 | 9 |
| Volatile Organic Compounds (VOCs) | | | | | | | | |
| Tetrachloroethene | 0.00E+00 | 2.63E+01 | 0.00E+00 | 2.10E-03 | 0.00E+00 | < 1% | 6.00E-03 | 0 |
| Trichloroethene | 0.00E+00 | 1.78E+00 | 0.00E+00 | 4.60E-02 | 0.00E+00 | < 1% | 5.00E-04 | 0 |
| Semivolatile Organic Compounds (SVOCs) | | | | | | | | |
| DEHP | 0.00E+00 | 7.71E+01 | 1.00E-01 | 1.40E-02 | 4.56E-07 | < 1% | 2.00E-02 | 1 |
| | | | | | | | | |

| Description | Variable | Value | e | Equations: | | | | |
|--|----------------------------|-------------------|-----------------------------|----------------------------|---|--|---|---------------------|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specifie | c unitless | Cancer Risk: | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 |) yrs | Nonmutagens: | Dick - SE | $(C_{soil} \times ED_a \times ED_b)$ | | EC × A |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | | 3 yrs | | | - 5000 00 | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | | 3 yrs | | $\times \left($ | $4F_a \times SA_a$ T × 365 day/y | | $AF_c \times EI$ |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | | l yrs | | $(BW_a \times AT)$ | $T_c \times 365 day/y$ | rear ' $BW_c \times$ | $AT_c \times C$ |
| Soil-to-Skin Adherence Fraction | AF_{a} | 0.07 | 7 mg/cm ² -day | Mutagens: | (| | | |
| Soil-to-Skin Adherence Fraction | AF_{c} | 0.2 | 2 mg/cm ² -day | $Risk = SF_d$ | $\times \left\{ C_{soil} \times EF_r \times \right.$ | $\times EV \times EC \times AE$ | $SS_d \times CF_d$ | |
| Averaging Time, Carcinogens | AT _c | 70 |) yrs | | | | | |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 20 |) yrs | | $\times \left \frac{ED_{0-2} \times D_{0-2}}{2} \right $ | $AF_c \times SA_c \times AL$ | $\frac{DAF_{0-2}}{2} + \frac{ED}{2}$ | $_{2-6} \times A$ |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | (| 5 yrs | | $BW_c \times A$ | T _c × 365 days | /year BV | $N_c \times AT$ |
| Body Weight, adult | \mathbf{BW}_{a} | 80 |) kg | | $\perp \frac{ED_{6-16} \times 1}{ED_{6-16}}$ | $AF_c \times SA_c \times AI$ $T_c \times 365 \ days$ $AF_a \times SA_a \times AI$ $AT_c \times 365 \ days$ | $DAF_{6-16} \perp \frac{E}{2}$ | D_{16-26} |
| Body Weight, child | BW_{c} | 1: | 5 kg | | $\overline{W}_a \times A$ | $AT_c \times 365 \ days$ | /year [–] | $BW_a \times$ |
| Conversion Factor | CF_d | 1E-06 | 6 kg/mg | Vinyl Chloride: | $Risk = SF_d \times \{$ | | | |
| COPC Concentration in Soil | C_{soil} | chemical-specifie | c mg/kg | | $RISK = SF_d \times f$ | $C_{soil} \times EF_r \times E$ | EV X EL X AI | $SS_d \times C$ |
| Fraction of EV in Contact with Soil | EC | | l unitless | | $\begin{bmatrix} EE \\ EE \end{bmatrix} (ED_{a})$ | $\times AF_c \times SA_c$ | $ED_a \times AF_a$ | $\times SA_a$ |
| Exposure Duration, child 0-2 | ED ₀₋₂ | | 2 yrs | | $EF_r \times ($ | BW_c + | $\overline{BW_a}$ | <u> </u> |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 2 | 4 yrs | | $\times \left \frac{EF_r \times \left(\frac{ED_c}{2} \right)}{2} \right $ | $AT \times 365^{0}$ | days | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 |) yrs | | L | $H_c \times 505$ | year | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 |) yrs | | | | | |
| Exposure Duration, adult | ED_{a} | |) yrs | Noncancer Hazard | l: | | | |
| Exposure Duration, child | ED_{c} | | 5 yrs | | 1 1[c | $(AF_c \times ABS_d)$ | $X \times ED_{nc,c} \times EF_r \times E$ | EV×EC×SA |
| Exposure Frequency, resident | $\mathrm{EF_{r}}$ | | 5 days/yr | Haz | $ard = \frac{1}{RfD_d} \times \left[C_{d}\right]$ | soil × (| $BW_c \times AT_{ncc} \times 36$ | $55\frac{day}{day}$ |
| Event Frequency | EV | | l events/day | | L | X | - 110,0 | year |
| Reference Dose Adjusted for GI Absorption | RfD_d | chemical-specific | | | | | | |
| Exposed Body Surface Area, adult | SA_a | | 2 cm^2 | | | | | |
| Exposed Body Surface Area, child | SA_c | 2373 | 3 cm^2 | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specific | c (mg/kg-day) ⁻¹ | | | | | |
| | | | | | Cancer | | | No |
| | | C _{soil} | ABS _d | $\mathbf{SF}_{\mathbf{d}}$ | Risk | % of | RfD _d | Haza |
| Analyte | Mutagen? | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dime |

Table A-12.4 Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: RME, Sandblast Area AOPC

| | | | | | Cancer | | | N |
|---|----------|-------------------|------------------|---------------------------|-----------------|-------|-------------------------|--------|
| | | C _{soil} | ABS _d | SF _d | Risk | % of | RfD _d | Haz |
| Analyte | Mutagen? | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (din |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | |
| Benzo[a]anthracene | М | 2.45E+00 | 1.30E-01 | 7.30E-01 | 4.06E-06 | 6% | | |
| Benzo[a]pyrene | М | 2.39E+00 | 1.30E-01 | 7.30E+00 | 3.96E-05 | 60% | | |
| Benzo[b]fluoranthene | М | 2.21E+00 | 1.30E-01 | 7.30E-01 | 3.67E-06 | 6% | | |
| Benzo[g,h,i]perylene | 0.00E+00 | 1.05E+00 | 1.30E-01 | | | | 3.00E-02 | 1 |
| Benzofluoranthenes, Total | М | 6.36E+00 | 1.30E-01 | 7.30E-01 | 1.06E-05 | 16% | | |
| Dibenz[a,h]anthracene | Μ | 2.38E-01 | 1.30E-01 | 7.30E+00 | 3.95E-06 | 6% | | |
| Indeno[1,2,3-cd]pyrene | М | 1.16E+00 | 1.30E-01 | 7.30E-01 | 1.92E-06 | 3% | | |
| | | | | | Cancer Risk | | | Hazard |
| | | | | Pathway Sum: | 6.6E-05 | | Pathway Sur | |

$$\begin{array}{c}
ABS_d \times CF_d \\
\underline{ED_c \times SA_c} \\
\times 365 \ day/year)
\end{array}$$

$$\begin{array}{c}
AF_c \times SA_c \times ADAF_{2-6} \\
AT_c \times 365 \ days/year \\
\underline{6 \times AF_a \times SA_a \times ADAF_{16-26}} \\
\times AT_c \times 365 \ days/year \\
\end{array}$$

$$\begin{array}{c}
F_d \\
\hline + \frac{AF_c \times SA_c}{BW_c} \\
\hline + \frac{BW_c}{BW_c} \\
\end{array}$$
Noncancer Noncancer Rard (Adult) (dimensionless)

$$\begin{array}{c}
\hline -- & -- \\
\hline -- & -$$

Table A-13.NI Nursing Infant: Hypothetical Fishing Platform User: CTE Summary - Sandblast Area AOPC

| Definition | Variable | Value | Source | Equations | |
|--|--------------------|-------------------|-----------------|---------------------------|---------------------|
| Infant Risk Adjustment Factor | IRAF | Chemical Specific | ODEQ 2010 | | |
| Carcinogenic IRAFc | | | | | |
| Total PCB | IRAFc_pcb | 0.6 | ODEQ 2010 | Infant Cancer Risk = | Mother Risk x IRAFc |
| Noncancer IRAFnc | | | | | |
| Total PCB | IRAFnc_pcb | 4 | ODEQ 2010 | Infant Noncancer Hazard = | Child HQ x IRAFnc |
| | | Mot | her | Infa | nt |
| | C _{soilt} | Cancer Risk | Hazard Quotient | Cancer Risk | Hazard Quotient |
| Analyte | (mg/kg) | (adult) | (child) | | |
| Total PCBs as Congeners (KM-based, capped) | 4E-01 | 2.4E-07 | 6.6E-02 | 1.4E-07 | 0.26 |
| | | | | | |

 Table A-13.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: CTE, Sandblast Area AOPC

| Definition | Variable | Valu | ie | Equations | Value | 2 | | | |
|--|--------------------------------------|-------------------|---------------------------|--|----------------------------------|--|--|-----------------------|-------------------------------------|
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 | yrs | Cancer Risk: | | | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 | yrs | Nonmutagens: | | | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 | yrs | 1 | | / IRS. | $\times ED_{a}$ | IRS | $\times ED_c$ \] |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 | yrs | $Risk = SF_o \times C_{soi} $ | $l \times EF_r \times FI$ | $\times CF_{O} \times \left(\frac{IRS_{O}}{BW_{a} \times AT}\right)$ | $\frac{d}{day} + \frac{d}{day}$ | | day |
| Averaging Time, carcinogens | AT _c | 70 | yrs | | | $\left(BW_{a} \times AT \right)$ | $_c \times 365 \overline{\text{year}} B$ | $W_c \times AI_c$ | $\times 365 \overline{\text{year}}$ |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 3 | yrs | Mutagens: | | Υ. | | | / - |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | 6 | yrs | $Risk = SF_O \times \left[C_{so}\right]$ | $a \times EE \times E$ | $I \times CF_{\circ}$ | | | |
| Body Weight | BW_a | 80 | kg | L | | | | | |
| Body Weight | BWc | 15 | kg | × | $(IRS_c \times EL)$ | $\frac{D_{0-2} \times ADAF_{0-2}}{\times 365 days/year} + $ | $IRS_c \times ED_{2-6} \times$ | $ADAF_{2-6}$ | |
| Conversion Factor | CF | 1E-06 | kg/mg | | (| | | | |
| COPC Concentration in soil | C _{soil} | chemical-specific | mg/kg | + - | $IRS_a \times ED_6$ | $\frac{-16 \times ADAF_{6-16}}{365 days/year} + \frac{1}{E}$ | $RS_a \times ED_{16-26} \times D_{16-26}$ | $ADAF_{16-2}$ | <u>.6</u>) |
| Exposure Duration, child 0-2 | ED ₀₋₂ | 1 | yrs | | $BW_a \times AT_c \times$ | : 365 days/year ' E | $W_a \times AT_c \times 365 a$ | lays/yea | r /] |
| Exposure Duration, child 2-6 | ED ₂₋₆ | | yrs | Vinyl Chloride: | | | IRS FD VIR | 5) | 1) |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 | yrs | | a | $CF_O \times \left \frac{EF_r \times \left(\frac{ED_c \times BV}{BV} \right)}{AT_c} \right $ | $\frac{D_a \wedge T_k}{W_c} + \frac{D_a \wedge T_k}{BW_a}$ | $\frac{\sigma_a}{IR}$ | S_c |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 | yrs | $Risk = SF_0 \times \{$ | $C_{soil} \times FI \times I$ | $CF_0 \times $ | days | $-+ \frac{1}{BV}$ | |
| Exposure Duration, adult | ED_a | 3 | yrs | (| | | x 365 year | | ·]) |
| Exposure Duration, child | ED_{c} | 6 | yrs | | | | | | |
| Exposure Frequency, resident | EFr | | days/yr | Noncancer Hazard: | | | | | |
| Fraction Contaminated Soil Ingested | FI | | unitless | . 1 | [_ <i>(II</i> | $RS_c \times ED_c \times EF_r \times F_r$ | $I \times CF_{o}$ | | |
| Soil Ingestion Rate (adult) | IRS _a | | mg/day | $Hazard = \frac{RfD_0}{RfD_0} \times$ | $C_{soil} \times (\overline{B})$ | $RS_c \times ED_c \times EF_r \times F_r$ $W_c \times AT_{ncc} \times 365 da$ | v/year | | |
| Soil Ingestion Rate (child) Oral Reference Dose | IRS _c RfD _O | chemical-specific | mg/day | , 0 | | <i>c nc,c</i> | 577 71 | | |
| Oral Slope Factor | SF ₀ | chemical-specific | | | | | | | |
| | 3F ₀ | chemical-specific | (IIIg/Kg-day) | | | | | | |
| | | C | SF | Cancer Risk | % of | DfD | Noncancer | % of | Noncancer |
| A | M49 | C _{soil} | , , | | | RfD _o | Hazard (Child) | | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total | (dimensionless) |
| Inorganic Constituents | | 0.515.00 | 1.505.00 | 0.100.07 | 50/ | 2.005.04 | 5 205 02 | 2004 | 5.055.02 |
| Arsenic | | 9.71E+00 | 1.50E+00 | 2.18E-06 | 5% | 3.00E-04 | 5.39E-02 | 30% | 5.05E-03 |
| Chromium (III) | | 5.79E+02 | No Toxicity Value | | | 1.50E+00 | 1.07E-03 | < 1% | 1.00E-04 |
| Lead | | 4.18E+02 | | | | No Toxicity Value | | | |
| Nickel | | 2.51E+02 | No Toxicity Value | | | 2.00E-02 | 3.48E-02 | 19% | 3.27E-03 |
| Polychlorinated Biphenyls (PCBs) - Mixture | es | | | | | | | | |
| Total PCBs | | 4.19E-01 | 2.00E+00 | 2.09E-07 | < 1% | 2.00E-05 | 5.82E-02 | 32% | 5.45E-03 |
| Volatile Organic Compounds (VOCs) | | | | | | | | | |
| Tetrachloroethene | | 2.63E+01 | 2.10E-03 | 1.38E-08 | < 1% | 6.00E-03 | 1.22E-02 | 7% | 1.14E-03 |
| Trichloroethene | | 1.78E+00 | 4.60E-02 | 2.04E-08 | < 1% | 5.00E-04 | 9.88E-03 | 5% | 9.27E-04 |
| | | | | | | | | | |
| Semivolatile Organic Compounds (SVOCs) | | | | | | | | | |

 Table A-13.1

 Risk and Hazard Estimates: Incidental Ingestion of Soil, Hypothetical Fishing Platform User: CTE, Sandblast Area AOPC

| Definition | Variable | Value | | Equations | Valu | e | | | |
|---|-----------------------|----------------------|---------------------------|-------------------------------------|--|---|---|-----------------------------|--|
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 ул | s | Cancer Risk: | | | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 ул | S | Nonmutagens: | | | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 ул | 'S | Г | | / IRS. | $x \times ED_{a}$ | IRS | $\times ED_{a}$ \] |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 yı | 'S | $Risk = SF_o \times C_s$ | $_{oil} \times EF_r \times FI$ | $\times CF_{O} \times \left(\frac{IRS_{O}}{BW_{a} \times AT}\right)$ | $\frac{d}{d} = \frac{d}{d} + \frac{d}{d} + \frac{d}{d}$ | | dav |
| Averaging Time, carcinogens | AT _c | 70 yı | 'S | | | $BW_a \times AT$ | $c \times 365 \frac{1}{year} B$ | $W_c \times AT_c$ | $\times 365 \overline{\text{year}}$ |
| Averaging Time, noncarcinogens, adult | AT _{nc.a} | 3 yı | 'S | Mutagens: | | × × | | | /] |
| Averaging Time, noncarcinogens, child | AT _{nc.c} | 6 yı | S | $Rick - SE \times C$ | $ \vee FF \vee F$ | IVCE | | | |
| Body Weight | BW _a | 80 k | g | $Risk = SF_0 \times \left[C\right]$ | | | | | |
| Body Weight | BWc | 15 k | | × | $\left(\frac{IRS_c \times EI}{1}\right)$ | $\frac{D_{0-2} \times ADAF_{0-2}}{\times 365 days/year} +$ | $IRS_c \times ED_{2-6} \times$ | ADAF ₂₋ | 6 |
| Conversion Factor | CF | 1E-06 k | | | | | | | |
| COPC Concentration in soil | C _{soil} | chemical-specific m | | + | $IRS_a \times ED_6$ | $\frac{-16 \times ADAF_{6-16}}{365 days/year} + \frac{I}{H}$ | $RS_a \times ED_{16-26} \times D_{16-26}$ | ADAF ₁₆₋ | 26 |
| Exposure Duration, child 0-2 | ED ₀₋₂ | 2 yr | 0 0 | I | $BW_a \times AT_c \times$ | : 365 days/year ' E | $W_a \times AT_c \times 365 a$ | days/yed | ar]] |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 4 yi | | Vinyl Chloride: | / | | | | 1\ |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 yı | | | (| $EF_r \times \left(\frac{ED_c}{R}\right)$ | $\frac{ED_a \times IR}{W} + \frac{ED_a \times IR}{BW}$ | $\frac{S_a}{IR}$ | S. D |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 yı | | $Risk = SF_0 \times \cdot$ | $C_{soil} \times FI \times$ | $CF_O \times \left[\frac{EF_r \times \left(\frac{ED_c >}{B} \right)}{AT_c} \right]$ | $\frac{davs}{davs}$ | $\frac{-}{R} + \frac{m}{R}$ | $\left \frac{\partial c}{\partial L}\right $ |
| Exposure Duration, adult | EDa | 3 у | -s | | (| AT _c | $\times 365 \frac{days}{year}$ | 21 | [c] |
| Exposure Duration, child | ED_{c} | бул | s | | | | - | | |
| Exposure Frequency, resident | EF_r | 152 da | ays/yr | Noncancer Hazard | 1: | | | | |
| Fraction Contaminated Soil Ingested | FI | 1.0 u | nitless | 1 | [<i>(1</i> | $RS_{-} \times ED_{-} \times EE_{-} \times F$ | $I \times CE_{1}$ | | |
| Soil Ingestion Rate (adult) | IRS _a | 50 m | ig/day | Hazard = $\frac{1}{P f P}$ | $\times C_{soil} \times (\frac{1}{p}) $ | $\frac{RS_c \times ED_c \times EF_r \times F}{W_c \times AT_{mc,c} \times 365 dc}$ | $\frac{1 \times 01_0}{1 \times 1000}$ | | |
| Soil Ingestion Rate (child) | IRS _c | 100 m | ig/day | $K J D_0$ | | $W_c \times AI_{nc,c} \times 505 uc$ | iy/yeur/] | | |
| Oral Reference Dose | RfD _o | chemical-specific (r | ng/kg-day) | | | | | | |
| Oral Slope Factor | SFo | chemical-specific (r | ng/kg-day) ⁻¹ | | | | | | |
| | | | | Cancer | | | Noncancer | | Noncancer |
| | | C _{soil} | SFo | Risk | % of | RfDo | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total | (dimensionless) |
| Polycyclic Aromatic Hydrocarbons (PAHs) |) | | | | | | | | |
| Benzo[a]anthracene | Μ | 2.45E+00 | 7.30E-01 | 2.54E-06 | 6% | No Toxicity Value | | | |
| Benzo[a]pyrene | М | 2.39E+00 | 7.30E+00 | 2.47E-05 | 58% | No Toxicity Value | | | |
| Benzo[b]fluoranthene | М | 2.21E+00 | 7.30E-01 | 2.29E-06 | 5% | No Toxicity Value | | | |
| Benzo[g,h,i]perylene | | 1.05E+00 | No Toxicity Value | | | 3.00E-02 | 9.72E-05 | < 1% | 9.11E-06 |
| Benzofluoranthenes, Total | М | 6.36E+00 | 7.30E-01 | 6.58E-06 | 15% | No Toxicity Value | | | |
| Dibenz[a,h]anthracene | М | 2.38E-01 | 7.30E+00 | 2.46E-06 | 6% | No Toxicity Value | | | |
| Indeno[1,2,3-cd]pyrene | М | 1.16E+00 | 7.30E-01 | 1.20E-06 | 3% | No Toxicity Value | | | |
| | | | | | | | | | |
| | | | Dathmon Com | Cancer Risk | | | Hazard Index (Child |) | Hazard Index (Adul |
| | | | Pathway Sum | : 4.2E-05 | | Pathway Sum: | 0.18 | | 0.017 |

 Table A-13.2

 Derivation of Inhalation Factors for Inhalation Exposures, Hypothetical Fishing Platform User: CTE, Sandblast Area AOPC

| Description | Variable | Value | Units VF Derivation | |
|---|-----------------------|--|---------------------|---|
| Unit conversion factor | CFi | $1.00E-04 \text{ m}^2/\text{cm}^2$ | | |
| Apparent Diffusivity | D_A | derived cm ² /s | VFs = 0 | $C_{vol} \times \frac{(3.14 \times D_A \times T_{resident})^{1/2}}{(2 \times \rho_b \times D_A)} \times CF_i$ |
| Diffusivity in air | D_i | chemical-specific cm ² /s | (13 2) | $(2 \times \rho_b \times D_A) \qquad (2 \times \rho_b \times D_A)$ |
| Diffusivity in water | D_w | chemical-specific cm ² /s | | |
| Default organic-carbon content | F _{OC} | 0.006 g/g | | |
| Henry's law constant | H' | chemical-specific dimensionles | 38 | |
| Soil-water partition coefficient | K _d | chemical-specific cm3/g | | |
| for organics: Kd= K_{OC} × F_{OC} | | | based on: | |
| Soil-organic carbon partition coefficient | K _{OC} | chemical-specific cm3/g | г | (0, 10/3 m m + 0, 10/3 m)/(-2) |
| Total soil porosity | n | 0.43 L _{porespace} /L _{soil} | $D_A = \frac{1}{2}$ | $\frac{\left(\theta_a^{10/3}D_iH' + \theta_w^{10/3}D_w\right)/n^2\right]}{\rho_n K_d + \theta_w + \theta_n H'}$ |
| Particulate Emission Factor (non-VOCs), default | PEF | 1.360E+09 m ³ /kg | | $ \rho_B \kappa_d + \theta_W + \theta_a H' $ |
| Inverse of the mean concentration at the center | Q/C _{vol} | 73.44 g/m ² -s per kg | g/m ³ | |
| of a 0.5-acre square source in Los Angeles | | | | |
| Dry soil bulk density | ρ_b | 1.5 g/cm ³ | | |
| Air-filled soil porosity | θ_{a} | 0.28 L _{air} /L _{soil} | | |
| Water-filled soil porosity | $\theta_{\rm w}$ | $0.15 L_{water}/L_{soil}$ | | |
| Exposure interval, 30-yr resident | T _{Resident} | 2.80E+08 seconds | | |
| Volatilization Factor for soil | VFs | derived m ³ /kg | | |

| | | | | | | | VFs (VOCs) | PEF |
|---|----------------------|-----------------------------|---------------------------------------|--|---|-------------------------------|--------------------|--------------------|
| | | D _i | D _w | H' | K _{oc} | K _d D _A | Residential | (non-VOCs) |
| Analyte | (cm ² /s) | Reference | (cm ² /s) Reference | (dimensionless) Reference | (cm ³ /g) Reference | (cm^{3}/g) (cm^{2}/s) | m ³ /kg | m ³ /kg |
| Inorganic Constituents | | | | | | | | |
| Arsenic | 0 | 0 | 0 0 | 31.61052561 Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 29 0 | non-VOC | 1.360E+09 |
| Chromium (III) | 0 | 0 | 0 0 | 0 0 | 0 0 | 1800000 0 | non-VOC | 1.360E+09 |
| Lead | 0 | 0 | 0 0 | 1.001885999 Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 900 0 | non-VOC | 1.360E+09 |
| Nickel | 0 | 0 | 0 0 | 1.001885999 Bond Method estimate, EPISuite v4.10 (USEPA 2011a) | 13.22 EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 65 0 | non-VOC | 1.360E+09 |
| Polychlorinated Biphenyls (PCBs) - Mixtures | | | | | | | | |
| Total PCBs | 0.040076 | RSLs ("WATER9"; USEPA 2012) | 4.683E-06 RSLs ("WATER9"; USEPA 2012) | 0.011572806 Experimental, EPISuite v4.10 (USEPA 2011a) | 130500 EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 783 3.1E-08 | non-VOC | 1.360E+09 |
| Volatile Organic Compounds (VOCs) | | | | | | | | / / |
| Tetrachloroethene | 0.0504664 | RSLs ("WATER9"; USEPA 2012) | 9.455E-06 RSLs ("WATER9"; USEPA 2012) | 0.723811518 Experimental, EPISuite v4.10 (USEPA 2011a) | 94.94 EPISuite v4.10 (USEPA, 2011a); MCI-estimated | 0.56964 0.00235 | 1497.209542 | use VF |
| Trichloroethene | 0.0686618 | RSLs ("WATER9"; USEPA 2012) | 0.0000102 RSLs ("WATER9"; USEPA 2012) | 0.402799065 Experimental, EPISuite v4.10 (USEPA 2011a) | 60.7 EPISuite v4.10 (USEPA 2011a); MCI-estimated | 0.3642 0.00266 | 1408.678568 | use VF |
| Semivolatile Organic Compounds (SVOCs) | | | | | | | | |
| DEHP | 0.0173403 | RSLs ("WATER9"; USEPA 2012) | 4.181E-06 RSLs ("WATER9"; USEPA 2012) | 1.10412E-05 Experimental, EPISuite v4.10 (USEPA 2011a) | 119600 EPISuite v4.10 (USEPA 2011a); MCI-estimated | 717.6 5.1E-11 | non-VOC | 1.360E+09 |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | / / |
| Benzo[a]anthracene | 0.0508647 | RSLs ("WATER9"; USEPA 2012) | 5.943E-06 RSLs ("WATER9"; USEPA 2012) | 0.00049072 Experimental, EPISuite v4.10 (USEPA 2011a) | 176900 EPISuite v4.10 (USEPA 2011a); MCI-estimated | 1061.4 1.3E-09 | non-VOC | 1.360E+09 |
| Benzo[a]pyrene | 0.0475831 | RSLs ("WATER9"; USEPA 2012) | 5.56E-06 RSLs ("WATER9"; USEPA 2012) | 1.86882E-05 Experimental, EPISuite v4.10 (USEPA 2011a) | 587400 EPISuite v4.10 (USEPA 2011a); MCI-estimated | 3524.4 2.3E-11 | non-VOC | 1.360E+09 |
| Benzo[b]fluoranthene | 0.0475831 | RSLs ("WATER9"; USEPA 2012) | 5.56E-06 RSLs ("WATER9"; USEPA 2012) | 2.68669E-05 Experimental, EPISuite v4.10 (USEPA 2011a) | 599400 EPISuite v4.10 (USEPA 2011a); MCI-estimated | 3596.4 2.8E-11 | non-VOC | 1.360E+09 |
| Benzo[g,h,i]perylene | 0 | 0 | 0 0 | 1.35357E-05 Experimental, EPISuite v4.10 (USEPA 2011a) | 1951000 EPISuite v4.10 (USEPA 2011a); MCI-estimated | 11706 0 | non-VOC | 1.360E+09 |
| Benzofluoranthenes, Total | 0.0475831 | RSLs ("WATER9"; USEPA 2012) | 5.56E-06 RSLs ("WATER9"; USEPA 2012) | 2.68669E-05 Experimental, EPISuite v4.10 (USEPA 2011a) | 599400 EPISuite v4.10 (USEPA 2011a); MCI-estimated | 3596.4 2.8E-11 | non-VOC | 1.360E+09 |
| Dibenz[a,h]anthracene | 0.0445672 | RSLs ("WATER9"; USEPA 2012) | 5.207E-06 RSLs ("WATER9"; USEPA 2012) | 5.76596E-06 Experimental, EPISuite v4.10 (USEPA 2011a) | 1912000 EPISuite v4.10 (USEPA 2011a); MCI-estimated | 11472 4.1E-12 | non-VOC | 1.360E+09 |
| Indeno[1,2,3-cd]pyrene | 0.0447842 | RSLs ("WATER9"; USEPA 2012) | 5.233E-06 RSLs ("WATER9"; USEPA 2012) | 0.0000656 EPISuite v4.10 (USEPA 2011a) | 3470000 EPISuite v4.10 (USEPA 2011a) | 20820 8.9E-12 | non-VOC | 1.360E+09 |

 Table A-13.3

 Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Sandblast Area AOPC

| Description | Variable | Valu | e | Equations | | | | | | | |
|--|-----------------------------|--|--|--|---|--|---|--|--|--|--|
| Averaging Time, carcinogens | AT _c | 7 | 0 yrs | Cancer Risk: | | | | | | | |
| Averaging Time, noncarcinogens, resident | AT _{nc,r} | 1 | 9 yrs | Nonmutagens: | | | | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 1 | 0 yrs | | $Risk - IIIR \times C$ | $ED_r \times \int \frac{ED_r \times D_r}{ED_r}$ | $\langle EF_r \times ET_r$ | | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | | 3 yrs | | $Risk = IOR \times C_0$ | $AT_{-} \times 365$ | $\frac{day}{day} \times 24$ | <u>hour</u> | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | : | 3 yrs | | ſ | | year `` _ ` | day J | | | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | | 1 yrs | Mutagens: | $Risk = IUR \times C_{o}$ $Risk = IUR \times \left[C_{o}\right]$ | $_{ir} \times EF_r \times ET_r$ | | | | | |
| Averaging Time, carcinogens | AT _c | | 0 yrs | | L | (<i>ED</i> ₀₋ | $_2 \times ADAF_0$ | $\frac{-2}{4 hours/day} + \frac{1}{AT_c}$ | $ED_{2-6} \times A$ | DAF_{2-6} | |
| COPC Concentration in air | $\mathbf{C}_{\mathrm{air}}$ | chemical-specifi | 10 | | × | $\frac{1}{AT_c \times 365 \text{ days}}$ | $\frac{1}{2}$ /year $\times 2^{4}$ | $\frac{1}{4 hours/day} + \frac{1}{AT_c}$ | × 365 days/yea | $r \times 24 h$ | ours/day |
| COPC Concentration in soil | C_{soil} | chemical-specifi | 0 0 | | | ED ₆₋₁₆ | $\times ADAF_{6-}$ | $\frac{16}{hours/day} + \frac{1}{AT_c \times AT_c}$ | $ED_{16-26} \times AL$ | OAF_{16-26} | , ji |
| Exposure Duration, child 0-2 | ED ₀₋₂ | | 2 yrs | | + | $AT_{\star} \times 365 days/$ | $vear \times 24$ | $\frac{10}{hours/day} + \frac{1}{AT_e \times}$ | 365 days/year | $\times 24 ho$ | urs/day |
| Exposure Duration, child 2-6 | ED ₂₋₆ | | 4 yrs | | | | <i>you.</i> <u>2</u> 1 | 100001070009 111200 | | | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 1 | 0 yrs | Vinyl Chloride | 1 | (| $ED_r \times EF_r$ | $\times ET_r$ |)] | | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | | 0 yrs | | $Risk = IUR \times C_a$ | $_{ir} \times \left\{ \frac{1}{AT_{*} \times 365} \right\}$ | lavs/vear | $\frac{\times ET_r}{\times 24 \text{ hours/day}} + 1$ | 1} | | |
| Exposure Duration, child | ED _c | | 6 yrs | | L | (| | <u>_</u> 1 110 al 5 au y | 71 | | |
| Exposure Duration, resident | ED _r | | 9 yrs | | | | | | | | |
| Exposure Frequency | EF_r | | 2 days/yr | | | | | | | | |
| Exposure Time | ET_r | | 4 hours/day | Noncancer Haz | | / EI | $D_a \times EE_a \times$ | ET_{rr} | | | |
| Inhalation Unit-Risk Factor | IUR | chemical-specifi | $c (\mu g/m^3)^{-1}$ | | $Hazard = \frac{1}{RfC} \times$ | $C_{air} \times ($ | dav | $\frac{br_r}{bour}$ | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+0 | 9 m ³ /kg | | ŊC | $\left(AT_{nc,r} \times \right)$ | $365 \frac{uay}{year}$ | $\times 24 \frac{100 \text{ m}}{\text{day}}$ | | | |
| Inhalation Reference Concentration | RfC | chemical-specifi | | | | ` | | Capil | ца | | |
| Volatilization Factor for VOCs | VF | chemical-specifi | c m³/kg | Where: | | | | $C_{air} = \frac{C_{SDII}}{VF \text{ or } P_{s}}$ | $\frac{\mu g}{mg} \times 1000 \frac{\mu g}{mg}$ | | |
| | | | VF (VOCs) or | | | Cancer | | | Noncancer | | Noncancer |
| | | C_{soil} | PEF (non-VOCs | s) C _{air} | IUR | Risk | % of | RfC | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (m^3/kg) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total | (dimensionless) |
| | | | (111 / Kg) | (µg/m) | (µg/m) | (unitensioness) | Iouui | (µg/m) | (unnensioness) | Total | () |
| Inorganic Constituents | | | (m/kg) | (μg/m) | (µg/m) | (unitensioness) | Total | (μg/m) | (dimensionless) | Total | (|
| Inorganic Constituents Arsenic | | 9.71E+00 | 1.360E+09 | (µg/m) 7.14E-06 | (µg/m) 4.30E-03 | 9.86E-10 | < 1% | (µg/m) 1.50E-02 | 1.19E-04 | < 1% | 1.19E-04 |
| 0 | | 9.71E+00 5.79E+02 | × 0/ | | | × / | | <u> </u> | × / | | × / |
| Arsenic | | | 1.360E+09 | 7.14E-06 | 4.30E-03 | 9.86E-10 | < 1% | 1.50E-02 | 1.19E-04 | < 1% | 1.19E-04 |
| Arsenic Chromium (III) | | 5.79E+02 | 1.360E+09 1.360E+09 | 7.14E-06 4.26E-04 | 4.30E-03 No Toxicity Value | 9.86E-10 | < 1% | 1.50E-02 No Toxicity Value | 1.19E-04 | < 1% | 1.19E-04 |
| Arsenic Chromium (III) Lead | res | 5.79E+02 4.18E+02 | 1.360E+09 1.360E+09 1.360E+09 | 7.14E-06 4.26E-04 3.07E-04 | 4.30E-03 No Toxicity Value | 9.86E-10 | < 1% | 1.50E-02 No Toxicity Value No Toxicity Value | 1.19E-04 | < 1% | 1.19E-04 |
| Arsenic Chromium (III) Lead Nickel | res | 5.79E+02 4.18E+02 | 1.360E+09 1.360E+09 1.360E+09 | 7.14E-06 4.26E-04 3.07E-04 | 4.30E-03 No Toxicity Value | 9.86E-10 | < 1% | 1.50E-02 No Toxicity Value No Toxicity Value | 1.19E-04 | < 1% | 1.19E-04 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtur | res | 5.79E+02 4.18E+02 2.51E+02 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 | 4.30E-03 No Toxicity Value 2.60E-04 | 9.86E-10 2.57E-09 | <1% <1% | 1.50E-02 No Toxicity Value No Toxicity Value 9.00E-02 | 1.19E-04 8.54E-04 | < 1% < 1% | 1.19E-04 8.54E-04 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtur Total PCBs | res | 5.79E+02 4.18E+02 2.51E+02 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 | 4.30E-03 No Toxicity Value 2.60E-04 | 9.86E-10 2.57E-09 | <1% <1% | 1.50E-02 No Toxicity Value No Toxicity Value 9.00E-02 | 1.19E-04 8.54E-04 | < 1% < 1% | 1.19E-04 8.54E-04 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtun Total PCBs Volatile Organic Compounds (VOCs) | res | 5.79E+02 4.18E+02 2.51E+02 4.19E-01 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 3.08E-07 | 4.30E-03 No Toxicity Value 2.60E-04 5.70E-04 | 9.86E-10 2.57E-09 9.40E-12 | < 1% < 1% < 1% | 1.50E-02 No Toxicity Value No Toxicity Value 9.00E-02 No Toxicity Value | 1.19E-04 8.54E-04 | < 1% < 1% | 1.19E-04 8.54E-04 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtun Total PCBs Volatile Organic Compounds (VOCs) Tetrachloroethene | | 5.79E+02 4.18E+02 2.51E+02 4.19E-01 2.63E+01 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.497E+03 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 3.08E-07 1.76E+01 | 4.30E-03 No Toxicity Value 2.60E-04 5.70E-04 2.60E-07 | 9.86E-10 2.57E-09 9.40E-12 2.45E-07 | <1% <1% <1% 46% | 1.50E-02 No Toxicity Value 9.00E-02 No Toxicity Value 4.00E+01 | 1.19E-04 8.54E-04 1.83E-01 | < 1% < 1% 41% | 1.19E-04 8.54E-04 1.83E-01 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtun Total PCBs Volatile Organic Compounds (VOCs) Tetrachloroethene Trichloroethene | | 5.79E+02 4.18E+02 2.51E+02 4.19E-01 2.63E+01 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.497E+03 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 3.08E-07 1.76E+01 | 4.30E-03 No Toxicity Value 2.60E-04 5.70E-04 2.60E-07 | 9.86E-10 2.57E-09 9.40E-12 2.45E-07 | <1% <1% <1% 46% | 1.50E-02 No Toxicity Value 9.00E-02 No Toxicity Value 4.00E+01 | 1.19E-04 8.54E-04 1.83E-01 | < 1% < 1% 41% | 1.19E-04 8.54E-04 1.83E-01 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtun Total PCBs Volatile Organic Compounds (VOCs) Tetrachloroethene Trichloroethene Semivolatile Organic Compounds (SVOCs) | , | 5.79E+02 4.18E+02 2.51E+02 4.19E-01 2.63E+01 1.78E+00 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.497E+03 1.409E+03 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 3.08E-07 1.76E+01 1.26E+00 | 4.30E-03 No Toxicity Value 2.60E-04 5.70E-04 2.60E-07 4.10E-06 | 9.86E-10 2.57E-09 9.40E-12 2.45E-07 2.77E-07 | <1% <1% <1% 46% 53% | 1.50E-02 No Toxicity Value 9.00E-02 No Toxicity Value 4.00E+01 2.00E+00 | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 | <1% <1% 41% 59% | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtur Total PCBs Volatile Organic Compounds (VOCs) Tetrachloroethene Trichloroethene Semivolatile Organic Compounds (SVOCs) DEHP | , | 5.79E+02 4.18E+02 2.51E+02 4.19E-01 2.63E+01 1.78E+00 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.497E+03 1.409E+03 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 3.08E-07 1.76E+01 1.26E+00 | 4.30E-03 No Toxicity Value 2.60E-04 5.70E-04 2.60E-07 4.10E-06 | 9.86E-10 2.57E-09 9.40E-12 2.45E-07 2.77E-07 | <1% <1% <1% 46% 53% | 1.50E-02 No Toxicity Value 9.00E-02 No Toxicity Value 4.00E+01 2.00E+00 | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 | <1% <1% 41% 59% | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtur Total PCBs Volatile Organic Compounds (VOCs) Tetrachloroethene Trichloroethene Semivolatile Organic Compounds (SVOCs) DEHP Polycyclic Aromatic Hydrocarbons (PAHs) |) | 5.79E+02 4.18E+02 2.51E+02 4.19E-01 2.63E+01 1.78E+00 7.71E+01 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.497E+03 1.409E+03 1.360E+09 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 3.08E-07 1.76E+01 1.26E+00 5.67E-05 | 4.30E-03 No Toxicity Value 2.60E-04 5.70E-04 2.60E-07 4.10E-06 2.40E-06 | 9.86E-10 2.57E-09 9.40E-12 2.45E-07 2.77E-07 7.28E-12 | <1% <1% <1% 46% 53% <1% | 1.50E-02 No Toxicity Value No Toxicity Value 9.00E-02 No Toxicity Value 4.00E+01 2.00E+00 No Toxicity Value | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 | <1% <1% 41% 59% | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtur Total PCBs Volatile Organic Compounds (VOCs) Tetrachloroethene Trichloroethene Semivolatile Organic Compounds (SVOCs) DEHP Polycyclic Aromatic Hydrocarbons (PAHs) Benzo[a]anthracene |) M | 5.79E+02 4.18E+02 2.51E+02 4.19E-01 2.63E+01 1.78E+00 7.71E+01 2.45E+00 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.497E+03 1.409E+03 1.360E+09 1.360E+09 1.360E+09 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 3.08E-07 1.76E+01 1.26E+00 5.67E-05 1.80E-06 | 4.30E-03 No Toxicity Value 2.60E-04 5.70E-04 2.60E-07 4.10E-06 2.40E-06 1.10E-04 | 9.86E-10 2.57E-09 9.40E-12 2.45E-07 2.77E-07 7.28E-12 8.49E-11 | <1% <1% <1% 46% 53% <1% <1% | 1.50E-02 No Toxicity Value No Toxicity Value 9.00E-02 No Toxicity Value 4.00E+01 2.00E+00 No Toxicity Value No Toxicity Value | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 | <1% | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 |
| Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtur Total PCBs Volatile Organic Compounds (VOCs) Tetrachloroethene Trichloroethene Semivolatile Organic Compounds (SVOCs) DEHP Polycyclic Aromatic Hydrocarbons (PAHs) Benzo[a]anthracene Benzo[a]pyrene |) M M | 5.79E+02 4.18E+02 2.51E+02 4.19E-01 2.63E+01 1.78E+00 7.71E+01 2.45E+00 2.39E+00 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.497E+03 1.409E+03 1.360E+09 1.360E+09 1.360E+09 1.360E+09 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 3.08E-07 1.76E+01 1.26E+00 5.67E-05 1.80E-06 1.76E-06 | 4.30E-03 No Toxicity Value 2.60E-04 5.70E-04 2.60E-07 4.10E-06 2.40E-06 1.10E-04 1.10E-03 | 9.86E-10 2.57E-09 9.40E-12 2.45E-07 2.77E-07 7.28E-12 8.49E-11 8.28E-10 | <1% <1% <1% 46% 53% <1% <1% <1% | 1.50E-02 No Toxicity Value 9.00E-02 No Toxicity Value 4.00E+01 2.00E+00 No Toxicity Value No Toxicity Value No Toxicity Value | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 | <1% <1% 41% 59% | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 |
| Arsenic Arsenic Chromium (III) Lead Nickel Polychlorinated Biphenyls (PCBs) - Mixtuu Total PCBs Volatile Organic Compounds (VOCs) Tetrachloroethene Trichloroethene Semivolatile Organic Compounds (SVOCs) DEHP Polycyclic Aromatic Hydrocarbons (PAHs) Benzo[a]anthracene Benzo[b]pfluoranthene |) M M | 5.79E+02 4.18E+02 2.51E+02 4.19E-01 2.63E+01 1.78E+00 7.71E+01 2.45E+00 2.39E+00 2.21E+00 | 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.497E+03 1.409E+03 1.360E+09 1.360E+09 1.360E+09 1.360E+09 1.360E+09 | 7.14E-06 4.26E-04 3.07E-04 1.85E-04 3.08E-07 1.76E+01 1.26E+00 5.67E-05 5.67E-05 1.80E-06 1.76E-06 1.63E-06 | 4.30E-03 No Toxicity Value 2.60E-04 5.70E-04 2.60E-07 4.10E-06 2.40E-06 2.40E-06 1.10E-04 1.10E-03 1.10E-04 | 9.86E-10 2.57E-09 9.40E-12 2.45E-07 2.77E-07 7.28E-12 8.49E-11 8.28E-10 7.66E-11 | <1% <1% <1% 46% 53% <1% <1% <1% <1% | 1.50E-02 No Toxicity Value 9.00E-02 No Toxicity Value 4.00E+01 2.00E+00 No Toxicity Value No Toxicity Value No Toxicity Value No Toxicity Value | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 | <1% <1% 41% 59% | 1.19E-04 8.54E-04 1.83E-01 2.63E-01 |

 Table A-13.3

 Risk and Hazard Estimates: Inhalation, Hypothetical Fishing Platform User: Sandblast Area AOPC

| Description | Variable | Valu | e | Equations | | | | | | | |
|--|-----------------------------|-------------------|--------------------------------------|----------------------|--------------------------|--|-----------------------------|---|--|---------------|------------------------------|
| Averaging Time, carcinogens | AT _c | 7 | 0 yrs | Cancer Risk: | | | | | | | |
| Averaging Time, noncarcinogens, resident | AT _{nc,r} | | 9 yrs | Nonmutagens: | | | | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 1 | 0 yrs | | Rick - IIIR × (| $C \to \sqrt{\frac{ED_r}{2}}$ | $\times EF_r \times ET_r$ | | | | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | | 3 yrs | | $M_{SK} = 10K \times 0$ | $AT_{\star} \times 365$ | $\frac{day}{day} \times 24$ | <u>hour</u> | | | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | | 3 yrs | | ſ | | year `` _ ` | day / | | | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | | 1 yrs | Mutagens: | $Risk = IUR \times C$ | $C_{air} \times \left(\frac{ED_r}{AT_c \times 365} \right)$ $C_{air} \times EF_r \times ET_r$ | | | | | |
| Averaging Time, carcinogens | AT_{c} | 7 | 0 yrs | | L | (ED. | $a \times ADAF_{a}$ | 2 | $ED_{n} \prec XA$ | DAF | |
| COPC Concentration in air | $\mathbf{C}_{\mathrm{air}}$ | chemical-specifi | c μg/m ³ | | × | $\left\{\frac{1}{4T \times 365 days}\right\}$ | lypar x 24 | $\frac{-2}{4 hours/day} + \frac{1}{AT_c}$ | × 365 days/yea | $r \times 24$ | ours / day |
| COPC Concentration in soil | C _{soil} | chemical-specifi | c mg/kg | | | $(H_c \times 305 uuy)$ | $\times \Delta D \Delta F$ | filours/uuy Fic | $FD \rightarrow XAD$ |) A E | .)] |
| Exposure Duration, child 0-2 | ED ₀₋₂ | | 2 yrs | | + | $+\frac{L D_{6-16}}{4T}$ | ALAGM X 24 | $\frac{16}{hours/day} + \frac{16}{AT_c}$ | 26E dava /waar | × 24 h | <u>6</u> |
| Exposure Duration, child 2-6 | ED ₂₋₆ | | 4 yrs | | | $AI_c \times 505 uuys/$ | yeur x 24 | nours/uuy Arc | x sos uuys/yeur | × 24 m | Jursfuuy]] |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 1 | 0 yrs | Vinyl Chloride: | 1 | (| FD × FF | \times FT |)] | | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 1 | 0 yrs | | $Risk = IUR \times C$ | $air \times \frac{1}{4T} \times 26E$ | dans mar | $\frac{\times ET_r}{\times 24 \text{ hours/day}} +$ | 1 | | |
| Exposure Duration, child | ED_{c} | | 6 yrs | | L | $(H_c \times 303)$ | uuys/yeur | × 24 nours/uuy | 71 | | |
| Exposure Duration, resident | ED_r | | 9 yrs | | | | | | | | |
| Exposure Frequency | EF_r | 15 | 2 days/yr | | | | | | | | |
| Exposure Time | ET_r | 2 | 4 hours/day | Noncancer Haza | ırd: 1 | / F | | | | | |
| Inhalation Unit-Risk Factor | IUR | chemical-specifi | c (µg/m ³) ⁻¹ | | $Hazard = \frac{1}{PfC}$ | $\times C_{air} \times \left(\frac{EI}{AT_{nc,r} \times I}\right)$ | day | hour | | | |
| Particulate emission factor (non-VOCs) | PEF | 1.360E+0 | 9 m ³ /kg | | КJС | $AT_{nc,r} \times$ | : 365 (uuy) > | $\times 24 \frac{hour}{dav}$ | | | |
| Inhalation Reference Concentration | RfC | chemical-specifi | c μg/m ³ | | | ` | | · / | | | |
| Volatilization Factor for VOCs | VF | chemical-specifi | c m [°] /kg | Where: | | | | $C_{air} = \frac{C_{SOL}}{VF \text{ or } F}$ | $\frac{l}{DEF} \times 1000 \frac{\mu g}{mg}$ | | |
| | | | VF (VOCs) or | | | Cancer | | | Noncancer | | Noncancer |
| | | C _{soil} | PEF (non-VOCs) | C _{air} | IUR | Risk | % of | RfC | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (m ³ /kg) | (µg/m ³) | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | (µg/m ³) | (dimensionless) | Total | (dimensionless) |
| Indeno[1,2,3-cd]pyrene | М | 1.16E+00 | 1.360E+09 | 8.53E-07 | 1.10E-04 | 4.02E-11 | < 1% | No Toxicity Value | | | |
| | | | | | Pathway Sum: | Cancer Risk 5.3E-07 | | Pathway Sum | Hazard Index (Child : 4.5E-01 |) | Hazard Index (Adu 4.5E-01 |
| | | | | | r auiway Suili: | 5.5E-07 | | r aurway Sulli | . 4.5E-01 | | 4.50-01 |

 Table A-13.4

 Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: CTE, Sandblast Area AOPC

| Description | Variable | Value | | Equations | | | | | | |
|--|----------------------------|-------------------|---------------------------|---------------------------|--|-----------------------------------|---|--|----------------------|-----------------|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specific | unitless | Cancer Risk: | | | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 | yrs | Nonmutagens | Rick - SE, X | [C × FC | $) \rightarrow FF \rightarrow FV$ | $\times EC \times ABS_d \times Cd$ | F. | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 | yrs | | u | 2 3011 | u | u | u | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 | yrs | | $\times \left(\frac{A}{A} \right)$ | $F_a \times SA_a$ | + | $\frac{AF_c \times ED_c \times SA_c}{\times AT_c \times 365 \ day}$ | <u>c)</u>] | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 | yrs | | $\langle BW_a \times AT \rangle$ | $c_c \times 365 day$ | $y/year BW_c$ | $\times AT_c \times 365 day$ | /year]] | |
| Soil-to-Skin Adherence Fraction | AF _a | 0.01 | mg/cm ² -day | Mutagens | | | | | | |
| Soil-to-Skin Adherence Fraction | AF_{c} | 0.04 | mg/cm ² -day | Risk = S | $F_d \times \left\{ C_{soil} \times EF_r \right\}$ | $. \times EV \times EC$ | $C \times ABS_d \times CF_d$ | d | | |
| Averaging Time, Carcinogens | AT _c | 70 | yrs | | (| | | | | A E |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | 3 | yrs | | $\times \left \frac{ED_{0-2}}{DW} \right $ | $\times Ar_c \times SA_c$ | $\frac{c \times ADAF_{0-2}}{c} +$ | $-\frac{ED_{2-6} \times AF_c \times S}{BW_c \times AT_c \times 3}$ | $A_c \times AD$ | AF_{2-6} |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | 6 | yrs | | | | | | | |
| Body Weight, adult | \mathbf{BW}_{a} | 80 | kg | | $+\frac{ED_{6-16}}{ED_{6-16}}$ | $\times AF_a \times SA$ | $a \times ADAF_{6-16}$ | $+\frac{ED_{16-26} \times AF_a}{BW_a \times AT_c}$ | $\times SA_a \times$ | $ADAF_{16-26}$ |
| Body Weight, child | BW_c | 15 | kg | | $BW_a >$ | $\langle AT_c \times 365 \rangle$ | days/year | $BW_a \times AT_c$ | × 365 da | ys/year]] |
| Conversion Factor | CF _d | 1E-06 | kg/mg | Vinyl Chloride | $Risk = SF_d \times$ | | | ADS VCE | | |
| COPC Concentration in Soil | C_{soil} | chemical-specific | mg/kg | | $Risk = 5\Gamma_d \wedge$ | $U_{SOII} \wedge DT_{\gamma}$ | | $ADJ_d \wedge CT_d$ | | |
| Fraction of EV in Contact with Soil | EC | 1 | unitless | | $\begin{bmatrix} EE \\ EE \end{bmatrix} (ED)$ | $_{c} \times AF_{c} \times SL$ | $A_c \downarrow ED_a \times AF$ | $F_a \times SA_a$ | 1) | |
| Exposure Duration, child 0-2 | ED ₀₋₂ | 2 | yrs | | \times | BW_c | — – — ВИ | V_a) $AF_c >$ | $\langle SA_c ($ | |
| Exposure Duration, child 2-6 | ED ₂₋₆ | 4 | yrs | | ^ | $AT \times 3$ | $\frac{days}{days}$ | $\left(\frac{A_a \times SA_a}{V_a}\right) + \frac{AF_c \times BV}{BV}$ | V_c | |
| Exposure Duration, child 6-16 | ED ₆₋₁₆ | 10 | yrs | | L | m _c x c | year | | IJ | |
| Exposure Duration, adult 16-26 | ED ₁₆₋₂₆ | 10 | yrs | | | | | | | |
| Exposure Duration, adult | ED_a | 3 | yrs | Noncancer Haz | ard: | | | | | |
| Exposure Duration, child | ED_{c} | 6 | yrs | | , 1 | $\left[c \right] $ | c×ABS _d ×ED _{nc.c} ×E | EF _r ×EV×EC×SA _c ×CF | 1)] | |
| Exposure Frequency, resident | EF_r | 152 | days/yr | E. E. E. | $azara = \frac{1}{RfD_d} \times$ | $c_{soil} \times (-$ | $BW_c \times AT_{nc}$ | $EF_r \times EV \times EC \times SA_c \times CF_c$ $C_c \times 365 \frac{day}{mar}$ | <u>)</u>] | |
| Event Frequency | EV | | events/day | | | L \ | | yeur | /] | |
| Refererenc Dose Adjusted for GI Absorption | RfD _d | chemical-specific | | | | | | | | |
| Exposed Body Surface Area, adult | SA _a | 6032 | | | | | | | | |
| Exposed Body Surface Area, child | SA_c | 2373 | | | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specific | (mg/kg-day) ⁻¹ | | | | | | | |
| | | | | | Cancer | | | Noncancer | | Noncancer |
| | | C _{soil} | ABS _d | SF _d | Risk | % of | RfD _d | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total | (dimensionless) |
| Inorganic Constituents | | | | | | | | | | |
| Arsenic | | 9.71E+00 | 3.00E-02 | 1.50E+00 | 6.27E-08 | 1% | 3.00E-04 | 1.54E-03 | 8% | 1.83E-04 |
| Chromium (III) | | 5.79E+02 | 1.00E-02 | | | | 1.95E-02 | 7.82E-04 | 4% | 9.32E-05 |
| Lead | | 4.18E+02 | 1.00E-02 | | | | | | | |
| Nickel | | 2.51E+02 | 1.00E-02 | | | | 8.00E-04 | 8.27E-03 | 43% | 9.85E-04 |
| Polychlorinated Biphenyls (PCBs) - Mixtures | | | | | | | | | | |
| Total PCBs | | 4.19E-01 | 1.40E-01 | 2.00E+00 | 2.81E-08 | < 1% | 2.00E-05 | 7.73E-03 | 40% | 9.21E-04 |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | |

2.63E+01

1.78E+00

0.00E+00

0.00E+00

2.10E-03

4.60E-02

0.00E+00

0.00E+00

< 1%

< 1%

6.00E-03

5.00E-04

0.00E+00

0.00E+00

< 1%

< 1%

0.00E+00

0.00E+00

Tetrachloroethene

Trichloroethene

Æ

 Table A-13.4

 Risk and Hazard Estimates: Dermal Contact with Soil, Hypothetical Fishing Platform User: CTE, Sandblast Area AOPC

| Description | Variable | Value | 2 | Equations: | | | | | | |
|---|--|-------------------|-----------------------------|---------------------------|--|---|---|--|------------------------|-----------------|
| Dermal Soil Absorption Fraction | ABS _d | chemical-specific | c unitless | Cancer Risk: | | | | | | |
| Age-dependent Adjustment Factor, 0-2 | ADAF ₀₋₂ | 10 |) yrs | Nonmutagens: | Dick - SE V | | | $\langle EC \times ABS_d \times Cd$ | C | |
| Age-dependent Adjustment Factor, 2-6 | ADAF ₂₋₆ | 3 | 3 yrs | | u | 2 3011 | u | u | u | |
| Age-dependent Adjustment Factor, 6-16 | ADAF ₆₋₁₆ | 3 | 3 yrs | | $\times \left(\frac{A}{A} \right)$ | $F_a \times SA_a$ | + | $\frac{AF_c \times ED_c \times SA_c}{\times AT_c \times 365 \ day}$ | <u> </u> | |
| Age-dependent Adjustment Factor, 16-26 | ADAF ₁₆₋₂₆ | 1 | l yrs | | $\langle BW_a \times AT \rangle$ | $_c \times 365 day$ | y/year BW _c | $\times AT_c \times 365 day$ | /year]] | |
| Soil-to-Skin Adherence Fraction | AF_a | | mg/cm ² -day | Mutagens: | | | | | | |
| Soil-to-Skin Adherence Fraction | AF _c | 0.04 | 4 mg/cm ² -day | Risk = Sk | $F_d \times \left\{ C_{soil} \times EF_r \right\}$ | $\times EV \times EC$ | $C \times ABS_d \times CF_d$ | d | | |
| Averaging Time, Carcinogens | AT _c | 70 |) yrs | | | | | | $A \times AD$ | AF- |
| Averaging Time, noncarcinogens, adult | AT _{nc,a} | | 3 yrs | | $\times \left \frac{LD_{0-2}}{DW} \right $ | | $\frac{c}{dava}$ | $\frac{ED_{2-6} \times AF_c \times S}{BW_c \times AT_c \times 3}$ | E dava | <u>112-6</u> |
| Averaging Time, noncarcinogens, child | AT _{nc,c} | | 5 yrs | | | $\sim AI_C \times 505$ $\sim AE \sim SA$ | x ADAE | $DW_C \times AI_C \times SC$ | $\sim SA \sim$ | Yeur ADAE 1) |
| Body Weight, adult | \mathbf{BW}_{a} | |) kg | | $+\frac{ED_{6-16}}{DW}$ | $\wedge Ar_a \wedge SA$ | $a \times ADAF_{6-16}$ | $+\frac{ED_{16-26} \times AF_a}{BW_a \times AT_c}$ | × SH _a × | $ADAF_{16-26}$ |
| Body Weight, child | BW _c | | 5 kg | | | $AI_c \times 365$ | aays/year | $BW_a \times AI_c$ | × 365 aa | ys/year]) |
| Conversion Factor | CF _d | | 5 kg/mg | Vinyl Chloride: | $Risk = SF_d \times$ | $\{C_{soil} \times EF_r\}$ | $\times EV \times EC \times$ | $ABS_d \times CF_d$ | | |
| COPC Concentration in Soil | C _{soil} | chemical-specific | 0 0 | | - (PD | | | u u | - | |
| Fraction of EV in Contact with Soil | EC | | l unitless | | $EF_r \times \left(\frac{ED_r}{2}\right)$ | $_{c} \times AF_{c} \times SL$ | $\frac{A_c}{C} + \frac{ED_a \times AF}{DM}$ | $\left(\frac{A}{A} \times SA_{a}\right) A F \sim$ | · s A]) | |
| Exposure Duration, child 0-2 | ED ₀₋₂ | | 2 yrs | | × | DVVC | days | $\frac{A_{a}}{A_{c}} + \frac{A_{c}}{B_{c}}$ | $\left \right\rangle$ | |
| Exposure Duration, child 2-6 | ED ₂₋₆ | | 4 yrs | | | $AT_c \times 3$ | $365 \frac{uuys}{vear}$ | $\left(\frac{A \times SA_a}{V_a}\right) + \frac{AF_c > BV}{BV}$ | v_c | |
| Exposure Duration, child 6-16 Exposure Duration, adult 16-26 | ED ₆₋₁₆ | |) yrs | | L | | year | | 11 | |
| Exposure Duration, adult | ED ₁₆₋₂₆ ED _a | |) yrs 3 yrs | Noncancer Haz | ard | | | | | |
| Exposure Duration, adult Exposure Duration, child | ED_a ED_c | | ó yrs | | | | | | | |
| Exposure Frequency, resident | EF, | | 2 days/yr | H | $azard = \frac{1}{RED} \times$ | $C_{soil} \times \left(\frac{AF_{c}}{2}\right)$ | $c \times ABS_d \times ED_{nc,c} \times E$ | $F_r \times EV \times EC \times SA_c \times CF_c$ $F_c \times 365 \frac{day}{year}$ | <u>"</u>)] | |
| Event Frequency | EV | 1 | l events/day | | nj Da | | BWC×AInd | year | Л | |
| Refererenc Dose Adjusted for GI Absorption | RfD_d | chemical-specific | c mg/kg-day | | | | | | | |
| Exposed Body Surface Area, adult | SA_a | 6032 | 2 cm^2 | | | | | | | |
| Exposed Body Surface Area, child | SA_c | 2373 | 3 cm^2 | | | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specific | c (mg/kg-day) ⁻¹ | | | | | | | |
| | | | | | Cancer | | | Noncancer | | Noncancer |
| | | C _{soil} | ABS _d | SF_d | Risk | % of | RfD _d | Hazard (Child) | % of | Hazard (Adult) |
| Analyte | Mutagen? | (mg/kg) | (unitless) | (mg/kg-day) ⁻¹ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total | (dimensionless) |
| Semivolatile Organic Compounds (SVOCs) | | | | | | | | | | |
| DEHP | | 7.71E+01 | 1.00E-01 | 1.40E-02 | 2.58E-08 | < 1% | 2.00E-02 | 1.02E-03 | 5% | 1.21E-04 |
| Polycyclic Aromatic Hydrocarbons (PAHs) | | | | | | | | | | |
| Benzo[a]anthracene | М | 2.45E+00 | 1.30E-01 | 7.30E-01 | 3.22E-07 | 6% | | | | |
| Benzo[a]pyrene | М | 2.39E+00 | 1.30E-01 | 7.30E+00 | 3.14E-06 | 61% | | | | |
| Benzo[b]fluoranthene | М | 2.21E+00 | 1.30E-01 | 7.30E-01 | 2.90E-07 | 6% | | | | |
| Benzo[g,h,i]perylene | | 1.05E+00 | 1.30E-01 | | | | 3.00E-02 | 1.20E-05 | < 1% | 1.43E-06 |
| Benzofluoranthenes, Total | М | 6.36E+00 | 1.30E-01 | 7.30E-01 | 8.35E-07 | 16% | | | | |
| Dibenz[a,h]anthracene | М | 2.38E-01 | 1.30E-01 | 7.30E+00 | 3.13E-07 | 6% | | | | |
| Dibenz[a,n]antinacene | | | | | | | | | | |

| Pathway Sum: | Cancer Risk | Hazard Index (Child) | Hazard Index (Adul |
|--------------|-----------------|----------------------|--------------------|
| | 5.2E-06 Pathway | y Sum: 0.019 | 0.0023 |

 Table A-14.1

 Risk and Hazard Estimates - Ingestion of Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC

| Definitions | Variable | Value | | Equations | | | | |
|--|-----------------------------|--------------------|---------------------------|-------------------|-------------------------------------|--|---|-------|
| Averaging Time, carcinogens | AT _c | 70 | yrs | Carcinogenic: | | | | |
| Averaging Time, noncarcinogens, worker | $AT_{nc,w}$ | 1 | yrs | | Г | | | |
| Body Weight, adult | \mathbf{BW}_{a} | 80 | kg | $Risk = SF_{a} >$ | $\langle C_{\text{water}} \times $ | $\underline{IRW_{w} \times EF_{w} \times E}$ | $\frac{D_{w} \times FI_{w} \times CF_{o}}{365 day / year}$ | |
| Conversion Factor, ingestion | CFo | 1.0E-03 | mg/µg | 0 | water | $BW_a \times AT_c \times 3$ | 365day/year) | |
| COPC Concentration in Water | C _{water} | chemical-specific | µg/L | | | | | |
| Exposure Duration, worker | ED_{w} | 1 | yrs | | | | | |
| Exposure Frequency, worker | EF_{w} | 9 | days/yr | Noncarcinogenic: | | | | |
| Fraction Contaminated Water Ingested, worker | FI_{w} | 1.0 | unitless | | Г | (| | Л |
| Water Ingestion Rate, worker | $\mathrm{IRW}_{\mathrm{w}}$ | | L/day | Hazard= | $\frac{1}{C} \times C_{max}$ | $ \frac{IRW_{w} \times EF_{w}}{V} $ | $\frac{\times ED_{w} \times FI_{w} \times CF_{o}}{\frac{1}{2}}$ | |
| Oral Reference Dose | RfD _O | chemical-specific | | Rj | ${}^{t}D_{o}$ | $e^{r} \left(BW_a \times AT_{nc} \right)$ | w×365day/year | |
| Oral Slope Factor | SFo | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Noncancer | |
| | | C _{water} | SFo | Risk | % of | RfD _o | Hazard | % of |
| Analyte | | $(\mu g/L)$ | $(mg/kg-day)^{-1}$ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | |
| Vanadium | | 2.23E-01 | No Toxicity Value | | | 5.00E-03 | 2.75E-07 | < 1% |
| Volatile Organic Compounds (VOCs) | | | | | | | | |
| Acetone | | 3.88E+00 | No Toxicity Value | | | 9.00E-01 | 2.66E-08 | < 1% |
| Benzene | | 1.37E-01 | 5.50E-02 | 6.64E-13 | < 1% | 4.00E-03 | 2.11E-07 | < 1% |
| Carbon disulfide | | 2.55E-01 | No Toxicity Value | | | 1.00E-01 | 1.57E-08 | < 1% |
| Chloroform | | 1.74E-01 | 3.10E-02 | 4.75E-13 | < 1% | 1.00E-02 | 1.07E-07 | < 1% |
| 1,1-Dichloroethane | | 5.00E+00 | 5.70E-03 | 2.51E-12 | < 1% | 2.00E-01 | 1.54E-07 | < 1% |
| 1,1-Dichloroethene | | 1.16E+00 | No Toxicity Value | | | 5.00E-02 | 1.43E-07 | < 1% |
| cis-1,2-Dichloroethene | | 6.60E+02 | No Toxicity Value | | | 2.00E-03 | 2.03E-03 | 77% |
| trans-1,2-Dichloroethene | | 1.80E+00 | No Toxicity Value | | | 2.00E-02 | 5.55E-07 | < 1% |
| 1,2-Dichloropropane | | 1.79E-01 | 3.60E-02 | 5.67E-13 | < 1% | 9.00E-02 | 1.23E-08 | < 1% |
| Ethylbenzene | | 4.47E-02 | 1.10E-02 | 4.33E-14 | < 1% | 1.00E-01 | 2.76E-09 | < 1% |
| Tetrachloroethene | | 5.45E+01 | 2.10E-03 | 1.01E-11 | 2% | 6.00E-03 | 5.60E-05 | 2% |
| Toluene | | 6.40E-01 | No Toxicity Value | | | 8.00E-02 | 4.93E-08 | < 1% |
| 1,1,1-Trichloroethane | | 2.22E+00 | No Toxicity Value | | | 2.00E+00 | 6.84E-09 | < 1% |
| Trichloroethene | | 4.37E+01 | 4.60E-02 | 1.77E-10 | 39% | 5.00E-04 | 5.39E-04 | 20% |

 Table A-14.1

 Risk and Hazard Estimates - Ingestion of Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC

| Definitions | Variable | Value | | Equations | | | | |
|--|-----------------------------|--------------------|--------------------|--|------------------------|--|--|-------|
| Averaging Time, carcinogens | AT_{c} | 70 | yrs | Carcinogenic: | | | | |
| Averaging Time, noncarcinogens, worker | $AT_{nc,w}$ | 1 | yrs | | Г | | | |
| Body Weight, adult | BW_a | 80 kg | | $Risk = SF_o \times \left[C_{water} \times \left(\frac{IRW_w \times EF_w \times ED_w \times FI_w \times CF_o}{BW_a \times AT_c \times 365 day / year} \right) \right]$ | | | | |
| Conversion Factor, ingestion | CFo | 1.0E-03 mg/µg | | | | | | |
| COPC Concentration in Water | C _{water} | chemical-specific | µg/L | | | | | |
| Exposure Duration, worker | ED_{w} | 1 | yrs | | | | | |
| Exposure Frequency, worker | EF_{w} | 9 days/yr | | Noncarcinogenic: | | | | |
| Fraction Contaminated Water Ingested, worker | FI_{w} | 1.0 | unitless | | Г | (| | ν, Γ |
| Water Ingestion Rate, worker | $\mathrm{IRW}_{\mathrm{w}}$ | 0.02 | L/day | Hazard= | $\frac{1}{-} \times C$ | $\times \left(\frac{IRW_{w} \times EF_{w}}{} \right)$ | $\frac{\langle ED_{w} \times FI_{w} \times CF_{o} \rangle}{\langle w \times 365 day / year \rangle}$ | |
| Oral Reference Dose | RfD ₀ | chemical-specific | mg/kg-day | Rf | \mathcal{D}_o | $e^{r} \left(BW_a \times AT_{nc,} \right)$ | $_{w}$ ×365day/ year | |
| Oral Slope Factor | SFo | chemical-specific | $(mg/kg-day)^{-1}$ | | _ | | | _ |
| | | | | Cancer | | | Noncancer | |
| | | C _{water} | SFo | Risk | % of | RfD _o | Hazard | % of |
| Analyte | | $(\mu g/L)$ | $(mg/kg-day)^{-1}$ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Vinyl Chloride | | 4.10E+00 | 7.20E-01 | 2.60E-10 | 58% | 3.00E-03 | 8.42E-06 | < 1% |
| Xylenes | | 1.32E-01 | No Toxicity Value | | | 2.00E-01 | 4.07E-09 | < 1% |
| 1,2,4-trimethylbenzene | | 4.85E-02 | No Toxicity Value | | | No Toxicity Value | | |
| o-xylene | | 7.35E-02 | No Toxicity Value | | | 2.00E-01 | 2.27E-09 | < 1% |
| Naphthalene | | 4.52E-02 | No Toxicity Value | | | 2.00E-02 | 1.39E-08 | < 1% |
| | | | Pathway Sum | Cancer Risk : 4.5E-10 | | | Hazard Index 2.6E-03 | |

 Table A-14.2

 and Hazard Estimates - Inhalation of Volatile Compounds from Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AC

| Definition | Variable | e Value | Э | Risk and Ha | zard Equations | | | | | |
|--|-----------------------------|--------------------|---------------------------|------------------|----------------------------------|--------------------------------|-------------------|-----------------------------|-----------------|-------|
| Averaging Time, carcinogens | AT _c | 70 |) yrs | Cancer Risk | | | | | | |
| Averaging Time, noncarcinogens, worker | AT _{nc,w} | | l yrs | | | [(| EF | $\times ED \times E$ | T)] | |
| Body Weight, adult | \mathbf{BW}_{a} | 80 |) kg | F | Risk = URF > | $\langle C_{air} \times -$ | $T \sim 2$ | $\frac{w}{265}$ days (), a | | |
| COPC Concentration in air | $\mathbf{C}_{\mathrm{air}}$ | chemical-specific | c μg/m ³ | | | | $I_c \times 2$ | sos aays / yee | $r \times 24$ | |
| COPC Concentration in water | C _{water} | chemical-specific | cμg/L | Noncancer H | Hazard: | | | | | |
| Exposure Duration, worker | ED_w | 1 | l yrs | | . Г | (| | | 7 | |
| Exposure Frequency, worker | EF_{w} | | 9 days/yr | H | $azard = \frac{1}{RfC} \times C$ | ' × | $EF_{w} \times E$ | $ED_{w} \times ET_{w}$ |]] | |
| Exposure Time, worker | ET_{w} | |) hours | 110 | RfC | $AT_{nc,w} \times 36$ | 5 days/ | year×24hours/ a | lay) | |
| Air Inhalation Rate, worker | IRA_w | 20 |) m ³ /workday | | L | | | | ~] | |
| Inhalation Reference Concentration | RfC | chemical-specific | c μg/m ³ | | | | | | | |
| Inhalation Unit-Risk Factor | URF | chemical-specific | $(\mu g/m^3)^{-1}$ | where: | $C_{air} = C_{water} \times V$ | /F | | | | |
| Volatilization Factor | VF | chemical-specific | $c L/m^3$ | | $C_{air} = C_{water} \times V$ | | | | | |
| | | | | | | Cancer | | | Hazard | |
| | | C _{water} | VF | C _{air} | URF _i | Risk | % of | RfC | Quotient | % of |
| Analyte | | $(\mu g/L)$ | (L/m^{3}) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | | | |
| Vanadium | | 2.23E-01 | non-VOC | | No Toxicity Value | | | 1.00E-01 | | |
| Volatile Organic Compounds (VOCs) | | | | | | | | | | |
| Acetone | | 3.88E+00 | 4.12E+00 | 1.60E+01 | No Toxicity Value | | | 3.10E+04 | 2.12E-06 | < 1% |
| Benzene | | 1.37E-01 | 9.35E+00 | 1.28E+00 | 7.80E-06 | 5.86E-10 | < 1% | 3.00E+01 | 1.75E-04 | < 1% |
| Carbon disulfide | | 2.55E-01 | 9.55E+00 | 2.44E+00 | No Toxicity Value | | | 7.00E+02 | 1.43E-05 | < 1% |
| Chloroform | | 1.74E-01 | 7.53E+00 | 1.31E+00 | 2.30E-05 | 1.77E-09 | 2% | 9.80E+01 | 5.49E-05 | < 1% |
| 1,1-Dichloroethane | | 5.00E+00 | 8.31E+00 | 4.15E+01 | 1.60E-06 | 3.90E-09 | 4% | No Toxicity Value | | |
| 1,1-Dichloroethene | | 1.16E+00 | 8.46E+00 | 9.82E+00 | No Toxicity Value | | | 2.00E+02 | 2.02E-04 | < 1% |
| cis-1,2-Dichloroethene | | 6.60E+02 | 8.36E+00 | 5.52E+03 | No Toxicity Value | | | No Toxicity Value | | |
| trans-1,2-Dichloroethene | | 1.80E+00 | 8.43E+00 | 1.52E+01 | No Toxicity Value | | | 6.00E+01 | 1.04E-03 | < 1% |
| 1,2-Dichloropropane | | 1.79E-01 | 7.70E+00 | 1.38E+00 | 1.00E-05 | 8.09E-10 | < 1% | 4.00E+00 | 1.42E-03 | < 1% |
| Ethylbenzene | | 4.47E-02 | 8.05E+00 | 3.60E-01 | 2.50E-06 | 5.28E-11 | < 1% | 1.00E+03 | 1.48E-06 | < 1% |
| Tetrachloroethene | | 5.45E+01 | | 3.52E+02 | 2.60E-07 | 5.38E-09 | 5% | 4.00E+01 | 3.62E-02 | 5% |
| Toluene | | 6.40E-01 | | 5.52E+00 | No Toxicity Value | | | 5.00E+03 | 4.54E-06 | < 1% |
| 1,1,1-Trichloroethane | | 2.22E+00 | 7.21E+00 | | No Toxicity Value | | | 5.00E+03 | 1.32E-05 | < 1% |
| Trichloroethene | | 4.37E+01 | | 3.17E+02 | 4.10E-06 | 7.62E-08 | 76% | 2.00E+00 | 6.51E-01 | 94% |
| Vinyl Chloride | | 4.10E+00 | | 4.32E+01 | 4.40E-06 | 1.12E-08 | 11% | 1.00E+02 | 1.78E-03 | < 1% |
| | | | | | | | | | | |

 Table A-14.2

 and Hazard Estimates - Inhalation of Volatile Compounds from Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AC

| Definition | Variable | e Value | | Risk and Ha | azard Equations | | | | | |
|--|-----------------------------|--------------------|-------------------------|---------------|----------------------------------|--|----------------------|---|-------------------------|-------|
| Averaging Time, carcinogens | AT _c | 70 | yrs | Cancer Risk | : | | | | | |
| Averaging Time, noncarcinogens, worker | $AT_{nc,w}$ | 1 | yrs | | | Γ (| EF | $\times ED \times D$ | ET | |
| Body Weight, adult | BW_a | 80 | kg | I | Risk = URF > | $\langle C_{air} \times -$ | | $\frac{w}{c} = \frac{1}{w} \frac{1}{c}$ | | |
| COPC Concentration in air | $\mathbf{C}_{\mathrm{air}}$ | chemical-specific | $\mu g/m^3$ | | | | $\prod_{c} \times 3$ | 65 aays / ye | $ear \times 24$ | |
| COPC Concentration in water | Cwater | chemical-specific | μg/L | Noncancer I | Hazard: | | | | | |
| Exposure Duration, worker | ED_w | 1 | yrs | | Г | (| | | 7 | |
| Exposure Frequency, worker | EF_{w} | 9 | days/yr | H | $azard = \frac{1}{RfC} \times C$ | · | $EF_{W} \times E$ | $CD_{w} \times ET_{w}$ |) | |
| Exposure Time, worker | ET_{w} | 4.00 | hours | 110 | $\frac{dzuru}{RfC} C$ | $a_{air} \wedge \overline{AT} \times 36$ | 5 days/ | year×24hours | day | |
| Air Inhalation Rate, worker | IRA_w | 20 | m ³ /workday | | ° L | nc,w | | | | |
| Inhalation Reference Concentration | RfC | chemical-specific | $\mu g/m^3$ | | | | | | | |
| Inhalation Unit-Risk Factor | URF | chemical-specific | $(\mu g/m^3)^{-1}$ | where: | $C_{air} = C_{water} \times V$ | VE | | | | |
| Volatilization Factor | VF | chemical-specific | L/m ³ | | $C_{air} = C_{water} \times V$ | / F | | | | |
| | | | | | | Cancer | | | Hazard | |
| | | C _{water} | VF | Cair | URF _i | Risk | % of | RfC | Quotient | % of |
| Analyte | | $(\mu g/L)$ | (L/m^{3}) | $(\mu g/m^3)$ | $(\mu g/m^3)^{-1}$ | (dimensionless) | Total | $(\mu g/m^3)$ | (dimensionless) | Total |
| 1,2,4-trimethylbenzene | | 4.85E-02 | 7.55E+00 | 3.66E-01 | No Toxicity Value | | | 7.00E+00 | 2.15E-04 | < 1% |
| o-xylene | | 7.35E-02 | 8.02E+00 | 5.89E-01 | No Toxicity Value | | | 1.00E+02 | 2.42E-05 | < 1% |
| Naphthalene | | 4.52E-02 | 6.60E+00 | 2.98E-01 | 3.40E-05 | 5.96E-10 | < 1% | 3.00E+00 | 4.09E-04 | < 1% |
| | | | | | Pathway Sums: | Cancer Risk 1.0E-07 | | | Hazard Index 6.9E-01 | |

 Table A-14.3

 Estimation of Kp, Tau(event), B, and t* for Organic Compounds: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC

| Definition | Variable | e Value | | Equations | | | | | |
|--|-----------------|-------------------------|--------------------|---|----------------------------------|---|-----------------|-----------------|----------|
| Permeability coefficient from water | K _p | Calculated (Equation 1) | cm/hr | $\frac{1}{V}$ 10 | (-2.8+0.66 Log K) | -0.0056 MW) | | | |
| Octanol:water partition coefficient | K | Chemical-specific | dimensionless | $\mathbf{K}_{p} = 10$ | | , , , , , , , , , , , , , , , , , , , | | | |
| Molecular weight | MW | Chemical-specific | g/mole | 2) $(1)^2$ | | | | | |
| Lag time per event | τ_{event} | Calculated (Equation 2) | hr/event | $\tau_{event} = \frac{(l_{sc})^2}{6D_{sc}}$ | - | | | | |
| Thickness of the strateum corneum | l _{sc} | 0.001 | cm | event 6D | | | | | |
| | | Calculated (Equation 2) | | 30 | | | | | |
| Effective diffusion coefficient, through the stratum corneum | D _{sc} | | | where: | $= l_{sc} \times 10^{(-2)}$ | .8-0.0056 MW) | | | |
| Relative contribution of permeability coefficients in stratum corneum and viable epidermis | В | Calculated (Equation 3) | dimensionless | D_{sc} | $-\iota_{sc}$ × 10 | | | | |
| Time it takes to reach steady state | t* | Calculated (Equation 4) | | 3) | MW | | | | |
| Correlation coefficient | b | Calculated (Equation 4) | dimensionless | $B = K_p$ | $\frac{\sqrt{MW}}{2.6}$ | (as an approximatio | on) | | |
| Correlation coefficient | с | Calculated (Equation 4) | dimensionless | | 2.0 | | | | |
| | | | | 4) If $B \leq 0$ | $0.6: t^* = 2.4$ | lτ _{event} | | | |
| | | | | If $B > 0$ | $0.6: t^* = 6\tau$ | $_{event}\left(b-\sqrt{b^2-c^2}\right)$ |) | | |
| | | | | | | | - | | |
| | | | | <i>c</i> = - | $\frac{1+3B+3B^2}{3\times(1+B)}$ | | | | |
| | | | | | $\frac{2\times(1+B)^2}{2}-c$ | | | | |
| | | | | b = - | $\frac{1}{\pi}$ - c | | | | |
| | | MW | LogK _{OW} | Kp ^a | τ _{event} | В | с | b | t* |
| Analyte | | (g/mole) | (dimensionless) | (cm/hr) | (hr/event) | (dimensionless) | (dimensionless) | (dimensionless) | (hr) |
| Inorganic Constituents | | | | | | | | | |
| Vanadium | | 50.94 | 2.30E-01 | 1.00E-03 | 2.03E-01 | 2.75E-03 | 3.35E-01 | 3.05E-01 | 4.87E-01 |
| Volatile Organic Compounds (VOCs) | | | | | | | | | |
| Acetone | | 58.08 | -2.40E-01 | 5.20E-04 | 2.22E-01 | 1.53E-03 | 3.34E-01 | 3.04E-01 | 5.34E-01 |
| Benzene | | 78.11 | 2.13E+00 | 1.47E-02 | 2.88E-01 | 5.01E-02 | 3.68E-01 | 3.34E-01 | 6.91E-01 |
| Carbon disulfide | | 76.13 | 1.94E+00 | 1.13E-02 | 2.81E-01 | 3.80E-02 | 3.59E-01 | 3.27E-01 | 6.74E-01 |
| Chloroform | | 119.38 | 1.97E+00 | 6.79E-03 | 4.90E-01 | 2.85E-02 | 3.53E-01 | 3.21E-01 | 1.18E+00 |
| 1,1-Dichloroethane | | 98.96 | 1.79E+00 | 6.72E-03 | 3.77E-01 | 2.57E-02 | 3.51E-01 | 3.19E-01 | 9.04E-01 |
| 1,1-Dichloroethene | | 96.94 | 2.13E+00 | 1.16E-02 | 3.67E-01 | 4.38E-02 | 3.63E-01 | 3.30E-01 | 8.81E-01 |
| cis-1,2-Dichloroethene | | 96.94 | 1.86E+00 | 7.67E-03 | 3.67E-01 | 2.90E-02 | 3.53E-01 | 3.21E-01 | 8.81E-01 |
| trans-1,2-Dichloroethene | | 96.94 | 2.09E+00 | 1.09E-02 | 3.67E-01 | 4.12E-02 | 3.61E-01 | 3.29E-01 | 8.81E-01 |
| 1,2-Dichloropropane | | 112.99 | 1.98E+00 | 7.48E-03 | 4.51E-01 | 3.06E-02 | 3.54E-01 | 3.22E-01 | 1.08E+00 |
| Ethylbenzene | | 106.17 | 3.15E+00 | 4.84E-02 | 4.13E-01 | 1.92E-01 | 4.71E-01 | 4.33E-01 | 9.92E-01 |
| Tetrachloroethene | | 165.83 | 3.40E+00 | 3.28E-02 | 8.92E-01 | 1.62E-01 | 4.49E-01 | 4.11E-01 | 2.14E+00 |
| Toluene | | 92.14 | 2.73E+00 | 3.06E-02 | 3.45E-01 | 1.13E-01 | 4.12E-01 | 3.76E-01 | 8.28E-01 |
| 1,1,1-Trichloroethane | | 133.41 | 2.49E+00 | 1.25E-02 | 5.87E-01 | 5.55E-02 | 3.71E-01 | 3.38E-01 | 1.41E+00 |
| Trichloroethene | | 131.39 | 2.42E+00 | 1.15E-02 | 5.72E-01 | 5.08E-02 | 3.68E-01 | 3.35E-01 | 1.37E+00 |
| Vinyl Chloride | | 62.5 | 1.62E+00 | 8.30E-03 | 2.35E-01 | 2.52E-02 | 3.50E-01 | 3.19E-01 | 5.65E-01 |
| Xylenes | | 106.17 | 3.16E+00 | 4.91E-02 | 4.13E-01 | 1.95E-01 | 4.74E-01 | 4.35E-01 | 9.92E-01 |
| 1,2,4-trimethylbenzene | | 120.2 | 3.63E+00 | 8.37E-02 | 4.95E-01 | 3.53E-01 | 5.99E-01 | 5.66E-01 | 1.19E+00 |
| o-xylene | | 106.17 | 3.12E+00 | 4.62E-02 | 4.13E-01 | 1.83E-01 | 4.65E-01 | 4.26E-01 | 9.92E-01 |
| Naphthalene | | 128.18 | 3.30E+00 | 4.57E-02 | 5.49E-01 | 1.99E-01 | 4.77E-01 | 4.38E-01 | 1.32E+00 |

 Table A-14.3

 Estimation of Kp, Tau(event), B, and t* for Organic Compounds: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC

| Definition | Variable | e Value | | Equation | S | | | | |
|--|-----------------|-------------------------|---------------------|--------------------------|------------------------------------|---|-----------------|-----------------|------|
| Permeability coefficient from water | K _p | Calculated (Equation 1) | cm/hr | ¹⁾ <i>v</i> – | 10 ^{(-2.8+0.66 Log I} | $K_{OW} = -0.0056 MW$ | | | - |
| Octanol:water partition coefficient | K _{ow} | Chemical-specific | dimensionless | \mathbf{n}_p – | 10 | | | | ł |
| Molecular weight | MW | Chemical-specific | g/mole | 2) | $(1)^{2}$ | | | | ł |
| Lag time per event | τ_{event} | Calculated (Equation 2) | hr/event | $\tau_{event} = -$ | <u>sc</u> / | | | | ľ |
| Thickness of the strateum corneum | l_{sc} | 0.001 | cm | | D_{sc} | | | | ł |
| Effective diffusion coefficient, through the stratum corneum | D _{sc} | Calculated (Equation 2) | cm ² /hr | where: | 1 10(- | 28 = 0.0056 MW | | | ľ |
| Relative contribution of permeability coefficients in stratum corneum and viable epidermis | В | Calculated (Equation 3) | dimensionless | D | $l_{sc} = l_{sc} \times 10^{(-1)}$ | 2.0 0.0000 11 (7) | | | |
| Time it takes to reach steady state | t* | Calculated (Equation 4) | hr | 3) | \sqrt{MW} | | | | ł |
| Correlation coefficient | b | Calculated (Equation 4) | dimensionless | B = I | $K_p \frac{\sqrt{MW}}{26}$ | (as an approximatio | n) | | ľ |
| Correlation coefficient | с | Calculated (Equation 4) | dimensionless | | 2.0 | | | | ľ |
| | | | | | $\leq 0.6: t^* = 2$ | | | | |
| | | | | If B | $> 0.6: t^* = 6$ | $	au_{event} \left(b - \sqrt{b^2 - c^2} \right)$ |) | | ſ |
| | | | | where: | $=\frac{1+3B+3B^2}{3\times(1+B)}$ | | | | ľ |
| | | | | c | $3 \times (1+B)$ | | | | |
| | | | | h | $=\frac{2\times(1+B)^2}{2}$ | c | | | |
| | | | | υ | π | | | | |
| | | MW | LogK _{OW} | Kp ^a | τ _{event} | В | с | b | t* |
| Analyte | | (g/mole) | (dimensionless) | (cm/hr) | (hr/event) | (dimensionless) | (dimensionless) | (dimensionless) | (hr) |

a NRP = Not reliably predicted; the compound's chemical properties fall outside the Effective Prediction Domain for Kp (Equations 3.9 and 3.10; USEPA 2004a).

 Table A-14.4

 Calculation of Dose Absorbed Per Unit Area Per Event: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC

| Definition | Variable | Value | | Equations | | | |
|--|--------------------------------|----------------------|---------------------------------|------------------------------------|-------------------------------------|---|---------------------|
| Relative contribution of permeability coefficients | В | Chemical-specific | dimensionless | Organics: | | | |
| in stratum corneum and viable epidermis | | | | If $t_{event} \le t^*$, then: | | <i></i> | |
| Concentration of chemical in water | C _{water} | Measured | μg/L | DA = 2FA | $\times K \times (C)$ | $\times CF_d$) $\times \left(\sqrt{\frac{6\tau_{event}}{\pi}}\right)$ | t _{event} |
| Conversion Factor | CF_d | 1.0E-06 | $(mg L) / (\mu g cm^3)$ | | | | |
| Dose absorbed per unit area per event | DA _{event} | | mg/cm ² -event | If $t_{event} > t^*$, then: | | $\times \left[\frac{t_{event}}{1+B} + 2\tau_{event} \left(\frac{1+B}{B}\right)\right]$ | 2)] |
| Fraction Absorbed, worker | FA_w | Chemical-specific | dimensionless | $DA_{max} = FA_m \times K$ | $\times (C_{m} \times CF_{d})$ | $\times \left \frac{t_{event}}{t_{event}} + 2\tau \right = \left \frac{1+\tau}{t_{event}} \right $ | $3B+3B^2$ |
| Permeability coefficient from water | K _p | Chemical-specific | cm/hr | event | p (• waier • • • a) | 1+B | $(1+B)^2$ |
| Lag time per event | τ_{event} | Chemical-specific | hr/event | | | L 、 | |
| Duration of event | t _{event} | - | hr/event | Inorganics: | | | |
| Time it takes to reach steady state | t* | Chemical-specific | | 0 | | | |
| The reactor steady state | ť | chemieur speerne | | $DA_{event} = K_p \times (e^{-1})$ | $C_{water} \times CF_d) \times t_d$ | event | |
| | | | | | | | |
| | $\mathbf{FA}_{\mathbf{W}}^{a}$ | $\mathbf{K_{p}}^{b}$ | C _{water} ^c | $\mathbf{\tau_{event}}^{d}$ | t* ^d | \mathbf{B}^{d} | DA _{event} |
| Analyte | (dimensionless) | (cm/hr) | $(\mu g/L)$ | (hr/event) | (hr) | (dimensionless) | $(mg/cm^2 - event)$ |
| Inorganic Constituents | | | | | | | |
| Vanadium | | 1.00E-03 | 2.23E-01 | 2.03E-01 | 4.87E-01 | 2.75E-03 | 4.46E-10 |
| Volatile Organic Compounds (VOCs) | | | | | | | |
| Acetone | 1 | 5.20E-04 | 3.88E+00 | 2.22E-01 | 5.34E-01 | 1.53E-03 | 4.93E-09 |
| Benzene | 1 | 1.47E-02 | 1.37E-01 | 2.88E-01 | 6.91E-01 | 5.01E-02 | 5.07E-09 |
| Carbon disulfide | 1 | 1.13E-02 | 2.55E-01 | 2.81E-01 | 6.74E-01 | 3.80E-02 | 7.25E-09 |
| Chloroform | 1 | 6.79E-03 | 1.74E-01 | 4.90E-01 | 1.18E+00 | 2.85E-02 | 3.49E-09 |
| 1,1-Dichloroethane | 1 | 6.72E-03 | 5.00E+00 | 3.77E-01 | 9.04E-01 | 2.57E-02 | 9.15E-08 |
| 1,1-Dichloroethene | 1 | 1.16E-02 | 1.16E+00 | 3.67E-01 | 8.81E-01 | 4.38E-02 | 3.60E-08 |
| cis-1,2-Dichloroethene | 1 | 7.67E-03 | 6.60E+02 | 3.67E-01 | 8.81E-01 | 2.90E-02 | 1.37E-05 |
| trans-1,2-Dichloroethene | 1 | 1.09E-02 | 1.80E+00 | 3.67E-01 | 8.81E-01 | 4.12E-02 | 5.26E-08 |
| 1,2-Dichloropropane | 1 | 7.48E-03 | 1.79E-01 | 4.51E-01 | 1.08E+00 | 3.06E-02 | 3.85E-09 |
| Ethylbenzene | 1 | 4.84E-02 | 4.47E-02 | 4.13E-01 | 9.92E-01 | 1.92E-01 | 5.75E-09 |
| Tetrachloroethene | 1 | 3.28E-02 | 5.45E+01 | 8.92E-01 | 2.14E+00 | 1.62E-01 | 6.59E-06 |
| Toluene | 1 | 3.06E-02 | 6.40E-01 | 3.45E-01 | 8.28E-01 | 1.13E-01 | 5.02E-08 |
| 1,1,1-Trichloroethane | 1 | 1.25E-02 | 2.22E+00 | 5.87E-01 | 1.41E+00 | 5.55E-02 | 8.69E-08 |
| Trichloroethene | 1 | 1.15E-02 | 4.37E+01 | 5.72E-01 | 1.37E+00 | 5.08E-02 | 1.56E-06 |
| Vinyl Chloride | 1 | 8.30E-03 | 4.10E+00 | 2.35E-01 | 5.65E-01 | 2.52E-02 | 8.28E-08 |
| Xylenes | 1 | 4.91E-02 | 1.32E-01 | 4.13E-01 | 9.92E-01 | 1.95E-01 | 1.72E-08 |
| 1,2,4-trimethylbenzene | 1 | 8.37E-02 | 4.85E-02 | 4.95E-01 | 1.19E+00 | 3.53E-01 | 1.13E-08 |
| o-xylene | 1 | 4.62E-02 | 7.35E-02 | 4.13E-01 | 9.92E-01 | 1.83E-01 | 9.05E-09 |
| Naphthalene | 1 | 4.57E-02 | 4.52E-02 | 5.49E-01 | 1.32E+00 | 1.99E-01 | 6.16E-09 |

Table A-14.4

Calculation of Dose Absorbed Per Unit Area Per Event: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC

FAr for organic chemicals is from Exhibit B-3, USEPA (2004. RAGS Part E. EPA/540/R/99/005); for chemicals not listed in Exhibit B-3, a default value of 1.0 was used.

 $K_{\rm p}$ for inorganics is from USEPA (2004. RAGS Part E. EPA/540/R/99/005). $K_{\rm p}$ for organics is from Table A-18.3

NRP = Not Reliably Predicted (see Table A-18.3).

from Table A-18.3

from Table A-18.3.

 Table A-14.5

 Risk and Hazard Estimates: Dermal Contact with Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC

| Definition | Variable | Value | | Equations | | | | |
|--|----------------------------|---------------------|---------------------------|--|--------------------------|--|---|-------|
| Averaging Time, carcinogens | AT_{c} | 70 | yrs | Carcinogenic: | | | | |
| Averaging Time, noncarcinogens, worker | $AT_{nc,w}$ | 1 | yrs | $\mathbf{D}_{inlt} = \mathbf{S}\mathbf{I}$ | | $_{t} \times \left(\frac{EF_{w} \times ED_{w} \times ED_{w} \times BW_{a} \times AT_{c} \times 36}{BW_{a} \times AT_{c} \times 36}\right)$ | $EV_{w} \times SA_{w}$ | |
| Body Weight, adult | \mathbf{BW}_{a} | 80 | kg | Klsk = SF | $d \times DA_{even}$ | $t \times \left(\overline{BW_a \times AT_c \times 36} \right)$ | 65 day / year | |
| COPC Absorbed Dose per Event | DA _{event} | chemical-specific | mg/cm ² -event | | L | · | | |
| Exposure Duration, worker | ED_w | 1 | yrs | Noncarcinogenic: | 1 | $(FF \sim 1)$ | $FD \rightarrow FV \rightarrow SA$ | 7] |
| Exposure Frequency, worker | EF_{w} | 9 | days/yr | Hazard = | $\frac{1}{PfD} \times L$ | $PA_{event} \times \left \frac{E\Gamma_w \times T}{PW \times AT} \right $ | $\frac{ED_{w} \times EV_{w} \times SA_{w}}{C_{nc,w} \times 365 \ day / \ year}$ | |
| Event Frequency, worker | EV_{w} | 2 | events/day | | $K_{J}D_{d}$ | $\left(DW_{a} \times AI \right)$ | $_{nc,w}$ × 505 aug / year | ノ」 |
| Oral Reference Dose Adjusted for GI Absorption | RfD_d | chemical-specific | | | | | | |
| Exposed Body Surface Area, worker | SA _a | 5700 | cm ² | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF _d | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Hazard | |
| | | DA _{event} | SFD | Risk | % of | RfD _D | Quotient | % of |
| Analyte | | $(mg/cm^2$ -event) | $(mg/kg-day)^{-1}$ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| Inorganic Constituents | | | | | | | | |
| Vanadium | | 4.46E-10 | | | | 1.30E-04 | 1.21E-05 | < 1% |
| Volatile Organic Compounds (VOCs) | | | | | | | | |
| Acetone | | 4.93E-09 | | | | 9.00E-01 | 1.93E-08 | < 1% |
| Benzene | | 5.07E-09 | 5.50E-02 | 1.40E-11 | < 1% | 4.00E-03 | 4.45E-06 | < 1% |
| Carbon disulfide | | 7.25E-09 | | | | 1.00E-01 | 2.55E-07 | < 1% |
| Chloroform | | 3.49E-09 | 3.10E-02 | 5.43E-12 | < 1% | 1.00E-02 | 1.23E-06 | < 1% |
| 1,1-Dichloroethane | | 9.15E-08 | 5.70E-03 | 2.62E-11 | < 1% | 2.00E-01 | 1.61E-06 | < 1% |
| 1,1-Dichloroethene | | 3.60E-08 | | | | 5.00E-02 | 2.53E-06 | < 1% |
| cis-1,2-Dichloroethene | | 1.37E-05 | | | | 2.00E-03 | 2.40E-02 | 62% |
| trans-1,2-Dichloroethene | | 5.26E-08 | | | | 2.00E-02 | 9.24E-06 | < 1% |
| 1,2-Dichloropropane | | 3.85E-09 | 3.60E-02 | 6.95E-12 | < 1% | 9.00E-02 | 1.50E-07 | < 1% |
| Ethylbenzene | | 5.75E-09 | 1.10E-02 | 3.17E-12 | < 1% | 1.00E-01 | 2.02E-07 | < 1% |
| Tetrachloroethene | | 6.59E-06 | 2.10E-03 | 6.95E-10 | 9% | 6.00E-03 | 3.86E-03 | 10% |
| Toluene | | 5.02E-08 | | | | 8.00E-02 | 2.21E-06 | < 1% |
| 1,1,1-Trichloroethane | | 8.69E-08 | | | | 2.00E+00 | 1.53E-07 | < 1% |
| Trichloroethene | | 1.56E-06 | 4.60E-02 | 3.61E-09 | 49% | 5.00E-04 | 1.10E-02 | 28% |
| Vinyl Chloride | | 8.28E-08 | 7.20E-01 | 2.99E-09 | 41% | 3.00E-03 | 9.70E-05 | < 1% |
| Xylenes | | 1.72E-08 | | | | 2.00E-01 | 3.03E-07 | < 1% |

 Table A-14.5

 Risk and Hazard Estimates: Dermal Contact with Water: Excavation/Trench Worker, Exposure to Groundwater - RME, Sandblast Area AOPC

| Definition | Variable | Value | | Equations | | | | |
|--|----------------------------|---------------------|---------------------------|---|--------------------------|--|--------------------------------------|-------|
| Averaging Time, carcinogens | AT _c | 70 | yrs | Carcinogenic: | | | | |
| Averaging Time, noncarcinogens, worker | $AT_{nc,w}$ | 1 | yrs | $\mathbf{B} := \mathbf{L} = \mathbf{S} \mathbf{E} \left[\mathbf{D} \mathbf{A} = \left[\mathbf{E} \mathbf{F}_{w} \times \mathbf{E} \mathbf{D}_{w} \times \mathbf{E} \mathbf{V}_{w} \times \mathbf{S} \mathbf{A}_{w} \right] \right]$ | | | | |
| Body Weight, adult | \mathbf{BW}_{a} | 80 | kg | $Risk = SF_d \times \left DA_{event} \times \left(\frac{EF_w \times ED_w \times EV_w \times SA_w}{BW_a \times AT_c \times 365 \ day / year} \right) \right $ | | | | |
| COPC Absorbed Dose per Event | DA _{event} | chemical-specific | | | L | | L . | |
| Exposure Duration, worker | ED_w | 1 | yrs | Noncarcinogenic: Hazard = | 1 [| | $ED \rightarrow EV \rightarrow SA$ | |
| Exposure Frequency, worker | EF_{w} | 9 | days/yr | Hazard = | $\frac{1}{PfD} \times I$ | $DA_{event} \times \left \frac{L\Gamma_{w} \times I}{PW \times AT} \right $ | $ED_{w} \times EV_{w} \times SA_{w}$ | |
| Event Frequency, worker | EV_{w} | 2 | events/day | | KJD_d | $\left(DW_{a} \times AI \right)$ | $_{nc,w}$ × 505 ady / year | ノ |
| Oral Reference Dose Adjusted for GI Absorption | RfD_d | chemical-specific | (mg/kg-day) | | | | | |
| Exposed Body Surface Area, worker | SA_a | 5700 | cm ² | | | | | |
| Oral Slope Factor Adjusted for GI Absorption | SF_d | chemical-specific | (mg/kg-day) ⁻¹ | | | | | |
| | | | | Cancer | | | Hazard | |
| | | DA _{event} | SFD | Risk | % of | RfD _D | Quotient | % of |
| Analyte | | $(mg/cm^2$ -event) | $(mg/kg-day)^{-1}$ | (dimensionless) | Total | (mg/kg-day) | (dimensionless) | Total |
| 1,2,4-trimethylbenzene | | 1.13E-08 | | | | | | |
| o-xylene | | 9.05E-09 | | | | 2.00E-01 | 1.59E-07 | < 1% |
| Naphthalene | | 6.16E-09 | | | | 2.00E-02 | 1.08E-06 | < 1% |
| | | | - | Cancer Risk | | | Hazard Index | |
| | | | Pathway Sums | 7.4E-09 | | | 3.9E-02 | |

APPENDIX B Models Used in the BHHRA

Appendix B Models Used in the BHHRA

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Appendix B-1 Virginia DEQ Trench Model
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- Appendix B-2 USEPA Adult Lead Model (ALM): Occupational
- Appendix B-3 ALM and IUEBK: Residential (Hypothetical Fishing Platform User)
- Appendix B-4 USEPA Johnson and Ettinger Vapor Intrusion Model (JEM): Attenuation Factor Derivation

GENERAL ABBREVIATIONS

(Additional Model-Specific Abbreviations Defined in Individual Sub-Appendices)

```
\% = percent
-- = data not available or not calculated
\mu g = microgram
atm = atom
ACH = air changes per hour
bgs = below ground surface
CF = Conversion Factor
CF_i = unit conversion factor
Cg = concentration gas
Cgw = concentration of contaminant in groundwater
Ctrench = concentration of contaminant in trench
cm = centimeter
cm^3 = cubic centimeter
cm^2 = square centimeter
COPC = chemical of potential concern
dL = deciliter
EE/CA = Engineering Evaluation/Cost Analysis
F = fraction of floor through which contaminant can enter
ft = feet
g = gram
GI = gastrointestinal
GSD = geometric standard deviation
H2O = hydrogen peroxide
Hi = Henry's law constant for contaminant
hr = hour
IEUBK = Integrated Exposure Uptake Biokinetic
K = Kelvin (temperature unit)
```



kg = kilogram Ki = overall mass transfer coefficient of contaminant KiG = gas-phase mass transfer coefficient of i KiL = liquid-phase mass transfer coefficient of I L = literm = meter $m^2 = square meter$ $m^3 = cubic meter$ mg = milligrammol = moleMW = molecular weight Mwi = molecular weight inhaled O2 = dioxidePh = leadPbB = blood lead concentrationppm = parts per million R = ideal gas constant s = secondT = average temperatureVF = volatilization factor yr = year

SOURCES

Appendix B-1

Virginia Department of Environmental Quality (VDEQ) Trench Model. Website Accessed May 13, 2014:

http://deq.state.va.us/Programs/LandProtectionRevitalization/RemediationProgram/VoluntaryRemediationProgram/VRPRiskAssessmentGuidance/Guidance.aspx

EPA Adult Lead Model (ALM) and integrated exposure uptake biokinetic (IEUBK) model for lead in children. Both models available from the following EPA Website: http://www2.epa.gov/superfund/lead-superfund-sites-software-and-users-manuals

Appendix B-2:

USEPA. 1996. Recommendations of the Technical Review Workgroup for Lead for an Interim Approach to Assessing Risks Associated with Adult Exposures to Lead in Soil



Appendix B-3:

- USEPA. 1994. Revised Interim Soil Lead Guidance for CERCLA Sites and RCRA Corrective Action Facilities. Washington, DC: Office of Emergency and Remedial Response. July. OSWER Directive #9355.4-12. July.
- USEPA. 2003. Recommendations of the Technical Review Workgroup for Lead for an Approach to Assessing Risks Associated with Adult Exposures to Lead in Soil [EPA-540-R-03-001, OSWER Dir #9285.7-54. January.
- USEPA. 2010. Integrated Exposure Uptake Biokinetic Model for Lead in Children, IEUBKwin32 V1.1 Build 11, Software Program. Developed by USEPA. February.

Appendix B-4

- USEPA. 2002. OSWER Draft Guidance for Evaluating the Vapor Intrusion to Indoor Air Pathway from Groundwater and Soils (Subsurface Vapor Intrusion Guidance) November 2002. EPA530-D-02-004.
- USEPA. 2004. Johnson and Ettinger Model. The USEPA SG-ADV Version 3.1 (dated 02/04) modified for multiple chemicals. The original model is available at the EPA website: http://www.epa.gov/oswer/riskassessment/airmodel/johnson_ettinger.htm
- USEPA. 2013. Oswer Final Guidance for Assessing And Mitigating The Vapor Intrusion Pathway From Subsurface Sources To Indoor Air. April.



Appendix B-1 Virginia DEQ Trench Model

INDEX OF TABLES

Table B-1.1: VDEQ Trench Model Settings

Table B-1.2: VDEQ Results for Landfill AOPC

Table B-1.3: VDEQ Results for Sandblast Area AOPC

TRENCH MODEL INPUT DERIVATION EVALUATION

Virginia Department of Environmental Quality (VDEQ) Trench Model

- The Virginia Department of Environmental Quality (VDEQ) Trench Model
- Accessed May 13, 2014: <u>http://deq.state.va.us/Programs/LandProtectionRevitalization/RemediationProgram/Volu</u> <u>ntaryRemediationProgram/VRPRiskAssessmentGuidance/Guidance.aspx</u>
- Recommend using the maximum detected VOCs concentration in groundwater (GW)
- Landfill: the maximum detected VOCs in monitoring well (MW) were used.
- **Sandblast:** the higher of the maximum VOC detected in either MW or temporary monitoring well (i.e., direct push wells) were used.
- Surrogate: 1,2-dichloroprpane used for 2,2-dichloropropane
- Surrogate: total xylenes used for m,p-xylenes
- VOC analysis for naphthalene used
- Depth to GW at both Landfill and Sandblast Area AOPCs met the criteria of potentially pooling in trenches 8 ft deep;

Landfill: shallowest GW elevation listed as **7.3 ft below top of casing** (Appendix D of RI; Table D-1; MW-1; 10/24/2001),

Sandblast: **6.41 ft BTC** (MW-11; 4/15/2008)

- The depth to width ratio (trench shape) is the most significant parameter to affect the attenuation factor (AF) as a width greater than depth changes the Air Exchange Rate (ACH) from 2 per hour (similar to an indoor room) to 360/hr. This means the overall volume of the trench does not make a difference.
- The VDEQ model defaults were used in calculating the chemical specific AFs. Key ones listed (See Attachment C for complete list)

Temperature = 77 f (F) Fraction of Floor through which contaminant can enter (unitless) = 1 Length = 8 ft Width = 3 ft Depth = 8 ft Model Calculated width/depth ratio is 0.38 Model Calculated ACH = 2/hr

Table B-1.1VDEQ Trench Model Settings

| For Mass-Transfer Coefficients | | For Emission Flux and Concentration in Trench | | | Trench dimensions | | | |
|--------------------------------|----------|---|-----|----------|-------------------|-------------|--------|----|
| Kg,H2O | 0.833 | cm/s | CF1 | 1.00E-03 | L/cm3 | Length | 8 | ft |
| MWH2O | 18 | | CF2 | 1.00E+04 | cm2/m2 | | 2.44 1 | m |
| K1,O2 | 0.002 | cm/s | CF3 | 3600 | s/hr | Width | 3 | ft |
| MWO2 | 32 | | F | 1 | | | 0.91 | m |
| Т | 77 | F | ACH | 2 | hr-1 | Depth | 8 1 | ft |
| Т | 298 | K | | | | | 2.44 | m |
| R | 8.20E-05 | atm-m3/mol-K | | | | Width/Depth | 0.38 | |

Table B-1.2

VDEQ Results for Landfill AOPC

(Accessed 02/05/2014)

| Table 3.8 Exposure-point concentrations (inhalation) for construction/utility workers in a trench: Groundwater less than 15 feet deep revised 10/5/07 | CAS No. | Molecular Weight MWi g/mol | Henry's Law Constant Hi atm-m3/mol | Gas-Phase Mass Transfer Coefficient KiG cm/s | Liquid-Phase Mass Transfer Coefficient KiL cm/s | Overall Mass Transfer Coefficient Ki cm/s | Concentration of Contaminant in Groundwater Cgw ug/L | | Concentration of Contaminant in Trench Ctrench ug/m3 | Concentration of Contaminant in Trench Ctrench mg/m3 |
|---|----------|-------------------------------------|---|--|---|---|--|----------|--|--|
| Acetone | 67-64-1 | 58.08 | 3.88E-05 | 5.63E-01 | 1.48E-03 | 5.58E-04 | 1.54E+01 | 4.12E+00 | 6.34E+01 | 6.34E-02 |
| Chloroform | 67-66-3 | 119.38 | 3.67E-03 | 4.42E-01 | 1.04E-03 | 1.02E-03 | 3.70E+00 | 7.53E+00 | 2.78E+01 | 2.78E-02 |
| Isopropylbenzene | 98-82-8 | 120.19 | 1.16E+00 | 4.41E-01 | 1.03E-03 | 1.03E-03 | 4.60E+00 | 7.62E+00 | 3.50E+01 | 3.50E-02 |
| Tetrachloroethene | 127-18-4 | 165.83 | 1.84E-02 | 3.96E-01 | 8.79E-04 | 8.76E-04 | 8.78E+00 | 6.47E+00 | 5.68E+01 | 5.68E-02 |
| Vinyl Chloride | 75-01-4 | 62.50 | 2.70E-02 | 5.49E-01 | 1.43E-03 | 1.43E-03 | 9.55E-01 | 1.05E+01 | 1.01E+01 | 1.01E-02 |
| n-propylbenzene | 103-65-1 | 120.19 | 1.05E-02 | 4.41E-01 | 1.03E-03 | 1.03E-03 | 2.00E+00 | 7.58E+00 | 1.52E+01 | 1.52E-02 |
| 1,2,4-trimethylbenzene | 95-63-6 | 120.19 | 6.16E-03 | 4.41E-01 | 1.03E-03 | 1.02E-03 | 5.20E+00 | 7.55E+00 | 3.92E+01 | 3.92E-02 |

Table B-1.3VDEQ Results for Sandblast Area AOPC

(Accessed 02/05/2014)

| Table 3.8 Exposure-point concentrations (inhalation) for construction/utility workers in a trench: Groundwater less than 15 feet deep revised 10/5/07 | CAS No. | Molecular Weight MWi g/mol | Henry's Law Constant Hi atm-m3/mol | Gas-Phase Mass Transfer Coefficient KiG cm/s | Liquid-Phase Mass Transfer Coefficient KiL cm/s | Overall Mass Transfer Coefficient Ki cm/s | Concentration of Contaminant in Groundwater Cgw ug/L | Volatilization Factor VF L/m3 | Concentration of Contaminant in Trench Ctrench ug/m3 | Concentration of Contaminant in Trench Ctrench mg/m3 |
|---|-----------|-------------------------------------|---|--|---|---|--|--|--|--|
| TAL Inorganics | | | | | | | | | | |
| Acetone | 67-64-1 | 58.08 | 3.88E-05 | 5.63E-01 | 1.48E-03 | 5.58E-04 | 3.88E+00 | 4.12E+00 | 1.60E+01 | 1.60E-02 |
| Benzene | 71-43-2 | 78.11 | 5.55E-03 | 5.09E-01 | 1.28E-03 | 1.27E-03 | 1.37E-01 | 9.35E+00 | 1.28E+00 | 1.28E-03 |
| Carbon disulfide | 75-15-0 | 76.14 | 3.03E-02 | 5.14E-01 | 1.30E-03 | 1.29E-03 | 2.55E-01 | 9.55E+00 | 2.44E+00 | 2.44E-03 |
| Chloroform | 67-66-3 | 119.38 | 3.67E-03 | 4.42E-01 | 1.04E-03 | 1.02E-03 | 1.74E-01 | 7.53E+00 | 1.31E+00 | 1.31E-03 |
| 1,1-Dichloroethane | 75-34-3 | 98.96 | 5.62E-03 | 4.71E-01 | 1.14E-03 | 1.13E-03 | 5.00E+00 | 8.31E+00 | 4.15E+01 | 4.15E-02 |
| 1,1-Dichloroethene | 75-35-4 | 96.94 | 2.61E-02 | 4.74E-01 | 1.15E-03 | 1.15E-03 | 1.16E+00 | 8.46E+00 | 9.82E+00 | 9.82E-03 |
| cis-1,2-Dichloroethene | 156-59-2 | 96.94 | 4.08E-03 | 4.74E-01 | 1.15E-03 | 1.13E-03 | 6.60E+02 | 8.36E+00 | 5.52E+03 | 5.52E+00 |
| trans-1,2-Dichloroethene | 156-60-5 | 96.94 | 9.38E-03 | 4.74E-01 | 1.15E-03 | 1.14E-03 | 1.80E+00 | 8.43E+00 | 1.52E+01 | 1.52E-02 |
| 1,2-Dichloropropane | 78-87-5 | 112.99 | 2.80E-03 | 4.50E-01 | 1.06E-03 | 1.04E-03 | 1.79E-01 | 7.70E+00 | 1.38E+00 | 1.38E-03 |
| Ethylbenzene | 100-41-4 | 106.17 | 7.88E-03 | 4.60E-01 | 1.10E-03 | 1.09E-03 | 4.47E-02 | 8.05E+00 | 3.60E-01 | 3.60E-04 |
| Tetrachloroethene | 127-18-4 | 165.83 | 1.84E-02 | 3.96E-01 | 8.79E-04 | 8.76E-04 | 5.45E+01 | 6.47E+00 | 3.52E+02 | 3.52E-01 |
| Toluene | 108-88-3 | 92.14 | 6.64E-03 | 4.82E-01 | 1.18E-03 | 1.17E-03 | 6.40E-01 | 8.62E+00 | 5.52E+00 | 5.52E-03 |
| 1,1,1-Trichloroethane | 71-55-6 | 133.40 | 1.72E-02 | 4.26E-01 | 9.80E-04 | 9.76E-04 | 2.22E+00 | 7.21E+00 | 1.60E+01 | 1.60E-02 |
| Trichloroethene | 79-01-6 | 131.39 | 1.03E-02 | 4.28E-01 | 9.87E-04 | 9.82E-04 | 4.37E+01 | 7.25E+00 | 3.17E+02 | 3.17E-01 |
| Vinyl Chloride | 75-01-4 | 62.50 | 2.70E-02 | 5.49E-01 | 1.43E-03 | 1.43E-03 | 4.10E+00 | 1.05E+01 | 4.32E+01 | 4.32E-02 |
| Total Xylenes | 1330-20-7 | 106.16 | 5.18E-03 | 4.60E-01 | 1.10E-03 | 1.09E-03 | 1.32E-01 | 8.02E+00 | 1.06E+00 | 1.06E-03 |
| 1,2,4-trimethylbenzene | 95-63-6 | 120.19 | 6.16E-03 | 4.41E-01 | 1.03E-03 | 1.02E-03 | 4.85E-02 | 7.55E+00 | 3.66E-01 | 3.66E-04 |
| o-xylene | 95-47-6 | 106.17 | 5.19E-03 | 4.60E-01 | 1.10E-03 | 1.09E-03 | 7.35E-02 | 8.02E+00 | 5.89E-01 | 5.89E-04 |
| Naphthalene | 91-20-3 | 128.17 | 4.83E-04 | 4.32E-01 | 9.99E-04 | 8.95E-04 | 4.52E-02 | 6.60E+00 | 2.98E-01 | 2.98E-04 |

Appendix B-2 USEPA Adult Lead Model (ALM)

Index of Tables

ALM Results for Adult Worker Receptors (Target 10 ug/L)

- Table B-2.1: Landfill AOPC 0-1 foot
- Table B-2.2: Landfill AOPC 0-3 foot
- Table B-2.3: Sandblast Area AOPC 0-1 foot
- Table B-2.4: Sandblast Area AOPC 0-3 foot
- ALM Results for Adult Worker Receptors (Target 5ug/L)
- Table B-2.5: Landfill AOPC 0-1 foot: Target 5 ug/L
- Table B-2.6: Landfill AOPC 0-3 foot: Target 5 ug/L
- Table B-2.7: Sandblast Area AOPC 0-1 foot: Target 5 ug/L
- Table B-2.8: Sandblast Area AOPC 0-3 foot: Target 5 ug/L

Table B-2.1 Landfill AOPC

Soil - unsieved (0-1 foot bgs) Calculations of Blood Lead Concentrations (PbBs) U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 211 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _S | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | | 0.12 |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 219 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.3 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult workers | ug/dL | 3.1 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 10.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.013% |

Table B-2.2 Landfill AOPC

Soil - unsieved (0-3 feet bgs) Calculations of Blood Lead Concentrations (PbBs) U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 342 |
| $R_{fetal/maternal}$ | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _s | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | | 0.12 |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 219 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.5 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult workers | ug/dL | 3.5 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 10.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.032% |

Table B-2.3 Sandblast AOPC

Soil - sieved <250 um (0-1 foot bgs) Calculations of Blood Lead Concentrations (PbBs) U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 300 |
| $R_{fetal/maternal}$ | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _s | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | | 0.12 |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 219 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.4 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult workers | ug/dL | 3.4 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 10.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.025% |

Table B-2.4 Sandblast AOPC

Soil - sieved <250 um (0-3 feet bgs) Calculations of Blood Lead Concentrations (PbBs) U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 202 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _S | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | | 0.12 |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 219 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.3 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult workers | ug/dL | 3.1 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 10.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.013% |

Table B-2.5 Landfill AOPC

Soil - unsieved (0-1 foot bgs) Calculations of Blood Lead Concentrations (PbBs): Target 5ug/L U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 211 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _s | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| Ws | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| $AF_{S, D}$ | Absorption fraction (same for soil and dust) | | 0.12 |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 219 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.3 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult workers | ug/dL | 3.1 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 5.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.68% |

Table B-2.6 Landfill AOPC

Soil - unsieved (0-3 feet bgs) Calculations of Blood Lead Concentrations (PbBs): Target 5 ug/L U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 342 |
| $R_{fetal/maternal}$ | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _s | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _s | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| $AF_{S, D}$ | Absorption fraction (same for soil and dust) | | 0.12 |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 219 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.5 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult workers | ug/dL | 3.5 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 5.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 1.3% |

Table B-2.7 Sandblast AOPC

Soil - sieved <250 um (0-1 foot bgs) Calculations of Blood Lead Concentrations (PbBs): Target 5 ug/L U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 300 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB_0 | Baseline PbB | ug/dL | 1.0 |
| IR _S | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| $AF_{S,D}$ | Absorption fraction (same for soil and dust) | | 0.12 |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 219 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.4 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult workers | ug/dL | 3.4 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 5.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 1.1% |

Table B-2.8 Sandblast AOPC

Soil - sieved <250 um (0-3 feet bgs) Calculations of Blood Lead Concentrations (PbBs): Target 5 ug/L U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

Version date 6/21/09

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 202 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _s | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| Ws | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | | 0.12 |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 219 |
| AT _{S,D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.3 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult workers | ug/dL | 3.1 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 5.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.65% |

Appendix B-3 ALM and IUEBK: Residential (Hypothetical Fishing Platform User)

Table B-3. Summary of Lead Evaluations

ALM Output

Landfill AOPC Soil – Unsieved (0-1 foot bgs) Sandblast Area AOPC – Sieved <250 um (0-1 foot bgs) Pistol Range AOPC Soil – Unsieved Pistol Range AOPC Sediment – Unsieved Bulb Slope AOPC – Unsieved IEUBK Model Input / Output Input Parameters Output: Landfill AOPC, Unsieved (0-1 foot bgs) Output: Sandblast Area AOPC, Sieved (0-1 foot bgs)

Output: Pistol Range AOPC, Unsieved Soil

Output: Pistol Range AOPC, Unsieved Sediment

Output: Bulb Slope AOPC, Unsieved

Table B-3Summary of Lead EvaluationFishing Platform ReceptorsPredicted Blood Lead Levels (PbB) in Child Receptors and Fetuses of Adult Receptors

| AOPC | Medium | EPC (mg/kg) | Receptor | Probability ^(a) | Probability Exceeds 5% ^(b) |
|----------------|--------------------|----------------|----------------|----------------------------|--|
| Landfill | Unsieved soil | 211 | Fetus of Adult | 0.034% | No |
| Landini | (0-1 ft bgs) | 211 | Child | 0.36% | No |
| Sandblast Area | Soil - sieved <250 | 300 | Fetus of Adult | 0.075% | No |
| Sandolast Alea | um (0-1 ft bgs) | 300 | Child | 1.52% | No |
| Pistol Range | Unsieved soil | 208 | Fetus of Adult | 0.033% | No |
| Fistor Kange | Unsieved som | 208 | Child | 0.34% | No |
| Pistol Range | Unsieved sediment | 27.1 | Fetus of Adult | 0.0033% | No |
| Fistor Kange | Unsieved sediment | 27.1 | Child | < 0.075% | No |
| Dulh Slope | Unsieved soil | 222 | Fetus of Adult | 0.038% | No |
| Bulb Slope | Unsieved som | | Child | 0.45% | No |

Notes

(a) Probability that the receptor would have a PbB exceeding the $10 \,\mu g/dL$ level of concern.

(b) USEPA's target is to limit the risk to a typical child or fetus of an adult exposed at a site to no

more than a 5% chance of exceeding the 10 μ g/dL PbB of concern (USEPA 1994, 2003).

PbB was also evaluated for probability of exceeding the CDC recommended level of concern of 5 ug/dl, and all results were less than 5% (CDC 2012, USEPA 2010).

% = percent μ g/dL = micrograms of lead per deciliter of blood AOPC = area of potential concern bgs = below ground surface EPC = exposure point concentration ft = foot mg/kg = milligrams per kilogram PbB = blood lead level um = micrometers

Sources

CDC. 2012. Low Level Lead Exposure Harms Children: A Renewed Call for Primary Prevention. January.

USEPA. 1994. Revised Interim Soil Lead Guidance for CERCLA Sites and RCRA Corrective Action Facilities. Washington, DC: Office of Emergency and Remedial Response. July. OSWER Directive #9355.4-12. July.

USEPA. 2003. Recommendations of the Technical Review Workgroup for Lead for an Approach to Assessing Risks Associated with Adult Exposures to Lead in Soil [EPA-540-R-03-001, OSWER Dir #9285.7-54. January.

USEPA. 2010. Integrated Exposure Uptake Biokinetic Model for Lead in Children, IEUBK win v1.1 Build 11, Software Program. Developed by U.S. EPA. February.

Landfill AOPC

Soil - unsieved (0-1 foot bgs) Calculations of Blood Lead Concentrations (PbBs) U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

Version date 6/21/09

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 211 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _s | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | | 0.12 |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 365 |
| AT _{S,D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.5 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult residents | ug/dL | 3.6 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 10.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.034% |

Sandblast AOPC

Soil - sieved <250 um (0-1 foot bgs) Calculations of Blood Lead Concentrations (PbBs) U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

Version date 6/21/09

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 300 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _S | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | 0.12 | |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 365 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.7 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult residents | ug/dL | 4.1 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 10.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.075% |

ALM: Pistol Range AOPC (Soil)

Soil - unsieved Calculations of Blood Lead Concentrations (PbBs) U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

Version date 6/21/09

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 208 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _S | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | 0.12 | |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 365 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.5 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult residents | ug/dL | 3.5 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 10.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.033% |

.

ALM: Piston Range AOPC (Sediment)

Sediment - unsieved Calculations of Blood Lead Concentrations (PbBs) U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

Version date 6/21/09

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|-------------------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 27.1 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _S | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | (same for soil and dust) 0.12 | |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 365 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.1 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult residents | ug/dL | 2.5 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 10.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.003% |

ALM: Bulb Slope AOPC

Soil - unsieved Calculations of Blood Lead Concentrations (PbBs) U.S. EPA Technical Review Workgroup for Lead, Adult Lead Committee

Version date 6/21/09

| Variable | Description of Variable | Units | GSDi and PbBo from Analysis of NHANES 1999- 2004 |
|-----------------------------|---|---------------------|--|
| PbS | Soil lead concentration | ug/g or ppm | 222 |
| R _{fetal/maternal} | Fetal/maternal PbB ratio | | 0.9 |
| BKSF | Biokinetic Slope Factor | ug/dL per ug/day | 0.4 |
| GSD _i | Geometric standard deviation PbB | | 1.8 |
| PbB ₀ | Baseline PbB | ug/dL | 1.0 |
| IR _S | Soil ingestion rate (including soil-derived indoor dust) | g/day | 0.050 |
| IR _{S+D} | Total ingestion rate of outdoor soil and indoor dust | g/day | |
| W _S | Weighting factor; fraction of IR_{S+D} ingested as outdoor soil | | |
| K _{SD} | Mass fraction of soil in dust | | |
| AF _{S, D} | Absorption fraction (same for soil and dust) | 0.12 | |
| EF _{S, D} | Exposure frequency (same for soil and dust) | days/yr | 365 |
| AT _{S, D} | Averaging time (same for soil and dust) | days/yr | 365 |
| PbB _{adult} | PbB of adult worker, geometric mean | ug/dL | 1.5 |
| PbB _{fetal, 0.95} | 95th percentile PbB among fetuses of adult residents | ug/dL | 3.6 |
| PbB _t | Target PbB level of concern (e.g., 10 ug/dL) | ug/dL | 10.0 |
| $P(PbB_{fetal} > PbB_t)$ | Probability that fetal PbB > PbB _t , assuming lognormal distribution | % | 0.038% |

IEUBK Model Parameters Used to Predict Blood Lead Levels Child Residents Soil and Sediments

| Medium/Parameter | Defau | It Parameters and Va | lues ⁽¹⁾ | Varied Values |
|----------------------------------|------------------------|------------------------------|---------------------|---------------------------|
| Air Data at the Home: | | | | Default values not varied |
| Vary Air Concentration By Year? | No | | | |
| Outdoor Air Lead Concentration: | $0.1 \ \mu g \ Pb/m^3$ | | | |
| Indoor Air Lead Concentration | | | | |
| (Percentage of Outdoor Air): | 30% | | | |
| Lung Absorption: | 32% | | | |
| | | Ventilation | Time Spent | |
| Age-Specific Data | Age | Rate | Outdoors | |
| | 0-1 | $2.0 \text{ m}^3/\text{day}$ | 1 hour/day | |
| | 1-2 | $3.0 \text{ m}^3/\text{day}$ | 2 hour/day | |
| | 2-3 | $5.0 \text{ m}^3/\text{day}$ | 3 hour/day | |
| | 3-4 | $5.0 \text{ m}^3/\text{day}$ | 4 hour/day | |
| | 4-5 | $5.0 \text{ m}^3/\text{day}$ | 4 hour/day | |
| | 5-6 | 7.0 m ³ /day | 4 hour/day | |
| | 6-7 | 7.0 m ³ /day | 4 hour/day | |
| Drinking Water Data at the Home: | | | | Default values not varied |
| Lead Concentration in Drinking W | ater | | | |
| Constant: | $4 \ \mu g/L$ | | | |
| Drinking Water Intake: | Age | Water Consumption | | |
| | 0-1 | 0.20 L/day | | |
| | 1-2 | 0.50 L/day | | |
| | 2-3 | 0.52 L/day | | |
| | 3-4 | 0.53 L/day | | |
| | 4-5 | 0.55 L/day | | |
| | 5-6 | 0.58 L/day | | |
| | 6-7 | 0.59 L/day | | |
| Use Alternative Water Values? | No | | | |
| Absorption Fraction Percent | 50% | | | |
| GI Method/Bioavailability | Non-Linear Ac | ctive Passive Method: | | |
| - | Fraction Passiv | ve/Total Accessible | 0.20 | |
| | | n Level (µg/day) | 100 | |

IEUBK Model Parameters Used to Predict Blood Lead Levels Child Residents Soil and Sediments

| Medium/Parameter | Defau | lt Parameters and Values | (1) | Varied Values |
|----------------------------------|----------------------|-----------------------------|-------|--|
| Diet Data at the Home: | | | | |
| Dietary Lead Intake | Age | Diet Intake | | Default values not varied |
| | 0-1 | 2.26 µg Pb/day | | |
| | 1-2 | 1.96 µg Pb/day | | |
| | 2-3 | 2.13 µg Pb/day | | |
| | 3-4 | 2.04 µg Pb/day | | |
| | 4-5 | 1.95 μg Pb/day | | |
| | 5-6 | 2.05 µg Pb/day | | |
| | 6-7 | 2.22 µg Pb/day | | |
| Absorption Fraction Percent | 50% | | | |
| GI Method/Bioavailability | | ctive Passive Method: | | Default values not varied |
| | | ve/Total Accessible | 0.20 | |
| | Half-Saturatio | n Level (µg/day) | 100 | |
| Alternate Diet Data: | Not used | | | |
| Yard Soil and Indoor House Dus | st Data (soil and du | st at the home): | | Default values not varied |
| Yard Soil Lead Level: | | | | |
| Constant - | Not used | | | |
| Variable - | AOPC-specific | c exposure point concentrat | tions | |
| Indoor House Dust Lead Level: | | | | |
| Constant - | Not used | | | |
| Variable - | Not used | | | |
| | | | | to lead in indoor household dust (0.7) |
| Contribution of outdoor airborne | | oor household dust (10 mg | /kg) | |
| Soil/Dust Ingestion Weighting F | | | | |
| (percent soil): | 45% | | | |

IEUBK Model Parameters Used to Predict Blood Lead Levels Child Residents Soil and Sediments

| Medium/Parameter | Defaul | t Parameters and Values | (1) | Varied Values |
|--|-------------------|--|-------------|---------------------------|
| | | Total House Dust & Yard Soil Intake (default | | |
| Amount Ingested Daily: | Age | values) | | |
| | 0-1 | 0.085 | | |
| | 1-2 | 0.135 | | |
| | 2-3 | 0.135 | | |
| | 3-4 | 0.135 | | |
| | 4-5 | 0.100 | | |
| | 5-6 | 0.090 | | |
| | 6-7 | 0.085 | | |
| Absorption Fraction Percent, Yard | | | | |
| Soil | 30% | | | |
| Absorption Fraction Percent, Hous | | | | |
| Dust | 30% | | | |
| GI Method/Bioavailability: Yard Soil | Fraction Passiv | tive Passive Method: e/Total Accessible 1 Level (µg/day) | 0.20 100 | |
| | | | | |
| GI Method/Bioavailability: House Dust | | tive Passive Method: | | |
| Dust | | e/Total Accessible | 0.20 | |
| | | i Level (µg/day) | 100 | |
| Maternal Data | Hall-Saturation | i Levei (µg/uay) | 100 | |
| Mother's blood lead concentration | at childbirth | 1.0 μg/dL | | Default value not varied |
| Alternate Source Data - non used | | | | |
| GI Method/Bioavailability: Soil/1 | fai Non-Linear Ac | tive Passive Method | | |
| ST method Brouvandonity. 501/1 | | re/Total Accessible | 0.20 | Default values not varied |
| | | i Level (µg/day) | 100 | |
| GSD | 1.6 | | | Default values not varied |
| Cutoff | 10 μg Pb/dL | | | Default values not varied |

IEUBK Model Parameters Used to Predict Blood Lead Levels Child Residents Soil and Sediments

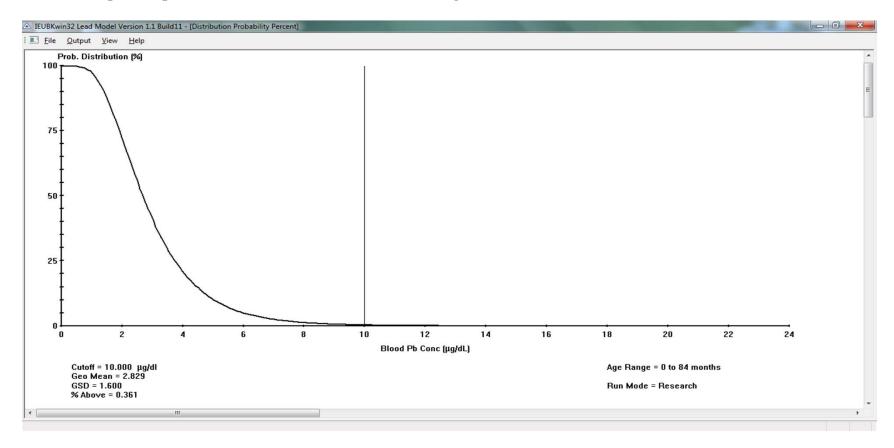
Notes: EE/CA = Engineering Evaluation/Cost Analysis g = gram GI = gastrointestinal GSD = geometric standard deviation IEUBK = Integrated Exposure Uptake Biokinetic kg = kilograms L/day = liter per day m³/day = meters cubed per day mg = milligram mg/kg = milligram per kilogram

μg = microgram μg Pb/day = microgram lead per day μg Pb/dL = microgram lead per deciliter μg Pb/m³ = microgram lead per meter cubed μg/day = microgram per day μg/dL = microgram per deciliter μg/L = micrograms per liter

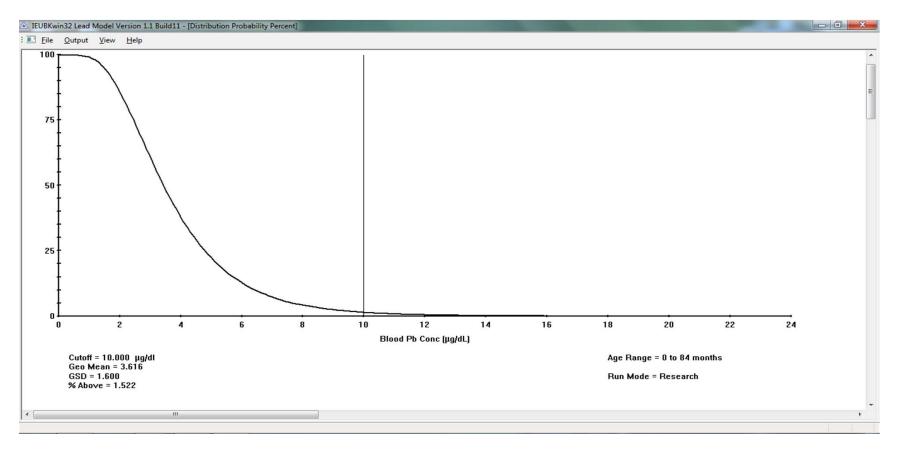
(1) Parameters and values were used in the IEUBK Lead Model (USEPA 2010) Version IEUBKWin32 Version 1.1 Build 11 to estimate blood lead levels and risks.

References:

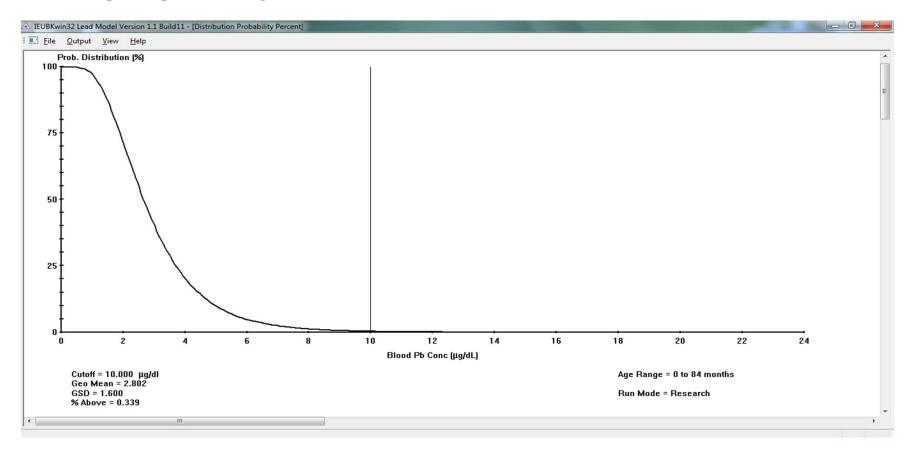
IEUBK Output Graph: Landfill AOPC, Unsieved Soil (0-1 ft bgs)



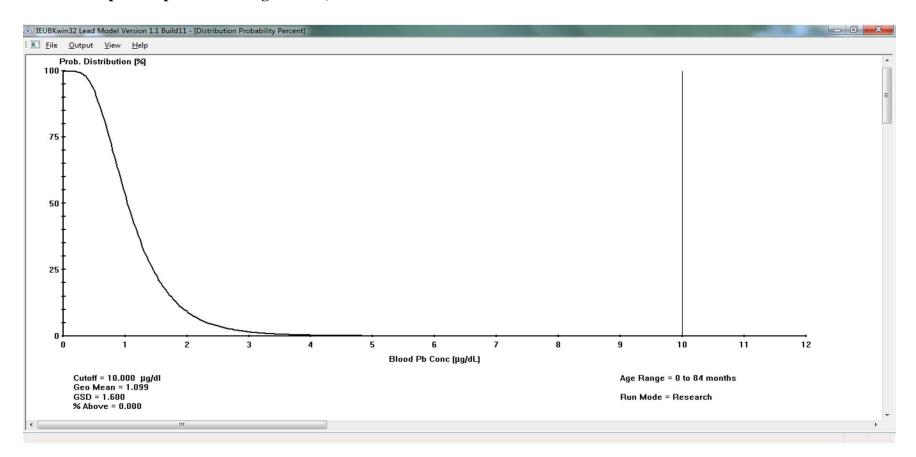




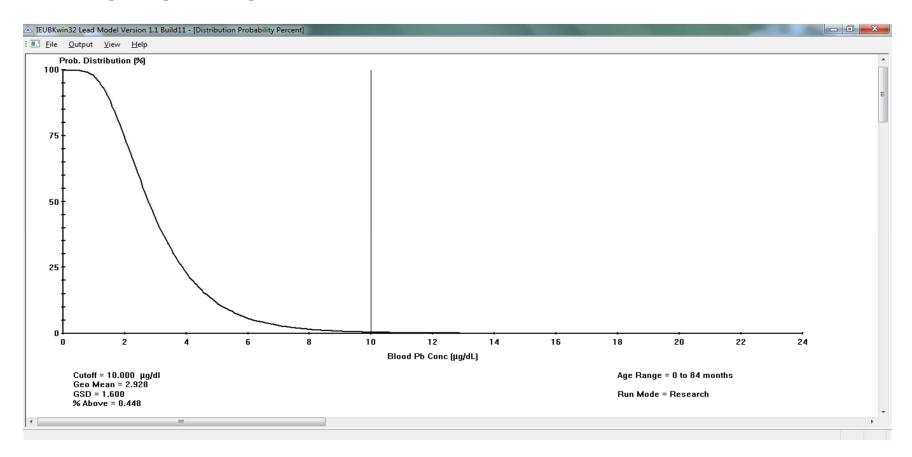
IEUBK Output Graph: Pistol Range AOPC, Unsieved Soil



IEUBK Output Graph: Pistol Range AOPC, Unsieved Sediment



IEUBK Output Graph: Bulb Slope AOPC, Unsieved Soil



Appendix B-4 USEPA Johnson and Ettinger Vapor Intrusion Model (JEM)

Index of JEM Model Pages

- B-4.1 Inputs for Current Scenario: Equipment Building
- B-4.2 Inter-calculations for Current Scenario: Equipment Building
- B-4.3 Inputs for Current Scenario: Service Building
- B-4.4 Inter-calculations for Current Scenario: Service Building
- B-4.5 Inputs for Future Scenario Sandblast Area AOPC
- B-4.6 Inter-calculations for Future Scenario Sandblast Area AOPC

USEPA Johnson and Ettinger Model Input Derivations

- The USEPA SG-ADV Version 3.1 (dated 02/04) modified for multiple chemicals. The original model is available at the EPA website: <u>http://www.epa.gov/oswer/riskassessment/airmodel/johnson_ettinger.htm</u>
- Soil Gas concentrations (ug/m³) were taken from Table 1-6
- Based on USEPA (2013), buildings within 100 ft of the VI source (either lateral or vertical) are considered "near" or within the "buffer zone"
- Based on (email) communications by the USACE, the Sandblast Building (outline shown on figures for Sandblast Area AOPC) has been demolished with no other structures or plans in its place. The occupied portion of the Equipment Building is a portable trailer that is approximately 12 ft x 30 ft. The Bradford Island Service Building sits just outside the Sandblast AOPC boundary but is also included due to being within the 'buffer zone''.
- <u>Current Scenario</u>: Using current building dimensions for existing office space and floor/building descriptions and closest soil gas sampling locations.
 - Current scenario Bradford Island Service Building (RME, CTE)
 - Current Scenario Equipment Building (RME, CTE)
 - The JEM uses a Qsoil default value of 5 L/m. The derivation of Qsoil utilizes the building perimeter; therefore a scaling factor was applied to account for the difference in the BISB building perimeter from the default for the current scenario. The scaling factor derived for the Equipment Building was less than the default value of 5 L/min and therefore the default value was used to maintain the conservative estimates for the Qsoil parameter, which is identified as having "high uncertainty" and "sensitive" parameter (USEPA, 2002). The following equation was used for scaling the model default Qsoil:

Building Perimeter Scaling Equation for Qsoil:

- o Qsoil (L/min)= 5 L/min x Perimeter (cm)/4000 cm
- Qsoil = average vapor flow rate into building (default of 5 L/min)
- \circ L/min = liters per minute

- Cm = centimeter
- <u>Hypothetical Future Scenario</u>: Using the maximum detected concentrations from the Sandblast Area AOPC and using JEM building dimension defaults (USEPA recommended). This is done for the RME and CTE.

B-4.1 Inputs for Current Scenario: Equipment Building

| SG-ADV |
|--------------------|
| Version 3.1; 02/04 |

| | | ENTER | | ENTER | ENTER |
|-------|--------------------------------|--------|----|----------------------|----------------|
| | | Soil | | Soil | |
| | | gas | | gas | Chemical |
| | | conc., | OR | conc., | CAS No. |
| | | C_g | | C_{g} | (numbers only, |
| | Chemical | (ppmv) | | (mg/m ³) | no dashes) |
| | 1,1,1-Trichloroethane | | Г | | 71556 |
| | 1,2,4-Trimethylbenzene | | | | 95636 |
| | 1,3,5-Trimethylbenzene | | | | 108678 |
| | 1,3-Butadiene | | | | 106990 |
| | Methylethylketone (2-butanone) | | | | 78933 |
| | Acetone | | | | 67641 |
| | Benzene | | | | 71432 |
| | Carbon disulfide | | | | 75150 |
| | cis-1,2-Dichloroethylene | | | 2.10E+01 | 156592 |
| | Dichlorodifluoromethane | | | | 75718 |
| | Ethylbenzene | | | | 100414 |
| | Hexane | | | | 110543 |
| | Cumene | | | | 98828 |
| | n-Propylbenzene | | | | 103651 |
| | o-Xylene | | | | 95476 |
| SB-14 | Tetrachloroethylene | | | 7.30E+02 | 127184 |
| | Toluene | | | | 108883 |
| SB-14 | Trichloroethylene | | | 4.10E+01 | 79016 |
| | 1,1-Dichloroethylene | | | 2.70E+00 | 75354 |
| | trans-1,2-Dichloroethylene | | | 5.80E+00 | 156605 |
| | Vinyl chloride (chloroethene) | | | 3.00E+00 | 75014 |

| | | Enter son gas concer | itration above. | | | | | | | | |
|------|--------------------|----------------------|-----------------|-------------------|--------------------------|-----------------------------|-----------------------|-------------------|-------------------------|---------------|---|
| | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | | ENTER | | |
| MORE | Depth | | | Totals m | ust add up to value of l | | Soil | | | | |
| ê | below grade | Soil gas | | | Thickness | Thickness | stratum A | | User-defined | | |
| | to bottom | sampling | Average | Thickness | of soil | of soil | SCS | | stratum A | | |
| | of enclosed | depth | soil | of soil | stratum B, | stratum C, | soil type | | soil vapor | | |
| | space floor, | below grade, | temperature, | stratum A, | (Enter value or 0) | (Enter value or 0) | (used to estimate | OR | permeability, | | |
| | L _F | | Ts | h _A | h _B | h _C | soil vapor | | k _v | | |
| | (cm) | (cm) | (°C) | (cm) | (cm) | (cm) | permeability) | | (cm ²) | | |
| | (ciii) | (ciii) | (C) | (ciii) | (ciii) | (ciii) | permeability) | | (сш) | | |
| | 15 | 106.7 | 10 | 106.7 | | | S | | | | |
| | | (3.5 ft) | | | | | | | | | |
| | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | 1 |
| MORE | Stratum A | Stratum A | Stratum A | Stratum A | Stratum B | Stratum B | Stratum B | Stratum B | Stratum C | Stratum C | S |
| ê | SCS | soil dry | soil total | soil water-filled | SCS | soil dry | soil total | soil water-filled | SCS | soil dry | |
| C | soil type | bulk density, | porosity, | porosity, | soil type | bulk density, | porosity, | porosity, | soil type | bulk density, | |
| | son type | | | | son type | | | | son type | | ł |
| | | r_b^A | n ^A | q_w^A | | r _b ^B | n ^B | q_w^B | | r_b^C | |
| | | (g/cm^3) | (unitless) | (cm^3/cm^3) | | (g/cm^3) | (unitless) | (cm^3/cm^3) | | (g/cm^3) | (|
| | | | | | | | | | | | |
| | S | 1.66 | 0.375 | 0.054 | | | | | | | |
| | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | | ENTER | | |
| | Enclosed | | Enclosed | Enclosed | | | | | Average vapor | | |
| MORE | space | Soil-bldg. | space | space | Enclosed | Floor-wall | Indoor | | flow rate into bldg. | | |
| ê | floor | pressure | floor | floor | space | seam crack | air exchange | | OR | | |
| e | thickness, | differential, | length, | width, | height, | width, | rate, | | Leave blank to calculat | P | |
| | , | DP | - | W _B | H _B | widdii, W | ER | | | c | |
| | L _{crack} | | L _B | | | | | | Q _{soil} | | |
| | (cm) | $(g/cm-s^2)$ | (cm) | (cm) | (cm) | (cm) | (1/h) | | (L/m) | | |
| | 10 | 40 | 914.4 | 365.76 | 366 | 0.1 | 1 | | | | |
| | 10 | | 30 ft | 12 ft | default | 0.1 | 0.25 Resident | | 5 | | |
| | ENTER | ENTER | ENTER | ENTER | ueraun | | 1.0 Commercial | | | | |
| | | | ENTER | ENTER | | | 1.0 Commerciai | | | | |
| | Averaging | Averaging | F | | | | | | | | |
| | time for | time for | Exposure | Exposure | | | | | | | |
| | carcinogens, | noncarcinogens, | duration, | frequency, | | | | | | | |
| | AT _C | AT _{NC} | ED | EF | | Scaling for building Pe | rimeter | | | | |
| | (yrs) | (yrs) | (yrs) | (days/yr) | = | 0.64008 | | | | | |
| | | • | | | - | Scaled Qsoil based on 5 | L/min default | | | | |
| | 70 | 30 | 30 | 350 | J | 3.2004 | | | | | |
| END | | | | | | Less than the default; | therefore use 5 L/min | | | | |

END

| ENTER | ENTER |
|----------------|-------------------|
| Stratum C | Stratum C |
| soil total | soil water-filled |
| porosity, | porosity, |
| n ^C | q_w^C |
| (unitless) | (cm^{3}/cm^{3}) |
| | |
| | |

INTERMEDIATE CALCULATIONS SHEET

B-4.2 Intercalculations for Current Scenario: Equipment Building

| | c. | Stratum A | Stratum B | Stratum C | Stratum A | Stratum A | Stratum A | Stratum A | Floor- | | |
|----------------------|------------------------|--|--|--|---|--------------------------------------|--------------------------------|-------------------------------|----------------------------|-------------------------------|---|
| Exposure | Source- building | soil air-filled | soil air-filled | soil air-filled | effective total fluid | soil intrinsic | soil relative air | soil effective vapor | wall seam | Soil | Bldg. ventilation |
| duration, | separation, | porosity, | porosity, B | porosity, | saturation, | permeability, | permeability, | permeability, | perimeter, | gas | rate, |
| t (sec) | L _T (cm) | q_a^A (cm ³ /cm ³) | q_a^B (cm ³ /cm ³) | q_a^C (cm ³ /cm ³) | S_{te} (cm ³ /cm ³) | k _i (cm ²) | k_{rg} (cm ²) | k_v (cm ²) | X _{crack} (cm) | conc. (mg/m ³) | Q_{buildinf} (cm ³ /s) |
| | | | | | | • | | | | - | · · · · |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 2,560 2,560 | 0.00E+00 0.00E+00 | 3.40E+04 3.40E+04 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 | 2,560 | 0.00E+00 | 3.40E+04 3.40E+04 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 2,560 | 0.00E+00 | 3.40E+04 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 0.998 | 9.91E-08 9.91E-08 | 2,560 2,560 | 0.00E+00 0.00E+00 | 3.40E+04 3.40E+04 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 2,560 | 0.00E+00 | 3.40E+04 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 0.998 | 9.91E-08 9.91E-08 | 2,560 2,560 | 0.00E+00 2.10E+01 | 3.40E+04 3.40E+04 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 2,560 | 0.00E+00 | 3.40E+04 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 2,560 | 0.00E+00 | 3.40E+04 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 2,560 2,560 | 0.00E+00 0.00E+00 | 3.40E+04 3.40E+04 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 2,560 | 0.00E+00 | 3.40E+04 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 2,560 2,560 | 0.00E+00 7.30E+02 | 3.40E+04 3.40E+04 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 2,560 | 0.00E+00 | 3.40E+04 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 2,560 | 4.10E+01 | 3.40E+04 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 2,560 2,560 | 2.70E+00 5.80E+00 | 3.40E+04 3.40E+04 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 2,560 | 3.00E+00 | 3.40E+04 |
| | | | | | | | | | | | |
| Area of | | | | | | | Stratum | Stratum | Stratum | Total | |
| enclosed | Crack- | Crack | Enthalpy of | Henry's law | Henry's law | Vapor | А | В | С | overall | |
| space below | to-total area | depth below | vaporization at ave. soil | constant at ave. soil | constant at ave. soil | viscosity at ave. soil | effective diffusion | effective diffusion | effective diffusion | effective diffusion | Diffusion path |
| grade, | ratio, | grade, | temperature, | temperature, | temperature, | temperature, | coefficient, | coefficient, | coefficient, | coefficient, | length, |
| A _B | h | Zcrack | $\mathrm{DH}_{\mathrm{v,TS}}$ | H _{TS} | H' _{TS} | m _{TS} | D^{eff}_{A} | ${\rm D}^{\rm eff}_{{\rm B}}$ | D^{eff}_{C} | $D^{eff}_{\ T}$ | L_d |
| (cm ²) | (unitless) | (cm) | (cal/mol) | (atm-m ³ /mol) | (unitless) | (g/cm-s) | (cm ² /s) | (cm ² /s) | (cm ² /s) | (cm^2/s) | (cm) |
| 3.73E+05 | 6.87E-04 | 15 | 7,885 | 8.48E-03 | 3.65E-01 | 1.75E-04 | 1.26E-02 | 0.00E+00 | 0.00E+00 | 1.26E-02 | 91.7 |
| 3.73E+05 | 6.87E-04 | 15 | 11,692 | 2.16E-03 | 9.30E-01 | 1.75E-04 | 9.80E-03 | 0.00E+00 | 0.00E+00 | 9.80E-03 | 91.7 |
| 3.73E+05 | 6.87E-04 | 15 | 11,678 | 2.07E-03 | 8.89E-02 | 1.75E-04 | 9.73E-03 | 0.00E+00 | 0.00E+00 | 9.73E-03 | 91.7 |
| 3.73E+05 3.73E+05 | 6.87E-04 6.87E-04 | 15 15 | 5,189 8,419 | 4.62E-02 2.63E-05 | 1.99E+00 1.13E-03 | 1.75E-04 1.75E-04 | 4.03E-02 1.31E-02 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 4.03E-02 1.31E-02 | 91.7 91.7 |
| 3.73E+05 | 6.87E-04 | 15 | 7,559 | 1.97E-05 | 8.47E-04 | 1.75E-04 | 2.01E-02 | 0.00E+00 | 0.00E+00 | 2.01E-02 | 91.7 |
| 3.73E+05 3.73E+05 | 6.87E-04 6.87E-04 | 15 15 | 8,122 6,682 | 2.68E-03 1.66E-02 | 1.15E-01 7.16E-01 | 1.75E-04 1.75E-04 | 1.42E-02 1.68E-02 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 1.42E-02 1.68E-02 | 91.7 91.7 |
| 3.73E+05 | 6.87E-04 6.87E-04 | 15 | 7,734 | 2.04E-03 | 8.77E-02 | 1.75E-04 1.75E-04 | 1.19E-02 | 0.00E+00 | 0.00E+00 0.00E+00 | 1.68E-02 1.19E-02 | 91.7 |
| 3.73E+05 | 6.87E-04 | 15 | 8,386 | 1.62E-01 | 6.96E+00 | 1.75E-04 | 1.08E-02 | 0.00E+00 | 0.00E+00 | 1.08E-02 | 91.7 |
| 3.73E+05 3.73E+05 | 6.87E-04 6.87E-04 | 15 15 | 10,155 7,737 | 3.17E-03 8.32E-01 | 1.36E-01 3.58E+01 | 1.75E-04 1.75E-04 | 1.21E-02 3.23E-02 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 1.21E-02 3.23E-02 | 91.7 91.7 |
| 3.73E+05 | 6.87E-04 | 15 | 12,644 | 3.74E-01 | 1.61E+01 | 1.75E-04 | 1.05E-02 | 0.00E+00 | 0.00E+00 | 1.05E-02 | 91.7 |
| 3.73E+05 3.73E+05 | 6.87E-04 6.87E-04 | 15 15 | 11,368 10,404 | 3.86E-03 2.04E-03 | 1.66E-01 | 1.75E-04 | 9.72E-03 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 9.72E-03 | 91.7 91.7 |
| 3.73E+05 | 6.87E-04 | 15 | 9,553 | 7.81E-03 | 8.79E-02 3.36E-01 | 1.75E-04 1.75E-04 | 1.41E-02 1.16E-02 | 0.00E+00 | 0.00E+00 0.00E+00 | 1.41E-02 1.16E-02 | 91.7 |
| 3.73E+05 | 6.87E-04 | 15 | 9,154 | 2.92E-03 | 1.26E-01 | 1.75E-04 | 1.41E-02 | 0.00E+00 | 0.00E+00 | 1.41E-02 | 91.7 |
| 3.73E+05 3.73E+05 | 6.87E-04 6.87E-04 | 15 15 | 8,557 6,392 | 4.78E-03 1.47E-02 | 2.06E-01 6.33E-01 | 1.75E-04 1.75E-04 | 1.28E-02 1.45E-02 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 1.28E-02 1.45E-02 | 91.7 91.7 |
| 3.73E+05 | 6.87E-04 | 15 | 7,136 | 4.94E-03 | 2.13E-01 | 1.75E-04 | 1.14E-02 | 0.00E+00 | 0.00E+00 | 1.14E-02 | 91.7 |
| 3.73E+05 | 6.87E-04 | 15 | 5,000 | 1.72E-02 | 7.41E-01 | 1.75E-04 | 1.71E-02 | 0.00E+00 | 0.00E+00 | 1.71E-02 | 91.7 |
| | | | | | | Exponent of | Infinite | | | | |
| ~ | ~ | | Average | Crack | | equivalent | source | Infinite | | | |
| Convection path | Source vapor | Crack | vapor flow rate | effective diffusion | Area of | foundation Peclet | indoor attenuation | source bldg. | Unit risk | Reference | |
| length, | conc., | radius, | into bldg., | coefficient, | crack, | number, | coefficient, | conc., | factor, | conc., | |
| L_p | C _{source} | r _{crack} | Q _{soil} | D ^{crack} | Acrack | exp(Pe ^f) | а | C _{building} | URF | RfC | |
| (cm) | (mg/m ³) | (cm) | (cm ³ /s) | (cm ² /s) | (cm ²) | (unitless) | (unitless) | (mg/m^3) | $(mg/m^3)^{-1}$ | (mg/m^3) | - |
| 15 | 0.00E+00 | 0.10 | 8.33E+01 | 1.26E-02 | 2.56E+02 | 1.27E+112 | 9.34E-04 | 0.00E+00 | NA | 2.2E+00 | 1,1,1-Trichloroethane |
| 15 | 0.00E+00 | 0.10 | 8.33E+01 | 9.80E-03 | 2.56E+02 | 1.95E+144 | 7.93E-04 | 0.00E+00 | NA | 6.0E-03 | 1,2,4-Trimethylbenzene |
| 15 15 | 0.00E+00 0.00E+00 | 0.10 0.10 | 8.33E+01 8.33E+01 | 9.73E-03 4.03E-02 | 2.56E+02 2.56E+02 | 1.77E+145 1.31E+35 | 7.89E-04 1.62E-03 | 0.00E+00 0.00E+00 | NA 3.0E-02 | 6.0E-03 2.0E-03 | 1,3,5-Trimethylbenzene 1,3-Butadiene |
| 15 | 0.00E+00 0.00E+00 | 0.10 | 8.33E+01 8.33E+01 | 1.31E-02 | 2.56E+02 2.56E+02 | 1.54E+108 | 9.54E-04 | 0.00E+00 | 3.0E-02 NA | 2.0E-03 5.0E+00 | Methylethylketone (2-butanone |
| 15 | 0.00E+00 | 0.10 | 8.33E+01 | 2.01E-02 | 2.56E+02 | 3.13E+70 | 1.21E-03 | 0.00E+00 | NA | 3.5E-01 | Acetone |
| 15 15 | 0.00E+00 0.00E+00 | 0.10 0.10 | 8.33E+01 8.33E+01 | 1.42E-02 1.68E-02 | 2.56E+02 2.56E+02 | 2.31E+99 1.19E+84 | 1.00E-03 1.10E-03 | 0.00E+00 0.00E+00 | 7.8E-06 NA | 3.0E-02 7.0E-01 | Benzene Carbon disulfide |
| 15 | 2.10E+01 | 0.10 | 8.33E+01 | 1.19E-02 | 2.56E+02 | 6.36E+118 | 9.00E-04 | 1.89E-02 | NA | 3.5E-02 | cis-1,2-Dichloroethylen |
| 15 15 | 0.00E+00 0.00E+00 | 0.10 | 8.33E+01 8.33E+01 | 1.08E-02 1.21E-02 | 2.56E+02 2.56E+02 | 3.08E+131 3.86E+116 | 8.43E-04 9.11E-04 | 0.00E+00 0.00E+00 | NA 2.5E-06 | 2.0E-01 1.0E+00 | Dichlorodifluoromethan Ethylbenzene |
| 15 | 0.00E+00 0.00E+00 | 0.10 | 8.33E+01 8.33E+01 | 3.23E-02 | 2.56E+02 2.56E+02 | 5.25E+43 | 9.11E-04 1.50E-03 | 0.00E+00 | 2.5E-06 NA | 2.0E-01 | Hexane |
| 15 | 0.00E+00 | 0.10 | 8.33E+01 | 1.05E-02 | 2.56E+02 | 3.33E+134 | 8.31E-04 | 0.00E+00 | NA | 4.0E-01 | Cumene |
| 15 15 | 0.00E+00 0.00E+00 | 0.10 0.10 | 8.33E+01 8.33E+01 | 9.72E-03 1.41E-02 | 2.56E+02 2.56E+02 | 3.09E+145 3.20E+100 | 7.88E-04 9.97E-04 | 0.00E+00 0.00E+00 | NA NA | 1.4E-01 1.0E-01 | n-Propylbenzene o-Xylene |
| 15 | 7.30E+02 | 0.10 | 8.33E+01 | 1.16E-02 | 2.56E+02 | 2.78E+121 | 8.88E-04 | 6.48E-01 | 5.9E-06 | 6.0E-01 | Tetrachloroethylen |
| 15 | 0.00E+00 | 0.10 | 8.33E+01 | 1.41E-02 | 2.56E+02 | 3.20E+100 | 9.97E-04 | 0.00E+00 | NA 1 1E 04 | 4.0E-01 | Toluene |
| 15 15 | 4.10E+01 2.70E+00 | 0.10 0.10 | 8.33E+01 8.33E+01 | 1.28E-02 1.45E-02 | 2.56E+02 2.56E+02 | 4.82E+110 1.43E+97 | 9.41E-04 1.02E-03 | 3.86E-02 2.75E-03 | 1.1E-04 NA | 4.0E-02 2.0E-01 | Trichloroethylene 1,1-Dichloroethylene |
| 15 | 5.80E+00 | 0.10 | 8.33E+01 | 1.14E-02 | 2.56E+02 | 4.75E+123 | 8.77E-04 | 5.09E-03 | NA | 7.0E-02 | trans-1,2-Dichloroethylen |
| 15 END | 3.00E+00 | 0.10 | 8.33E+01 | 1.71E-02 | 2.56E+02 | 3.09E+82 | 1.12E-03 | 3.35E-03 | 8.8E-06 | 1.0E-01 | Vinyl chloride (chloroethene) |
| END | L | | | | | | | | | | |

B-4.3 Inputs for Current Scenario: Service Building

| SG-ADV | | | | | | |
|--------------------|--|--|--|--|--|--|
| Version 3.1; 02/04 | | | | | | |

| | Se | oil Gas Concentration | Data | | | _ |
|-----------------|-----------------------|-----------------------|-----------|--|-----------|-------|
| ENTER | ENTER | | ENTER | | | |
| | Soil | | Soil | | | |
| Chemical | gas | | gas | | | |
| CAS No. | conc., | OR | conc., | | | |
| (numbers only, | C_{g} | | C_{g} | | | |
| no dashes) | (mg/m ³) | | (ppmv) | Chemical | | |
| 71556 | |) г | | 1.1.1-Trichloroethane | | _ |
| 71556 95636 | | | | 1,2,4-Trimethylbenzene | | |
| 95636 108678 | | | | 1,2,4-1 hmethylbenzene | | |
| | | | | | | |
| 106990 78933 | | | | 1,3-Butadiene | | |
| 67641 | | | | Methylethylketone (2-butano Acetone | ne) | |
| 71432 | | | | | | |
| | | | | Benzene Carbon disulfide | | |
| 75150 | 8.60E+00 | | | | | |
| 156592 | 8.00E+00 | | | cis-1,2-Dichloroethylene | | |
| 75718 | | | | Dichlorodifluoromethane | | |
| 100414 | | | | Ethylbenzene | | |
| 110543 | | | | Hexane | | |
| 98828 | | | | Cumene | | |
| 103651 | | | | n-Propylbenzene | | |
| 95476 | | | | o-Xylene | | |
| 127184 | 1.80E+03 | | | Tetrachloroethylene | | SB-13 |
| 108883 | 0.505.01 | | | Toluene | | |
| 79016 | 2.70E+01 | | | Trichloroethylene | | SB-13 |
| 75354 | 2.60E+00 | | | 1,1-Dichloroethylene | | _ |
| 156605 | 5.60E+00 | | | trans-1,2-Dichloroethylen | | 4 |
| 75014 | 2.90E+00 | | | Vinyl chloride (chloroethen | e) | |
| | Enter soil gas concer | ntration above. | | | | |
| ENTER | ENTER | ENTER | ENTER | ENTER ENTER | ENTER | |
| Depth | | | Totals mu | st add up to value of Ls (cell F24) | Soil | |
| below grade | Soil gas | | | Thickness Thickness | stratum A | |

| | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | | ENTER | | |
|------|--------------------|-----------------------------|-------------------------|-------------------|------------------------|-------------------------|-------------------|-------------------|------------------------|-----------------------------|---|
| MORE | Depth | | | Totals m | ust add up to value of | Ls (cell F24) | Soil | | | | |
| ê | below grade | Soil gas | | | Thickness | Thickness | stratum A | | User-defined | | |
| | to bottom | sampling | Average | Thickness | of soil | of soil | SCS | | stratum A | | |
| | of enclosed | depth | soil | of soil | stratum B, | stratum C, | soil type | | soil vapor | | |
| | space floor, | below grade, | temperature, | stratum A, | (Enter value or 0) | (Enter value or 0) | (used to estimate | OR | permeability, | | |
| | L_F | L _s | Ts | h _A | h _B | h _C | soil vapor | | k _v | | |
| | (cm) | (cm) | (°C) | (cm) | (cm) | (cm) | permeability) | | (cm^2) | | |
| | | | | | | | | | | | |
| | 15 | 106.7 | 10 | 106.7 | | | S | | | | |
| | | (3.5 ft) | | | | | | | | | |
| | | | | | | ENTER | | | | | |
| MODE | ENTER | ENTER | ENTER | ENTER | ENTER | | ENTER | ENTER | ENTER | ENTER | |
| MORE | Stratum A SCS | Stratum A | Stratum A soil total | Stratum A | Stratum B SCS | Stratum B | Stratum B | Stratum B | Stratum C SCS | Stratum C | |
| ê | | soil dry | | soil water-filled | | soil dry | soil total | soil water-filled | | soil dry | |
| | soil type | bulk density, | porosity, | porosity, | soil type | bulk density, | porosity, | porosity, | soil type | bulk density, | |
| | | r _b ^A | n ^A | q_w^A | | r_b^B | n ^B | q_w^{B} | | r _b ^C | |
| | | (g/cm^3) | (unitless) | (cm^3/cm^3) | | (g/cm^3) | (unitless) | (cm^{3}/cm^{3}) | | (g/cm^3) | |
| | - | | | | | | | | | | |
| | S | 1.66 | 0.375 | 0.054 | | | | | | | - |
| | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | | ENTER | | |
| | Enclosed | ENTER | Enclosed | Enclosed | LINIER | LIVILK | ENTER | | Average vapor | | |
| MORE | space | Soil-bldg. | space | space | Enclosed | Floor-wall | Indoor | | flow rate into bldg. | | |
| ê | floor | pressure | floor | floor | space | seam crack | air exchange | | OR | | |
| C | thickness, | differential, | length, | width, | height, | width, | rate, | | Leave blank to calcula | te | |
| | · · · · · · | DP | L _B | W _B | H _B | width, W | ER | | Q _{soil} | | |
| | L _{crack} | | | - | | | | | | | |
| | (cm) | $(g/cm-s^2)$ | (cm) | (cm) | (cm) | (cm) | (1/h) | | (L/m) | | |
| | 10 | 40 | 3810 | 2286 | 366 | 0.1 | 1 | | 15.24 | | |
| | | | 125 ft | 75 ft | default | | 0.25 Resident | | 13.24 | | |
| | ENTER | ENTER | ENTER | ENTER | doruun | | 1.0 Commercial | | | | |
| | Averaging | Averaging | | | | | | | | | |
| | time for | time for | Exposure | Exposure | | | | | | | |
| | carcinogens, | noncarcinogens, | duration, | frequency, | | | | | | | |
| | AT _C | AT _{NC} | ED | EF | | Scaling for building Pe | rimeter | | | | |
| | (yrs) | (yrs) | (yrs) | (days/yr) | | 3.048 | | | | | |
| | (313) | (315) | (315) | ((()))) | = | Scaled Qsoil based on 5 | L/min default | | | | |
| | 70 | 30 | 30 | 350 | 1 | 15.24 | | | | | |
| | | | | | - | | | | | | |
| END | | | | | | | | | | | |

END

| ENTER | ENTER |
|----------------|-------------------|
| Stratum C | Stratum C |
| soil total | soil water-filled |
| porosity, | porosity, |
| n ^C | q_w^c |
| (unitless) | (cm^3/cm^3) |
| | |
| | |

B-4.4 Inter-Calculations for Current Scenario: Service Building

INTERMEDIATE CALCULATIONS SHEET

| | | Stratum A | Stratum B | Stratum C | Stratum A | Stratum A | Stratum A | Stratum A | Floor- | | |
|-----------------------|-------------------------------|-------------------------------|------------------------------|-----------------------------|----------------------------|----------------------------|-------------------------------|----------------------------------|-------------------------------|---------------------------------|--|
| | Source- | soil | soil | soil | effective | soil | soil | soil | wall | | Bldg. |
| Exposure duration, | building separation, | air-filled porosity, | air-filled porosity, | air-filled porosity, | total fluid saturation, | intrinsic permeability, | relative air permeability, | effective vapor permeability, | seam perimeter, | Soil gas | ventilation rate, |
| t t | L _T | q _a ^A | q_a^B | q_a^C | Saturation, | k _i | k _{rg} | k _v | X _{crack} | conc. | Qbuildins |
| (sec) | (cm) | (cm^{3}/cm^{3}) | (cm^{3}/cm^{3}) | $(\text{cm}^3/\text{cm}^3)$ | (cm^3/cm^3) | (cm ²) | (cm ²) | (cm ²) | (cm) | (mg/m^3) | (cm ³ /s) |
| 0.46E+00 | 01.7 | 0.221 | EDDOD | EDDOD | 0.002 | 0.025.00 | 0.000 | 0.01E.09 | 12 102 | 0.000 | 9.955.05 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 12,192 12,192 | 0.00E+00 0.00E+00 | 8.85E+05 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 0.00E+00 | 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 0.00E+00 | 8.85E+05 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 12,192 12,192 | 0.00E+00 0.00E+00 | 8.85E+05 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 0.00E+00 | 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 0.00E+00 | 8.85E+05 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 12,192 12,192 | 8.60E+00 0.00E+00 | 8.85E+05 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 0.00E+00 | 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 0.00E+00 | 8.85E+05 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 0.998 | 9.91E-08 9.91E-08 | 12,192 12,192 | 0.00E+00 0.00E+00 | 8.85E+05 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 0.00E+00 | 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 1.80E+03 | 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 0.00E+00 | 8.85E+05 |
| 9.46E+08 9.46E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 12,192 12,192 | 2.70E+01 2.60E+00 | 8.85E+05 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 5.60E+00 | 8.85E+05 |
| 9.46E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 12,192 | 2.90E+00 | 8.85E+05 |
| | | | | | | | | | | | |
| Area of | | | | | | | Stratum | Stratum | Stratum | Total | |
| enclosed | Crack- | Crack | Enthalpy of | Henry's law | Henry's law | Vapor | A | B | C | overall | |
| space below | to-total area | depth below | vaporization at ave. soil | constant at ave. soil | constant at ave. soil | viscosity at ave. soil | effective diffusion | effective diffusion | effective diffusion | effective diffusion | Diffusion path |
| grade, | ratio, | grade, | temperature, | temperature, | temperature, | temperature, | coefficient, | coefficient, | coefficient, | coefficient, | length, |
| A_B | h | Z _{crack} | $DH_{v,TS}$ | H _{TS} | H' _{TS} | m _{TS} | D^{eff}_{A} | D^{eff}_{B} | D ^{eff} _C | $\mathbf{D}_{T}^{\mathrm{eff}}$ | L_d |
| (cm ²) | (unitless) | (cm) | (cal/mol) | (atm-m ³ /mol) | (unitless) | (g/cm-s) | (cm ² /s) | (cm ² /s) | (cm ² /s) | (cm ² /s) | (cm) |
| 8 90E + 06 | 1.27E.04 | 15 | 7 005 | 9 49E 02 | 2.65E.01 | 1.75E.04 | 1.26E.02 | 0.00E+00 | 0.00E+00 | 1.26E-02 | 01.7 |
| 8.89E+06 8.89E+06 | 1.37E-04 1.37E-04 | 15 15 | 7,885 11,692 | 8.48E-03 2.16E-03 | 3.65E-01 9.30E-02 | 1.75E-04 1.75E-04 | 1.26E-02 9.80E-03 | 0.00E+00 | 0.00E+00 0.00E+00 | 9.80E-02 | 91.7 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 11,678 | 2.07E-03 | 8.89E-02 | 1.75E-04 | 9.73E-03 | 0.00E+00 | 0.00E+00 | 9.73E-03 | 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 5,189 | 4.62E-02 | 1.99E+00 | 1.75E-04 | 4.03E-02 | 0.00E+00 | 0.00E+00 | 4.03E-02 | 91.7 |
| 8.89E+06 8.89E+06 | 1.37E-04 1.37E-04 | 15 15 | 8,419 7,559 | 2.63E-05 1.97E-05 | 1.13E-03 8.47E-04 | 1.75E-04 1.75E-04 | 1.31E-02 2.01E-02 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 1.31E-02 2.01E-02 | 91.7 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 8,122 | 2.68E-03 | 1.15E-01 | 1.75E-04 | 1.42E-02 | 0.00E+00 | 0.00E+00 | 1.42E-02 | 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 6,682 | 1.66E-02 | 7.16E-01 | 1.75E-04 | 1.68E-02 | 0.00E+00 | 0.00E+00 | 1.68E-02 | 91.7 |
| 8.89E+06 8.89E+06 | 1.37E-04 1.37E-04 | 15 15 | 7,734 8,386 | 2.04E-03 1.62E-01 | 8.77E-02 6.96E+00 | 1.75E-04 1.75E-04 | 1.19E-02 1.08E-02 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 1.19E-02 1.08E-02 | 91.7 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 10,155 | 3.17E-03 | 1.36E-01 | 1.75E-04 | 1.08E-02 1.21E-02 | 0.00E+00 | 0.00E+00 | 1.08E-02 1.21E-02 | 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 7,737 | 8.32E-01 | 3.58E+01 | 1.75E-04 | 3.23E-02 | 0.00E+00 | 0.00E+00 | 3.23E-02 | 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 12,644 | 3.74E-01 | 1.61E+01 | 1.75E-04 | 1.05E-02 | 0.00E+00 | 0.00E+00 | 1.05E-02 | 91.7 |
| 8.89E+06 8.89E+06 | 1.37E-04 1.37E-04 | 15 15 | 11,368 10,404 | 3.86E-03 2.04E-03 | 1.66E-01 8.79E-02 | 1.75E-04 1.75E-04 | 9.72E-03 1.41E-02 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 9.72E-03 1.41E-02 | 91.7 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 9,553 | 7.81E-03 | 3.36E-01 | 1.75E-04 | 1.16E-02 | 0.00E+00 | 0.00E+00 | 1.16E-02 | 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 9,154 | 2.92E-03 | 1.26E-01 | 1.75E-04 | 1.41E-02 | 0.00E+00 | 0.00E+00 | 1.41E-02 | 91.7 |
| 8.89E+06 8.89E+06 | 1.37E-04 1.37E-04 | 15 15 | 8,557 6,392 | 4.78E-03 1.47E-02 | 2.06E-01 6.33E-01 | 1.75E-04 1.75E-04 | 1.28E-02 1.45E-02 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 1.28E-02 1.45E-02 | 91.7 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 7,136 | 4.94E-03 | 2.13E-01 | 1.75E-04 | 1.14E-02 | 0.00E+00 | 0.00E+00 | 1.14E-02 | 91.7 |
| 8.89E+06 | 1.37E-04 | 15 | 5,000 | 1.72E-02 | 7.41E-01 | 1.75E-04 | 1.71E-02 | 0.00E+00 | 0.00E+00 | 1.71E-02 | 91.7 |
| | | | | | | Exponent of | Infinite | | | | |
| | | | Average | Crack | | equivalent | source | Infinite | | | |
| Convection | Source | <i>~</i> · | vapor | effective | | foundation | indoor | source | Unit | | |
| path length, | vapor conc., | Crack radius, | flow rate into bldg., | diffusion coefficient, | Area of crack, | Peclet number, | attenuation coefficient, | bldg. conc., | risk factor, | Reference conc., | |
| L _p | Conc., C _{source} | radius, r _{crack} | Q _{soil} | D ^{crack} | A _{crack} | exp(Pe ^f) | a | Conc., C _{building} | URF | RfC | |
| (cm) | (mg/m ³) | (cm) | (cm ³ /s) | (cm^2/s) | (cm ²) | (unitless) | (unitless) | (mg/m^3) | $(mg/m^3)^{-1}$ | (mg/m^3) | |
| | | <u> </u> | | | | - | | | 1 | | = |
| 15 15 | 0.00E+00 | 0.10 | 2.54E+02 | 1.26E-02 9.80E-03 | 1.22E+03 1.22E+03 | 5.68E+71 | 2.38E-04 | 0.00E+00 | NA | 2.2E+00 | 1,1,1-Trichloroethane |
| 15 | 0.00E+00 0.00E+00 | 0.10 0.10 | 2.54E+02 2.54E+02 | 9.80E-03 9.73E-03 | 1.22E+03 1.22E+03 | 2.27E+92 9.34E+92 | 2.26E-04 2.26E-04 | 0.00E+00 0.00E+00 | NA NA | 6.0E-03 6.0E-03 | 1,2,4-Trimethylbenzene 1,3,5-Trimethylbenzene |
| 15 | 0.00E+00 | 0.10 | 2.54E+02 | 4.03E-02 | 1.22E+03 | 3.00E+22 | 2.69E-04 | 0.00E+00 | 3.0E-02 | 2.0E-03 | 1,3-Butadiene |
| 15 | 0.00E+00 | 0.10 | 2.54E+02 | 1.31E-02 | 1.22E+03 | 1.77E+69 | 2.39E-04 | 0.00E+00 | NA | 5.0E+00 | Methylethylketone (2-butanone |
| 15 15 | 0.00E+00 0.00E+00 | 0.10 0.10 | 2.54E+02 2.54E+02 | 2.01E-02 1.42E-02 | 1.22E+03 1.22E+03 | 1.33E+45 3.99E+63 | 2.54E-04 2.42E-04 | 0.00E+00 0.00E+00 | NA 7.8E-06 | 3.5E-01 3.0E-02 | Acetone Benzene |
| 15 | 0.00E+00 | 0.10 | 2.54E+02 | 1.68E-02 | 1.22E+03 | 6.54E+53 | 2.48E-04 | 0.00E+00 | NA | 7.0E-01 | Carbon disulfide |
| 15 | 8.60E+00 | 0.10 | 2.54E+02 | 1.19E-02 | 1.22E+03 | 1.11E+76 | 2.35E-04 | 2.02E-03 | NA | 3.5E-02 | cis-1,2-Dichloroethylen |
| 15 15 | 0.00E+00 0.00E+00 | 0.10 0.10 | 2.54E+02 2.54E+02 | 1.08E-02 1.21E-02 | 1.22E+03 1.22E+03 | 1.46E+84 4.21E+74 | 2.31E-04 2.36E-04 | 0.00E+00 0.00E+00 | NA 2.5E-06 | 2.0E-01 1.0E+00 | Dichlorodifluoromethane Ethylbenzene |
| 15 | 0.00E+00 0.00E+00 | 0.10 | 2.54E+02 2.54E+02 | 3.23E-02 | 1.22E+03 1.22E+03 | 4.21E+74 9.64E+27 | 2.36E-04 2.65E-04 | 0.00E+00 0.00E+00 | 2.5E-06 NA | 2.0E-01 | Ethylbenzene Hexane |
| 15 | 0.00E+00 | 0.10 | 2.54E+02 | 1.05E-02 | 1.22E+03 | 1.27E+86 | 2.30E-04 | 0.00E+00 | NA | 4.0E-01 | Cumene |
| 15 | 0.00E+00 | 0.10 | 2.54E+02 | 9.72E-03 | 1.22E+03 | 1.33E+93 | 2.26E-04 | 0.00E+00 | NA | 1.4E-01 | n-Propylbenzene |
| 15 15 | 0.00E+00 1.80E+03 | 0.10 0.10 | 2.54E+02 2.54E+02 | 1.41E-02 1.16E-02 | 1.22E+03 1.22E+03 | 2.14E+64 5.42E+77 | 2.42E-04 2.34E-04 | 0.00E+00 4.21E-01 | NA 5.9E-06 | 1.0E-01 6.0E-01 | o-Xylene Tetrachloroethylene |
| 15 | 0.00E+00 | 0.10 | 2.54E+02 2.54E+02 | 1.41E-02 | 1.22E+03 | 2.15E+64 | 2.42E-04 | 0.00E+00 | 5.9E-00 NA | 4.0E-01 | Toluene |
| 15 | 2.70E+01 | 0.10 | 2.54E+02 | 1.28E-02 | 1.22E+03 | 7.02E+70 | 2.38E-04 | 6.43E-03 | 1.1E-04 | 4.0E-02 | Trichloroethylene |
| 15 15 | 2.60E+00 5.60E+00 | 0.10 0.10 | 2.54E+02 2.54E+02 | 1.45E-02 1.14E-02 | 1.22E+03 1.22E+03 | 1.54E+62 1.46E+79 | 2.43E-04 2.33E-04 | 6.32E-04 1.31E-03 | NA NA | 2.0E-01 7.0E-02 | 1,1-Dichloroethylen(trans-1,2-Dichloroethylen(|
| 15 | 2.90E+00 | 0.10 | 2.54E+02 2.54E+02 | 1.14E-02 1.71E-02 | 1.22E+03 1.22E+03 | 6.32E+52 | 2.33E-04 2.49E-04 | 7.22E-04 | NA 8.8E-06 | 1.0E-02 | Vinyl chloride (chloroethene |
| | | | - | • | | | | | | | |

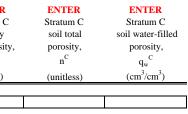
B-4.5 Inputs for Future Scenario Sandblast Area AOPC

SG-ADV Version 3.1; 02/04

| | S | oil Gas Concentratio | n Data | |
|----------------|------------|----------------------|---------|--------------------------------|
| ENTER | ENTER | | ENTER | |
| | Soil | | Soil | |
| Chemical | gas | | gas | |
| CAS No. | conc., | OR | conc., | |
| (numbers only, | C_{g} | | C_{g} | |
| no dashes) | (mg/m^3) | | (ppmv) | Chemical |
| | | | | |
| 75354 | 9.80E+01 | | | 1,1-Dichloroethylene |
| 156592 | 1.85E+04 | | | cis-1,2-Dichloroethylene |
| 156605 | 6.25E+03 | | | trans-1,2-Dichloroethylene |
| 75014 | 2.10E+02 | | | Vinyl chloride (chloroethene) |
| 78933 | 3.10E+01 | | | Methylethylketone (2-butanone) |
| 67641 | 9.70E+01 | | | Acetone |
| 71432 | 8.50E+01 | | | Benzene |
| 75150 | 4.20E+01 | | | Carbon disulfide |
| 156592 | 3.30E+02 | | | cis-1,2-Dichloroethylene |
| 75718 | 3.20E+00 | | | Dichlorodifluoromethane |
| 100414 | 1.55E+03 | | | Ethylbenzene |
| 110543 | 1.10E+02 | | | Hexane |
| 98828 | 1.10E+02 | | | Cumene |
| 103651 | 2.30E+03 | | | n-Propylbenzene |
| 95476 | 2.80E+03 | | | o-Xylene |
| 127184 | 3.40E+04 | | | Tetrachloroethylene |
| 108883 | 4.75E+04 | | | Toluene |
| 79016 | 6.10E+02 | | | Trichloroethylene |

| | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | | ENTER | |
|------|--------------------|-----------------------------|----------------|-----------------------------|--------------------------|-----------------------------|----------------------|-------------------|--------------------------|-----------------------------|
| MORE | Depth | | | Totals m | ust add up to value of I | ls (cell F24) | Soil | | | |
| ê | below grade | Soil gas | | | Thickness | Thickness | stratum A | | User-defined | |
| | to bottom | sampling | Average | Thickness | of soil | of soil | SCS | | stratum A | |
| | of enclosed | depth | soil | of soil | stratum B, | stratum C, | soil type | | soil vapor | |
| | space floor, | below grade, | temperature, | stratum A, | (Enter value or 0) | (Enter value or 0) | (used to estimate | OR | permeability, | |
| | $L_{\rm F}$ | L_s | Ts | h _A | h _B | h _C | soil vapor | | k _v | |
| | (cm) | (cm) | (°C) | (cm) | (cm) | (cm) | permeability) | | (cm^2) | |
| | `` / | × / | | | × / | × / | | | · · · · · · | |
| | 15 | 106.7 | 10 | 106.7 | | | S | | | |
| | default for Slab | 3.5 ft | default | 3.5 ft | | • | Based on boring logs | | • | |
| | | Table 6-3j | | | | | (2004) | | | |
| | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER |
| MORE | Stratum A | Stratum A | Stratum A | Stratum A | Stratum B | Stratum B | Stratum B | Stratum B | Stratum C | Stratum C |
| ê | SUAUIII A | soil dry | soil total | soil water-filled | SCS | soil dry | soil total | soil water-filled | SCS | soil dry |
| е | soil type | bulk density, | porosity, | porosity, | soil type | bulk density, | porosity, | porosity, | soil type | bulk density, |
| | son type | | n ^A | | son type | | n ^B | | son type | |
| | | r _b ^A | | q_w^A | | r _b ^B | | q_w^B | | r _b ^C |
| | | (g/cm^3) | (unitless) | $(\text{cm}^3/\text{cm}^3)$ | | (g/cm^3) | (unitless) | (cm^3/cm^3) | | (g/cm^3) |
| | | | 1 | 1 | | | 1 | | 1 | |
| | S | 1.66 | 0.375 | 0.054 | | | | | | |
| | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | ENTER | | ENTER | |
| | Enclosed | | Enclosed | Enclosed | | | | | Average vapor | |
| MORE | space | Soil-bldg. | space | space | Enclosed | Floor-wall | Indoor | | flow rate into bldg. | |
| ê | floor | pressure | floor | floor | space | seam crack | air exchange | | OR | |
| | thickness, | differential, | length, | width, | height, | width, | rate, | | Leave blank to calculate | • |
| | L _{crack} | DP | L _B | WB | H _B | W | ER | | Q _{soil} | |
| | (cm) | $(g/cm-s^2)$ | (cm) | (cm) | (cm) | (cm) | (1/h) | | (L/m) | |
| | (em) | (8.1 | (em) | (em) | (em) | (em) | (1/11) | | (1) | |
| | 10 | 40 | 1000 | 1000 | 366 | 0.1 | 1 | | 5 | |
| | Default | defaut | default | default | default | default | 0.25 Resident | | | |
| | ENTER | ENTER | ENTER | ENTER | | | 1.0 Commercial | | | |
| | Averaging | Averaging | | | | | | | | |
| | time for | time for | Exposure | Exposure | | | | | | |
| | carcinogens, | noncarcinogens, | duration, | frequency, | | | | | | |
| | AT _C | AT _{NC} | ED | EF | | | | | | |
| | (yrs) | (yrs) | (yrs) | (days/yr) | _ | | | | | |
| | | | | | - | | | | | |
| | 70 | 25 | 25 | 250 | | | | | | |
| | | | | | | | | | | |

END



INTERMEDIATE CALCULATIONS SHEET

B-4.6 Inter-Calculations for Future Scenario Sandblast Area AOPC

| Exposure duration, | Source- building separation, | Stratum A soil air-filled porosity, q_a^A | Stratum B soil air-filled porosity, q _a ^B | Stratum C soil air-filled porosity, q_a^C | Stratum A effective total fluid saturation, | Stratum A soil intrinsic permeability, | Stratum A soil relative air permeability, | Stratum A soil effective vapor permeability, | Floor- wall seam perimeter, | Soil gas | Bldg. ventilation rate, |
|--------------------------------------|--------------------------------------|---|---|---|---|--|--|---|---|---|--|
| t (sec) | L _T (cm) | q_a (cm ³ /cm ³) | q_a (cm ³ /cm ³) | q_a (cm ³ /cm ³) | S_{te} (cm ³ /cm ³) | k _i (cm ²) | k _{rg} (cm ²) | k_v (cm ²) | X _{crack} (cm) | conc. (mg/m^3) | Q _{buildin} (cm ³ /s) |
| ` <i>`</i> | · · · · | · · · · · · · · · · · · · · · · · · · | · · · · · · | · · · · · · | × / | · / | × / | × / | <u>```</u> | × 0 / | |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 9.80E+01 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 1.85E+04 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 6.25E+03 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 2.10E+02 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 3.10E+01 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 9.70E+01 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 8.50E+01 | 1.02E+05 |
| 7.88E+08 7.88E+08 | 91.7 91.7 | 0.321 0.321 | ERROR ERROR | ERROR ERROR | 0.003 | 9.92E-08 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 4,000 4,000 | 4.20E+01 3.30E+02 | 1.02E+05 1.02E+05 |
| 7.88E+08 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 4,000 | 3.30E+02 3.20E+00 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 9.91E-08 | 4,000 | 3.20E+00 1.55E+03 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 1.10E+02 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 1.10E+02 1.10E+02 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 2.30E+03 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 2.80E+03 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 3.40E+04 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 4.75E+04 | 1.02E+05 |
| 7.88E+08 | 91.7 | 0.321 | ERROR | ERROR | 0.003 | 9.92E-08 | 0.998 | 9.91E-08 | 4,000 | 6.10E+02 | 1.02E+05 |
| enclosed space below grade, | Crack- to-total area ratio, | Crack depth below grade, | Enthalpy of vaporization at ave. soil temperature, | Henry's law constant at ave. soil temperature, | Henry's law constant at ave. soil temperature, | Vapor viscosity at ave. soil temperature, | A effective diffusion coefficient, | B effective diffusion coefficient, | C effective diffusion coefficient, | overall effective diffusion coefficient, | Diffusion path length, |
| A _B | h | Z _{crack} | DH _{v.TS} | H _{TS} | H' _{TS} | m _{TS} | D^{eff}_{A} | D ^{eff} _B | D ^{eff} _C | D_{T}^{eff} | L_d |
| (cm^2) | (unitless) | (cm) | (cal/mol) | (atm-m ³ /mol) | (unitless) | (g/cm-s) | (cm^2/s) | (cm^2/s) | (cm^2/s) | (cm^2/s) | (cm) |
| (<i>i</i> | | | | | | 6 | | | | | |
| 1.06E+06 | 3.77E-04 | 15 | 6,392 | 1.47E-02 | 6.33E-01 | 1.75E-04 | 1.45E-02 | 0.00E+00 | 0.00E+00 | 1.45E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 7,734 | 2.04E-03 | 8.77E-02 | 1.75E-04 | 1.19E-02 | 0.00E+00 | 0.00E+00 | 1.19E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 7,136 | 4.94E-03 | 2.13E-01 | 1.75E-04 | 1.14E-02 | 0.00E+00 | 0.00E+00 | 1.14E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 5,000 | 1.72E-02 | 7.41E-01 | 1.75E-04 | 1.71E-02 | 0.00E+00 | 0.00E+00 | 1.71E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 8,419 | 2.63E-05 | 1.13E-03 | 1.75E-04 | 1.31E-02 | 0.00E+00 | 0.00E+00 | 1.31E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 7,559 | 1.97E-05 | 8.47E-04 | 1.75E-04 | 2.01E-02 | 0.00E+00 | 0.00E+00 | 2.01E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 8,122 | 2.68E-03 | 1.15E-01 | 1.75E-04 | 1.42E-02 | 0.00E+00 | 0.00E+00 | 1.42E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 6,682 | 1.66E-02 | 7.16E-01 | 1.75E-04 | 1.68E-02 | 0.00E+00 | 0.00E+00 | 1.68E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 3.77E-04 | 15 15 | 7,734 8,386 | 2.04E-03 | 8.77E-02 6.96E+00 | 1.75E-04 1.75E-04 | 1.19E-02 | 0.00E+00 0.00E+00 | 0.00E+00 | 1.19E-02 | 91.7 91.7 |
| 1.06E+06 1.06E+06 | 3.77E-04 3.77E-04 | 15 | 8,386 | 1.62E-01 3.17E-03 | 6.96E+00 1.36E-01 | 1.75E-04 1.75E-04 | 1.08E-02 1.21E-02 | 0.00E+00 0.00E+00 | 0.00E+00 0.00E+00 | 1.08E-02 1.21E-02 | 91.7 |
| 1.06E+06 1.06E+06 | 3.77E-04 3.77E-04 | 15 | 7,737 | 8.32E-01 | 3.58E+01 | 1.75E-04 | 3.23E-02 | 0.00E+00 | 0.00E+00 0.00E+00 | 3.23E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 3.77E-04 | 15 | 12,644 | 3.74E-01 | 1.61E+01 | 1.75E-04 | 1.05E-02 | 0.00E+00 | 0.00E+00 | 1.05E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 3.77E-04 | 15 | 11,368 | 3.86E-03 | 1.66E-01 | 1.75E-04 | 9.72E-03 | 0.00E+00 | 0.00E+00 | 9.72E-03 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 10,404 | 2.04E-03 | 8.79E-02 | 1.75E-04 | 1.41E-02 | 0.00E+00 | 0.00E+00 | 1.41E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 9,553 | 7.81E-03 | 3.36E-01 | 1.75E-04 | 1.16E-02 | 0.00E+00 | 0.00E+00 | 1.16E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 9,154 | 2.92E-03 | 1.26E-01 | 1.75E-04 | 1.41E-02 | 0.00E+00 | 0.00E+00 | 1.41E-02 | 91.7 |
| 1.06E+06 | 3.77E-04 | 15 | 8,557 | 4.78E-03 | 2.06E-01 | 1.75E-04 | 1.28E-02 | 0.00E+00 | 0.00E+00 | 1.28E-02 | 91.7 |
| | | | | Create | | Exponent of | Infinite | Infinito | | | |

| | | | | | | Exponent of | Infinite | | | | |
|------------|----------------------|--------------------|----------------------|----------------------|--------------------|-----------------------|--------------|-----------------------|-----------------|------------|-------------------------------|
| | | | Average | Crack | | equivalent | source | Infinite | | | |
| Convection | Source | | vapor | effective | | foundation | indoor | source | Unit | | |
| path | vapor | Crack | flow rate | diffusion | Area of | Peclet | attenuation | bldg. | risk | Reference | |
| length, | conc., | radius, | into bldg., | coefficient, | crack, | number, | coefficient, | conc., | factor, | conc., | |
| L_p | C _{source} | r _{crack} | Q _{soil} | D ^{crack} | Acrack | exp(Pe ^f) | а | C_{building} | URF | RfC | |
| (cm) | (mg/m ³) | (cm) | (cm ³ /s) | (cm ² /s) | (cm ²) | (unitless) | (unitless) | (mg/m ³) | $(mg/m^3)^{-1}$ | (mg/m^3) | _ |
| | | | | | | | | | | | |
| 15 | 9.80E+01 | 0.10 | 8.33E+01 | 1.45E-02 | 4.00E+02 | 1.54E+62 | 5.48E-04 | 5.37E-02 | NA | 2.0E-01 | 1,1-Dichloroethylene |
| 15 | 1.85E+04 | 0.10 | 8.33E+01 | 1.19E-02 | 4.00E+02 | 1.11E+76 | 5.10E-04 | 9.44E+00 | NA | 3.5E-02 | cis-1,2-Dichloroethylen |
| 15 | 6.25E+03 | 0.10 | 8.33E+01 | 1.14E-02 | 4.00E+02 | 1.46E+79 | 5.03E-04 | 3.14E+00 | NA | 7.0E-02 | trans-1,2-Dichloroethylen |
| 15 | 2.10E+02 | 0.10 | 8.33E+01 | 1.71E-02 | 4.00E+02 | 6.32E+52 | 5.77E-04 | 1.21E-01 | 8.8E-06 | 1.0E-01 | Vinyl chloride (chloroethene |
| 15 | 3.10E+01 | 0.10 | 8.33E+01 | 1.31E-02 | 4.00E+02 | 1.77E+69 | 5.28E-04 | 1.64E-02 | NA | 5.0E+00 | Methylethylketone (2-butanone |
| 15 | 9.70E+01 | 0.10 | 8.33E+01 | 2.01E-02 | 4.00E+02 | 1.33E+45 | 6.03E-04 | 5.85E-02 | NA | 3.5E-01 | Acetone |
| 15 | 8.50E+01 | 0.10 | 8.33E+01 | 1.42E-02 | 4.00E+02 | 3.99E+63 | 5.44E-04 | 4.62E-02 | 7.8E-06 | 3.0E-02 | Benzene |
| 15 | 4.20E+01 | 0.10 | 8.33E+01 | 1.68E-02 | 4.00E+02 | 6.54E+53 | 5.74E-04 | 2.41E-02 | NA | 7.0E-01 | Carbon disulfide |
| 15 | 3.30E+02 | 0.10 | 8.33E+01 | 1.19E-02 | 4.00E+02 | 1.11E+76 | 5.10E-04 | 1.68E-01 | NA | 3.5E-02 | cis-1,2-Dichloroethylen |
| 15 | 3.20E+00 | 0.10 | 8.33E+01 | 1.08E-02 | 4.00E+02 | 1.46E+84 | 4.91E-04 | 1.57E-03 | NA | 2.0E-01 | Dichlorodifluoromethan |
| 15 | 1.55E+03 | 0.10 | 8.33E+01 | 1.21E-02 | 4.00E+02 | 4.21E+74 | 5.14E-04 | 7.97E-01 | 2.5E-06 | 1.0E+00 | Ethylbenzene |
| 15 | 1.10E+02 | 0.10 | 8.33E+01 | 3.23E-02 | 4.00E+02 | 9.64E+27 | 6.70E-04 | 7.37E-02 | NA | 2.0E-01 | Hexane |
| 15 | 1.10E+02 | 0.10 | 8.33E+01 | 1.05E-02 | 4.00E+02 | 1.27E+86 | 4.86E-04 | 5.35E-02 | NA | 4.0E-01 | Cumene |
| 15 | 2.30E+03 | 0.10 | 8.33E+01 | 9.72E-03 | 4.00E+02 | 1.33E+93 | 4.71E-04 | 1.08E+00 | NA | 1.4E-01 | n-Propylbenzene |
| 15 | 2.80E+03 | 0.10 | 8.33E+01 | 1.41E-02 | 4.00E+02 | 2.14E+64 | 5.42E-04 | 1.52E+00 | NA | 1.0E-01 | o-Xylene |
| 15 | 3.40E+04 | 0.10 | 8.33E+01 | 1.16E-02 | 4.00E+02 | 5.42E+77 | 5.06E-04 | 1.72E+01 | 2.6E-07 | 6.0E-01 | Tetrachloroethylene |
| 15 | 4.75E+04 | 0.10 | 8.33E+01 | 1.41E-02 | 4.00E+02 | 2.15E+64 | 5.42E-04 | 2.57E+01 | NA | 4.0E-01 | Toluene |
| 15 | 6.10E+02 | 0.10 | 8.33E+01 | 1.28E-02 | 4.00E+02 | 7.02E+70 | 5.24E-04 | 3.20E-01 | 4.1E-06 | 4.0E-02 | Trichloroethylene |

END

APPENDIX C Work Plan Updates

Technical Memorandum

UPDATE TO RISK ASSESSMENT WORK PLAN, UPLAND OPERABLE UNIT

Bradford Island Cascade Locks, Oregon

March 6, 2014

Prepared for:



U.S. Army Corps of Engineers Portland District 333 S.W. First Avenue Portland, Oregon 97208-2946

Prepared by:

111 S.W. Columbia, Suite 1500 Portland, Oregon 97201-5850 25697026

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ABBREVIATIONS

| % | percent |
|----------------------|--|
| Ah-R | aryl hydrocarbon |
| AOPC | area of potential concern |
| ATSDR | Agency for Toxic Substances and Disease Registry |
| BAF | bioaccumulation factor |
| BHHRA | baseline human health risk assessment |
| BERA | baseline ecological risk assessment |
| bgs | below ground surface |
| CERCLA | Comprehensive Environmental Response, Compensation and Liability Act |
| cfs | cubic feet per second |
| COI | contaminant of interest |
| COPC | contaminant of potential concern |
| cPAH | carcinogenic polycyclic aromatic hydrocarbon |
| CPEC | contaminant of potential ecological concern |
| CSM | conceptual site model |
| CTE | central tendency exposure |
| CVOC | chlorinated volatile organic compound |
| DEQ | (Oregon) Department of Environmental Quality |
| DQO | data quality objective |
| DRO | diesel range organics |
| EcoSSLs | Ecological Soil Screening Levels |
| ECSI | Environmental Cleanup Site Information |
| EPA | Environmental Protection Agency |
| EPC | exposure point concentration |
| FS | feasibility study |
| GRO | gasoline range organics |
| HEAST | Health Effects Assessment Summary Tables |
| HI | hazard index |
| HMSA | hazardous material storage area |
| HPAH | high molecular weight polycyclic aromatic hydrocarbon |
| HQ | hazard quotient |
| IRIS | Integrated Risk Information System |
| JEM | Johnson and Ettinger (Vapor Intrusion) Model |
| kg | kilogram(s) |
| L | liter(s) |
| LOAEL | lowest-observable-adverse-effect-level |
| LOEC | lowest-observable-effect-concentration |
| log K _{ow} | octanol-water partition coefficient |
| LPAH | low molecular weight polycyclic aromatic hydrocarbon |
| μg m ³ | microgram(s) |
| | cubic meter milligram(g) |
| mg MD | milligram(s) Managament Plan |
| MP NOAEI | Management Plan |
| NOAEL | no-observable-adverse-effect-level |
| NOEC | no-observable-effect-concentration |

| OU | operable unit |
|-------|---|
| РАН | polycyclic aromatic hydrocarbon |
| PCB | polychlorinated biphenyl |
| PM | project manager |
| PPRTV | provisional peer-reviewed toxicity value |
| RA | Risk Assessment |
| RAGS | Risk Assessment Guidance for Superfund |
| RAIS | Risk Assessment Information System |
| RBC | Risk-Based Concentrations |
| RI | remedial investigation |
| RI/FS | remedial investigation/ feasibility study |
| RM | river mile |
| RME | Reasonable Maximum Exposure |
| RRO | residual range organics |
| RSL | Regional Screening Levels |
| SLV | screening level value |
| SVOC | semivolatile organic compound |
| TAG | technical advisory group |
| TCE | trichloroethene |
| TEF | toxicity equivalence factors |
| TPH | total petroleum hydrocarbons |
| TRV | Toxicity Reference Value |
| UCL | upper confidence limit |
| URS | URS Corporation |
| USACE | United States Army Corps of Engineers |
| USEPA | United States Environmental Protection Agency |
| VCP | Voluntary Cleanup Program |
| VDEQ | Virginia Department of Environmental Quality |
| VOC | volatile organic compound |
| WP | Work Plan |

1.0 INTRODUCTION

This report documents proposed updates to methodologies to conduct risk assessments at Bradford Island, Bonneville Dam. The Portland District of the United States Army Corps of Engineers (USACE) has characterized and evaluated the contamination arising from historical practices at Bradford Island in Oregon. Bradford Island is part of the Bonneville Dam complex, which is located on the Columbia River at river mile (RM) 146.1, approximately 40 miles east of Portland, Oregon (Figure 1-1). The site is a multipurpose facility that consists of the First and Second Powerhouses, the old and new navigation locks, and a spillway with a capacity of 1.6 million cubic feet per second (cfs) (USACE 2000). Figure 1-2 shows features of the Bonneville Dam complex.

Site investigations on Bradford Island began with the Landfill. The Landfill was used from the early 1940s until the early 1980s. The USACE informed the U.S. Environmental Protection Agency (USEPA) and the Oregon Department of Environmental Quality (DEQ) of the presence of the Landfill in 1996. The Landfill was added to the DEQ Environmental Cleanup Site Information (ECSI) database in April 1997, and the Bonneville Dam Project Manager (PM) signed a DEQ Voluntary Cleanup Agreement letter for the Landfill in February 18, 1998 under the DEQ Voluntary Cleanup Program (VCP). In 2014, USACE will continue investigation of Bradford Island under the Comprehensive Environmental Response, Compensation and Liability Act (CERCLA). The USACE is currently working with the DEQ to address the state's concerns regarding this investigation and any associated cleanup.

Numerous investigations have been performed by the USACE and their contractors since 1997, focusing on two Operable Units (OUs), the Upland OU, and the River OU (Figure 1-3). A review of site records for the Upland OU, including employee interviews, site environmental audits, and environmental investigations resulted in the identification of four areas of potential concern (AOPCs): the Landfill, Sandblast Area, Pistol Range, and the Bulb Slope (Figure 1-4). The primary contaminants of interest (COIs) that have been identified in soil and/or groundwater in the four AOPCs include certain metals; polychlorinated biphenyl (PCBs); semivolatile organic compounds (SVOCs), including polycyclic aromatic hydrocarbon (PAHs); butyltins; volatile organic compounds (VOCs); and a few pesticides and herbicides.

In 2007, the USACE submitted a Remedial Investigation/Feasibility Study (RI/FS) Management Plan (MP) (URS 2007), which defined the objectives of the remedial investigation (RI) and described the work to be performed to meet the project objectives. Using the Data Quality Objectives (DQO) approach (USEPA 2006), the RI/FS MP identified data gaps and described plans for extensive data collection to fill the identified data gaps for the site soils and groundwater (Upland OU) and for the offshore sediments, surface water, and tissues of various aquatic species (River OU). The RI/FS MP also described how the collected data would be used to delineate the nature and extent of contamination, evaluate the potential risks to human and ecological receptors, and support decision-making. The USACE and the external stakeholders for the project, which are collectively referred to as the Technical Advisory Group (TAG) and include the DEQ, conducted extensive internal and external review of the RI/FS MP, and the document was finalized in September 2007.

The Final RI report (URS 2012) documented the investigation, identified source areas at Bradford Island, defined the nature and extent of the environmental contamination, and identified the contaminants of potential concern (COPCs) for human health and contaminants of potential



ecological concern (CPECs) in the media from the two OUs. Based on the screening level risk assessments (RAs) completed for each of the AOPCs, the RI report made the following recommendations:

Landfill AOPC

• Perform a site-specific baseline human health risk assessment (BHHRA) and a baseline ecological risk assessment (BERA) to determine if risks to human and ecological receptors are unacceptable.

Sandblast Area AOPC

- Perform a site-specific BHHRA and a BERA to determine if risks to human and ecological receptors are unacceptable.
- Evaluate the feasibility of a using a vapor extraction system or other remedial techniques to achieve acceptable soil gas VOC concentrations.

Pistol Range AOPC

• Perform a site-specific BERA to determine if exposure to lead in soil poses unacceptable risk to ecological receptors.

Bulb Slope AOPC

• Perform a site-specific BERA to determine if exposure to lead and mercury in soil poses unacceptable risk to ecological receptors.

The BHHRA and BERA will build on the data and findings of the Final RI (URS 2012). Only upland exposure pathways will be addressed in the Upland OU baseline RAs. The Upland OU to River OU pathways (i.e., potential mass wasting, soil erosion, and groundwater discharge) will not be addressed in the Upland OU baseline RAs. These possible pathways will be considered in the Upland Feasibility Study (FS), the River FS, or a Separate Source Control Document.

1.1 Objectives and Organization

This RA Work Plan (WP) Update Technical Memorandum provides updates and changes to the methodologies for conducting the BHHRAs and BERAs at the Upland OU since the RI/FS MP (URS 2007). Most of the updates to the RA procedures are driven by the following:

- USACE's adopting the USEPA/CERCLA approach for the baseline RAs.
- DEQ comments on the Bradford Island RI (Appendix P of URS 2012) that are to be addressed in the baseline RAs.
- Regulatory (USEPA and DEQ) updates to guidances.
- Other technical updates.

1.2 Project Schedule

The following table presents the expected current and potential future Upland OU project milestones with expected completion dates.



| Project Milestones | Estimated Date |
|--|-----------------------|
| Data Evaluation Technical Meeting | March 2014 |
| Technical Advisory Group (TAG) Meeting | March-April 2014 |
| Draft Upland OU Baseline RAs | May 2014 |
| TAG Meeting | May-June 2014 |
| TAG Meeting | July 2014 |
| Final Upland OU Baseline RAs | July 2014 |
| Draft Upland FS for Review | October-November 2014 |
| TAG Meeting | November 2014 |
| Final FS Published | January-February 2015 |
| Upland OU Record of Decision | To Be Determined |



2.0 BASELINE HUMAN HEALTH RISK ASSESSMENT APPROACH

This section presents the updates to the BHHRA methodology that will be performed for the Upland OU. Most of the BHHRA process is described in the RI/FS MP (URS 2007), with only subsequent procedural updates described herein. The two main reasons for adjusting the approach for the BHHRA described in the RI/FS MP are to:

- 1. Reflect updates to HHRA guidance documents, and
- 2. Incorporate responses to DEQ's comments on the Final RI in which the USACE agreed to certain elements of the approach to the BHHRA (see the responses to DEQ's comments presented in Appendix P of the Final RI, URS 2012).
- 3. One additional potential receptor who was identified after the completion of the RI report has been added to the BHHRA.

2.1 Outstanding DEQ Comments to Be Addressed in BHHRA

Most of the comments related to the HHRA in the draft RI Report were addressed in the Final RI Report. There were two comments from DEQ that significantly affected the HHRA. One was DEQ's general comment regarding data management (e.g., censoring limits, inclusion of historic data). This comment was accepted and addressed in the Final RI Report and requires no further effort in the BHHRA.

The second major comment (Specific Comment #12) related to the process to identify the COPCs to be included in the BHHRA. In response to DEQ's comments, the COPC selection process was limited only to the mechanical process outlined in DEQ's HHRA guidance (DEQ 2010a), and nuances related to site-specific considerations were omitted. This resulted in a slightly longer list of COPCs for each environmental medium at the Landfill AOPC and the Sandblast AOPC (Tables 11-1 and 11-2 in the Final RI, based on Tables M-12 and M-20).

Addressing the two types of comments above did not change the recommendation to conduct a BHHRA for the Landfill and Sandblast AOPCs.

Addressing Specific Comment #12—Inclusion of PAHs as COPCs in groundwater for the hypothetical potable use exposure pathway at the Sandblast Area AOPC—was deferred to the BHHRA. There is low confidence in the PAH data since they were collected from direct-push groundwater samples. The DEQ has since agreed that the potable use pathway is incomplete for Bradford Island and does not need to be included in the BHHRA (e-mail communication from Bob Schwarz, January 16, 2014). Therefore, this comment is no longer relevant to the BHHRA.

The screening Assessment in the RI identified COPCs in Upland OU groundwater potentially discharging to the River OU, and COPCs in Upland OU soil that could migrate to the River OU through soil erosion and/or mass wasting were identified as warranting additional evaluation for human receptors in the River OU. Further assessment of these migration pathways will be conducted in the Upland FS, River FS, or a separate Source Control Document.

2.2 Regulatory Updates

The RI/FS MP (URS 2007) detailed the proposed approach to the BHHRA (Appendix B of the RI/FS MP). Since then, several regulatory updates that are relevant to the HHRA have been



published. Since USACE has elected to follow a CERCLA model for the execution of the BHHRA but Oregon DEQ remains the reviewing agency, updates from both agencies need to be considered with primary consideration given to USEPA guidance.

Relevant updates from USEPA and Oregon DEQ which may require some modifications to the approach for the BHHRA include the following:

<u>USEPA</u>

- Regional Screening Levels (USEPA 2013a)
- Updated Toxicity Values (USEPA IRIS and other sources, current)
- Final Guidance for Assessing and Mitigating the Vapor Intrusion Pathway from Subsurface Sources to Indoor Air External Review Draft. (USEPA 2013b)
- EPA National Risk Management Research Laboratory Engineering Issue: Indoor Air Vapor Intrusion Mitigation Approaches (USEPA 2008)

DEQ

- Human Health Risk Assessment Guidance (DEQ 2010a)
- Guidance for Assessing and Remediating Vapor Intrusion in Buildings (DEQ 2010b)

Modifications to the BHHRA methodology based on these updates are included in the section below.

2.3 Summary of BHHRA Approach

The approach for the BHHRA will follow the methodology presented in Appendix B of the RI/FS MP (URS 2007) with updates and changes discussed in this section.

2.3.1 AOPCs, Media, and COPCs

The Final RI (URS 2012) recommended the Landfill and Sandblast Area AOPCs for further evaluation in a BHHRA. The contaminated media are soil and groundwater with soil gas being an issue for the buildings on the Sandblast Area. The media-specific chemical groups of COPCs are as follows:

Media and COPCs for Landfill AOPC

- Soil: Metals, PAHs, Chlorinated VOCs (CVOCs)
- Groundwater (incidental exposure in trench setting for construction workers only): Metals, phthalates

Media and COPCs for Sandblast Area AOPC

- Soil: Metals, PAHs, chlorinated VOCs (CVOCs)
- Groundwater (incidental exposure in trench setting for Construction Workers only): Metals
- Soil Gas: CVOCs



Tables 2-1 and 2-2 list the individual COPCs recommended for further evaluation in the BHHRA by receptors and media for the Landfill and Sandblast Area AOPCs, respectively.

2.3.2 Receptors and Pathways

The eastern portion of Bradford Island continues to be a restricted access area that is managed by USACE. Due to the industrial nature of land use at the site, residential receptors are absent. Onsite groundwater is not used for any purpose at Bradford Island. Along the shore of the Bradford Island, tribal fishermen may fish from fishing platforms erected along the shoreline.

Activities that are typically carried out by employees on the eastern side of the Island include grounds maintenance, vegetation clearing, painting, building maintenance and administrative activities. The grounds crew typically works up to three days per week during peak season. The building crews typically work up to 40 hours per week. The workers are supplied with drinking water. Construction activities (less than one-year duration) and short-term excavation involving utility repair or other types of soil-disturbing activities in a trench setting may be present.

The conceptual site models (CSMs) depicted in Figures 2-1 and 2-2 present potential exposures of humans to various media (i.e., soil gas, soil, and groundwater). Potential exposures by human receptors who may be exposed to COPCs at the upland AOPCs are described below.

Landfill AOPC

The landfill continues to be a fully vegetated area with no bare soils. Vegetation clearing occurs along a 3-4 foot-wide strip along roadsides when needed. There are no built or occupied structures at the landfill or in its vicinity. There are no plans at this time for construction of enclosed structures at this AOPC nor is it considered feasible from an engineering standpoint. Therefore, vapor intrusion pathways for indoor inhalation are incomplete and will remain incomplete.

Receptors who may be exposed to COPCs in soil and groundwater (incidental exposure in trench setting only) at the Landfill AOPC and will be evaluated in the BHHRA include the following:

- Adult outdoor site maintenance worker engaged in activities that do not involve a significant degree of soil disturbance (e.g., landscape workers). These receptors at Bradford Island may be exposed to COPCs in surface soil (0-3 feet below the ground surface [bgs]) by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil.
- Construction workers may be exposed to COPCs in surface and subsurface soil (0-10 feet bgs) by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil. Occasionally, construction workers and short-term excavation workers also may be exposed to COPCs in shallow groundwater by incidental ingestion and dermal contact, if they undertake activities at depths that may bring them in contact with shallow groundwater.



Sandblast AOPC

The Sandblast AOPC is a mixture of vegetated and unvegetated areas. The former Sandblast building has been demolished and the former Hazardous Material Storage Area (HMSA) exists only as a lean-to area that is open to the outdoors. The two remaining buildings in the vicinity are the Service Building and the Equipment Building and both are in active occupational use. The Service Building provides work space for approximately 27 people and may support additional full-time employees in the future. The Equipment Building includes office space for two people.

- Adult outdoor site maintenance worker engaged in activities that do not involve a significant degree of soil disturbance (e.g., landscape workers). These receptors at Bradford Island may be exposed to COPCs in surface soil (0-3 feet bgs) by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil.
- Construction workers and short-term excavation workers may be exposed to COPCs in surface and subsurface soil (0-10 feet bgs) by incidental ingestion, inhalation (dusts and vapors), or dermal uptake of contaminants from soil. Occasionally, they also may be exposed to COPCs in shallow groundwater by incidental ingestion and dermal contact.
- Adult indoor workers may be exposed to VOCs emanating from soil gas and entering the indoor environment by vapor intrusion.

Fishing from Shore Platforms

Fishing on platforms used by tribal fishermen and their children is evaluated within the Upland OU. Tribal fishers and their children may utilize fishing platforms along the shoreline of the Island to fish for species that may be consumed or sold. The fishing platforms are typically accessed by boat and the actual fishing activity may include overnight stays at the platforms or in the Upland OU portion of Island itself.

Each of the four Upland AOPCs (Landfill, Sandblast Area, Pistol Range, Bub Slope) will be considered. Exposure pathways will include adult and child exposures to shallow soil (0-3 feet bgs) at each AOPC, for incidental ingestion, dermal contact, and inhalation of dusts and vapors. A screening step will be followed by forward risk assessment only for COIs that exceed the screening level values (SLVs) for the fishing platform exposure scenarios.

USACE has provided some general information regarding the nature of the receptor activity. The key assumptions used to characterize these receptors are as follows:

- There are two types of users of the fishing platforms:
 - 1) Tribal fishers who are allowed to sell their catch and use platforms for only 2 months a year. They may be there in excess of 24 hours during this period.
 - 2) Sustenance fishers who cannot sell their catch but are allowed to fish any time throughout the year.
- The tribal fishermen target salmon in the salmon migration path along the south shore of the dam complex.



- For purposes of the risk assessment of fishing platforms only, it is assumed that fishing platforms could be constructed and fishers should occupy any area of Bradford Island, including any of the four AOPCs. The tribal fishermen and their children will arrive and depart by small boat.
- Only direct exposure to contaminants in soil or sediment are evaluated (i.e., subsistence or fish consumption is not evaluated since that will be included in the River OU assessment).

To further refine assessment of the fishing platform exposure, an additional screening step will be conducted in a two-level process to assess possible exposures by commercial or sustenance fishers on Bradford Island. In Level 1 of the screening, the soil data for COIs at each AOPC will be compared to the DEQ Residential Risk-Based Concentrations (RBCs) and USEPA's Residential RSLs. Those AOPCs where all COIs are lower than the screening values will be dropped from further evaluation. AOPCs where there are exceedences will be carried to a second screening level. For the Pistol Range AOPC, sediment data and soil data will be grouped separately.

In the Level 2 screening, residential RBCs and RSLs will be multiplied by a factor that represents an exposure duration as a fraction of default residential exposure duration. For example, it may be assumed conservatively that an adult or child will spend several days a week for 8 months of the year for 30 years at the Upland AOPCs.

If any COIs from any AOPC fail the residential screening in Level 1, they will be compared to the Platform Fishing SLVs developed in Level 2. If all values are below the Level 1 and/or Level 2 SLVs, there will be no need for further evaluation and the screening exercise will be documented in the Baseline HHRA.

If any COIs exceed the fishing platform SLVs, they will be recommended for site-specific forward risk assessment calculations, similar to the other receptors already included in the BHHRA.

Minor Pathways

Since the DEQ RBCs only consider dermal contact in their groundwater exposure in a trench setting, potential inhalation of VOCs in a trench setting may be a complete but minor pathway for the construction worker at the Landfill and Sandblast Area AOPCs. The uncertainty assessment will include an evaluation of inhalation of VOCs from groundwater in a trench setting even though they were not identified as dermal COPCs in groundwater for the trench setting (Section 2.3.1).

2.3.3 Exposure Assessment

Quantifying exposure involves estimating chemical intake rates based on the evaluation of chemical releases from the site and estimation of exposure point concentrations (EPCs) for specific pathways.

The methods for calculating potential chemical intakes from soil, groundwater, and soil gas (vapor intrusion into indoor air) for the populations and exposure pathways selected for quantitative evaluation will primarily follow USEPA guidance and will also consider Oregon DEQ guidance, as appropriate. Exposure factor values are drawn from the Risk Assessment



Guidance for Superfund (RAGS) Part A (USEPA 1989) and all succeeding guidance documents as listed in Appendix B of the RI/FS MP (URS 2007) and in Section 2.2 above. In addition, Oregon DEQ's HHRA guidance, Vapor Intrusion guidance and current tables for Calculating Risk-Based Concentrations for Individual Chemicals were also consulted (DEQ 2010a,b, 2012).

There is generally close agreement between USEPA and DEQ for the exposure factor values. Where exposure factor values are not available from USEPA, DEQ's recommended values will be used.

The EPC is a chemical-specific and media-specific value that represents a central tendency exposure (CTE) or Reasonable Maximum Exposure (RME) estimate of the concentration to which a receptor is exposed. In accordance with USEPA guidance (USEPA 1989), and to address comments received on the Final RI (URS 2012), both RME and CTE estimates will be used in the BHHRA. As described in Appendix A of the Remedial Investigation Management Plan (URS 2007) and in accordance with the most recent USEPA guidance regarding statistical methodology to be used in EPC estimation (USEPA 2002), the 95th percentile (%) upper confidence limit (UCL) on the mean of values in a medium (95% UCL) (soil, groundwater for trench setting, soil gas) will be used as the EPC representing the RME. Where samples sizes are less than eight, the maximum or single location data may be used, as appropriate. As previously agreed, the 95% UCL is also acceptable to DEQ and is more conservative than DEQ's suggested 90% UCL. The EPC representing the CTE scenario will be based on the mean value. The RME and CTE EPCs will represent a single AOPC and will be estimated using statistical methods and values recommended by USEPA's ProUCL software (USEPA 2013c). EPCs for soil will be based on 0- to 3- foot bgs depth interval for outdoor workers, and 0- to 10-foot-bgs depth interval for construction workers and excavation workers.

Table 2-3 lists the exposure factors that will be used in the BHHRA for each of the receptors noted above. The selection of RME and CTE values generally exceeds the likely exposures of on-island personnel with respect to exposure duration, exposure frequency, and other uptake parameters. For example, although there are both part-time and full-time employees at Bradford Island, it is doubtful that they will have an exposure duration of 25 years. If warranted, the proposed exposure factors may be modified as part of the uncertainty assessment to provide context to the findings of the BHHRA.

Soil - Outdoor Maintenance Worker, Construction Worker

Soil-related pathways for the outdoor worker (0-3 feet bgs) and construction worker (0-10 feet bgs) are common to the Landfill and Sandblast Area AOPCs. The equations for estimating direct contact with soil are presented in the RI/FS MP and are not repeated here.

Soil Gas - Indoor Office Worker

The vapor intrusion pathway for the indoor office worker at the Sandblast Area AOPC will be evaluated by using soil gas data. USEPA's current versions of the Johnson and Ettinger Model (JEM) for soil gas data may be used to estimate site-specific risks for the vapor intrusion pathway (USEPA 2004a,b, 2013c). The current two buildings at the Sandblast Area AOPC and a hypothetical future building may be modeled using JEM. A combination of USEPA default and site-specific values (i.e., dimensions of existing inhabited buildings and available site specific soil parameters) will be used to represent building and soil properties. It is recognized that



DEQ's Vapor Intrusion guidance (DEQ 2010) de-emphasizes the use of site-specific modeling. The uncertainty assessment will include a discussion of the soil, groundwater and soil gas data in comparison to DEQ RBCs, the magnitude of exceedances, and other lines of evidence such as spatial distributions and proximity to source areas.

Groundwater – Construction Worker (Incidental exposures only)

There is no consumption of groundwater on the Island. Occasionally, construction workers and excavation workers may be exposed to groundwater within an excavation due to the occurrence of groundwater at depths as shallow as 5 feet bgs (as noted in the Final RI Appendix B logs and D tables [URS 2012]).

At the Landfill AOPC, construction workers in a trench setting may come into dermal contact with metals and phthalate COPCs in groundwater and may experience some minimal incidental ingestion of groundwater. These pathways will be evaluated by assuming that a portion of the overall exposure duration of construction workers (e.g., 25 of 250 days) may be spent in a trench setting where incidental contact with groundwater may occur.

At the Sandblast Area AOPC, the risk assessment will evaluate incidental contact with groundwater within a trench scenario for the construction worker.

As noted in Section 2.3.2, the DEQ RBCs only consider dermal contact in their groundwater exposure in a trench setting. The uncertainty assessment for the construction worker at the Landfill and Sandblast AOPCs will include an evaluation of inhalation of VOCs from groundwater in a trench setting even though they were not identified as dermal COPCs in groundwater for the trench setting. The Virginia Department of Environmental Quality (VDEQ 2014) trench model may be used to model trench air concentrations for VOCs in groundwater.

2.3.4 Toxicity Values

The selection of toxicity values will follow the hierarchy of sources that is currently recommended by USEPA (2013a), as follows:

- Integrated Risk Information System (IRIS) (USEPA)
- Provisional Peer-Reviewed Toxicity Values (PPRTVs) (USEPA)
- Minimal Risk Levels (Agency for Toxic Substances Disease Registry [ATSDR])
- Chronic Reference Exposure Levels (California Environmental Protection Agency [EPA])
- Appendices to PPRTVs (USEPA)
- Health Effects Assessment Summary Tables (HEAST)

This hierarchy is generally similar to Oregon DEQ's hierarchy with some minor variations. Toxicity values for the majority of the COPCs are available from IRIS or as PPRTVs. A few selected COPCs are discussed in more detail below.

2.3.4.1 Lead

Exposure to lead in Sandblast AOPC soils will be evaluated using the most current version of the Adult Lead Model, as recommended by USEPA (2014).



2.3.4.2 Polycyclic Aromatic Hydrocarbons

cPAHs will be evaluated by the use of Toxicity Equivalence Factors (TEFs) relative to benzo(a)pyrene, as listed below from USEPA (2013a).

| Compound | TEF |
|-------------------------|-------|
| Benzo(a)pyrene | 1.0 |
| Benz(a)anthracene | 0.1 |
| Benzo(b)fluoranthene | 0.1 |
| Benzo(k)fluoranthene | 0.01 |
| Chrysene | 0.001 |
| Dibenz(a,h)anthracene | 1.0 |
| Indeno(1,2,3-c,d)pyrene | 0.1 |

2.3.4.3 Trichloroethylene (TCE)

Toxicity values for TCE will be based on USEPA's current recommendations as listed in IRIS and USEPA (2013a). These include the following:

- Oral Slope Factor 4.6E-02 (milligrams per kilogram [mg/kg]-day)-1
- Inhalation Unit Risk 4.1E-06 (micrograms per cubic meter $[\mu g/m^3]$)-1
- Oral Reference Dose 5.0E-04 mg/kg/day
- Inhalation Reference Concentration 2.0E-03 mg/m³

2.3.5 Risk Characterization

As noted in in the RI/FS MP (URS 2007), both excess lifetime cancer risk and non-cancer hazard will be estimated for all carcinogenic COPCs, for each combination of receptor, exposure medium and exposure scenario. For non-carcinogenic chemicals, only non-cancer hazards will be estimated. The results will be summed to provide quantitative estimates of multi-pathway and multi-media risks and hazards for each receptor. If initial hazard indices are greater than 1, then segregation of hazard quotients by target organ or system specific analysis will also be performed. The estimated risks, hazard quotients, and hazard indices will be presented in the context of USEPA's acceptable risk range of 1E-06 to 1E-04 for cancer and hazard quotient or hazard index of 1.0 for noncancer COPCs (USEPA 1991).

An Uncertainty Assessment section describing the sources of uncertainty in the BHHRA and their potential effect on the underestimation or overestimation of risk and hazard will be included.



3.0 BASELINE ECOLOGICAL RISK ASSESSMENT APPROACH

This section presents the BERA methodology that will be implemented for the Upland OU. Most of the BERA process is described in the RI/FS MP (URS 2007), with only subsequent procedural updates described herein. The two main reasons for adjusting the approach for the BERA described in the RI/FS MP are to:

- 1. Reflect updates to ERA guidance documents, and
- 2. Incorporate responses to DEQ's comments on the Final RI (URS 2012) in which the USACE agreed to certain elements of the approach going forward with the BERA (see the responses to DEQ's comments presented in Appendix P of the Final RI, URS 2012).

The only relevant guidance update is the latest USEPA Ecological Soil Screening Levels (EcoSSLs) and associated methodology (USEPA 2005-2008). As described in the following subsections, USEPA's EcoSSL document now has priority in the hierarchy of toxicity reference values (TRVs) and bioaccumulation factors (BAFs) that will be used to assess risk.

Two responses to DEQ's comments on the Final RI require incorporation into the BERA: responses to Specific Comments 22 (add American mink, *Neovison vison*, as an Upland receptor) and 42 (consider Region 5 TRVs for birds and mammals). Both of these responses have been implemented into the updated approach for the BERA, as presented below.

3.1 AOPCs, Media, and CPECs

All four AOPCs in the Upland OU are retained for evaluation in the BERA. Only soil was identified as a medium of concern for terrestrial ecological receptors. Table 3-1 provides the list of CPECs that will be evaluated for each AOPC in the BERA. These CPECs include metals, total high-molecular-weight PAHs (HPAHs), tributyltin, organochlorine pesticides, VOCs, and SVOCs for the Landfill and Sandblast Area AOPCs, lead for the Pistol Range AOPC, and lead and mercury for the Bulb Slope AOPC. Risk estimates will be calculated for each CPEC for all receptors potential present at a given AOPC. For the receptors with large home ranges that will be assumed to forage in the combined four AOPCs (discussed in Section 3.3.1), all CPECs shown in Table 3-1 will be evaluated for this "combined AOPC" exposure unit.

3.2 Receptors and Exposure Pathways

The terrestrial receptors and exposure pathways identified in the RI/FS MP (URS 2007) will be evaluated in the BERA, and the American mink will be added to represent a hypothetical upland large mammal, as requested by the DEQ (Specific Comments 22 on the Final RI). The assessment and measurement endpoints originally presented in the RI/FS MP (URS 2007) are shown in Table 3-2, and an assessment endpoint associated with the mink was added.

The following list of receptors and exposure pathways will be included in the Upland BERA:

- Terrestrial plants and soil invertebrates exposed through direct contact with surface and shallow soil.
- Vagrant shrew (*Sorex vagrans*) exposed through incidental ingestion of surface and shallow soil, prey (100% soil invertebrates), and water.



- American mink exposed through incidental ingestion of surface soil, Upland prey (15% small mammals, and water.
- American robin (*Turdus migratorius*) exposed through incidental ingestion of surface soil, prey (100% soil invertebrates), and water.
- Canada goose (*Branta canadensis*) exposed through incidental ingestion of surface soil, prey (100% plants), and water.
- American kestrel (*Falco sparverius*) exposed through incidental ingestion of surface soil, prey (100% small mammals), and water.

Based on the available data, surface soil is defined as 0 to 1 foot bgs for all AOPCs with the exception of the Pistol Range, for which surface soil is defined as 0 to 1.5 feet bgs. Shallow soil is defined as 0 to 3 feet bgs, and this depth interval only applies to the Landfill and Sandblast Area AOPCs.

As described in detail in the response to DEQ's Specific Comment #22 on the Final RI (URS 2012), the mink was selected to represent large mammals in the Upland OU due to the absence of mammals on the island with the exception of rodents and feral cats. Mink are also known to be sensitive to environmental contaminants. The minimal amount of available habitat (approximately 12 acres) makes the island unsuitable for supporting viable populations of wildlife species with larger home ranges (URS 2002). Mink are present in the area and could feasibly access the island and forage there on rodents. Several sources of information on the mink's diet were consulted while developing the response to DEQ's comment, and the consensus is that mink's diet primarily consists of crayfish, fish, and other aquatic-related prey. Typically, 10% or less of their diet is comprised of terrestrial prey (e.g., birds and small mammals) (USEPA 1993, 1995).

To be conservative, a dietary composition of 15% small mammals from the Upland OU will be assumed as a first step in the evaluation for the mink, even though a lower proportion may actually be more realistic for site conditions (i.e., permanent water source; riverine habitat). The remaining 85% of the mink's diet is assumed to be comprised of prey from the River OU (i.e., fish and crayfish), which will be evaluated in the BERA for the River OU.

For the water ingestion pathway, highly mobile wildlife receptors (i.e., kestrel and mink) will be assumed to ingest water from the river, while less mobile receptors (i.e., robin and shrew) and the Canada goose will be assumed to ingest water that puddles in the Upland OU (via a simple equilibrium partitioning calculation).

3.3 Exposure Assessment

The foodweb model for the Upland OU and the ecological CSMs for each AOPC are provided on Figures 3-1 through 3-5, which were most recently presented in the Final RI (URS 2012). Minor updates to the foodweb model and CSMs presented in the Final RI (URS 2012) were made for purposes of the Upland OU BERA. These figures illustrate the potentially complete and significant ecological exposure pathways for the Upland OU, including those associated with transport to the River OU. In the Screening Assessment that was conducted in the RI, CPECs in Upland OU groundwater potentially discharging to the River OU did not pose a risk to aquatic receptors. CPECs in Upland OU soil that could migrate to the River OU through soil erosion and/or mass wasting did pose an unacceptable risk to aquatic receptors. Further assessment of



this migration pathway will be conducted in the Upland FS, River FS, or a Separate Source Control Document.

3.3.1 Exposure Factors

No updates to the life history parameters developed in the RI/FS MP (URS 2007) were needed with the exception of the dietary composition adjustment for the mink described in Section 3.2 and the home range for the mink discussed below. Table 3-3 presents the life history parameters for each terrestrial receptor that will be evaluated in the Upland BERA.

For the animal receptors with small home ranges (i.e., shrew and robin), it is feasible for an individual to forage within one specific AOPC. Although the Canada goose has a potentially large home range (Table 3-3), part of Bradford Island is managed as goose habitat, and therefore this receptor could spend most of its time on the island. For the animal receptors with large home ranges (i.e., kestrel and mink), it was assumed that these receptors could forage at each individual AOPC (albeit unlikely) and within all four AOPCs combined. Receptor-specific area use factors will be calculated in the BERA as the AOPC (and/or combined AOPC) site size divided by the size of the home range.

The home range size of 1.85 km for the mink that was originally developed in the RI/FS MP (URS 2007) for the River OU was developed based on the idea that this receptor would be foraging along the river banks on and near Bradford Island. The shape of mink home ranges depends on the type of habitat available. In riverine environments, home ranges are linear, whereas those in marsh habitats tend to be more circular (USEPA 1993). In addition, home ranges for adult males are generally larger than adult female home ranges, especially during the mating season. Because the purpose of selecting the mink as a receptor in the Upland OU is to address the portion of the daily dose attributed to Upland prey, the smallest home range for an adult female in riverine habitat expressed as an area (7.8 hectare, or approximately19 acres) was selected for use in the Upland BERA (USEPA 1993). This home range will be used to calculate AOPC-specific area use factors for the mink.

3.3.2 Exposure Point Concentrations

The 95% UCLs calculated for soil in the RI will be used as the EPCs for all terrestrial receptors. In the Screening ERA presented in the Final RI Report (URS 2012), maximum detected concentrations were used for terrestrial plants and invertebrates, and the 95% UCLs were used for birds and mammals. Given that the objective of the BERA is to provide more realistic, site-specific estimates of risk, the 95% UCL will be used for plants and invertebrates as well as birds and mammals. This is due to the lack of protected plant and invertebrate species in the Upland OU and the goal of estimating risk to these receptor groups at the community level.

3.3.3 Bioaccumulation Factors

As described in the RI/FS MP (URS 2007), a combination of regression-derived BAFs, median BAFs, and octanol-water partition coefficient ($logK_{ow}$) based BAFs from the literature will be used to predict tissue concentrations in the BERA. The regression-based approach is typically preferred because it provides a more site-specific prediction of a CPEC concentration in a certain dietary tissue, as it incorporates the site EPC.



The following two sources of BAFs are the primary documents that will be consulted for the BERA:

- USEPA's Eco-SSLs guidance document, Attachment 4-1 (2005a, last updated 2007)
- Oak Ridge National Laboratory's Risk Assessment Information System (RAIS) database

Other sources in the literature may also be reviewed in the absence of BAFs in the Eco-SSLs guidance or, secondarily, the RAIS database. In the absence of empirically-derived BAFs, equations to estimate BAFs for organic chemicals are typically based on the logK_{ow} of the chemical and the organic carbon content of the soil. For soil invertebrates, estimated lipid content (e.g., earthworms) is also incorporated into the BAF development.

3.4 Toxicity Values

Screening levels, or SLVs, are expressed as concentrations in media (i.e., mg of chemical/kg of soil). Although "screening levels" are typically associated with exposure via direct contact, and are commonly referred to as direct toxicity benchmarks, limited sources of generic media-based screening levels address both direct contact and dietary exposure for birds and mammals. In contrast, diet-based TRVs protective of birds and mammals are expressed as a daily dose normalized to body weight (mg of chemical/kg of body weight/day).

Both no-observable-adverse-effect-levels (NOAELs)/no-observable-effect-concentrations (NOECs) and lowest-observable-adverse-effect-levels (LOAELs)/lowest-observable-effect-concentrations (LOECs) will be selected for each receptor group in order to develop a range of hazard quotients (HQs) for consideration by risk managers. An uncertainty factor of 5 will be used to adjust chronic LOAELs/LOECs to NOAELs/NOECs and vice-versa, when necessary. Ideally, NOAELs and LOAELs from the same dose-response studies will be selected. As noted in Section 3.0, the TRV selection process was changed based on guidance updates and DEQ comments on the Final RI (URS 2012).

3.4.1 Screening Levels for Terrestrial Plants and Soil Invertebrates

The following hierarchy of sources of terrestrial plant and soil invertebrate SLVs will be consulted for the BERA:

- USEPA's Eco-SSLs guidance document, Attachment 4-1 (2005a, last updated 2007)
- DEQ's Level II SLVs (DEQ 2001)
- Oak Ridge National Laboratory's Risk Assessment Information System (RAIS) database

Other sources of the literature may also be review in the absence of SLVs from the hierarchy listed above.

3.4.2 Toxicity Reference Values for Birds and Mammals

The following hierarchy of sources of TRVs for birds and mammals will be consulted for the BERA:

- USEPA's Eco-SSLs (USEPA 2005-2008)
- Draft Portland Harbor RI/FS (Lower Willamette Group 2011)



- Oak Ridge National Laboratory's Risk Assessment Information System (RAIS) database
- USEPA Region 5 Ecological Screening Levels (USEPA 2003)
- USEPA's IRIS database

Other sources in the literature may also be reviewed in the absence of TRVs from the hierarchy listed above. To avoid the need to consider allometric scaling factors, TRVs based on large mammal studies will be selected for the mink. This is the primary reason for reviewing the mammalian TRVs provided in the Draft Portland Harbor RI/FS report (Lower Willamette Group 2011).

To address DEQ's Specific Comment 42 on the RI, USEPA Region 5 Ecological Screening Levels will be reviewed to identify the TRVs that were used to develop these screening levels. These TRVs would be considered in the absence of TRVs from the first few sources listed in the hierarchy above.

3.5 Risk Characterization

The two types of TRVs listed in Section 3.4 will be incorporated into the analysis: one based on a NOAEL/NOEC and a second based on an observed adverse effect in a test species (LOAEL/LOEC). If the HQ based on the NOAEL TRV is less than 1, adverse effects are unlikely because of the inherent conservatism (protectiveness) built into the exposure and effects assessments. If the HQ derived using the LOAEL TRV is greater than 1, it indicates that exposure exceeds a known effect concentration for a test organism; such exposure pathways may warrant attention with respect to potential risk management measures.

If the estimated exposure exceeds the NOAEL TRV (i.e., the NOAEL TRV-based HQ is >1.0) but is less than the LOAEL TRV, (i.e., the LOAEL TRV-based HQ is <1.0), then that complete exposure pathway will be considered in greater detail before conclusions about the likelihood that a risk or hazard is present is made. For non-listed (common) terrestrial species known to be present on Bradford Island, more emphasis will be placed on CPECs that have LOAEL TRV-based HQs above 1. However, the range of HQs developed for each receptor and CPEC will be considered in the risk characterization.

Hazard indices (HIs) are ideally calculated for the appropriate CPEC groups identified as those chemicals demonstrating similar modes of toxicity, or those that affect the same target organ. HIs are estimated by calculating the summation of HQs for COI groups that meet these criteria. The implications of HQs greater than or less than 1.0 discussed above, will also be applied to HIs.

Due to a lack of data regarding additive effects associated with exposure to multiple chemicals for nonhuman receptors, professional judgment will be used in the development of HIs. For the BERA, HIs will be calculated based on the following chemical groupings in an effort to derive more meaningful HIs than by simply adding all HQs (as cited in the RI/FS MP, URS 2007):

- Inorganics (including butyltins) Adequate avian and mammalian toxicity data may be available for several inorganics detected at the site, and literature citations will support development of the HI for inorganics. If no information is available for a particular inorganic, it will automatically be included in the HI for this chemical class.
- Low-molecular-weight PAHs (LPAHs) The toxicity of PAHs is highly variable and is driven by the number of ring structures and molecular weight (Eisler 1987). LPAHs



consist of fewer than four fused benzene rings and have molecular weights less than 200. LPAHs are more soluble and bioavailable in aqueous solution than HPAHs and are associated with acute toxicological effects on biota. LPAHs are much less persistent in the environment, are not considered bioaccumulative, and are readily metabolized by birds and mammals. For these reasons, it is not likely that they will affect reproduction and development. Although they may act through different mechanisms of toxicity, the HQs for LPAHs will be added separately from the HPAHs to generate an HI.

- HPAHs These PAHs consist of four or more fused benzene rings and have a molecular weight greater than 200. HPAHs are more persistent than LPAHs, which is attributed to their higher hydrophobicity. Due to their size, HPAHs are relatively immobile and exhibit extremely low volatility and solubility (Eisler 1987). Bioaccumulation potential tends to increase as the molecular weight of PAHs increases and HPAHs are generally associated with carcinogenic effects. Chronic exposure to these cPAHs may also produce non-carcinogenic effects by destroying hematopoietic, lymphoid, and reproductive tissues (Eisler 1987). Generating separate risk estimates for HPAHs and LPAHs is further supported by the recently (2007) published USEPA Eco-SSL for PAHs (USEPA 2005-2008). The HQs for HPAHs will be added separately from the LPAHs to generate an HI.
- Organochlorine pesticides and herbicides and PCBs Persistent, lipophilic chlorinated compounds such as chlorinated pesticides, herbicides, and PCBs are associated with a variety of ecotoxicological effects. Effects include mortality and neurotoxicity at high doses. At lower doses, effects include endocrine disruption and impairment of reproductive, developmental and neurological functions. Dioxin-like PCBs can alter gene expression by the activation of the aryl hydrocarbon (Ah-R) receptor. Both dioxin-like and non-dioxin-like PCBs may also act through many other modes of toxicity, the most common of which as characterized as narcosis (a non-specific mechanism of toxicity associated with organic compounds). Chlorinated pesticides such as DDTs can interfere with the nervous system and may impair the endocrine system by exerting estrogenic effects. PCBs and chlorinated pesticides are highly persistent, have the potential to biomagnify in the food web, and are known to cause similar developmental and reproductive effects in birds and mammals. Although they may act through different mechanisms of toxicity, their HQs will be added to generate an HI for persistent and bioaccumulative chlorinated compounds.
- Non-PAH SVOCs These compounds are known to exist in soils of the Upland OU and will be included in the food web evaluation if they demonstrate a potential for bioaccumulation. Phthalates are known to be present in Upland OU soils and have the potential to bioaccumulate in terrestrial ecosystems. Phthalates tend to have low water solubility, adsorb to sediments, and dissolve easily in oils. Phthalate esters metabolize to monoesters, which deregulate cellular activity by mimicking endogenous ligands (Wexler 1998). Experiments with fish have not found extensive bioaccumulation of these compounds, although rat studies have found fetotoxcity and teratogenic effects. Increased incidences of carcinomas and adenomas observed in rats caused USEPA to label bis(2-ethylhexyl)phthalate a probable human carcinogen. The HQs for phthalates will be added to generate an HI for total phthalates. The decision to include the HQs for other SVOCs in HI summations will be evaluated on a case by case basis.



• VOCs – Those VOCs with the potential to bioaccumulate, as defined in the Final RI (URS 2012) (log Kow exceeding 3.5, with an optimum range between 3.5 and 5.5; Suter 1993) will also be considered in the evaluation of dietary exposure by wildlife. For those VOCs that are not potentially bioaccumulative, direct exposure by birds and mammals through evaluation of the incidental soil ingestion pathway. The HQs for VOCs will be added to generate an HI for this chemical class.

If the HIs generated for the chemical classes discussed above exceed 1.0, then further assessment of the appropriateness to add the HQs for individual chemicals may be conducted for the specific receptor.



4.0 SUMMARY

The BHHRA and BERA will build on the findings of the Final RI (URS 2012). Only Upland OU exposure pathways will be addressed in the baseline RAs. The Upland OU to River OU pathways (i.e., potential mass wasting and erodible soils and groundwater discharge) will not be addressed in the Upland OU baseline RAs. These pathways will be considered in the Upland FS, in the River FS, or in a Separate Source Control Document.

This Technical Memorandum documents the BHHRA and BERA methodology and highlights updates subsequent to the RI/FS MP (URS 2007). These BHHRA and BERA procedural updates were driven by the USACE adopting the USEPA/CERCLA approach for the baseline RAs, DEQ comments on the Bradford Island RI (URS 2012) that are to be addressed in the baseline RAs, regulatory (USEPA and DEQ) updates to guidances, and other technical updates.

The RI/FS MP (URS 2007) and this Technical Memorandum, with the methodology updates, will be used to conduct the Upland OU BHHRAs and BERAs.



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TABLES



Table 2-1Summary of COPCs - Landfill AOPCBradford Island -Upland Operable Unit

| Media | Receptor | COPCs |
|--|--------------------------------|---|
| Soil | Adult Outdoor Worker | Arsenic Benzo(a)anthracene Benzo(a)pyrene Benzo(b)fluoranthene Benzofluoranthenes, Total Dibenz(a,h)anthracene Indeno(1,2,3-cd)pyrene Lead [#] |
| Jon | Construction Worker | Benzo(a)pyrene Tetrachloroethene (PCE) 1,1-Dichloroethene [#] cis-1,2-Dichloroethene [#] trans-1,2-Dichloroethene [#] Trichloroethene (TCE) Vinyl Chloride |
| Groundwater (Incidental Contact in Trench Setting Only) | Construction/Excavation Worker | Antimony Iron Mercury Thallium # Zinc Di-n-octyl Phthalate |

= non-carcinogenic COPC

COPC = chemicals of potential concern

Shaded indicates COPC selection is not based on a direct exceedance of the screening level (SL) but other factors (i.e., multi-media, C/SL>0.1, degradation product, or No SL)

Table 2-2Summary of COPCs - Sandblast Area AOPCBradford Island -Upland Operable Unit

| Media | Receptor | COPCs |
|--|--------------------------------|---|
| Soil | Occupational Outdoor Worker | Arsenic Benzo(a)pyrene Benzofluoranthenes, Total Chromium Lead [#] Tetrachloroethene (PCE) 1,1-Dichloroethene [#] cis-1,2-Dichloroethene [#] Dibenz(a,h)anthracene trans-1,2-Dichloroethene [#] Trichloroethene (TCE) Vinyl Chloride |
| | Construction Worker | Benzo(a)pyrene Tetrachloroethene (PCE) 1,1-Dichloroethene [#] cis-1,2-Dichloroethene [#] trans-1,2-Dichloroethene [#] Trichloroethene (TCE) Vinyl Chloride |
| Soil Gas | Indoor Worker | Tetrachloroethene (PCE) Trichloroethene (TCE) 1,1-Dichloroethene [#] cis-1,2-Dichloroethene [#] trans-1,2-Dichloroethene [#] Vinyl Chloride |
| Groundwater (Incidental Contact in Trench Setting Only) | Construction/Excavation Worker | Vanadium # |

= non-carcinogenic COPC

COPC = chemicals of potential concern

Shaded indicates COPC selection is not based on a direct exceedance of the screening level (SL) but other factors (i.e., multi-media, C/SL>0.1, degradation product, or No SL)

Table 2-3 **Exposure Factors for Human Health** Bradford Island - Upland Operable Unit

Reasonable Maximum Exposure (RME)

| | | | Indoor Wo | rker (IW) | Outdoo | r Worker | Construction V | Worker (CW) | Excavation | n Worker |
|---|----------------------|---------------------------------|-------------|--------------------|-------------|-----------------|----------------|-----------------|---------------|-----------------|
| Definitions | Variable | Units | USEPA, 2013 | DEQ, 2010, 2013 | USEPA, 2013 | DEQ, 2010, 2013 | USEPA, 2013 | DEQ, 2010, 2013 | USEPA 2013 | DEQ, 2010, 2013 |
| Averaging Time, Carcinogens | ATc | yrs | 70 | | 70 | | 70 | | | 70 |
| Averaging Time, Noncarcinogens, adult | ATnc,a | yrs | 25 | | 25 | | 1 | | | 1a |
| Body Weight, adult | BW _{adult} | kg | 70 | | 70 | | 70 | | | 70 |
| Exposure Duration | ED _w | yrs | 25 | | 25 | | 1 | | | 1 |
| Exposure Frequency | EF_w | days/yr | 250 | | 225 | 250 | 250 | | NA | 9b |
| Exposure Time | ET_w | hours/day | 8 | | 8 | | 8 | | Not Available | Not Available |
| Event Frequency (contact with groundwater) | EV_w | events/day | NA | NA | NA | NA | NA | 2c | Not Available | 2c |
| Event Time (contact with groundwater) | t _{event} | hr/event | NA | NA | NA | NA | NA | 2 | NA | 2 |
| Soil Ingestion Rate | IR _{soil,w} | mg/day | 50 | NA | 100 | | 330 | | Not Available | 330 |
| Particulate Emission Factor (non-VOCs), default | PEF | m ³ /kg | NA | NA | 1.360E+09 | 1.320E+09 | 1.360E+06 | Not Available | Not Available | 1.320E+09 |
| Soil Adherence Factor | AF_{w} | mg/cm ² -event | 0.1d | | 0.2 | 0.1 | 0.3 | | Not Available | 0.3e |
| Exposed Body Surface Area (soil) | SAs | cm ² | 3300 | | 3300 | | 3300 | | | 3300 |
| Exposed Body Surface Area (groundwater/seep) | SA_w | cm ² | NA | NA | NA | NA | NA | 5700 | Not Available | 5700 |
| Duration of event | t _{event} | hr/event | NA | NA | NA | NA | NA | 2g | Not Available | 2g |
| Conversion Factor - dermal | CF _d | kg/mg | 1.0E-06 | | 1.0E-06 | | 1.0E-06 | | NA | 1.0E-06 |
| Conversion Factor - inhalation | CF _i | m ² /cm ² | 1.0E-04 | | 1.0E-04 | | 1.0E-04 | | NA | 1.0E-04 |
| Conversion Factor - ingestion | CF _o | kg/mg | 1.0E-06 | | 1.0E-06 | | 1.0E-06 | | NA | 1.0E-06 |

Central Tendency Exposure (CTE)

| | | | Indoor Worker | Outdoor | Construction | |
|---|----------------------|---------------------------|---------------|-----------|--------------|----------------------|
| | | | (IW) | Worker | Worker (CW) | Trench Worker |
| Definitions | Variable | Units | DEQ, 2010 | DEQ, 2010 | DEQ, 2010 | DEQ, 2010 |
| Averaging Time, Carcinogens | ATc | yrs | 70 | 70 | 70 | 70 |
| Averaging Time, Noncarcinogens, adult | ATnc,a | yrs | 6 | 6 | 0.5 | 1 |
| Body Weight, adult | BW _{adult} | kg | 70 | 70 | 70 | 70 |
| Exposure Duration | ED_{w} | yrs | 6 | 6 | 0.5 | 1 |
| Exposure Frequency | EF_w | days/yr | 250 | 250 | 250 | 9 |
| Event Frequency (contact soil/groundwater) | EV_w | events/day | NA | NA | 2 | 2 |
| Event Time (contact with groundwater) | t _{event} | hr/event | NA | NA | 2 | 2 |
| Soil Ingestion Rate | IR _{soil,w} | mg/day | NA | 50a | 100a | 100a |
| Particulate Emission Factor (non-VOCs), default | PEF | m ³ /kg | NA | 1.320E+09 | 1.360E+06 | 1.360E+06 |
| Soil Adherence Factor | AF_w | mg/cm ² -event | NA | 0.02 | 0.1i | 0.1i |
| Exposed Body Surface Area (soil) | SAs | cm ² | NA | 3300 | 3300 | 3300 |
| Exposed Body Surface Area (groundwater/seep) | SAw | cm ² | NA | NA | 5700 | 5700 |

Bolded values are values selected for use in BHHRA

-- = Same

NA = Not Applicable

a = EPA 2002

b = The value of 9 days per year was based on standard dimensions for a residential excavation site from DEQ (1997a) and construction worker excavation statistics from EPA and Means Heavy Construction Cost Data, 8th Annual Edition, R.S. Means Company, Inc., Kingston, MA. c = Assumes that direct soil contact or groundwater contact activities occur twice a day (morning, afternoon) for a total of four hours per day.

d = EPA 2004: Exhibit 3-3, mean for commercial gardener used to represent upper end commercial exposure.

e = EPA 2004: Exhibit 3-3, 95th percentile construction worker

f = EPA 1997, adjusted for time spent at work (8 hours / 24 hours)

g = Assumes that direct soil contact or groundwater contact activities occur twice a day (morning, afternoon) for a total of four hours per day.

h = EPA 2004: Exhibit 3-3, mean for commercial groundskeeper

i = EPA 2004: Exhibit 3-3, mean for construction worker

DEQ, 2010: Human Health Risk Assessment Guidance

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 $USEPA \ Regional \ Screening \ Values \ (RSL): http://www.epa.gov/reg3hwmd/risk/human/rb-concentration_table/usersguide.htm$

Table 3-1CPECs to be Evaluated in the BERABradford Island - Upland Operable Unit

| | | | | Boiaccumulative COIs without SLVs or |
|--------------|-------|---|---|---|
| AOPC | Media | Inorganic CPECs | Organic CPECs | SLVs not Bioaccumulation-based |
| Landfill | Soil | Antimony, chromium, copper, lead, mercury, and nickel | Total HPAHs | Tributyltin, chlordane-technical, heptachlor, 1,2,4-trimethylbenzene, bis(2-ethylhexyl)phthalate, butylbenzylphthalate, carbazole, dibenzofuran, and di-n-butylphthalate |
| Sandblast | Soil | | Bis(2-ethylhexyl)phthalate and total HPAHs | Tributyltin, BHC-delta, BHC-gamma, chlordane-alpha, chlordane- gamma, endosulfan I, endosulfan II, endosulfan sulfate, endrin, endrin aldehyde, endrin ketone, heptachlor, methoxychlor, 1,2,4- trumethylbenzene, 4-isopropyltoluene, n-propylbenzene, butylbenzylphthalate, carbazole, dibenzofuran, di-n-butylphthalate, and di-n-octylphthalate |
| Pistol Range | Soil | Lead | | |
| Bulb Slope | Soil | Lead and mercury | | |

AOPC = area of potential concern

BERA = baseline ecological risk assessment

BHC = hexachlorocyclohexane

COI = contaminant of interest

CPEC = contaminant of potential ecological concern

HPAH = high molecular weight PAH

PAH = polycyclic aromatic hydrocarbons

SLV = screening level value

Table 3-2Assessment and Measurement Endpoints for the BERABradford Island - Upland Operable Unit

| | Measurement Endpoints | | | | | |
|--|---|---|--|--|--|--|
| Assessment Endpoints | Measures of Exposure | Measures of Effect | | | | |
| Protection of the terrestrial plant community and soil invertebrate populations that may be exposed to COIs in soil to maintain species diversity, abundance, and nutrient cycling. | Measured concentrations in surface and shallow soils that reduce survival, growth, and/or productivity of the plant or soil invertebrate communities. | Potential toxicity due to exceedances of screening values related to maintenance of the terrestrial plant community, based on a 20% reduction (or greater) in growth or yield (DEQ 2001; Efroymson et al. 1997a). | | | | |
| | | Potential toxicity due to exceedances of screening values related to maintenance of the soil-dwelling invertebrate community, based on a 20% reduction (or greater) in growth, reproduction, or activity (DEQ 2001; Efroymson et al. 1997b). | | | | |
| Protection of herbivorous birds (trophic level 1), represented by the Canada goose , with no unacceptable effects on reproduction, growth, or development on a population level due to COIs in soil and terrestrial plants. | Measured concentrations in surface soils, estimated water concentrations in upland puddles, and estimated concentrations in terrestrial plant tissues that reduce reproduction, health, and/or survival of populations of avian herbivores. | Potential or observed toxicity due to exceedances of screening values and/or acceptable hazard quotients related to survival, growth, and reproduction of resident populations of Canada geese (DEQ 2001; Efroymson et al. 1997c; USEPA 2005b; Sample et al. 1996). | | | | |
| Protection of invertivorous birds (trophic level 2), represented by the robin , with no unacceptable effects on reproduction, growth, or development on a population level due to COIs in soil and invertebrates. | Measured concentrations in surface soils, estimated water concentrations in upland puddles, and estimated concentrations in soil invertebrate tissues that reduce reproduction, health, and/or survival of populations of avian invertivores. | Potential or observed toxicity due to exceedances of screening values and/or acceptable hazard quotients related to survival, growth, and reproduction of resident populations of robins (DEQ 2001; Efroymson et al. 1997c; USEPA 2005b; Sample et al. 1996). | | | | |
| Protection of carnivorous small mammals (trophic level 2-3), represented by the vagrant shrew , with no unacceptable effects on reproduction, growth, or development on a population level due to COIs in soil and invertebrates. | Measured concentrations in surface and shallow soils, estimated concentrations in water from upland puddles, and estimated concentrations in soil invertebrate tissues that reduce reproduction, health, and/or survival of populations of carnivorous small mammals. | Potential or observed toxicity due to exceedances of screening values and/or acceptable hazard quotients related to survival, growth, and reproduction of resident populations of vagrant shrews (DEQ 2001; Efroymson et al. 1997c; USEPA 2005b; Sample et al. 1996). | | | | |
| Protection of top-level predatory birds (trophic level 3- 4), represented by the American kestrel , with no unacceptable effects on reproduction, growth, or development on a population level due to COIs in soil and small mammals. | Measured concentrations in surface soils, measured concentrations in river water, and estimated concentrations in small mammal tissues that reduce reproduction, health, and/or survival of populations of top-level predatory birds. | Potential or observed toxicity due to exceedances of screening values and/or acceptable hazard quotients related to survival, growth, and reproduction of resident populations of American kestrels (DEQ 2001; Efroymson et al. 1997c; USEPA 2005b; Sample et al. 1996). | | | | |

Table 3-2Assessment and Measurement Endpoints for the BERABradford Island - Upland Operable Unit

| Aggagement Endnaints | Measurement Endpoints | | | | |
|---|---|---|--|--|--|
| Assessment Endpoints | Measures of Exposure | Measures of Effect | | | |
| Protection of large carnivorous mammals (trophic level | Measured concentrations in surface soils, measured | Potential or observed toxicity due to exceedances of | | | |
| 3-4), represented by the mink , with no unacceptable | concentrations in river water, and estimated | screening values and/or acceptable hazard quotients | | | |
| effects on reproduction, growth, or development on a | concentrations in small mammal tissues that reduce | related to survival, growth, and reproduction of resident | | | |
| population level due to COIs in Upland OU soil and | reproduction, health, and/or survival of populations of | populations of mink (DEQ 2001; DEQ 2007; LWG | | | |
| Upland OU small mammals (15% of diet). | carnivorous large mammals. | 2007; USEPA 2005b; Sample et al. 1996). | | | |
| | | | | | |

Notes:

Surface soil = 0 to 1 foot bgs for all AOPCs, with the exception of the Pistol Range for which surface soil is defined as 0 to 1.5 feet bgs.

Shallow soil = 0 to 3 feet bgs (only applies to Landfill and Sandblast Area AOPCs).

USEPA = United States Environmental Protection Agency

DEQ = (Oregon) Department of Environmental Quality

LWG = Lower Willamette Group

Table 3-3 **Exposure Factors for Ecological Receptors** Bradford Island - Upland Operable Unit

| Parameter | Symbol | Units | American Robin | Vagrant Shrew | Canada Goose | American Kestrel | Mink | Reference/Comment |
|--|----------------------------|----------------------|--------------------------------------|--------------------------------------|--------------------------------------|--------------------------------------|--|---------------------------------------|
| Habitat | | | Terrestrial | Terrestrial | Terrestrial | Terrestrial | Aquatic-dependent | (A) |
| Operable Unit | OU | | Upland | Upland | Upland | Upland | River | |
| Trophic Level | | | Level 3 | Level 3 | Level 2 | Level 4 | Level 2-4 | (A) |
| Occurrence | | | Resident | Resident | Resident | Resident | Resident | |
| Status | | | None | None | None | None | None | (B) |
| Home Range | | acres or km | 0.37 acres (C) | 0.26 acres (A) | 2,430 acres (C) | 270 acres (D) | 19 acres (C) | |
| Area Use Factor | AUF | fraction of site | Receptor- and AOPC- specific | General equation (1) |
| Seasonality Factor | SF | fraction of one year | 1 | 1 | 1 | 1 | 1 | All are resident species |
| Body Weight | BW | kg | 0.0773 | 0.007 | 3.0 | 0.116 | 0.974 | (C) |
| Dietary Composition | | | Soil-dwelling invertebrates | Soil-dwelling invertebrates | Aboveground vegetation | Small mammals | Small mammals, benthic invertebrates, and fish | (A,C) Professional judgment also used |
| Diet - Plant material | df_i | fraction of diet | 0 | 0 | 1 | 0 | 0 | (A,C) Professional judgment also used |
| Diet - Soil-dwelling invertebrates | df_i | fraction of diet | 1 | 1 | 0 | 0 | 0 | (A,C) Professional judgment also used |
| Diet - Small mammals | df_i | fraction of diet | 0 | 0 | 0 | 1 | 0.15 | (A,C) Professional judgment also used |
| Food Ingestion Rate | IR _{food} | kg/day dw | 0.013 | 0.0014 | 0.15 | 0.017 | 0.054 | (E) |
| Food ingestion Rate | m _{food} | kg/day ww | NA | NA | NA | NA | 0.16 | (E) |
| Fraction of Soil or Sediment in Diet | | fraction of diet | 0.104 | 0.04 | 0.082 | 0.02 | 0.094 | (F) (2) |
| Incidental Soil or Sediment Ingestion Rate | IR _{soil} | kg/day dw | 0.0013 | 0.000055 | 0.013 | 0.00033 | 0.0051 | General equation (3) |
| Incidental Surface Water Ingestion Rate | IR _{water} | L/day | 0.011 | 0.0011 | 0.12 | 0.014 | 0.097 | (C) |
| Concentration in Food - type <i>i</i> | $C_{food i}$ | mg/kg dw | Food-specific chemical concentration | Measured/modeled concentration |
| Concentration in Soil/Sediment | C _{soil/sed} | mg/kg dw | Chemical-specific | Chemical-specific | Chemical-specific | Chemical-specific | Chemical-specific | Measured concentration |
| Concentration in Surface Water | C _{water} | mg/L | Chemical-specific | Chemical-specific | Chemical-specific | Chemical-specific | Chemical-specific | Measured concentration |

Notes:

AOPC = area of potential concern BW = body weight

dw = dry weight

kg = kilogram

km = kilometer

L = liter

mg = milligram

NA = not applicable

SF = seasonality factor

ww = wet weight

(1) Area Use Factor (AUF) = area of the site divided by the area of home range

Because part of Bradford Island is managed as Canadian goose habitat, an AUF of 1.0 will be used for the goose regardless of its home range.

(2) The remaining 85% of the mink's diet is assumed to be comprised of prey from the River OU (i.e., fish and crayfish), which will be evaluated in the River OU Baseline ERA.

(3) Percent soil or sediment in diet reported for species with similar feeding habits used to develop soil and sediment ingestion rates (Beyer et al. 1994). For species without an adequate surrogate, the default percent soil in diet of 2% was used (Beyer et al., 1994). For the vagrant shrew, the percent soil in diet for a white-footed mouse (2%) was doubled to account for a higher soil ingestion rate expected for a shrew.

(4) $IR_{soil} = IR_{food_dw} * fraction of soil/sediment in diet$

(A) Zeiner, D.C., W.F. Laudenslayer, K.E. Mayer, and M. White (eds) 1990. California's Wildlife. Volumes I-III. Department of Fish and Game, The Resources Agency, State of California, Sacramento, CA. April., and Lower Willamette Group. 2007. Round 2 Ecological Risk Assessment for the Portland Harbor (Appendix G of the Portland Harbor Remedial Investigation/Feasibility Study).

(B) Oregon Department of Fish and Wildlife, Oregon List of Threatened and Endangered Fish and Wildlife Species: http://www.dfw.state.or.us/threatened_endangered/t_e.html

(C) United States Environmental Protection Agency (USEPA). (1993). Wildlife Exposure Factors Handbook. Office of Research and Development, Washington, DC. EPA/600/R-93/187a. December, 1993.

(D) California Wildlife Biology, Exposure Factor, and Toxicity Database (Cal/ECOTOX). 2002. Created by the California Office of Environmental Health Hazard Assessment and the University of California at Davis. http://www.oehha.ca.gov/cal_ecotox/default.htm

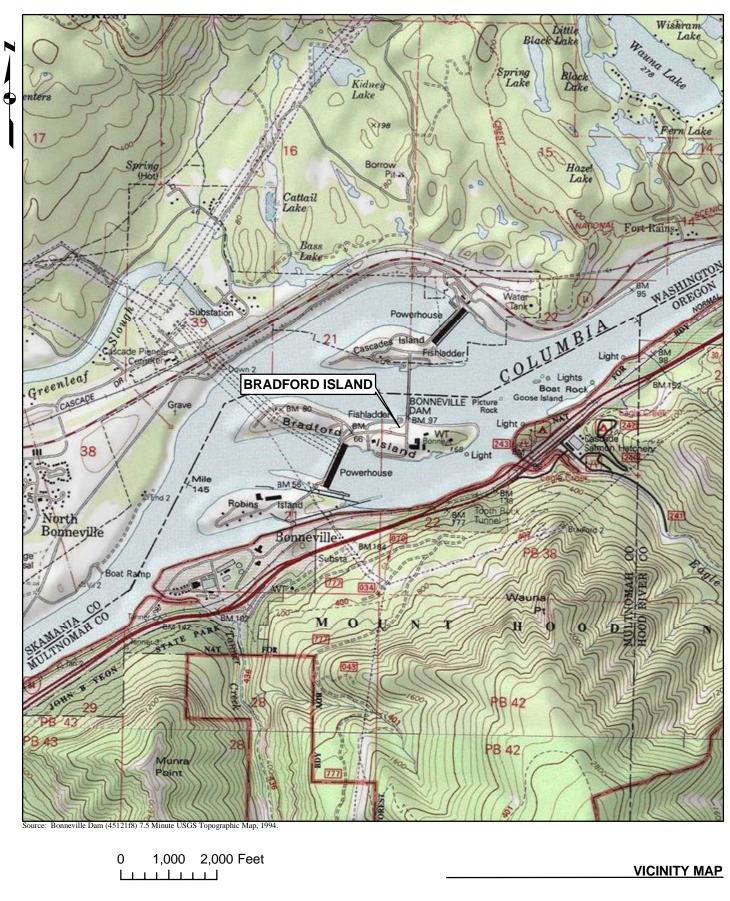
(E) Nagy KA. 2001. Food requirements of wild animals: predictive equations for free-living mammals, reptiles, and birds. Nutrition Abstracts and Reviews, Series B 71, 21R-31R.

If species-specific empirical data not available, then the generic equations for "all mammals" or "all birds" used to estimate IR food-

(F) Beyer, W.N., Connor, E.E. and Gerould, S. 1994. Estimates of soil ingestion by wildlife. Journal of Wildlife Management. 58:375-82.

FIGURES



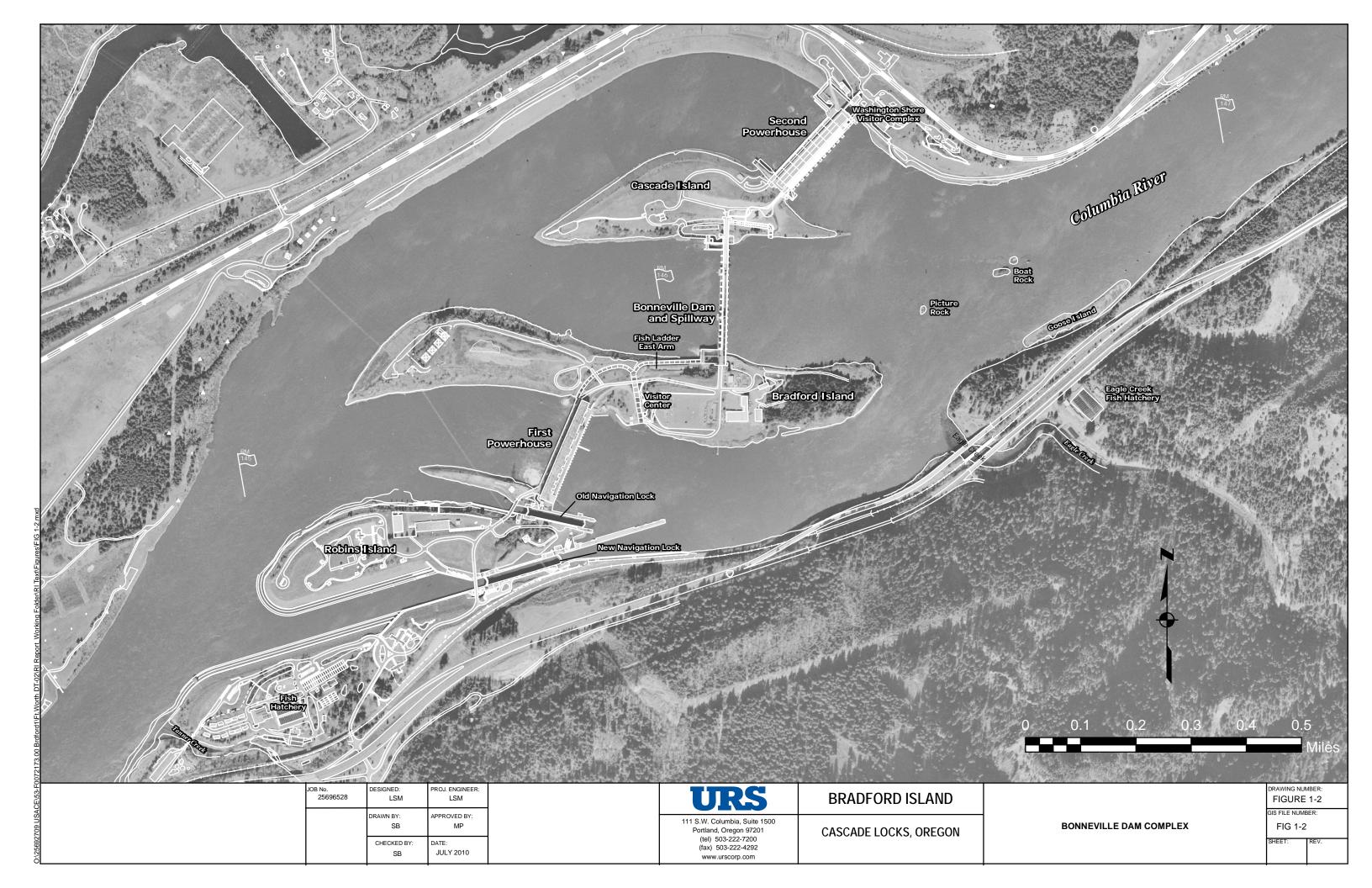


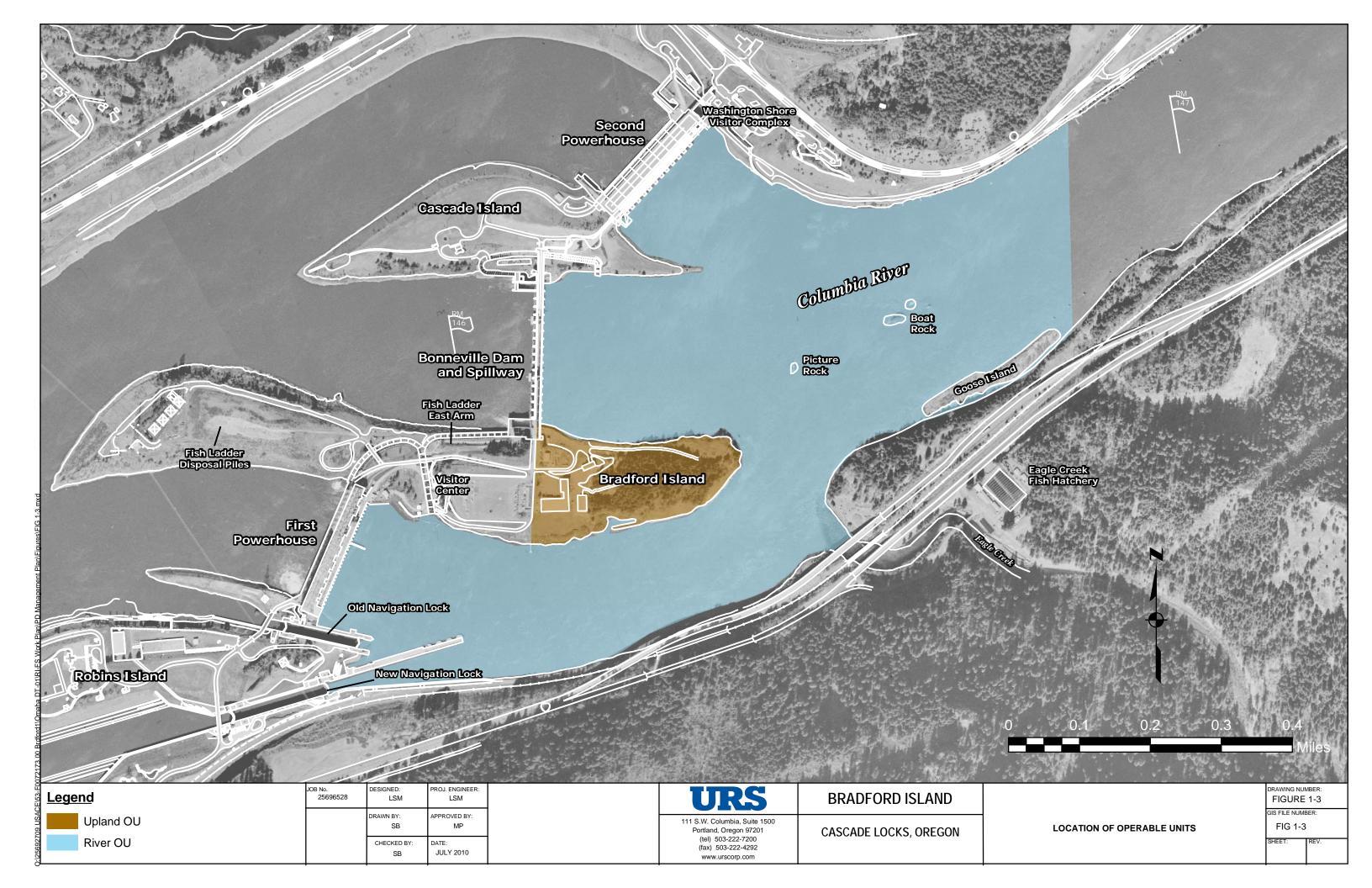


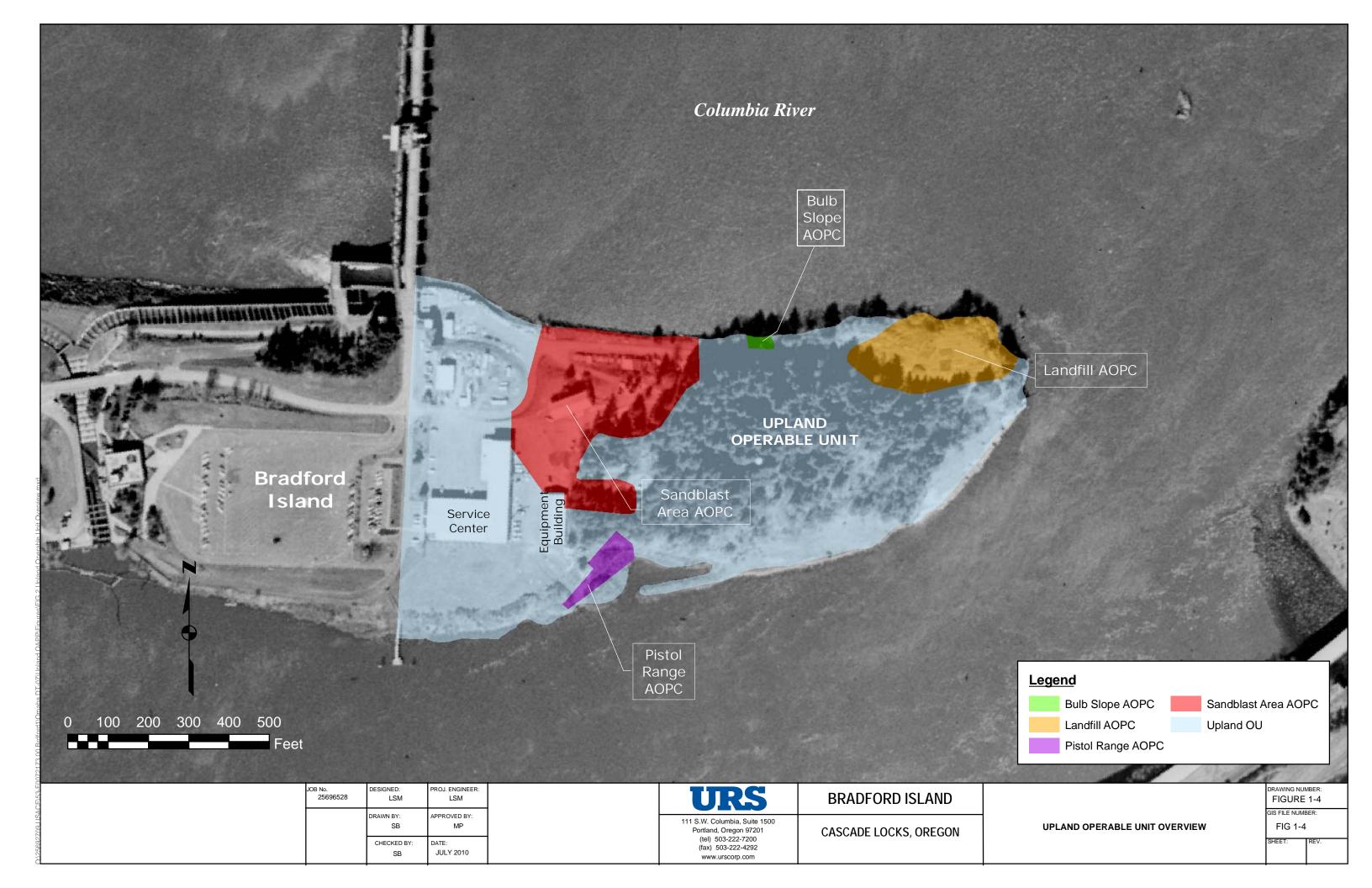
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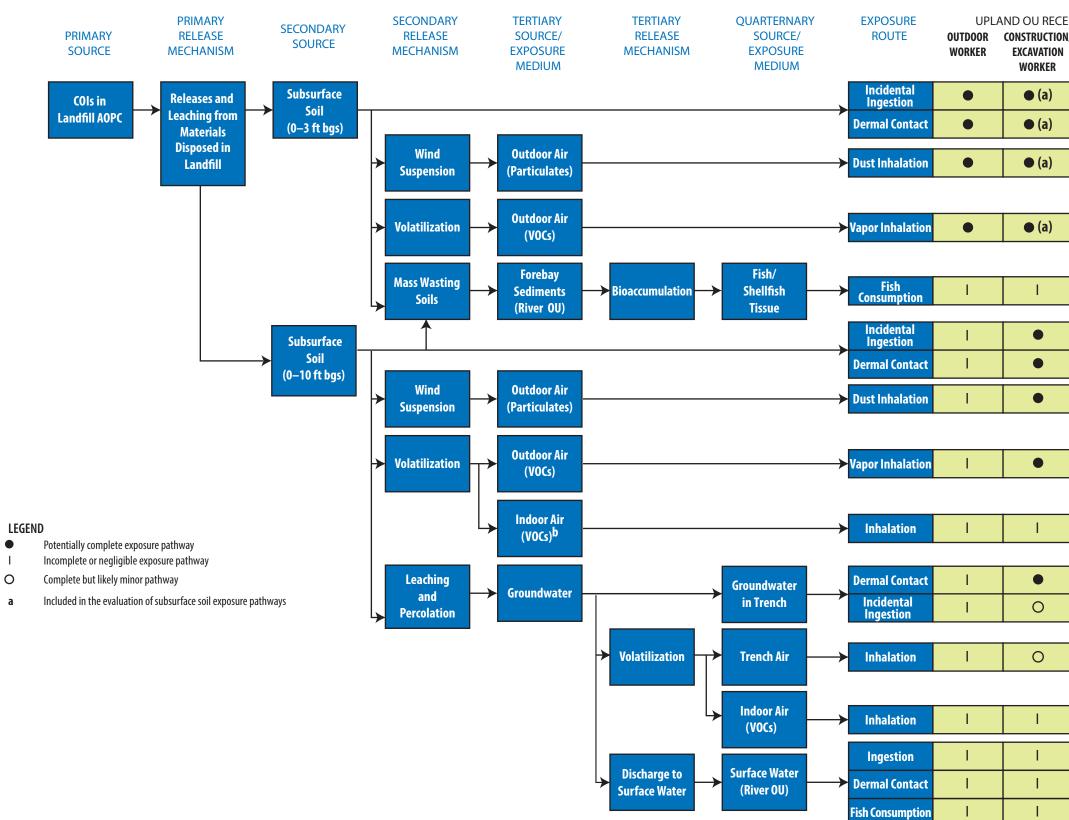
BRADFORD ISLAND CASCADE LOCKS, OREGON

FIGURE 1-1



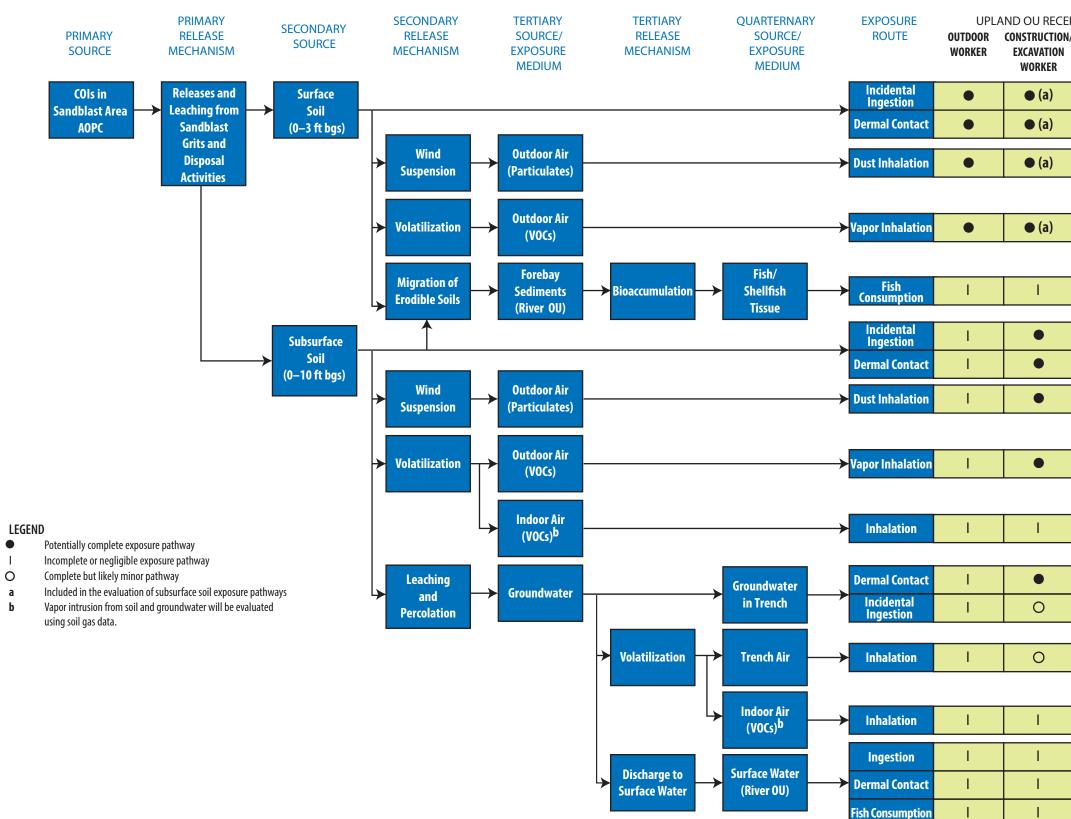






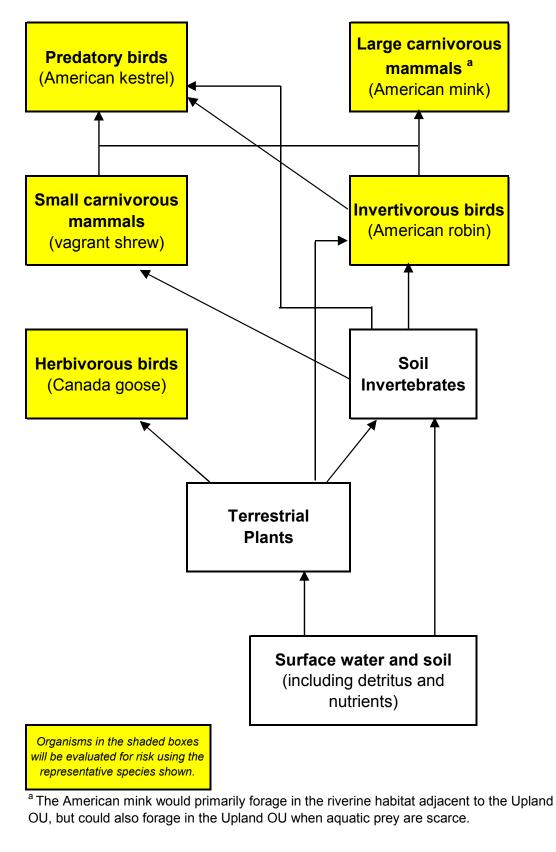
| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND |
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| | DRAWN BY: MS | APPROVED BY: MP | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 | CASCADE LOCKS, OREGON |
| | CHECKED BY: HP | DATE: JULY 2010 | (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | ,, |

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| | DRAWN BY: MS | APPROVED BY: MP | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 | CASCADE LOCKS, OREGON |
| CI | CHECKED BY: HP | DATE: JULY 2010 | (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | |

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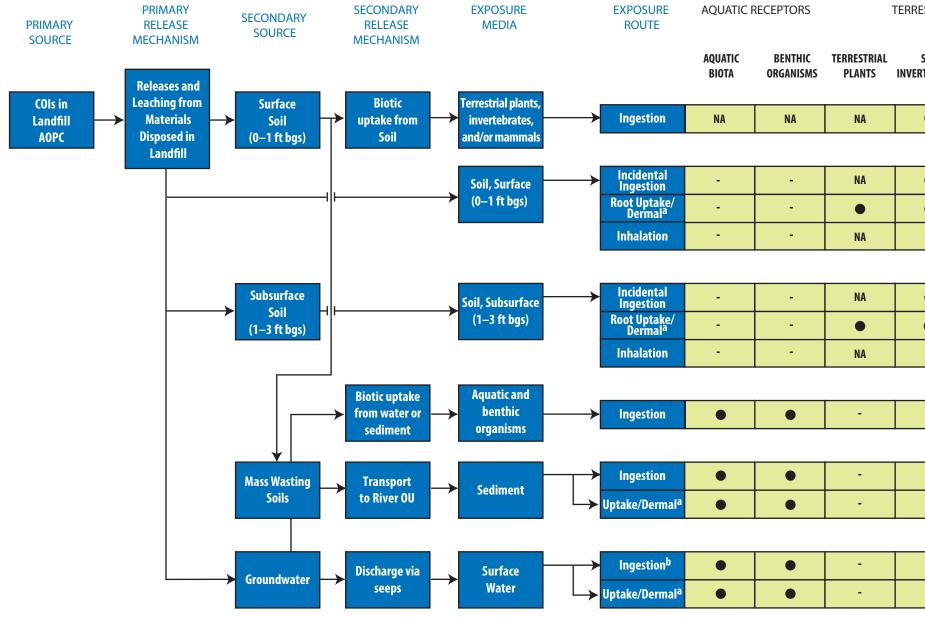




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BRADFORD ISLAND CASCADE LOCKS, OREGON

FIGURE 3-1



- Complete and potentially significant pathway
- O Complete but likely minor pathway
- Incomplete pathway
- NA Pathway not applicable to receptor

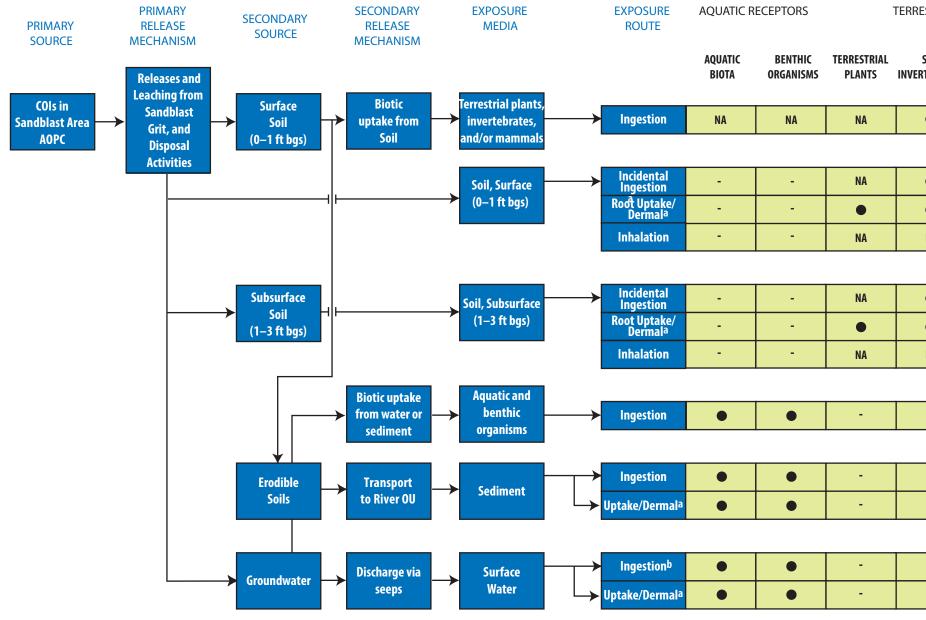
Aquatic Biota defined as aquatic plants, plankton, invertebrates, fish and aquatic-dependent wildlife.a. The dermal exposure pathway for wildlife is expected to be minor due to the barrier offered by fur and feathers.b. Highly mobile birds and mammals could ingest water from the river, while less mobile species would likely ingest water from puddles in the Upland OU.

| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | LANDFILL AOPC | DRAWING NU | 3-2 |
|---------------------|--------------------------------------|--|--|-----------------------|--------------------------------------|-----------------------------------|-----|
| | DRAWN BY: MS CHECKED BY: HP | APPROVED BY: MP DATE: JULY 2010 | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | CASCADE LOCKS, OREGON | ECOLOGICAL CONCEPTUAL EXPOSURE MODEL | GIS FILE NUM FIG 3-2 SHEET: | |

| BIRDS | MAMMALS | | |
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- Complete and potentially significant pathway
- O Complete but likely minor pathway
- Incomplete pathway
- NA Pathway not applicable to receptor

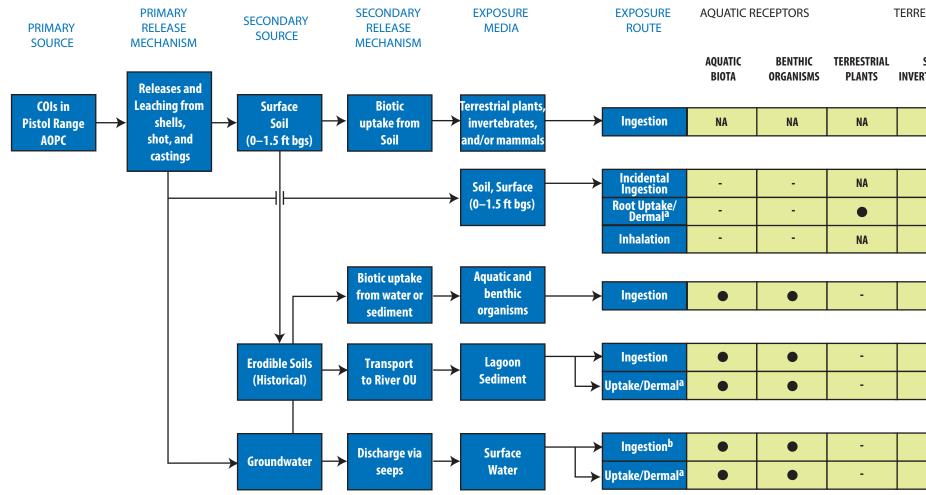
Aquatic Biota defined as aquatic plants, plankton, invertebrates, fish and aquatic-dependent wildlife.a. The dermal exposure pathway for wildlife is expected to be minor due to the barrier offered by fur and feathers.b. Highly mobile birds and mammals could ingest water from the river, while less mobile species would likely ingest water from puddles in the Upland OU.

| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | SANDBLAST AREA AOPC | DRAWING NU | | |
|---------------------|-------------------|------------------------|---|-----------------------|--------------------------------------|------------|-----------------------------|--|
| | DRAWN BY: MS | APPROVED BY: MP | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 | CASCADE LOCKS, OREGON | ECOLOGICAL CONCEPTUAL EXPOSURE MODEL | | GIS FILE NUMBER: FIG 3-3 | |
| | CHECKED BY: HP | DATE: JULY 2010 | (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | , | | SHEET: | REV. | |

| SOIL RTEBRATES | BIRDS | MAMMALS | | |
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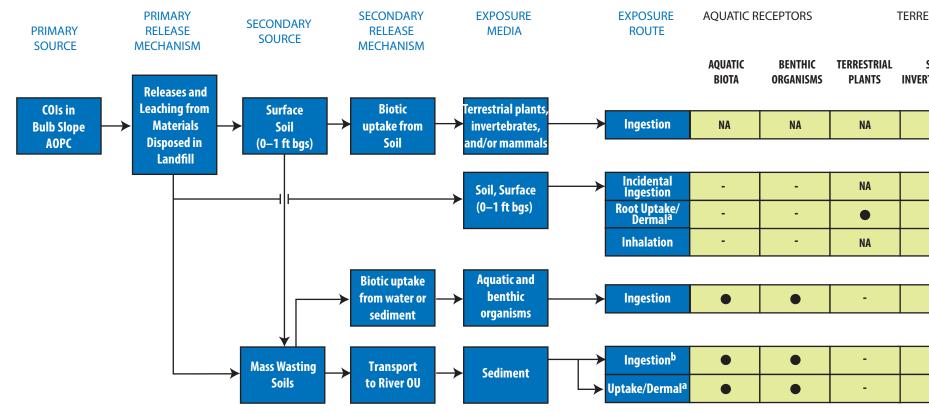


- Complete and potentially significant pathway
- O Complete but likely minor pathway
- Incomplete pathway
- NA Pathway not applicable to receptor

Aquatic Biota defined as aquatic plants, plankton, invertebrates, fish and aquatic-dependent wildlife.a. The dermal exposure pathway for wildlife is expected to be minor due to the barrier offered by fur and feathers.b. Highly mobile birds and mammals could ingest water from the river, while less mobile species would likely ingest water from puddles in the Upland OU.

| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | PISTOL RANGE AOPC | DRAWING NUMB | -4 | |
|---------------------|-------------------|------------------------|---|-----------------------|--------------------------------------|--------------|-----------------------------|--|
| | DRAWN BY: MS | APPROVED BY: MP | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 | CASCADE LOCKS, OREGON | ECOLOGICAL CONCEPTUAL EXPOSURE MODEL | | GIS FILE NUMBER: FIG 3-4 | |
| | CHECKED BY: HP | DATE: JULY 2010 | (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | | | SHEET: R | EV. | |

| SOIL RTEBRATES | BIRDS | MAMMALS |
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- Complete and potentially significant pathway
- O Complete but likely minor pathway
- Incomplete pathway
- NA Pathway not applicable to receptor

Aquatic Biota defined as aquatic plants, plankton, invertebrates, fish and aquatic-dependent wildlife.a. The dermal exposure pathway for wildlife is expected to be minor due to the barrier offered by fur and feathers.b. Highly mobile birds and mammals could ingest water from the river, while less mobile species would likely ingest water from puddles in the Upland OU.

| JOB No. 25696528 | DESIGNED: LSM | PROJ. ENGINEER: LSM | URS | BRADFORD ISLAND | BULB SLOPE AOPC | DRAWING NU | E 3-5 | |
|---------------------|-------------------|--|---|-----------------------|--------------------------------------|------------|-----------------------------|--|
| | DRAWN BY: MS | MS MP 111 S.W. Columb Portland, Ore | 111 S.W. Columbia, Suite 1500 Portland, Oregon 97201 | CASCADE LOCKS, OREGON | ECOLOGICAL CONCEPTUAL EXPOSURE MODEL | | GIS FILE NUMBER: FIG 3-5 | |
| | CHECKED BY: HP | DATE: JULY 2010 | (tel) 503-222-7200 (fax) 503-222-4292 www.urscorp.com | | | SHEET: | REV. | |

| SOIL RTEBRATES | BIRDS | MAMMALS |
|-------------------|-------|---------|
| • | • | • |
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APPENDIX D USACE Responses to ODEQ Comments on Draft Upland OU BHHERA

| Comment Number | Comments by Jennifer Peterson, Mike Poulsen and Bob Schwarz, DEQ Cleanup Program March 2, 2016 | US Army Corps of Engineers Response/Action Taken |
|-------------------|---|---|
| General Hu | iman Health Comments | |
| 1. | State Rules and Guidance . The report emphasizes federal risk assessment requirements and guidance over state rules and guidance. This includes use of | Comment noted. The US Army Corps of Engineers (USACE) will be making final decisions based on Comprehensive Environmental |
| | acceptable risk levels and evaluations of total petroleum hydrocarbons. DEQ will be making final decisions based on state rules. | Response, Compensation & Liability Act (CERCLA). Please see Section 1.3 of the Remedial Investigation (RI) Report (URS, 2012) for a detailed description of regulatory initiative and authority. Please also see response to Comment #4 regarding petroleum hydrocarbons. No revisions to the Baseline Human Health Risk Assessment (BHHRA) are necessary. |
| 2. | Fishing Platform User Exposure. The USACE agreed to include fishing platform users as a potential future exposure scenario. In the report there is an inappropriate distinction made between this exposure scenario and the other ones. The fishing platform scenario should be included in the conceptual site model with the other potential scenarios. It is appropriate to acknowledge the uncertainty associated with this potential future scenario. | Fishing platform scenario will be added to the conceptual exposure models (CEMs) for the Landfill and Sandblast AOPCs, Figures 2-1 and 2-2. The human health CEMs for Bulb Slope and Pistol Range AOPCs will be added to the BHHERA, and will also include the fishing platform scenario. USACE will retain the distinction that the Fishing Platform User scenario is a hypothetical future scenario. |
| Specific Hu | man Health Comments | |
| 3. | Page 2-3. DEQ uses 5 percent frequency of detection only as an indication that there may be an error in the measured concentration of a chemical. If data are found to be invalid, they should not be used in the risk assessment. For valid data, it is entirely reasonable to evaluate risks from chemicals that may be present (at potentially high concentrations) in a small area and are not detected elsewhere. | Comment noted. The detection frequency step of the screening process (used in the Remedial Investigation [RI] Report; URS 2012) was described in the RI/Feasibility (FS) Management Plan (MP) (URS 2007) and follows precedent established in risk assessments for other sites led by the Oregon Department of Environmental Quality (ODEQ), including the Level II Ecological Screening Assessment previously performed for the Landfill Area of Potential Concern (AOPC). The detection frequency topic was previously addressed in the Remedial Investigation (RI) Report (URS 2012). The potential for underestimation of human health risk based on detection frequency status was discussed in chemical-specific and AOPC-specific detail in Appendix O.2.1.3 in the RI Report. The results of the detection frequency evaluation provided in Appendix O.2.1.3 indicate that none of the contaminants of interest (COIs) eliminated as contaminants of potential concern (COPCs) in the Landfill and Sandblast AOPCs on the basis of frequency of detection represent a significant uncertainty in the HHRA. No revisions to the BHHRA are necessary. |

| Comment Number | Comments by Jennifer Peterson, Mike Poulsen and Bob Schwarz, DEQ Cleanup Program March 2, 2016 | US Army Corps of Engineers Response/Action Taken |
|-------------------|---|--|
| 4. | Page 2-4 . TPH was excluded from evaluation because there is a petroleum exclusion in CERCLA. TPH is a hazardous substance in Oregon, and should be evaluated accordingly. | Comment noted. USACE plans to continue to follow CERCLA. However, the total petroleum hydrocarbons (TPH) were not ignored in the BHHRA. TPH are included in the initial screening for the HHRA in the RI Report (Tables 2-1 to 2-4, URS 2012). The only TPH fraction exceeding the screening level was TPH-RRO. Although it was not retained as a COPC in the risk calculations due to the CERCLA exclusion, the potential risk associated with TPH-RRO is discussed in the Uncertainty Section, in Section 2.6.4. No revisions to BHHRA are necessary. |
| 5. | Page 2-9 . Using the 90% UCL as the exposure point concentration is a requirement of rule, although DEQ can accept other methods. We consider the 95% UCL to be similar and an acceptable basis for an EPC. No action is required. | Comment noted. No revisions to the BHHRA are necessary. |
| 6. | Page 2-19 and page 4-2 . The Corps states that " exceedance of the ODEQ risk thresholds or USEPA acceptable risk range does not automatically mean that adverse effects may have occurred or will occur (e.g., does not automatically mean that the COC will be recommended for further evaluation in the FS)." If the risk assessment shows unacceptable risk from exposure to chemicals, DEQ will require consideration of those chemicals in the feasibility study. | Comment noted. The BHHRA uses multiple lines of evidence to derive recommendations for further evaluation in the FS. No revisions to the BHHRA are necessary. |
| 7. | Pages 2-20 to 2-23. For naturally-occurring chemicals present above background levels, DEQ requires the evaluation of risk. If the excess cancer risk is greater than 1 x 10-6, that is unacceptable. DEQ does not consider the exceedance above background risk. However, both EPA and DEQ will not require remediation to below naturally occurring background concentrations. The difference is that DEQ allows chemicals to be screened out prior to conducting a risk assessment if the concentrations are below background levels. But both agencies require evaluating risk at the EPC concentration, not after subtracting out the background concentration or risk. DEQ considers background levels for hazardous substances that occur naturally. Typically this evaluation is limited to inorganic chemicals. Although | Consistent with USEPA and ODEQ requirements, risk <u>was</u> evaluated at the total exposure point concentration (EPC). In addition, the contributions from background levels of arsenic and incremental site- related arsenic to total risk were also presented, in compliance with U.S. Environmental Protection Agency's (USEPA) guidance regarding the role of background in risk assessment (USEPA 2002b,c) and to illustrate the magnitude of the contributions to risk from the two types of sources. The comparison was useful in indicating that the site-related contribution to arsenic risk was very minor for the occupational exposures. For the Fishing Platform user, the site-related contribution to arsenic risk was higher, and therefore, it was retained as a chemical of concern and included in Table 2-27. |
| | we acknowledge that cPAHs may occur naturally, we are not confident that the cPAHs measured in the reference area are naturally occurring. | The Reference Area background carcinogenic polycyclic aromatic hydrocarbon (cPAH) upper prediction level (UPL) was used only for concentration comparison purposes, but was not used to screen out cPAHs. As shown in Table 2-27, cPAHs were retained as contaminants |

| Comment Number | Comments by Jennifer Peterson, Mike Poulsen and Bob Schwarz, DEQ Cleanup Program March 2, 2016 | US Army Corps of Engineers Response/Action Taken |
|-------------------|---|---|
| | | of concern (COCs) and receptor-specific risk based concentrations (RBCs) were developed. |
| | | No revisions to the BHHRA are necessary. |
| 8. | Page 2-30, Section 2.6.3.2. The fishing platform exposure scenario is a potential future exposure, and one the Corps agreed to evaluate. The assumptions in the risk assessment should incorporate reasonable maximum exposure, not "impossible" or worst-case conditions. The likelihood of future exposure can be discussed in the uncertainty section. | Text in Section 2.6.3.2 will be changed from "impossible" to " <i>highly unlikely</i> ." Please note, Section 2.6.3.2 is already part of the Uncertainty Section (Section 2.6). |
| 9. | Page 2-31. To the extent that chemicals are detected in soil or groundwater, regardless of whether they are degradation products of PCE, their risks should be evaluated. | Clarification provided. As stated in Section 2.1 of the BHHERA, risks from tetrachloroethylene (PCE) and its degradation products, including both detections and non-detections, were evaluated in the BHHRA to meet ODEQ recommendations. See Section 2.5 and Tables 2-1, 2-2, 2- 11, 2-12, and 2-16 through 2-24 for quantitative risk evaluations of detected concentrations of PCE and its degradation products. For an evaluation of the non-detect concentrations of PCE degradation products see Section 2.6.3.4 and Tables A-3.1 to A11.4 of Appendix A. No revisions to the BHHRA are necessary. |
| 10. | Page 2-33 and Table 2-1 (notes for Tables 2-1 to 2-4) footnote ***.Petroleum hydrocarbons are hazardous substances in Oregon, and risks fromexposure to TPH should be evaluated accordingly. | Comment noted. Please see response to Comment #4. No revisions to the BHHRA are necessary. |
| 11. | Page 2-35. The Corps states that "This would imply that risk levels for individual chemicals that exceed 1 x 10-6 are acceptable as long as the cumulative multi-chemical risk level does not exceed 1 x 10-5." As stated, this is incorrect; both criteria are used. Excess cancer risk exceeding 1 x 10-6 for an individual carcinogen is unacceptable, regardless of the cumulative risk. However, it appears that results shown in tables are interpreted correctly. | The text will be revised to state: "Cumulative risks and hazards are also presented in the context of $ODEQ$'s acceptable multi-carcinogen risk level of 1×10^{-5} . However, even if the cumulative multi-chemical risk level does not exceed 1×10^{-5} , individual chemicals are still identified as a COC if they exceed the individual risk level of 1×10^{-6} (i.e., the RME Construction Worker at both the Landfill and Sandblast Area AOPCs have cumulative cancer risks less than ODEQ's threshold of 1×10^{-5} but benzo(a)pyrene is identified as a COC at both AOPCs due to exceeding the individual risk level of 1×10^{-6} ." |

| Comment Number | Comments by Jennifer Peterson, Mike Poulsen and Bob Schwarz, DEQ Cleanup Program March 2, 2016 | US Army Corps of Engineers Response/Action Taken |
|-------------------|--|--|
| 12. | Table 2-7a. Three values are shown for PEF. The correct value should be 1.36 x 10+9 or 1.32 x 10+9, not 1.36 x 10+6, m3/kg. DEQ uses the inverse value of 7.58 x 10-10 kg/m3 (which corresponds to 1.32 x 10+9 m3/kg). The CTE body weight for the indoor worker should be 80 kg, not 70 kg. | The following revisions will be made to Table 2-7a: 1) Outdoor Maintenance Worker - The particulate emission factor (PEF) value of 1.36E⁺⁰⁹ will be revised to arithmetic notation as 1.36 x 10⁺⁹ for the outdoor maintenance worker. This value (1.36E⁺⁰⁹) will be used because it is consistent with USEPA guidance and practice (USEPA 2002a, 2015a,b). 2) Excavation/Trench Worker - The excavation/ trench worker is |
| | | 2) Excavation/Trench Worker - The excavation/ trench worker is assumed to be only exposed incidentally to groundwater at the bottom of a trench. The PEF value of 1.32E+09, which was shown for the Excavation/ Trench Worker, will be deleted. The soil ingestion rate of 330 mg/day will also be deleted. Neither was used in the risk calculations. |
| | | 3) Construction Worker - The construction worker is assumed to be exposed only to soils. For the construction worker scenario, USEPA's Soil Screening guidance (USEPA 2002a) recommends that PEFs should reflect exposure to the higher levels of dust that would typically be expected for a construction scenario. However, USEPA does not provide default PEFs for construction activities. Neither does ODEQ guidance. Therefore, the default PEF (1.0 x 10⁺⁰⁶ m³/kg) recommended by the California Department of Toxic Substances Control (DTSC) for construction workers was used (DTSC, HHRA Note 2, 2014). The DTSC value is more conservative than ODEQ's current practice of using the same PEF for all types of receptors, and therefore assumed to be acceptable for the purposes of the BHHRA. 4) CTE body weight will be changed to 80 kg in the table; it was used in the risk calculations (not 70 kg). |
| 13. | Table 2-8. We could not confirm the OAF value of 0.013 for lead. | The oral absorption factor (OAF) value of 0.013 for lead will be deleted from Table 2-8 and replaced with a dash notation "". Because lead was evaluated separately, using the procedures described in Appendix B-2, the following footnote will be added to Table 2-8: " <i>Lead is evaluated independently using adult and child lead models</i> <i>because a linear adjustment of the RSL may not be appropriate, see</i> <i>Table 2-25 and Appendix B.</i> " |

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| 14. | Table 2-9. Rather than leave the evaluation of lead risk as a dash, it would be helpful to explain the risk evaluation, possibly in a footnote. This also applies to subsequent tables. | Agreed. The following footnote will be added to Tables 2-9, 2-10, 2-13, 2-14, 2-16, 2-17, 2-20 and 2-21: <i>"Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B."</i> |
| 15. | Table 2-13. The excess cancer risks for lead should not be presented as 0.00. Instead, provide a footnote stating that although lead is a carcinogen, the excess cancer risk was not quantitatively evaluated. | Agreed. The excess cancer risk value of "0.00" will be deleted from Table 2-13 and replaced with a dash notation "". Because lead was evaluated separately, the following footnote will be added to Table 2-13: <i>"Lead is evaluated independently using adult and child lead models because a linear adjustment of the RSL may not be appropriate, see Table 2-25 and Appendix B."</i> |
| 16. | Figures 2-1 and 2-2. Fishing platform users should be included in the CSM. Also, CSMs for the Firing Range and Bulb Slope should also be provided. It is important to identify relevant exposure pathways and receptors to select appropriate screening values, even if all chemicals are screened out. | The fishing platform scenario will be added to the CEMs and all four Upland OU human health CEMs will be included with the BHHRA. |
| General ER | A Comments | |
| 17. | Scope of BERA, Groundwater Seeps, Source control, Upland Soil: Groundwater to Surface Water, storm water and erosional soil / material pathways and associated data should be evaluated and discussed in the risk assessment. These pathways are presented in Figure 3-3, ERA CSM with the exception of storm water, which should be added to the figures where storm water outfalls are present. a. Groundwater, total, dissolved and seep water (some of the data presented in Table 1-7 and 1-8): i. Identify areas where groundwater and seep water are higher than reference, as was done in previous risk assessments ii. Identify areas where groundwater and seep water are above risk based levels using surface water criteria b. All contaminants of interest above risk based criteria (COPCs) in soil or water media need to be identified in the risk assessment. For example, in landfill soil, previous screening identified barium, iron, manganese, zinc, and BEHP above risk based criteria in landfill soil. Arsenic, lead, tributyltin, pentachlorophenol, 1,2,4-trimethylbenzene, isopropylbenzene, n-propylbenzene, di-n-octyl phthalate, 2- methylnapthalene, acenapthalene, phenantharene were detected and identified based on the bioaccumulation pathway with no SLV. COIs below background criteria, and therefore not included in the risk | The stormwater pathway will be added to the CEM (Figure 3-3). The upland-to-river pathways were evaluated at the screening level in the RI Report (URS 2012). As stated in Section 1.0 of the BHHERA (page 1-3), these pathways were not evaluated in the Upland OU Baseline Human Health and Ecological Risk Assessment (BHHERA) report but will be considered in the Upland FS or River FS. a. The groundwater to surface discharge scenario (part of the upland-to-river pathway) was evaluated in the screening level ecological risk assessment (SLERA) presented in the RI Report (URS, 2012). As stated in Section 1.0 of the BHHERA (page 1-3), further evaluation of the pathway will be considered in the Upland FS or River FS. b. As stated in Section 1 of the BHHERA, the baseline RAs build upon the findings of the RI Report and screening level RAs (URS 2012). All contaminants of potential ecological concern (CPECs) identified in the SLERA as warranting further evaluation were evaluated in the Baseline Ecological Risk Assessment (BERA), see responses to specific Questions 26-30. Contaminants of interest (COIs) below background criteria, and therefore not included in the BERA, were identified in the SLERA. |

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| | estimates, should also be identified on this basis. c. Soil PCBs risk estimates are not made in the upland risk assessment, although there are known detections in landfill soil. As noted in (b) of this comment, the reason for PCB exclusion should be clear. This comment should be extended to all detected COIs. | c. As stated in Section 1 of the BHHERA, the baseline RAs builds upon the findings of the RI Report and screening level RAs (URS 2012). All CPECs identified in the SLERA as warranting further evaluation were evaluated in the BERA (see responses to specific Questions 26- 30). Polychlorinated biphenyls (PCBs) were evaluated in the SLERA and were not found to be CPECs for upland receptors (see RI Report Table 12-1, and RI Report Appendix N Tables N-1 to N-4, N-5 to N- 9, N-11 to N-14, N-16 to N-19, N-35 to N-42, and N-43 to N-46). |
| 18. | Scope of BERA, Landfill: The heterogeneity of the landfill composition should be discussed relative to the risk identified in this document. PCBs need to be evaluated in upland soil, including the landfill. | The following discussion on the heterogeneity of the Landfill AOPC composition will be added to the BHHERA uncertainty sections (as new Sections 2.6.6 and 3.5.2.7): "For approximately 40 years (early 1940s until the early 1980s), USACE managed, stored and disposed of waste materials at the landfill in excavated pits or existing depressions. By 1982, the surface of the Landfill AOPC had been capped with soil cover. In 1989, approximately 8-inches of additional soil cover was placed on the Landfill site by the USACE (Hibbs, personnel comm. 2001). Because the waste was buried in separate pits, rather than one continuous pit, the heterogeneity of the landfill composition makes contamination characterize difficult. Although sampling targeted areas known (via historical aerial review or previously exposed areas) or suspected (via electrical resistivity data and seismic refraction data) as having the greatest landfilling activity, risk may be underestimated due to the heterogeneous nature of the landfill." PCBs were evaluated in the RI Report and SLERA (URS 2012) and were not found to be CPECs for upland receptors (see RI Report Table 12-1, and RI Report Appendix N Tables N-1 to N-4, N-5 to N-9, N-11 to N-14, N-16 to N-19, N-35 to N-42, and N-43 to N-46) |

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| Specific ER | A Comments | |
| 19. | Page ES-5 – ES-6: Pesticides: The landfill and sandblast area should be evaluated relative to pesticides and PCBs. Limited data should be presented relative to the in water fish tissue detections. Data gaps should be discussed. The heterogeneity of the landfill in particular should be discussed. | PCBs were evaluated in the screening level RAs presented in the RI Report (URS 2012) and were not identified as CPECs or contaminants of potential concern (COPCs) for human health. Detected pesticides in the Landfill and Sandblast Area AOPCs were evaluated in the BERA. |
| | | As stated in Section 1.0 of the BHHERA (page 1-3), the upland-to-river pathways evaluation will be part of the Upland or River FS; therefore, an evaluation of Upland concentrations relative to fish tissue detections will not be presented in the Upland OU BHHERA. |
| | | The discussion provided in the response to comment #18 on the heterogeneity of the Landfill AOPC composition will be added to the BHHERA uncertainty sections (as new Sections 2.6.6 and 3.5.2.7). |
| 20. | Section 3.2, Receptors and Exposure Pathways and Section 3.5.2.4: A carnivorous mammal should be represented in the risk assessment, assuming a diet of primarily terrestrial species. While a mink receptor is included, this species does not represent a relevant surrogate receptor for this measurement endpoint given that the risk assessment assumes the mink is primarily feeding in the river and not the terrestrial uplands. In order to represent the terrestrial species. Alternatively, a weasel or other similar carnivorous mammals should be included instead of the mink. While the two species are similar in chemical sensitivity, specifying the weasel as the representative upland receptor eliminates confusion regarding the assumed percentage of terrestrial dietary consumption. | The American mink was added to represent a hypothetical upland large mammal, as requested by the ODEQ (Specific Comments 22 on the Final RI), due to the absence of mammals on the island with the exception of rodents and feral cats. The upland receptors and their dietary assumptions, including a conservative percent diet of upland prey for mink (15%), was presented in the Update to Risk Assessment (RA) Work Plan (URS 2014, provided as Appendix C to the Upland OU BHHERA). The weasel was not previously requested by the ODEQ, nor is it part of the Update to RA Work Plan. No revisions to the BERA are necessary. |
| 21. | Section 3.2, Page 3-1: It is unclear what is meant between surface soil, defined as 0 to 1 ft bgs (all AOPCs), and shallow soil, defined as 0 to 3 ft bgs (Landfill and Sandblast Area). It does not appear that "shallow soil", which would incorporate a deeper exposure interval, was included in the risk assessment. | Soil deeper than 3 feet was not evaluated in the SLERA or the BERA. The 0 to 3 foot soil depth interval was established as the ecologically relevant depth for evaluations in Appendix C of the RI/FS MP (URS 2007a) and was evaluated in both the SLERA (see Section 12.3.2.3, page 12-10, of the RI Report; URS 2012) and BERA (see Section 3.2, page 3- 1 and 3-2, of the BHHERA). No revisions to the BERA are necessary. |
| 22. | Section 3.2, Receptors and Exposure Pathways, Figure 3-1: Terrestrial plants and invertebrates should be included in the risk assessment. This particularly true given the metals and other contaminants present in the landfill and sandblast grit area. | As stated in Section 3.2, terrestrial plants and invertebrates were included in the risk assessment (and are shaded in Figure 3-1 indicating their assessment). The shaded note in Figure 3-1 will be revised to say "Organisms in the shaded boxes will be evaluated for risk (representative species are shown for birds and mammals)." |

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| 23. | Section 3.3.3: Bioaccumulation Factors: EPA Estimation Programs Interface (EPI) should be used to select Kow and Koc values for the calculation of partitioning-based BAF values. EPI values for Kow and Koc are used in EPA's Regional Screening Level Tables, and the supporting documentation for these Tables also provides recommendations for selection | The logK _{ow} s used in the BERA are presented in Table J-6 in the Final RI (URS 2012), for which USEPA EPI Suite was used as the primary source. |
| | of experimental or estimated values. Parameters from the Regional Screening Level Tables also form the basis of the calculation of DEQ's human health risk based concentrations. | |
| 24. | Section 3.3, Exposure Assessment, Groundwater to Surface Water: Is the conclusion that groundwater pathways to the river do not pose a risk to aquatic receptors? | As stated in Section 1.0 of the BHHERA (page 1-3), the upland-to-river pathways evaluation was not part of the Upland OU BHHERA and no conclusions regarding it were made in the BERA. The upland-to-river potential pathway will be further evaluated in the FS. No revisions to the BERA are necessary. |
| 25. | Section 3.4.2 and Table 3-17, Toxicity Reference Values for Birds and Mammals: TRVs used in DEQ's guidance, from ORNL (cited last in the hierarchy here) should be selected second after EPA Eco SSLs. It is unclear why Portland Harbor TRVs would be selected. Out of the CPECs listed in Table 3-17 this will impact TRVs selected for mercury, tributyltin, and BHP. | The toxicity reference value (TRV) selection hierarchy was presented in the Update to RA Work Plan (URS 2014, provided as Appendix C to the Upland OU BHHERA). The BERA was consistent with the work plan. No revisions to the BERA are necessary. |
| 26. | Section 3.5.1.1, Landfill AOPC and Table 3-21 Summary: a. Terrestrial Plants: Zinc and lead should have been identified based on the previous risk assessment. b. Birds: Antimony, cadmium, MCPP, dichlorprop and BEHP should have been identified based on the previous risk assessment c. Mammals: Antimony, lead, dichlorprop, MCPP and dibenzofuran should have been identified based on the previous risk assessment d. Terrestrial Plants: Zinc and PCE based on the previous risk assessment e. Soil Invertebrates: Zinc based on the previous risk assessment f. Birds: Antimony and BEHP based on the previous risk assessment | Several of the CPECs (i.e., lead, antimony, BEHP, cadmium, dibenzofuran) were evaluated in the BERA. A few CPECS (i.e., zinc, MCPP, dichlorprop, and PCE) were not retained for further evaluation in the SLERA based upon the weight of evidence evaluation presented in the SLERA risk interpretation section (see Section 12.3.4.3 in the RI Report, URS 2012). No revisions to the BERA are necessary. |
| 27. | Section 3.5.1 and Table 3-21, AOPC Table Summaries: All COIs above risk based criteria should be identified in the risk assessment and added to the receptor specific tables and summaries. | Not all COIs above risk based criteria were recommended for further evaluation based upon the weight of evidence evaluation presented in the risk interpretation section of the SLERA (see Section 12.3.4.3 in the RI Report, URS 2012) or BERA (see Section 3.5.4 in the BHHERA). No revisions to the BERA are necessary. |

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| 28. | Section 3.5.1.2, Sandblast Area AOPC Table 3-21 Summary. Please add additional COCs to the receptor specific summaries based on the previous risk assessment, including antimony, zinc, PCE, BEHP, dibenzofuran and cadmium. | See responses to Questions 26 and 27. No revisions to the BERA are necessary. |
| 29. | Section 3.5.1.3, Pistol Range AOPC Table 3-21 Summary: a. Birds: Zinc should be identified based on the previous risk assessment | As discussed in Section 12.3.4.3.3 of the SLERA (URS 2012), "due to the low toxicity ratio for zinc (hazard quotient [HQ] of 2), absence of protected species of terrestrial birds, and because zinc is an essential nutrient for birds (metabolically regulated), no further evaluation is recommended for zinc for birds." In the SLERA, only lead at the Pistol Range AOPC was recommended for further evaluation in the BERA. See responses to Questions 26 and 27. No revisions to the BERA are necessary. |
| 30. | Section 3.5.1.4, Bulb Slope AOPC Table 3-21 Summary: a. Mammals: Lead should be identified based on the previous risk assessment | The no observable adverse effect level (NOAEL) and lowest observable adverse effect level (LOAEL) HQs were less than 1.0 for lead at the Bulb Slope AOPC for exposure to mammals (see Tables 3-14d and 3-14e in the BERA). No revisions to the BERA are necessary. |
| 31. | Section 3.5.3, Calculation of Site Specific Risk Based Concentrations: a. Reference Area UPLs should not be used in place of risk based concentrations, and the term reference should be replaced with background. According to DEQ guidance, metals should be eliminated if present at background (naturally occurring levels). The full risk should be analyzed if Site concentrations are above background values. Each Site sample should be compared to DEQ's background UPLs for this comparison (http://www.deq.state.or.us/lq/pubs/docs/cu/DebORbackgroundMetal.pdf). If any detection on the Site is above the UPL, the contaminant should be included in the risk assessment. b. Table 1-9 should be replaced with DEQ's background UPL values. | a. The term "background" is used throughout the BHHERA to describe the Reference Area. To be consistent, the term "reference area" in Section 3.5.3 will be changed to "background". Table 3-20 will be revised to show RBCs and Reference Area background UPLs separately, replacing the background values in the RBC columns with "<i>< bkg</i>" and a new column will be added to the right side of the table that shows the Reference Area background UPLs. Screening out of COIs below background was done in the RI and was based on a statistical population-to-population evaluation (see Section 8 of the RI Report, URS 2012). The BERA evaluated the risk to all CPECs carried forward from the SLERA and did not screen out CPECs based on comparison to Reference Area background UPLs prior to the Level III assessment. b. As stated in Section 6.1.5 of the RI Report (URS 2012), the objective of the upland Reference Area was to provide site-specific background concentrations of inorganic COIs in soil and groundwater (URS 2008a). The samples were also analyzed for selected organic analytes to evaluate the potential contribution, if any, of non-site-specific sources of organic COIs to site risk. USACE will continue to use the site-specific background UPLs established in the upland Reference Area (Table 1-9). |

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| 32. | Section 3.5.4.1, Landfill AOPC: Zinc, antimony, MCPP, dibenzofuran should be added to this list. It would also be helpful to indicate if detections of these additional CPECs are co-located with the areas identified in this section. | See response to Questions 26-27. No revisions to the BERA are necessary. |
| 33. | Section 3.5.4.2, Sandblast Area AOPC: Antimony, Zinc, BEHP and dibenzofuran should be added. It would also be helpful to indicate if detections of these additional CPECs are co-located with the areas identified in this section. | See response to Questions 26-28. No revisions to the BERA are necessary. |
| 34. | Section 3.5.4.4, Bulb Slop AOPC: Zinc should be added based on the previous risk assessment. | See response to Question 29. No revisions to the BERA are necessary. |
| 35. | Table 1-1 through 1-5, Statistical Summary for AOPC data: Please show all COIs and not just the COPCs identified in the 2012 RI Report. Plant and invertebrate exceedances of RBCs should also be added to the Figure set, similar to the wildlife CEC locations above RBCs (e.g., Figures 3-6 through 3-9). | Statistical summaries for all COIs in each AOPC are presented in the RI Report (URS 2012). Only the COPCs and CPECs carried forward from the screening level RAs for risk evaluation in the BHHERA are presented in the Section 1 Tables of the BHHERA in order to be clear what specific COIs were further evaluated. |
| | | Contaminants of ecological concern (CEC) locations above plant and invertebrate exceedances of RBCs are not depicted in Figures 3-6 through 3-8 because their RBCs (High screening level values [SLVs]) are lower than the wildlife RBCs, with the exception of chromium for plants at the Landfill and Sandblast Area AOPCs (as noted on Figures 3-6 and 3-7). The chromium RBC for plants is less than the site-specific Reference Area background UPL (28.1 mg/kg), and rather than show background chromium exceedances in the figures, the figures show the wildlife chromium RBC (177 mg/kg) exceedances. The notes on Figures 3-6 and 3-7 will be clarified to say, "* <i>Wildlife RBCs are depicted because all plant and invertebrate High SLVs are lower than wildlife RBCs except chromium. Plant/invert chromium risk is evaluated at background levels and therefore not shown."</i> |
| | | Figure 3-8 does not depict plant and soil invertebrate RBCs because they were not derived (see Table 3-20). As noted on Figure 3-8, "*Plant and invertebrate High SLV HQs < 1." |
| | | Figure 3-9 does depict soil invertebrate RBC exceedances. As noted on Figure 3-9, "*Wildlife and plant RBCs were not derived because |
| | | LOAEL/High SLV HQs < 1." |
| | | No revisions to the BHHERA are necessary. |

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| 36. | Table 3-4, Exposure Assumptions for Shrew, Food and Soil Ingestion Rates: Please use those food and soil ingestion rates calculated for shrew by EPA in the Ecological Soil Screening Level Values instead of using the white footed mouse. Mice and shrew have very different ingestion rates based on metabolism. | The food and soil ingestion rates for the shrew were calculated and established in the Update to RA Work Plan (provided as Appendix C to the Upland OU BHHERA). The food ingestion rate was based on the 'all mammal' equation in Nagy 2001 using the vagrant shrew body weight. As noted in Table 3-4, the white-footed mouse incidental soil ingestion rate was doubled in the absence of a shrew-specific value. USACE will continue to use the established exposure parameters from the work plan. No revisions to the BERA are necessary. |
| 37. | Table 3-16, Terrestrial Plant and Soil Invertebrate Toxicity Benchmarks: a. The low and high terrestrial SLVs are based multiplying soil screening levels by a factor of 5 (note b). This does not have any basis in toxicology (dose response). The original SLVs ("low SLVs) should be used to assess risk. This issue was also included in DEQ's Sept. 2011 comments (Specific Comment 24 and 25) "A Q=5 should not be used for plant and invertebrate SLVs based on MATCs (Maximum Acceptable Toxicant Concentrations). A Q of 5 can be used for SLVs served with a no observed adverse effect level (NOAEL) where threatened and endangered species are not present". | a. In the SLERA (URS 2012), risk to soil invertebrates and plants was evaluated at the individual level and conservatively based on Q=1, as per ODEQ's request. Evaluating risk to soil invertebrates and plants at the individual level was not warranted for the BERA, as there are no threatened and endangered (T&E) plant or invertebrate species in the Upland OU. As stated in the Update to RA Work Plan (Appendix C in the Upland OU BHHERA), the goal of the BERA is estimating risk to soil invertebrate and terrestrial plants at the community level (i.e., it is appropriate to assess risk at Q=5). In the BERA, the original SLV from the SLERA was conservatively used as the low TRV. In the absence of a high TRV, in some cases the low TRV was multiplied by 5, which results in a high TRV based on Q=5 in order to assess risk at the community level. The BERA was conservative by presenting risk at both Q=1 AND Q=5. Although many of the USEPA ecological soil screening levels (Eco-SSLs) for plants and invertebrates are derived from maximum acceptable threshold concentrations (MATCs), the Eco-SSLs are still considered to be used as cleanup levels" (USEPA 2005). Therefore, use of both Q values is appropriate for a BERA that is being conducted to support an FS. |
| | b. Additionally, invertebrates and plants are not mobile, and therefore the 95% UCL should not be used as an exposure point concentration. This issue was also included in DEQ's Sept. 2011 comments (Specific Comment 23 and 25). Specific Comment 23: <i>"The appropriate exposure point concentration for plants and invertebrates is point by point and therefore frequency of detection over a larger area does not apply."</i> Specific Comment 25: <i>"For plants, aquatic, and soil invertebrates that are immobile or nearly immobile, the maximum</i> | b. In the SLERA (URS 2012), the maximum detection as the EPC (and subsequent sample by sample analysis) was used for evaluation of the community receptors. Based on the results of the evaluation, specific CPECs were carried forward into the BERA. However, as stated in the Update to RA Work Plan, the 95% UCLs calculated for soil in the RI were used as the EPCs for all terrestrial receptors in the BERA, including plants and invertebrates due to the lack of protected plant and invertebrate species in the Upland OU and the goal of estimating |

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| | detected concentration (MDC) in soil or sediment is the appropriate value for comparison" to the screening level value". It is recommended instead to show maps exceeding criteria for each | risk to these receptor groups at the community level. Protection of the individual (i.e., use of maximum detection or sample by sample analysis) is not the goal of the BERA phase. |
| | sample location. >20% of the AOPC exposure area exceeding SLVs may be one way to indicate an unacceptable risk. c. Background UPLs are shown here, whereas the footnote states "reforence one 05% UPL concentrations are shown". See provides | c. See response to Question 31. |
| | "reference area 95% UPL concentrations are shown". See previous comments regarding the use of appropriate <u>DEQ background UPLs</u> . | No revisions to the BERA are necessary. |
| 38. | Table 3-17, Toxicity Reference Values for Birds and Mammals: The LWG-derived TRVs (footnote b) should be placed last in the hierarchy of sources. For example, this impacts TRVs selected for mercury and BEHP where the primary TRV source can be found in Sample 1996 (note "c" in the footnotes). | See response to Question 25. No revisions to the BERA are necessary. |
| 39. | Table 3-18a-d, Terrestrial and Invertebrate HQ: EPCs should be sample by sample. Summary statistics on how many samples are above HQs along and HQ magnitude should be presented. Risk displayed as an average concentration over the "low SLV" should be interpreted as significant exceedances given the large exposure area. | See responses to Question 37. No revisions to the BERA are necessary. |
| 40. | Figure 1-4, Upland Operable Unit Overview: The Sandblast Area, Bulb Slope, Landfill and Pistol Range AOPCs are shown as a subset of the larger Upland Operable Unit. Is the rest of the upland operable unit considered outside the locality of the facility? | The rest of the Upland OU is considered outside the locality of the risk assessments, but not outside the locality of the Bonneville Dam complex facility. The four AOPCs were evaluated based on previous site investigations and, as such, were the focus of the upland evaluations in the RI (including screening level RAs) and the BHHERA. See Section 3.0 of the RI (URS 2012) for further details on the site background, previous investigations, and establishment of the upland AOPCs. |
| 41. | Figure 3-2, Landfill AOPC CSM: Root/Dermal uptake for soil invertebrates should be complete at 1-3 ft, similar to 0-1ft. | The CEM (Figure 3-2) will be revised to indicate that the root/dermal updated for soil invertebrates is a complete pathway for the shallow soil. It was evaluated in the BERA. |
| 42. | Figure 3-3, Sandblast Area AOPC: Exposure to sandblast grit itself, not just soil, should be complete to all receptors. | As stated in Section 1 of the BHHERA, both Sandblast Area AOPC soil subgroups (one noted as having higher soil and the other as higher sandblast grit) were utilized as soil in the screening-level RAs in the Final RI and in these baseline RAs. In Figure 3-3, the box for soil at 0-1 ft bgs will be revised to " <i>soil and sandblast grit</i> ." |
| 43. | Figures 3-6 through 3-8 . Please revise AOPC figures that are inclusive of all locations above risk levels. Risk levels for all COPCs (as indicated by above comments) should be used. For example, the Figure 3-8 AOPC boundary of the pistol range shows two samples high in lead outside the boundary. | See response to Question 35. The AOPC boundary of the pistol range will be revised in Figure 3-8 to include the two locations north of the backstop that exceed RBCs. |

References:

DTSC, 2014. Human Health Risk Assessment (HHRA) Note 2. Available at https://www.dtsc.ca.gov/assessingrisk/humanrisk2.cfm.

- USEPA. 2002a. Supplemental Guidance for Developing Soil Screening levels for Superfund Sites. OSWER 9355.4-24. Solid Waste and Emergency Response. December.
- USEPA. 2002b. Guidance for Comparing Background and Chemical Concentrations in Soil at CERCLA Sites. USEPA 540-R-1-003. OSWER 9285.7-41. September.

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- USEPA. 2005. Guidance for Developing Ecological Soil Screening Levels Revised Draft. OSWER Directive 9285.7-55, USEPA, OSWER, February.
- USEPA. 2015a. Regional Screening Levels (RSL) for Chemical Contaminants at Superfund Sites. RSL Table Update. May.
- USEPA. 2015b. RSLs-User's Guide (webpage). USEPA website available at: <u>https://www.epa.gov/risk/regional-screening-levels-rsls-users-guide-november-2015</u> (accessed 03-22-2016). November.
- URS. 2007. Remedial Investigation/Feasibility Study (RI/FS) Management Plan (MP), Bradford Island, Bonneville Lock and Dam Project, Cascade Locks, Oregon. September.
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- URS. 2012. Upland and River Operable Units Remedial Investigation Report. Bradford Island, Bonneville Dam Forebay, Cascade Locks, Oregon. June.
- URS. 2014. Update to Risk Assessment Work Plan for the Upland OU. Bradford Island, Bonneville Dam Forebay, Cascade Locks, Oregon. April.

Hibbs, Don. 2001. USACE. Personal communication.

APPENDIX E

Yakima Nation Comments on Draft Upland/River OU BHHERAs

| Comment Number | Comments by Yakama Nation as forwarded to the US Army Corps of Engineers by letter, dated 28 January 2016 | US Army Corps of Engineers Response/Action Taken |
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| Concerns a | bout the Remedial Investigation | |
| 1. | 1. Contaminant Sources a. The RI does not adequately evaluate historic PCB use with respect to potential releases. The Yakama Nation requests the following information: i. The history of PCBs use at the Bonneville Dam complex that may have led to releases, besides in-water disposal of transformers. For example, historically many dams' power generating equipment was designed to slowly leak (PCB- containing) hydraulic oil; PCBs have also been used in the past as a component of paints at fish hatcheries, dust control oils, etc. | Early work on this project included record searches, review of aerial photographs, interviews of employees, etc. We are confident that any historical releases to the Bonneville Dam forebay have been identified. No attempt was made to broaden the investigation to include the Bonneville Dam project, as a whole (see Note 1, below). The downstream sediment samples, collected in 2008, do not indicate that historic or unknown releases of PCBs are impacting resources below the dam; see 3.b.viii, below). |
| | | No revisions to the Baseline Human Health or Ecological Risk Assessments (BHHRAs) are necessary. |
| 2. | Contaminant Sources The RI does not adequately evaluate historic PCB use with respect to potential releases. The Yakama Nation requests the following information: Efforts made to eliminate PCB use in Bonneville Dam power-generating equipment and any other sources. | Efforts made to eliminate PCB usage or impacts from legacy PCB usage at Bonneville Dam are not a part of the scope of this CERCLA project; however, contact information for the offices that are actively working this can be provided (see Note 1, below). |
| | | No revisions to the BHHRAs are necessary. |
| 3. | Contaminant Sources Upland: In general, the information supports the delineation of upland contamination in the former disposal areas. However, it is unclear how the areas of concern were initially identified to begin delineating the site uplands. It appears upland investigations may have been limited to a decision unit (i.e. the area surrounding the landfill) rather than a comprehensive site investigation. Outside of the eastern portion of the Bradford Island Upland OU, please provide a summary of evaluations conducted to identify recognized environmental conditions on the remaining Bonneville Dam complex and | In 1992, Portland District conducted its first Environmental Review of Government Operations (ERGO) assessment at Bonneville Dam. The intent of the assessment was to make sure that all applicable environmental laws were being followed and to identify any instances of non-compliance. In that 1992 report, the landfill on the eastern tip of the island was noted. A recommendation to take samples of soil in the landfill (completed in 1996) led to a determination that contaminants were present. |
| | surrounding nearshore mainland. | Work was then initiated to determine the nature and extent of any release of contaminants at/from the landfill. As the investigation moved forward, additional sites were identified, including the in-water location where disposal of capacitors had occurred, the bulb slope, the sandblast area and the pistol range (see 1.a.i., above; some of these additional sites were identified during the literature searches, some by interviews, some by our contractors during the course of working at the site, etc.). A decision was made to complete one Remedial Investigation for all of the sites as opposed to conducting distinct investigations for each one. |

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| 4. | Comments by Yakama Nation as forwarded to the US Army Corps of Engineers by letter, dated 28 January 2016 Contaminant Sources Civer: There appears to be a data gap regarding the location or source of PCBs, which will make remedy selection difficult. The older data (in Appendix G of the RI) include samples that show very high PCB concentrations in sediment samples from the forebay prior to any removal actions. The highest PCB concentrations tended to be associated with coarse sediments, with high fractions of gravel and coarse sand. High concentrations of PCBs in coarser sediments indicates the release of product directly to the sediments, rather than through upland releases. The data are inconclusive as to whether additional sources exist and further evaluation is needed. For example, no samples were collected in the deeper forebay in front of the dam or farther upstream. | US Army Corps of Engineers Response/Action Taken Once it was known that the release associated with the in-water disposal of capacitors had crossed state lines (PCBs in sediments on the north side of the forebay in Washington state), the decision was made to move under CERCLA (as opposed to working under the State of Oregon's Voluntary Cleanup Program.) No revisions to the BHHRAs are necessary. We agree that the high concentrations of PCBs in coarse sediment indicate a direct release of product. We attribute this to the release of PCB laden oils from the capacitors. Each of the 6 inerteen capacitors were capable of holding 2-3 gallons of oil. And, PCBs were present in the oil at a 20% concentration. Also, oil with PCB residue was released through the sandblast area outfalls (a known oil spill). We could find no record of any other direct release of product (see 1.a.i, above). And, the grid systems devised for the equipment removal in 2002 and the sediment removal in 2007 give us confidence that no additional equipment was missed on the river bottom. So, while we agree with your assessment that the high PCB levels are consistent with a release of product, as opposed to run off, we believe that the current condition of the site is consistent with the sources already identified. To summarize, we do not believe there is another, as yet, unidentified PCB source. |
| 5. | 2. Contaminant transport and migration: a. If PCB-laden product was released directly to sediment, it would be much denser than the river water and would flow into porous (coarse) sediments, cracks in bedrock, and down slope along any retaining barrier such as bedrock. The data analysis presented in the 2014 River OU Data Evaluation Tech Memo (Figure 3-1) shows a number of locations associated with the debris piles where the PCB concentrations exceeded 1,000 micrograms per kilogram (ug/kg) in the most riverward sample, indicating that contaminants have likely migrated farther beyond the area delineated. | No revisions to the BHHRAs are necessary. In 2003, seventeen depositional samples were collected from several areas around the forebay. A few of these are in deep water and in front of the powerhouses and spillway (as close as we could safely get). The samples in deep water were difficult to obtain (two locations were unsuccessful due to depth and strong currents). The majority of samples were non-detect for PCBs (Aroclor 1254). Three of the samples had low detections (the highest is 2.9 ug/kg). Our earlier discreet sampling efforts did not find a bright line for the extent of the release. Detection of PCBs at low levels is found in front of the spillway (on both the north and south sides), on the south side of Bradford Island, and near the west end of Goose Island. Levels of contamination are lower as you move away (following the currents) from the original sources of the release. In 2012, it was determined that |

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| | | enough information was already available (see Remedial Investigation report, June 2012) to determine that risks to the environment and human health were present and that further sampling would not further inform the remedial options that would be considered in the Feasibility Study. (Note: sampling and analysis takes approximately 2 years to complete from start to finish.) |
| | | We agree that further sampling may be warranted, or even preferred (to delineate the exact footprint for the remedial action.) But, additional samples would not further inform the risk assessments or feasibility study alternatives. |
| | | No revisions to the BHHRAs are necessary. |
| 6. | 2. Contaminant transport and migration:b. The Yakama Nation still requires hydraulic model information requestedDecember 15 and 28, 2015. The most recent data, indicate higherconcentrations of PCBs are located toward the eastern end of Bradford Island, | We are working to provide all requested materials. We agree that the eastern tip of Bradford Island, where Debris Pile #1 was located, is the most impacted location. |
| | possibly residual contamination deposits, but also possibly contributions of contaminated sediments carried along shore. | No revisions to the BHHRAs are necessary. |
| 7. | 3. Site delineation: a. The adequacy of the sampling to define the extent of contamination, both in space and over time, is a major concern. Without adequate site characterization and delineation, the effectiveness of any remedy is uncertain. | We believe that enough sampling has been done to determine that the forebay is the 'extent' of the release. And, that the north side and eastern tip of Bradford Island are the most impacted areas within the forebay. No other sources were found after 15 years of direct sampling and investigation. Also, the most contaminated locations are consistent with the location of the debris piles and the outfall. Therefore, the data we currently have is adequate for the purposes of conducting a baseline risk assessment and to select a remedial alternative. |
| | | No revisions to the BHHRAs are necessary. |
| 8. | 3. Site delineation: b. River: i. It is unclear how the River OU eastern delineation line was drawn, since there are no data upriver of this line besides the reference area (approximately 2 miles upstream). Based on RI information and the more recent River OU RA, Figure 2-7, sediment and tissue samples indicate HH risk extends across the entire unit, including locations at upriver boundary. | The eastern boundary for the In Water OU was determined with hydraulic models. Currents can flow upstream under certain operational scenarios (spillways closed). The western end of Goose Island is the most upstream point that the hydraulic model shows currents extending. The eastern end of Goose Island was chosen for the upstream boundary to be conservative. Although a few bass have been caught along Goose Island with high PCB concentrations, sediment and clam data from the 2008 sampling event are consistent with the setting of this eastern boundary (i.e. no elevated levels of contaminants east of the western tip |

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| | | of the island). |
| | | No revisions to the BHHRAs are necessary. |
| 9. | 3. Site delineation: b. River: ii. Goose Island Slough was constructed in 1989-1993 by removing a portion of the southeastern tip of Bradford Island. One of the most heavily impacted sediment locations is at the eastern tip of Bradford Island. At the time of construction this material was not evaluated for the presence of contamination; however, limited sampling of slough sediments and fish tissue post-construction has been completed. This data are inconclusive about whether or not contaminated sediments were re-located into this area. The Yakama Nation requests discussion of the hydrodynamic model and any older data that indicate the potential for borrowed sediments to have become contaminated. | To place material in water in the state of Oregon requires a 404(b) permit and a 401 permit. These permits would only have been issued if the soils / sediments met State and EPA guidelines for in-water placement. We can look for and share those records. No revisions to the BHHRAs are necessary. |
| 10. | 3. Site delineation: b. River: iii. The information available to date are not adequate to narrow down the focus area for cleanup decisions in the FS. The highest PCBs concentration in fish tissue are found in the forebay and Goose Island areas. In addition, very highPCB concentrations were observed at the upriver site boundary (e.g., 69,276 µg/kg PCB, equivalent to 7% PCBs in fish tissue sample 68), again leaving questions as to whether high or higher concentrations might be observed if fish from farther upstream were sampled. The higher concentration at Goose Island should be further considered to determine whether there is a proximate or upstream source of PCBs. | See Note 2, below. Also, sampling at Goose Island was completed in 2009 at the request of the TAG, after some Goose Island bass (fish caught in the slough in 2008) were found to have high PCB levels. No new source areas were found. It is our opinion that fish are an inadequate media for tracking down new sources of contamination. No revisions to the BHHRAs are necessary. |
| 11. | 3. Site delineation: b. River: iv. In the recent data (2011), the concentrations of PCBs in fish, clams, and sediments were still high along the north side of Bradford Island, with concentrations in all media increasing toward the eastern end of the island. The sampling at Bradford Island was limited to the north side of the island; however, it did not bound the extent of the high concentration area to the south around the end of the island, nor riverward. One of the debris piles was located near the eastern end of the island and river current flow is around the tip of the island at least some of the time (per conversation with USACE, December 15, 2015), providing a reasonable basis for residual high concentrations of PCBs to be found in this location. Considered together with the potential for heterogeneity in the PCB distributions, these data are simply too limited to confidently delineate the present distribution of PCB contamination. | The 2011 sampling event was done to supplement the 2008 sampling results. In 2008, samples were not taken in the removal areas (at the time, we did not believe there would be enough sediment present to obtain samples because the sediment removal action had just taken place a few months earlier). Taken together, the 2008 and 2011 sampling data adequately describes the forebay in the shallower shoreline areas (within 150 feet of the shoreline). Deeper areas, beyond 150 feet, were not sampled during this timeframe. Some samples were taken in 2003 (see 2.a., above). These samples were non-detect or had very low detections for PCBs. Also, these areas present a safety risk to divers and they are not normal habitat for bass, sculpin, etc. For these reasons, no additional deep water samples were taken in 2008 and 2011. |

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| | | No revisions to the BHHRAs are necessary. |
| 12. | 3. Site delineation: b. River: v. USACE assumes the primary source of residual PCBs is located in cracks of the river bottom disposal areas. The Yakama Nation requests information on direct support exists for this assumption. | We can provide the diver logs from the 2011 sampling event. They note that surface sediment was not available for collection and that they had to get sediment material from the cracks and crevices in the rocky river bottom. |
| | | No revisions to the BHHRAs are necessary. |
| 13. | 3. Site delineation: b. River: vi. The use of organisms with smaller home ranges (sculpin, crayfish, clams, periphyton and macroinvertebrates) for evaluation of ecological receptors might be more conclusive. | We prepared a comprehensive Work Plan (2007) for sampling, including media to be utilized. We agree that smaller home range organisms are desirable, but media at all levels of the food chain are required for a complete risk assessment. |
| | | No revisions to the BHHRAs are necessary. |
| 14. | 3. Site delineation: b. River: vii. The Yakama Nation requests information to support USACE's conclusion that sediments in general, and especially contaminated sediments, do not accumulate behind the Bonneville Dam in the forebay. | High velocity areas are not depositional. Hydraulic models will show that the area in front of the spillway is a high velocity area. Samples do exist on the Oregon side of powerhouse 1. These samples show non- detect for PCBs. No sampling was ever contemplated in front of powerhouse 2. Hydraulic modeling does not support the idea that sediments could reach this area. |
| | | No revisions to the BHHRAs are necessary. |
| 15. | 3. Site delineation: b. River: viii. We are concerned about the adequacy of the sediment investigation downstream of the Bonneville Dam. The Yakama Nation requests information on how the 4 downstream sediment samples listed in the RI were selected to evaluate downstream sediment impacts. In addition, we request all historic downstream data not used in the RI. | Hydraulic modeling was utilized to determine the most likely depositional areas downstream of the dam. Six locations (not 4) were sampled (see RI, page 8-10) for more on this topic. No other sampling events have occurred below Bonneville Dam. No revisions to the BHHRAs are necessary. |
| 16. | 3. Site delineation: b. River: ix. Older data: 1. Older data should be considered. The Yakama Nation disagrees with USACE that including historic data will add uncertainty to the RI (USACE response to Oregon Department of Environmental Quality comments, 2011). Because the forebay is a dynamic environment, understanding where contamination has come to be located over time will inform a cleanup selection and design. Not considering this information in the conceptual site model adds greater uncertainty. | At the request of ODEQ, most of the older data was included in the In Water baseline risk assessment. Please see the In Water OU Technical Memo (June 2014) for a full explanation of what data was added to the risk assessment data set. No revisions to the BHHRAs are necessary. |

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| 17. | 3. Site delineation: b. River: ix. Older data: 2. For example, these older data show that the PCB contamination was heterogeneously distributed in the sediments (e.g., Figure 3-1 of 2014 Data Evaluation Technical Memorandum, River OU), indicating that a high frequency sampling was necessary to identify the "hot spots" at that time. There is no reason to assume that the same heterogeneity is not present today, leaving the possibility that areas with higher contamination have just been missed in the limited recent sampling. | All sampling done after 2003 uses a compositing method. Up to 10 discreet samples are collected within a 50 foot gridded area and combined for analysis. We agree that discreet samples would give wildly varied results, which is why we started using the compositing method. No amount of sampling will be enough to locate every hot spot on the river bottom. But, we believe enough sampling has already been done to tie the hot spots to the debris piles and the outfalls (our PCB sources). No revisions to the BHHRAs are necessary. |
| 18. | 3. Site delineation: b. River: ix. Older data: 3. RI, appendix G (historical data not used) does not include a map or adequate location information. We request location information for this data. | See 3.b.ix.1, above. The RI, Appendix G no longer reflects the data used in the RA. Please refer to the In Water OU Technical Memo (June 2014) for more up to date information. No revisions to the BHHRAs are necessary. |
| 19. | 3. Site delineation: b. River: ix. Older data: 4. We understand that past efforts were made to sample clam tissue in the forebay close to the dam (early 2000s). We request information on this sampling event. | In 2003, sediment and clam tissue was collected (over 120 locations). However, the clam tissue samples were never analyzed. After 3 years, they were discarded. No revisions to the BHHRAs are necessary. |
| 20. | 3. Site delineation: c. Upland: i. The groundwater VOC plume needs further evaluation. Fairly high concentrations of volatile organic compounds (VOCs) were observed, as were apparent groundwater plumes of those substances. In general, VOCs do not accumulate in sediments or biota and are not highly toxic to aquatic biota. However, the VOC plumes could provide a vector (solvent) for the transport of other organic contaminants. | While groundwater data from borings delineated a PCE plume in 2004 with concentrations greater than 10 times the SLV, subsequent sampling in 2008 suggested that the mass of VOCs available to leach to groundwater is decreasing over time. The breakdown products of PCE (1,1-DCA, cis-1,2-DCE, and vinyl chloride) were also present in groundwater but at generally lower concentrations and with fewer SLV exceedances. |
| | | The hypothesis that the groundwater VOC plume can function as a vector for transport of other organic contaminants is legitimate but is not reflected in other groundwater and seep data. From data collected in support of the RI, limited or no detection of contaminants, including butyltins, herbicides, pesticides, PCBs, PAHs, and SVOCs, were identified in groundwater and seep water. |
| 21. | 3. Site delineation:c. Upland: | No revisions to the BHHRAs are necessary. Surface water was evaluated in the screening level ecological risk |
| | ii. The Yakama Nation requests information on how the FS will address | assessment and was not found to pose unacceptable risk. Surface water |

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| | groundwater to surface water contaminant contributions. This pathway was not carried forward into the RA for further evaluation, despite toxicity criteria exceedances. | was also evaluated as an exposure medium in the baseline human health risk assessment and was found to not pose unacceptable risk. Any contaminant contribution from groundwater to surface water is considered negligible given the results of the risk assessments. |
| | | No revisions to the BHHRAs are necessary. |
| 22. | 3. Site delineation: c. Upland: iii. The FS cleanup alternatives should still include preventative measures to eliminate current and future transport of upland soil and groundwater contamination to the river. In general, upland risks are limited in comparison to river OU PCBs and the RI reasonably discussed the possible inputs from the upland sites to the river, but the quantitative impact of those inputs has been poorly characterized, e.g., associated with slope failure. It also seems very unlikely that the upland sites are the source of the high PCB concentrations observed in the river. | Quantitative characterization of a landfill is not possible. Mass wasting of soil has been confirmed by a recent geotechnical report. Agree that the Upland Landfill alternatives need to address upland soil. No revisions to the BHHRAs are necessary. |
| Comments | on the Upland and River OUs Draft Risk Assessments | |
| 23. | 1. River and Upland Combined: a. With respect to these limited or focused RAs, observations are prefaced by noting that there are risks posed by a number of substances at this site including metals and a variety of organic chemicals. Although the risks posed by these substances are lower than the risks posed by the PCB contamination, the risks from all substances should be identified. | Risks, both cancer and non-cancer, for all identified COPCs and CEPCs were calculated in the baseline risk assessments. These calculated risks are presented in Tables 2-6.1 through 2-12 and 3-15 through 3-19 in the River OU baseline RA and Tables 2-9 through 2-25 and 3-11a through 3-19d in the Upland OU baseline RA. |
| | | No revisions to the BHHRA are necessary. |
| 24. | River and Upland Combined: Tribal exposure scenario: USACE made assumptions about future tribal use to estimate risk at Bradford Island and we appreciate effort to use conservative assumptions in its RA; however, Yakama Nation reiterates its request to have further dialogue on this issue. | We would welcome further dialogue. No revisions to the BHHRA are necessary. |
| 25. | 2. River OU: a. Lipid normalization should be done when comparing concentrations among species. Similarly, for the bass, age differences should be considered as well. It was not clear when the River OU Baseline Ecological Risk Assessment (BERA) states that the concentration in fish tissue were substantially greater than in clams or crayfish whether those measurements were lipid normalized. Lipid content has a major control over the concentrations and the fish have greater lipid content than the shellfish. | Lipid normalization was conducted for all calculations. As stated in the in River OU baseline RA, the median site-specific clam, crayfish, sculpin, and bass lipid contents are 2.6%, 0.73%, 4.1%, and 3%, respectively (page 3-5, River OU Baseline RA). These values were used calculate relevant tissue concentrations. No revisions to the BHHRA are necessary. |

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| 26. | 2. River OU: b. For the dietary uptake by mink it would be useful to discuss in the uncertainty section the effects of assuming the mink ate only the most contaminated food to demonstrate the possible upper limit of the risks. | The results for the mink are likely a conservative, over estimate of risk in both the River and Upland OU baseline RAs. Excerpts from the River OU baseline RA are provided below for reference. Based on the assumed dietary fractions identified in the baseline RA, the EPC (95% UCL on the mean) was used to calculate CPEC exposure. Use of the 95% UCL on the mean is in accordance with the most recent USEPA guidance regarding statistical methodology to be used in EPC estimation and more conservative than ODEQ's guidance. Because of this conservative methodology, USACE feels a discussion regarding the possible upper limit of risk is unwarranted. Several sources of information on the mink's diet were consulted while developing the response to ODEQ's Specific Comment #22 to the Final RI (URS 2012), and the consensus is that mink's diet primarily consists of crayfish, fish, and other aquatic-related prey. Typically, 10% or less of their diet is comprised of terrestrial prey (e.g., birds and small mammals) (USEPA 1993b, 1995). Although the Upland OU BERA assumed 15% small mammals for the upland RA, mink were conservatively assumed to have a dietary composition of 100% prey from the River OU for this River OU BERA. This was done because it is likely that mink preferentially use the River OU (i.e., permanent water source, riverine habitat) rather than the Upland OU to forage. Therefore, the uncertainty that risk may have been underestimated in the River OU BERA is minimal and unlikely to impact risk management decisions in the FS. In addition, although the mink was assumed to consume more the one type of riverine food item (i.e., crayfish, sculpin, and bass), the mink's diet was assumed to be comprised of each equally (i.e., 33.3% each). In actuality, mink dietary composition is likely to fluctuate with availability, season, and, potentially, animal preference. Additionally, in instances when an CPEC was not analyzed in all three tissues, the dietary composition was adjusted for that particular analyte based on available ti |

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| 27. | 2. River OU: c. Similarly, since there is uncertainty in the biomagnification factor (BMF) it would be helpful to discuss the range of possible BMF values compared to the resulting risk estimate. We request calculations using these BMF values to determine if there's a change in risk, both cancerous and non-cancerous. | The BMF values selected for the baseline RA were ODEQ's default values. As stated in the baseline RA, the study for which the default eagle BMFs are from (Buck 2004) noted that the PCB BMFs varied "quite markedly" among the Columbia River segments evaluated (i.e., ranged from 90 to 155) and used prey fish tissue from a wider range of prey items than the assumed dietary assumptions for this BERA (i.e., 100% bass). The use of the default BMFs, which showed variations among the study river segments from studies on which they are based, introduces a level of uncertainty that is hard to quantify and may over- or underestimate risk. However, combined with the dietary assumption of 100% bass (with highest detected CPEC concentrations), this likely conservatively skews the uncertainty toward overestimation. Given that the risk results provide a likely overestimation of risk, USACE feels a discussion regarding the possible upper limit of risk is unwarranted |
| 28. | 2. River OU: d. Table 3-13. It would be helpful to include the maximum concentrations observed in the reference areas in the table since the upper prediction limit (UPL) can exceed the maximum if the sample size is small and variable. | No revisions to the BHHRA are necessary. It is generally believed that the UPL is a more statistically reliable value and a better representation of the data set, as compared to the maximum detected value. However, maximum concentrations can be found in the risk assessments, Tables 1-1 through 1-8 for the River OU baseline RA and Tables 1-1 through 1-9 for the Upland OU baseline RA. The tables relevant to the reference area samples present both the UPL and maximum detected value. No revisions to the BHHRA are necessary. |
| 29. | 2. River OU: e. In comparison to data from the Portland Harbor BERA, the fish tissue no-observed-adverse-effect concentrations (NOAELs) used in the Bradford Island BERA are similar, but the Portland Harbor lowest-observed-adverse-effect levels (LOAELs) are usually less than 3 times the NOELs, rather than the 5 times used as the default in the BI BERA. For PCBs, the NOELs are virtually the same (Portland Harbor -0.42 milligrams/kilogram wet weight (mg/kg ww) versus Bradford Island -0.43 mg/kg ww), but the Portland Harbor LOEL is 0.93 mg/kg ww verses 2.2 mg/kg ww. For PCBs and other substances using the lower LOELs from Portland Harbor would increase some of the LOEL-based HQs, better characterizing the risks posed by those substances. | No revisions to the BHHRA are necessary. During a May 20, 2015 meeting between USACE and ODEQ, it was agreed ODEQ toxicity reference values (TRVs) would be the first choice in the hierarchy of TRV selection. This was done in response to some of ODEQ's expressed concerns regarding the TRVs utilized in the Portland Harbor RA. As a result, some TRVs results in both higher and lower characterized risks. Ultimately, PCBs are still identified as a COC/CEC with unacceptable risk warranting remedial action. No revisions to the BHHRA are necessary. |

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| 30. | 2. River OU: f. USACE needs a better data set to determine fish tissue contaminant concentration trends. Yakama Nation does not agree that PCB concentrations are decreasing in post-removal (2011) fish tissue. The river OU RA presents information that the highest concentrations were found in young fish and that the sampling efforts had a low sample size and low sample frequency. | Monitoring fish tissue concentrations will almost certainly be part of the recommended alternative. But, additional sampling is unnecessary to complete the In Water FS. No revisions to the BHHRA are necessary. |
| 31. | 3. Upland OU BERA: Erosion of the landfill at the north bank shoreline of Bradford Island, as identified in the 2015 slope stability analysis, is an ongoing concern and should be addressed in FS cleanup alternatives. | Agreed. This is being addressed in the Upland FS. No revisions to the BHHRA are necessary. |
| Global Con | nments | |
| 32. | Reference Area: The appropriateness of the reference area data is a concern and should be re-evaluated. It is close to the site (i.e., within small mouth bass home range of approximately 0.7 to 1 mile) and under the direct influence of multiple releases from waterfront cleanup sites with confirmed and suspected contaminant sources also found at Bradford Island, including: | Sample results from all reference area samples are statistically similar, no matter how close or how far from the cited locations of concern. Also, please note that approximately 8 of the reference area bass were actually caught several miles upstream (because they could not catch enough in the actual reference area). The results in these bass are comparable, if not slightly elevated, in PCBs, when compared to the other bass. We do understand the concern, but the sample data from the reference area does not indicate a problem. No revisions to the BHHRAs are necessary. |
| 33. | and Cascade, OR. These facilities have existing discharge permits to pollute the Columbia River. 1. Reference Area: b. It is inappropriate to eliminate COCs and CECs due to similar reference area (background) concentration ranges when the reference area is located in an | For the River OU reference area, the upstream reference area was selected based on modeling results characterizing the upstream extent of the river flow reversal caused by the powerhouses and the spillway that |

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| | industrial area directly influenced by releases from similar contaminant sources. Specifically mercury and arsenic (mentioned in Section 2 text) are of concern. Evaluation of these contaminants for elimination as a COC or CEC should be risk-based only. | could transport impacted sediment back upstream. The reference area was found to be statistically different from the source area. This reference area was subsequently used for comparative analyses in the Remedial Investigation and baseline RAs. |
| | | For the Upland OU reference area, the location was selected because it was upgradient of and unaffected by the site related waste handling activities. The reference site was also found to have samples that generally reflected background or ambient concentrations of all COIs. Lastly, the reference area exhibited similar physical soil characteristics relative to the soil sampled in the four AOPCs in the Upland OU. Because the reference area exhibits these characteristics, USACE believes it is appropriate to use the reference area as site background concentrations and apply these concentrations when determining COCs and CECs. CERCLA guidance states that it is generally not feasible to set cleanup levels below background and thus can be used for helping to screen contaminants for risk management purposes. For mercury and arsenic, the 95% reference UPL (0.06 and 5.5 mg/kg, respectively) are either in line or much lower the Oregon DEQ's regional background values for inorganics in soil. No revisions to the BHHRAs are necessary. |
| 34. | Reference Area: Yakama Nation is concerned about USACE's stated intentions to use reference area data as background values. It is inappropriate to set cleanup levels to a background value that is directly influenced by releases from contaminant sources. | As stated in response to Global Comment 1.b., the reference areas for the River and Upland OU were identified and justified with empirical information identifying the reference areas as statistically different from the source areas and justifiable as reference locations. There is no information supporting the idea that the reference areas are directly influenced by releases from contaminated sources. Further, the baseline RAs followed CERCLA guidance when developing and applying background concentrations to risk management decisions. "Under CERCLA, cleanup levels are not set at concentrations below natural background levels. Similarly, for anthropogenic contaminant concentrations, the CERCLA program normally does not set cleanup levels below anthropogenic background concentrations" (Reference: US EPA, 2002. Role of Background in the CERCLA Cleanup Program. OSWER 9285.6-07P). As such, USACE believes it is appropriate and in line with guidance to use the reference area to represent anthropogenic |

| Comment Number | Comments by Yakama Nation as forwarded to the US Army Corps of Engineers by letter, dated 28 January 2016 | US Army Corps of Engineers Response/Action Taken |
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| | | background. No revisions to the BHHRAs are necessary. |
| 35. | Reference Area: Comparisons of site data to those from reference area is principally a risk management issue. The presence of similar concentrations elsewhere does not reduce the risks at the site. | USACE agrees that in some cases deferring to background concentrations for select contaminants will not reduce risks to <i>de</i> <i>minimus</i> levels. This is because <i>de minimus</i> concentrations fall below background concentrations. It is infeasible to set preliminary remedial goals below background. This strategy is in line with both current CERCLA and ODEQ guidance regarding the consideration of background contamination concentrations. |
| 36. | 2. CEPCs and COPCs not retained: Yakama Nation wants to make it clear that monitoring and evaluation of the broader CEPCs and COPCs list of contaminants exceeding toxicity levels should be continued in future response action stages. | No revisions to the BHHRAs are necessary. Comment noted. This issue will be given full consideration when developing strategies for post construction confirmation sampling and long term monitoring. Risk driver contaminants, as well as the full set of identified COCs and CECs will be sampled for in post construction and long term monitoring sampling. CPECs and COPCs will be given consideration as warranted. No revisions to the BHHRAs are necessary. |

Note 1: We understand the concern that the entire Bonneville Dam complex is not a part of the scope of this project. However, please note that the Bradford Island project is not the primary vehicle for managing waste at Bonneville Dam. In fact, most actions are managed by an Environmental Compliance Team following Resource Conservation and Recovery Act (**RCRA**) processes and protocols. This is a fully staffed team that is integral to the routine operation and maintenance of the project. Their mission manages the majority of actions related to the generation, transportation, treatment, storage, and disposal of hazardous waste at the project. Their process begins with ERGO assessments, which are now conducted on an annual basis (the first ERGO assessment was completed in 1992) and ends when an item of non-compliance is corrected. Some past actions have included removal of underground storage tanks and removal of soils associated with known localized spills.

The sites that have become the 'Bradford Island' project, are being managed under the Comprehensive Environmental Response, Compensation, and Liability Act of 1980 (CERCLA). These sites are either too large to be managed as part of a routine program or fit the definition of a 'legacy' site.

Note 2: After the sediment removal work was completed in 2007, the sediments in the barges were sampled for final disposition. PCBs could not be detected in the samples. This is relevant because it shows that PCBs in sediments that were removed from the hottest locations could not be detected once handled by the removal contractor. In the case of Goose Island millions of tons of material was removed and placed to form Goose Island. It is highly unlikely that any contamination could be detected after the material went through such a removal and placement action, given the results in 2007.