

CLEANUP ACTION PLAN (CAP)

FORMER SCOTT PAPER COMPANY MILL SITE ANACORTES, WASHINGTON

WASHINGTON STATE DEPARTMENT OF ECOLOGY TOXICS CLEANUP PROGRAM LACEY, WASHINGTON

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ARAR applicable or relevant and appropriate requirement

BAF bioaccumulation factor
BGS below ground surface

BMP best management practice

CAP Cleanup Action Plan

cm Centimeter

CM1 Commercial Marine 1
CSL cleanup screening level

COPS chemicals of potential concern

cPAH carcinogenic polynuclear aromatic hydrocarbon CQAPP Construction Quality Assurance Project Plan

cy cubic yards

Ecology Washington State Department of Ecology

ft feet

K-C Kimberly-Clark Corporation

mg/kg milligrams per kilogram

mg/L milligrams per liter

MJB Properties

MHHW mean higher high water

MLLW mean lower low water

MTCA Model Toxics Control Act

NFA No Further Action

ng/kg nanograms per kilogram

OMMP Operations, Maintenance and Monitoring Plan

PAH polynuclear aromatic hydrocarbon

PCB polychlorinated biphenyl

Port Port of Anacortes

RA remedial action

RD remedial design

RI/FS remedial investigation/feasibility study

SAP Sampling and Analysis Plan SHS Sun Healthcare Systems, Inc.

SMS Sediment Management Standards

SQS sediment quality standards

TEE terrestrial ecological evaluation

TVS total volatile solids

μg/kg micrograms per kilogram

 μ g/L micrograms per liter

USACE U.S. Army Corps of Engineers

USEPA U.S. Environmental Protection Agency

VCP Voluntary Cleanup Program

WAC Washington Administrative Code

EXECUTIVE SUMMARY

This document presents the draft Cleanup Action Plan (CAP) for upland properties and aquatic lands at the former Scott Paper Mill site (the Site) located between 17th and 20th Streets, and East of Q Avenue in Anacortes, Washington. This CAP was prepared as a collaborative effort by the Washington State Department of Ecology (Ecology) and the entities responsible for cleanup of the various portions of the Site: Port of Anacortes (Port), Kimberly-Clark Corporation (K-C), and MJB Properties (MJB). This draft CAP has been prepared pursuant to a Consent Decree meeting the requirements of the Model Toxics Control Cleanup Act (MTCA) administered by Ecology under Chapter 173-340 of the Washington Administrative Code (WAC), and the requirements of the Sediment Management Standards (SMS) administered by Ecology under Chapter 173-204 WAC. This draft CAP describes Ecology's proposed site-wide cleanup action and sets forth functional requirements that the cleanup must meet, including follow-up monitoring.

Site Background

Commercial use of the Site began in 1890 with a lumber mill operation. In 1925, a pulp mill operation was added at the Site. In 1940, the Scott Paper Company (Scott) purchased the lumber and pulp operations. Scott discontinued operations at the lumber mill in 1955 and the pulp mill in 1978, and sold the mill properties in 1979.

The Site is currently divided into two main portions: north and south.

- North Portion. Historic features on the north portion of the Site included numerous buildings, sheds, piers, tailings ponds, boilers, fuel storage tanks, a smokestack, and burners. Scott operated this portion of the Site from 1940 until 1955. The Port of Anacortes purchased this portion in 1979 and used a part of the property as a log yard from 1990 to 1993. This portion was divided into three parcels in 1998, and Sun Healthcare Systems, Inc. purchased one of these parcels. In 1999, the Port of Anacortes and Sun Healthcare Systems, Inc. conducted an independent cleanup action on their parcel to remove petroleum-contaminated soil and wood debris. A 2-foot-thick soil cover and a soil containment wall along the shoreline were also installed.
- South Portion. The main features on the south portion of the Site were the pulp mill buildings, built around 1925. The pulp mill used waste wood from the lumber mill. Scott operated this portion from 1940 until 1978. This portion of the Site was purchased

in 1979, and used for several years as a staging area for assembly of portable buildings, boat manufacturing, and storage. MJB Properties purchased this portion of the Site in 1990.

Study Background

In 2008, a detailed Remedial Investigation (RI) and Feasibility Study (FS) were prepared by the Port, K-C and MJB under Ecology's direction. The RI utilized information about the history and environmental conditions of the Site gathered during prior investigations, supplemented with additional detailed investigations, to characterize the nature and extent of contamination. The RI identified contamination in soil, groundwater, and marine sediments at the Site as follows:

- Soil: Metals, petroleum hydrocarbons, carcinogenic polynuclear aromatic hydrocarbons (cPAHs), polychlorinated biphenyls (PCBs), and dioxins/furans were found at concentrations above preliminary cleanup levels.
- **Groundwater**: Sporadic slight exceedances of petroleum hydrocarbons, arsenic, sulfide, bis(2-ethylhexyl) phthalate, ammonia, and 4-methylphenol were found.
- Marine sediments: Metals, PCBs, and wood debris were found at concentrations above preliminary cleanup levels.

The follow-on FS developed and evaluated cleanup action alternatives for addressing contamination identified at the Site. More detailed information on the RI/FS, including the cleanup options that were evaluated, can be found on Ecology's Toxics Cleanup Website (http://www.ecy.wa.gov/programs/tcp/sites/scott_paper/scott_hp.html).

Cleanup Action Plan Overview

Based on the findings of the RI/FS, Ecology, the Port, K-C and MJB prepared this draft CAP, which provides the following:

- Identifies cleanup levels for soil, groundwater, and marine sediment.
- Recommends cleanup actions to achieve these cleanup levels from the options identified in the RI/FS, and describes these actions.
- Presents a schedule to carry out the cleanup.
- Identifies monitoring activities to demonstrate whether the cleanup was effective.

The following actions would address upland and shoreline soil and groundwater contamination at the Site:

- Excavate and test areas with contaminated soil.
- Within areas of the Site located inland (west) of the shoreline zone, approximately 18,600 cubic yards (cy) of contaminated soil will be removed from 13 areas. The depth of inland soil removal is based on the depth of contamination, which generally ranges between 2 and 15 feet below ground surface.
- Along the Site shoreline zone, approximately 10,600 cy of contaminated soil will be removed from 7 areas. Generally, soil in the shoreline zone will be removed to a depth of 10 feet below ground surface.
- Transport contaminated soil to approved disposal facilities.
- Backfill excavated areas with clean soil. Some of the near-surface soil in the excavation areas is clean and will be reused on-site as appropriate.
- Monitor groundwater for at least one year after the cleanup action is completed.
- Prepare environmental covenants as necessary to restrict future development and control any future soil disturbance where contamination may remain at the Site.

The following actions would address contamination in the shallow intertidal and deeper subtidal Site areas of Fidalgo Bay:

- Dredge approximately 30,200 cy of contaminated surface sediment from intertidal and subtidal areas of Fidalgo Bay to a minimum depth of 2 feet below mudline.
- Remove wood debris, brick, and pilings from these areas concurrent with the dredging.
- Transport the dredged sediments to off-site woody debris beneficial reuse facilities as appropriate, or to approved disposal facilities.
- Place structures offshore of the north portion of the Site to attenuate the impact of waves on the shoreline, to control potential future erosion of shoreline contaminated soils remaining at depth.
- Backfill dredged areas with clean sand and gravel. Place a minimum 2-foot-thick capping layer of clean sand, gravel, and armor stone as necessary along the shoreline in areas where contaminants will remain at depth.
- Cobble armor stone will be placed on the shoreline of the south portion of the property to control potential future erosion of shoreline contaminated soils remaining at depth.

Within this area, a minimum 0.5-ft-thick top-dressing of sand and gravel will be placed in the interstices of the cap armor stone.

• Replant damaged eelgrass bed habitat.

Introduction

1.0 INTRODUCTION

This document presents the Cleanup Action Plan (CAP) for upland properties and aquatic lands at the former Scott Paper Company Mill site (the Site) located in Anacortes, Washington (Figure 1). This CAP was prepared as a collaborative effort by the Washington State Department of Ecology (Ecology) and the entities responsible for cleanup of the various portions of the Site: Port of Anacortes (Port), Kimberly-Clark Corporation (K-C), and MJB Properties (MJB). It has been prepared pursuant to the requirements of the Model Toxics Control Cleanup Act (MTCA) administered by Ecology under Chapter 173-340 of the Washington Administrative Code (WAC), and the requirements of the Sediment Management Standards (SMS) administered by Ecology under Chapter 173-204 WAC. This CAP provides a general description of the proposed site-wide cleanup action and sets forth functional requirements that the cleanup must meet to achieve the cleanup action objectives for the Site.

1.1 Regulatory Framework

Between 2004 and 2008, detailed environmental investigations of the northern portion of the Site were performed pursuant to Consent Decree No. 03-2-00492-1 between the Port and Ecology (Consent Decree; Ecology 2003). The work required under the Consent Decree included preparation of a remedial investigation/feasibility study (RI/FS) that addressed soil at Port Parcels 1 and 3 (Figure 2), groundwater throughout the Port Uplands Area, and nearshore sediments adjacent to Port properties at the Site. Concurrent investigations of the southern portion of the Site were performed under Agreed Order No. DE 1783 between K-C and Ecology (Agreed Order; Ecology 2004). The work required under the K-C Agreed Order included preparation of an RI/FS for soil and groundwater at the MJB North Area and for marine sediments offshore of the MJB North Area. MJB (pursuant to agreements with K-C) performed upland soil and groundwater RI/FS tasks for the MJB North Area.

In addition to the required work described above, the Port Consent Decree and the K-C Agreed Order also required the Port and K-C, respectively, to address any remaining site-wide RI/FS issues. To ensure that site-wide issues were efficiently addressed, the Port, K-C, and MJB combined the various required elements of the Consent Decree and Agreed Order into a single site-wide RI/FS report. In August 2008, Ecology provided the draft final RI/FS report for public review. Responsiveness summaries to public and

stakeholder comments were provided by Ecology in October and November 2008, respectively. The final RI/FS report (GeoEngineers et al. 2008) was approved by Ecology on December 16, 2008.

1.2 Purpose

The purpose of this CAP is to:

- Describe the Site, including a summary of its history and extent of contamination
- Identify site-specific cleanup levels and points of compliance for each hazardous substance and medium of concern
- Identify applicable state and federal laws for the proposed cleanup action
- Identify and describe the selected cleanup action alternative for the Site
- Summarize the other cleanup action alternatives evaluated in the RI/FS
- Discuss environmental covenants and Site use restrictions
- Discuss compliance monitoring requirements
- Present the schedule for implementing the cleanup action plan

2.0 SUMMARY OF SITE CONDITIONS

Various investigation and cleanup activities have been conducted at the Site since approximately 1990. The final RI/FS report prepared in November 2008 describes investigations conducted between 2004 and 2008 at the Site (GeoEngineers et al. 2008). The purpose of the investigations was to collect, develop, and evaluate sufficient information to allow the selection of an appropriate cleanup action for the Site. Because the Site includes upland areas and aquatic lands, as shown in Figure 2, the media investigated included soil, soil vapors, groundwater, and sediment. In addition to the 2004 to 2008 investigations, the scope and results of previous investigations were also described in the RI/FS report to provide a comprehensive summary of Site conditions (Anchor et al. 2002; contained in Appendix B of the RI/FS report). More recent (fall 2008) soil and sediment sampling data collected following completion of the RI/FS are presented in Appendices A and B, respectively, of this CAP to provide further refinement of areas requiring cleanup. This section summarizes pertinent environmental conditions at the Site (i.e., nature and extent of contamination) and an overview of the conceptual site model. More detailed descriptions of Site conditions are provided in the RI/FS report and Appendices A and B of this CAP.

2.1 Site History

The former Scott Paper Mill was located in Anacortes, Washington, on the west shore of Fidalgo Bay. The development of the shoreline as an industrial area began in the late 1800s. Prior to development of the Site, the area was largely a shallow tideland. In 1892, a lumber mill was built at the Site that extended on pilings into Fidalgo Bay. The lumber mill was located in the area referred to as the Port Uplands Area (Figure 2). Wharves and offshore log rafts were present in much of the northern portion of the Marine Area (extending from the shoreline to about the inner harbor line) until the late 1940s (see photograph on the cover of this CAP). Between approximately 1890 and 1940, approximately 5 to 20 feet (ft) of fill materials including sawdust and mill refuse were placed throughout the former tide flat beneath and adjacent to the wharves, also extending into the MJB North Area.

In 1925, a pulp mill was constructed at the property referred to as the MJB North Area. Pulp was produced using an acid-sulfate process using byproducts from the lumber mill. In 1940, Scott Paper purchased the pulp and lumber mills, and operated the

facilities until 1955. Process improvements by Scott Paper included the conversion to an ammonium sulfite process in 1952, the construction of a 16-inch effluent pipeline to Guemes Channel and an on-site surge pond for the pipeline in May 1951, and the addition of pulp bleaching facilities in 1955. Effluent was discharged directly into Fidalgo Bay from 1925 to 1951. A knots and tailings pond was constructed in 1959, on what is now Port Parcel 2, to reduce settleable solids in the mill's effluent. Materials known to have been utilized at the former pulp mill include petroleum, sulfur, anhydrous ammonia, ammonium hydroxide, and chlorine. Bunker C and diesel fuels were used to generate power and operate equipment. The pulp mill closed in 1978. Scott Paper was acquired by K-C in December 1995.

The former Scott Paper Mill operations were bounded by Cap Sante Boat Haven to the north, Fidalgo Bay to the east, and Q Avenue to the west. To the south, the maximum extent of former Scott Paper Mill operations was approximately 20th Street. Site boundaries are depicted in Figure 2. In 1978 and 1979, the Port purchased the northern portion of the Site. The southern portion of the Site was purchased by the Snelson-Anvil Corporation in 1979, and has been owned by MJB since 1990. In 1999, Sun Healthcare Systems, Inc. (SHS) purchased Parcel 2 from the Port and, following initial cleanup and redevelopment (see below), subsequently subdivided and sold Parcel 2 into four sublots. In 2008, the Port acquired a narrow strip of the Marine Area between the Port and MJB properties.

2.2 Port Uplands Area Redevelopment

After closure of the mill in 1978, little activity occurred on the Port Uplands Area until 1990 when the Port constructed and operated a log storage facility on Parcels 1 and 2 (Figure 2). The log storage yard was in operation through 1993. As part of the log storage yard construction, approximately 30,000 cubic yards (cy) of wood debris was removed from the west side of Parcel 1 and the excavations were backfilled with dredged materials obtained from the 1968 U.S. Army Corps of Engineers (USACE) expansion and dredging of the Cap Sante Marina. In 1994, an additional 100,000 cy of dredged sand from the Swinomish Channel was delivered by the USACE and used as geotechnical preload to support redevelopment of the Port Uplands Area. Since placement of the preload material, Parcels 2 and 3 have been developed. Construction

of Seafarers' Memorial Park on Parcel 3 began in 1995. As part of the park construction, the park soil was capped with 0.5 to 1 ft of a topsoil mixture. The topsoil was prepared from the preload material mixed with wood debris remnants from the Port log storage yard. In 2000, SHS developed a campus/office park on Parcel 2.

2.3 MJB Property Redevelopment

The MJB North Area is the southern portion of the former Site operations area, and refers to the area between 17th and 20th Streets, east of R Avenue (Figure 2). In about 1982, the Snelson-Anvil Corporation reportedly removed wood debris and soft soils from much of the MJB North Area and backfilled the excavations with imported granular fill, particularly in areas where heavy Site operation loads were planned. Historical aerial photos indicate that portions of the MJB North Area, particularly in areas close to the shoreline, have not been significantly excavated since the time of the mill operations. Since 1982, the MJB North Area has been used for light industrial operations.

MJB North Channel 1, located immediately offshore of the MJB North Area, is an existing private navigation channel (Figure 2). North Channel 1 was originally dredged in 1975 to a depth of approximately 12 ft below mean lower low water (MLLW) for barge access and moorage of barges and other vessels. However, no maintenance dredging has been conducted in North Channel 1 since initial dredging.

2.4 Prior Cleanup Actions

Following detailed investigations of Port Parcel 2 (ThermoRetec 1999a) and subsequent preparation of a soil Cleanup Action Plan (CAP) for this area (ThermoRetec 1999b), cleanup at Parcel 2 was conducted by SHS, with oversight by Ecology under the MTCA Voluntary Cleanup Program (VCP). The Parcel 2 cleanup included, among other elements, removal and off-site landfill disposal of 3,469 tons of petroleum-contaminated soil (excavation areas are depicted in Figure 2), soil capping, and environmental covenants to prevent future exposure to subsurface soil at the property and to restrict groundwater use for drinking water. Work also included the installation of a sheetpile wall along the shoreline (near MW-112) for containment of residual contaminated soil, concurrently providing structural foundation support for the building constructed by

SHS. A project completion report for the Parcel 2 property was submitted to Ecology in 2000 (ThermoRetec 2000).

In 2000, Ecology issued a No Further Action (NFA) letter for diesel-range and oil-range petroleum hydrocarbons, polycyclic aromatic hydrocarbons (PAHs), polychlorinated biphenyls (PCBs), dioxins/furans, wood debris, and metals in soil at Parcel 2 (Ecology 2000). The NFA letter was conditional to long-term groundwater monitoring to ensure continued environmental protection. However, in 2005 Ecology modified the type of written opinions it provides under the VCP, and no longer provides NFA letters for a single medium such as soil (Ecology 2005). Accordingly, Ecology subsequently rescinded the NFA letter on September 26, 2006, as the completed cleanup did not address all contamination in all media at the Site. This CAP describes Ecology's selected site-wide cleanup remedy for the Site, including elements of the NFA letter as appropriate.

Storm-generated wave and current action has resulted in significant erosion at the filled shoreline since at least 1962, which has contributed to contaminant transport from the uplands to the marine area (GeoEngineers et al. 2008). The shoreline along parts of the Port and MJB properties has been temporarily reinforced to minimize this erosion, and protection of the shoreline has required routine maintenance by the Port and MJB. In February 2005, the Port completed a MTCA Bank Stabilization Interim Action along the Seafarers' Memorial Park shoreline under the Consent Decree (Landau Associates 2005).

In 2008, the Port installed two underground storage tanks at Parcel 3. An interim action was completed to address contaminated soils and wood debris excavated during the tank installation. An interim action completion report is currently being prepared by the Port to document the interim action activities (GeoEngineers 2009).

2.5 Summary of Environmental Conditions

This section summarizes environmental conditions at the Site for soil (including soil vapor), groundwater, and sediment media, based on the extensive RI/FS sampling and analysis efforts completed at the Site (GeoEngineers et al. 2008). Habitat features and

aquatic resources of the Marine Area are also summarized. Further details and sources of the information presented in this section are provided in the final RI/FS report.

2.5.1 Soils

Site soils consist of multiple layers of fill overlying native marine sediment and glacial deposits. Shallow soil is predominantly gravel and sand fill material with occasional mixed wood debris. The deeper subsurface fill contains a heterogeneous mixture of soil and wood debris. Relatively extensive wood debris deposits of varying thickness are present throughout much of the Port Uplands Area, extending from 5 to 20 ft below ground surface (BGS), and continuing into the nearshore (intertidal and shallow subtidal) area of Fidalgo Bay. Fill material sometimes containing wood debris is also found in the MJB North Area along the shoreline. The thickness of the wood-containing fill material in the MJB North Area ranges from less than a foot near the former mill surge ponds to nearly 15 ft at the shoreline.

An empirical demonstration presented in the final RI/FS report (GeoEngineers et al. 2008) verified that, with the exception of localized diesel-range and oil-range petroleum hydrocarbon contamination in portions of Port Parcels 2 and 3, chemical concentrations present in soil are protective of groundwater and surface water (see Section 2.5.2 below). Therefore, soil chemicals of potential concern (COPCs) were generally identified based on potential direct human contact and terrestrial ecological exposures.

A range of metals (antimony, arsenic, chromium, copper, lead, mercury, nickel, and zinc), diesel-range and motor oil-range petroleum hydrocarbons, PCBs, carcinogenic PAHs (cPAHs), and dioxins/furans have been detected at the Port Uplands Area at concentrations above the preliminary cleanup levels established for the Site (GeoEngineers et al. 2008). In addition, metals (antimony, arsenic, chromium, copper, lead, nickel, thallium, and zinc) and cPAHs were identified in soil at the MJB North Area at concentrations above the preliminary cleanup levels. The specific depth intervals and COPCs detected in soils at

concentrations exceeding preliminary cleanup levels are summarized in Figures 3 to 10.

The only constituent detected in shallow surface (0 to 2 ft BGS) soils in the Port Uplands Area above preliminary cleanup levels was arsenic, which was detected at concentrations higher than regional background levels (20 milligrams per kilogram [mg/kg]) at one location (LAI-S-4) in Seafarers' Memorial Park (Figure 3). Elevated arsenic, lead, and cPAH concentrations were also detected in the surface fill layer at a number of isolated locations throughout the MJB North Area (Figure 7).

At depths greater than 2 ft BGS, elevated concentrations of arsenic were detected in subsurface soils in an isolated area near the northeastern corner of Port Parcel 1 (Figure 5). Multiple constituents were detected in subsurface soil remaining at Port Parcel 2 at concentrations above the preliminary cleanup levels, including metals (antimony, arsenic, chromium, copper, lead, mercury, nickel, and zinc), diesel- and motor oil-range petroleum hydrocarbons, cPAHs, PCBs, and dioxins/furans. These exceedances were concentrated in two areas: 1) the rectangular parking lot near the center of Parcel 2 (elevated diesel- and motor oilrange petroleum hydrocarbon concentrations); and 2) in the area of the subsurface containment wall in the southeastern portion of Parcel 2. A similar range of COPCs was detected in subsurface soil at Parcel 3, particularly along parts of the shoreline of Seafarers' Memorial Park and near the present southern end of R Avenue (Figures 4 to 6). Likewise, elevated metal concentrations (antimony, arsenic, copper, lead, thallium, zinc) and cPAHs were detected in subsurface soil in the northeast and/or southeast portion of the MJB North Area, and appear to be limited to the wood layer and woody fill layers in the subsurface from roughly 4 to 10 ft BGS (Figures 8 to 10).

Soil vapors at Port Parcel 3, where subsurface wood debris accumulations are greatest at the Site, have been monitored for the presence of hydrogen sulfide and methane. No detectable concentrations of either of these gases were measured during the 2004 to 2008 investigations. Previous monitoring of soil

vapors at Port Parcel 3 identified low concentrations of hydrogen sulfide in ambient air samples collected in 1993.

Based on the information presented in the RI/FS report (GeoEngineers et al. 2008), soil in parts of Port Parcels 1 and 3 and the MJB North Area required evaluation of cleanup action alternatives due to the presence of COPCs at concentrations exceeding the preliminary cleanup levels determined to be protective of human health and terrestrial ecological receptors. As discussed in Section 2.4, cleanup actions have been previously evaluated and implemented by SHS for soil at Port Parcel 2. In addition to soil excavation and off-site landfill disposal actions, SHS placed an indicator layer and clean soil cap over Parcel 2, constructed and activated a methane control system, implemented infiltration controls, installed a subsurface containment wall, and implemented environmental covenants (ThermoRetec 2000). However, because soil containing COPCs at concentrations exceeding preliminary cleanup levels remains in portions of Port Parcel 2, the need for additional cleanup at Parcel 2 was evaluated further in the RI/FS.

2.5.2 Groundwater

Two hydrogeologic units have been identified at the Site: 1) a shallow water-bearing unit; and 2) a deeper confining unit. The shallow water-bearing unit occurs in the fill material, and the depth to groundwater in this unit ranges from 3 to 12 ft BGS (7 to 15 ft saturated thickness) across the Site. The confining unit, which underlies the shallow water-bearing unit, consists of native marine silts and clays. The thickness of the confining unit is greater than 2 to 10 ft throughout the Site.

Predominant groundwater flow directions are to the north toward Cap Sante Waterway in the northern portion of the Site, and to the east and southeast toward Fidalgo Bay in other areas of the Site. Groundwater flow directions do not appear to be significantly affected by tidal fluctuations. However, groundwater level data suggest that hydraulic gradients decrease or possibly reverse temporarily at high tide in the vicinity of some of the shoreline wells.

To identify groundwater COPCs, chemical analytical results were compared with preliminary cleanup criteria that are protective of potential exposure by aquatic organisms to hazardous substances in groundwater and ingestion by Site users of marine organisms potentially contaminated by releases of affected Site groundwater (GeoEngineers et al. 2008). Groundwater data for diesel-range and motor oil-range petroleum hydrocarbons were compared with MTCA Method A groundwater (drinking water-based) cleanup levels [in accordance with WAC Chapter 173-340-730(3)(b)(iii)(C)], because regulatory criteria protective of marine surface water have not been established for petroleum hydrocarbons. Based on detailed evaluations presented in the RI/FS report, the nature and extent of groundwater contamination at the Site can be summarized as follows:

- **Port Area Interior Monitoring Wells**. Groundwater at interior wells in most of the Port Uplands Area contains low concentrations of COPCs (below preliminary cleanup levels). There was one marginal exceedance of dissolved arsenic above drinking water criteria at well MW-111 (see Figure 2). This exceedance was isolated and is not representative of groundwater conditions. Total and/or dissolved arsenic was also detected at concentrations above drinking water criteria during four monitoring events at well MW-102. Diesel-range and motor oil-range petroleum hydrocarbons were detected at concentrations above MTCA Method A cleanup levels during one monitoring event at well MW-110, and free product was observed during two monitoring events at this same well at measured thicknesses of 0.03 ft and 0.6 ft. In addition, diesel- and motor oil-range petroleum hydrocarbons were detected at concentrations above MTCA Method A cleanup levels in an unfiltered groundwater grab sample (GEI24-W) obtained from a direct-push boring completed at the southeast corner of the Seafarers' Park Building as part of the 2008 Port Uplands supplemental soil investigation (see Appendix A).
- MJB Area Interior Monitoring Wells. Groundwater at interior well MW-7 does not contain COPCs at concentrations above preliminary cleanup

- levels. At interior well MW-4, dissolved arsenic was detected in groundwater at concentrations exceeding drinking water criteria.
- Port Area Shoreline Monitoring Wells. Groundwater at shoreline wells
 on Port property, located landward from the groundwater/surface water
 interface in the porewater discharge zone, has not been identified to
 contain detections of COPCs above preliminary cleanup levels. Although
 there were a few sporadic detections of ammonia, sulfide, and bis(2
 ethylhexyl)phthalate, these were isolated occurrences and are not
 representative of groundwater conditions.
- MJB Area Shoreline Monitoring Wells. Similarly, groundwater at shoreline wells of MJB North Area property landward from the groundwater/surface water interface in the porewater discharge zone also does not contain COPCs at concentrations above preliminary cleanup levels.

As detailed in the final RI/FS report (GeoEngineers et al. 2008), direct human ingestion of hazardous substances in groundwater is not a potential exposure pathway, because groundwater at the Site or potentially affected by the Site is not a current or reasonably likely future source of drinking water.

Groundwater in the shoreline area at the Site was determined to be protective of marine surface water. However, because there were some exceedances of preliminary cleanup levels for arsenic and diesel-range and motor oil-range petroleum hydrocarbons at interior wells (including free-phase petroleum product observed at MW-110), remedial options for groundwater at the Port Uplands Area were evaluated in the RI/FS (GeoEngineers et al. 2008). Groundwater throughout the MJB North Area landward of the shoreline was determined to be protective of marine surface water. Consequently, the RI/FS did not evaluate remedial options for groundwater at the MJB North Area.

2.5.3 Sediments

Relatively low rates of sediment deposition occur within the Site area. These results are consistent with observed wave action that contributes to periodic

sediment transport, maintaining a mixed sand/gravel/cobble intertidal substrate in much of the Site area (Antrim et al. 2000). The contact between fine-grained native sediments and overlying material occurs at an elevation ranging from approximately -3 to -8 ft MLLW. This is consistent with the filling of the tideflat that historically extended southward from the Cap Sante area. Overlying the native sediments in the northern portion of the Marine Area is fill that contains wood debris. This fill is thickest near the shoreline and tapers out in the Marine Area; the fill thickness in parts of the northern portion of the shoreline ranges from approximately 10 to 15 ft. Overlying the wood-containing fill along the northern portion of the shoreline is 10 to 15 ft of imported granular fill material consisting of poorly graded sand and silt or fine sand.

A debris field consisting of dimensional lumber, wood fragments, and other debris is present on the sediment surface, most extensively across the intertidal area of the northern Marine Area, extending to shoreline areas adjacent to the MJB North Area. The wood debris content (based on visual observations) of surface sediments ranges from greater than 75 percent near the shoreline to less than 5 percent near the inner harbor line. A number of decaying pilings are also present. Sediment bioassays were performed to develop site-specific cleanup levels for wood debris content and total volatile solids (TVS) that are protective of sediment habitats (see Section 3.3.3). The extent of site-specific cleanup level exceedances within the Marine Area is depicted in Figure 11.

Chemical analytical results for sediment samples collected throughout the Marine Area during 2004 to 2008 and previous investigations were compared in the RI/FS to SMS chemical criteria to identify COPCs for the offshore portions of the Site (GeoEngineers et al. 2008). Sediment samples collected from the intertidal beach area immediately offshore of the Site contained several metals (copper, lead, mercury, and zinc) and PCBs at concentrations above sediment quality standards (SQS) chemical criteria, which comprise the preliminary cleanup levels for sediments. The sampling data define a localized area of elevated metals and PCBs within the intertidal zone of the South Marine Area (Figure 11). As discussed in the RI/FS, an evaluation of available tissue sampling

data conducted by Ecology, the U.S. Environmental Protection Agency (USEPA) and others showed no evidence of bioaccumulation of mercury, PCBs, or dioxins/furans in crabs or shellfish within the portion of Fidalgo Bay potentially affected by Site releases.

To ensure protection of human health, the RI/FS considered potential bioaccumulation risks associated with residual mercury and PCB exposure that may remain in the Marine Area following completion of the cleanup action. The potential bioaccumulation risks were assessed in the RI/FS report, and revealed that remediation of those Site sediments exceeding SQS chemical criteria would be protective of potential human health mercury and PCB bioaccumulation risks (see Section 3.3.3).

Based on the findings of the RI/FS, surface sediments in upper intertidal portions of the Marine Area immediately adjacent to portions of Port Parcel 3 and the MJB North Area required evaluation of cleanup action alternatives due to the presence of COPCs exceeding preliminary cleanup levels (GeoEngineers et al. 2008). A likely source of these localized contaminated sediment deposits is historical and potentially ongoing erosion of adjacent upland fill material comprising the shoreline. These fill materials have been documented to contain elevated metal and PCB chemical concentrations, similar to those identified in the adjacent sediments. As discussed in the RI/FS, shoreline stabilization performed by the Port in this area appears to have reduced transport of metals and PCBs to the South Marine Area in the last several years (Figure 11). Surface and subsurface woody debris deposits in this area also required evaluation of cleanup action alternatives due to the presence of these potentially deleterious substances at concentrations exceeding the preliminary cleanup levels established to protect aquatic ecological receptors at the Site.

Limited areas of the upper intertidal zone adjacent to the Site containing a mix of sand and gravel may provide suitable spawning habitat for sandlance (*Ammodytes hexapterus*) or surf smelt (*Hypomesus pretiosus*) (Antrim et al. 2000). Offshore areas contain limited areas of eelgrass (*Zostera* spp.) of varying

densities. The eelgrass beds provide a number of ecological functions including support of prey species, substrate for spawning of Pacific herring (*Clupea pallasi*), and rearing for juvenile salmon and crab. As discussed in the RI/FS report (GeoEngineers et al. 2008), detailed eelgrass surveys of the areas offshore of the MJB North Area and Port Uplands Area were performed during August 2004 and August 2007, respectively, and the results are summarized in Figure 11.

2.6 Conceptual Site Model

This section summarizes the conceptual model for the fate and transport of contamination at the Site as described in the RI/FS (GeoEngineers et al. 2008). The conceptual site model also describes the contaminant exposure pathways identified for the Site and the potential risks posed to human health and the environment by hazardous and/or deleterious substances in soil, groundwater, and/or sediment.

Soil impacts at the Site resulted from past releases of hazardous substances to soils, primarily during the 1890 to 1940 period when fill materials including sawdust and mill refuse were placed throughout the former tide flat area. Soil contamination has also resulted from the release of petroleum to Site soils from facilities located at the Site.

Sediment impacts have resulted from the direct deposition of deleterious substances in the Marine Area (e.g., from log rafting activities), potential past transport of contaminants in groundwater, erosion of hazardous substances in shoreline soils in the Port Uplands and MJB North Areas to Fidalgo Bay, and/or decay of wood debris present below the surface sediment. The conceptual site model illustrating potential contaminant transport mechanisms is shown in Figure 12.

2.6.1 Soil

Potential upland soil exposure pathways at the Site include:

- Contact (dermal, incidental ingestion, or inhalation) by visitors, workers (including excavation workers), and potential future residents or other Site users with hazardous substances in soil
- Contact (dermal, incidental ingestion, or inhalation) by terrestrial wildlife with hazardous substances in soil

 Contact by terrestrial plants and soil biota and/or food-web exposure to hazardous substances in soil

Site areas where COPCs were detected in soils at concentrations above preliminary cleanup levels for protection of human and terrestrial ecological receptors are shown in Figures 3 through 10. The majority of the soil exceedances occur between 6 and 10 ft BGS along parts of the shoreline at Seafarers' Memorial Park, the central parking lot at Parcel 2 (petroleum hydrocarbons), the area of the subsurface containment wall in the southeastern portion of Parcel 2, and in wood debris fill layers between 4 and 10 ft BGS in the northeastern portion of the MJB North Area. Elevated soil concentrations also occur in several locations along "R" Avenue.

2.6.2 Groundwater

Although arsenic and/or petroleum constituents were detected at interior monitoring wells at concentrations above levels protective of marine surface water, concentrations of these constituents at or near the groundwater/surface water interface in the porewater discharge zone are protective of Site receptors. Even at interior wells, relatively few constituents were found at concentrations above drinking water-based cleanup levels.

Human ingestion of hazardous substances in groundwater is not a potential exposure pathway because groundwater at the Site or potentially affected by the Site is not a current or reasonable future source of drinking water. The MTCA regulation (WAC 173-340-720[2][d]) states that even if groundwater is classified as a potential future source of drinking water because it is present in sufficient quantity, contains less than 10,000 milligrams per liter (mg/L) total dissolved solids, and is not too deep to recover, the groundwater may still be classified as non-potable due to its proximity to marine surface water. To be classified as non-potable on the basis of its proximity to marine surface water, the following conditions must also be met:

• The groundwater does not serve as a current source of drinking water

- Contaminated groundwater will not migrate to groundwater that is a current or potential future source of drinking water
- There are known points of entry of the groundwater into surface water
- The surface water is not classified as a suitable domestic water supply source
- The groundwater is sufficiently hydraulically connected the surface water that the groundwater is not practicable to use as a drinking water source

The shallow groundwater at the Site meets at least four, and likely all five, of these conditions. First, groundwater at the Site is not a current source of drinking water. Second, the groundwater migrates toward marine surface water and discharges at seeps in the intertidal and/or subtidal zone (GeoEngineers et al. 2008). Third, the marine surface water offshore of the Site is not classified as a suitable domestic water supply. Fourth, the Site groundwater is hydraulically connected to marine surface water, as evidenced by the tidal influence on groundwater levels in wells near the shoreline. Finally, migration of shallow groundwater to a lower aquifer that is a current or potential future source of drinking water is unlikely, due to the presence of a confining native silt/clay unit at the base of the shallow water-bearing unit at the Site (see the RI/FS report for further information regarding Site hydrogeology). Consequently, the Site groundwater qualifies as a non-potable water source.

At interior well MW-110, diesel-range and motor oil-range petroleum hydrocarbons have been detected above MTCA Method A cleanup levels. In addition, free product has been observed during two monitoring events at MW-110, at measured thicknesses of 0.03 ft and 0.6 ft. Petroleum hydrocarbons were also detected above MTCA Method A cleanup levels in an unfiltered groundwater sample collected from a direct-push soil boring as part of the 2008 Port Uplands Area supplemental soil investigation reported in Appendix A. However, as discussed in the RI/FS (GeoEngineers et al. 2008), groundwater COPC concentrations in other areas of the Site do not appear to be migrating to Fidalgo Bay and/or Cap Sante Waterway.

Porewater (0 to 10 centimeters [cm] below mudline) in intertidal sediment deposits complies with preliminary groundwater cleanup levels. Ammonia and sulfide were rarely detected in sediment porewater, and were well below the preliminary cleanup level for ammonia and the screening level for sulfide. These data are consistent with tidal mixing and associated oxidation of sediment porewater that occurs near the sediment/water interface. In the presence of dissolved oxygen, ammonia and sulfide both rapidly undergo chemical and biological oxidation to nitrate and sulfate, respectively. Thus, tidal mixing and associated oxidation processes attenuate potential ammonia and sulfide risks to benthic infauna at the Site.

2.6.3 Sediments

As discussed above, wave and current modeling of the Site has shown that storm-generated wave and current action has resulted in significant erosion of the filled shoreline since at least 1962. The shoreline along parts of the Port and MJB properties has been temporarily reinforced to minimize this erosion, and protection of the shoreline has required routine maintenance by the Port and MJB. Net sediment transport along the western shore of Fidalgo Bay, in the vicinity of the Site, appears to be predominantly in a southerly direction from the Port Marine Area to the MJB North Marine Area (City of Anacortes 1999). In February 2005, the Port completed a temporary Bank Stabilization Interim Action along the Seafarers' Memorial Park shoreline under the Port's Consent Decree (Landau Associates 2005). Monitoring suggests that shoreline bank erosion has now ceased in the Interim Action area. However, pea gravel placed on the large rock armoring continues to erode and will likely require future maintenance if no further remedial actions were to be implemented. The northeastern shoreline of the MJB property has experienced continued erosion during MJB's time of ownership (since 1990), with an apparent increased rate of erosion within the past five years.

Sediment impacts at the Site are attributed to historical direct deposition of woody debris and associated deleterious substances in the Marine Area and shoreline erosion of hazardous substances in soils the Port Uplands Area and the

MJB North Area to Fidalgo Bay. Historical sources of woody debris at the Site include former log rafting operations, over-water storage of milled wood, placement of woody debris-containing fill materials (including sawdust, bark, and wood chips), and lumber/pilings remaining from the former pier structure. A range of surficial debris is present in the beach area, including dimensional lumber, bricks, and other construction materials. Debris accumulations are most evident within the intertidal and shallow subtidal zones extending from south of the Cap Sante Boat Haven breakwater to south of the existing kayak dock.

Nearshore chemical source areas exceeding sediment cleanup screening level (CSL) chemical criteria were depicted in the RI/FS report, and include soil in parts of the Port Uplands Area (Port Parcel 3) and a portion of the MJB North Area that contain elevated concentrations of metals and/or PCBs. Erosion of soil from these areas is the likely source of down-drift sediment contamination observed just to the south. Portions of these areas of the shoreline are currently armored with riprap.

Based on the available Site characterization data (summarized in RI/FS report), relatively extensive wood debris deposits are present throughout much of the upland areas of the Site, extending 10 to 30 ft BGS, and continuing into the nearshore (intertidal and shallow subtidal) area of Fidalgo Bay. Intertidal and shallow subtidal surface sediments offshore of the MJB North Area typically consist of a relatively thin layer of silt and sand sediments overlying the wood debris deposits. The thickness of the naturally developed sediment "cap" in this area of the Site is typically 0.5 to 1 ft at upper intertidal elevations, increasing in thickness at lower tidal elevations.

3.0 CLEANUP REQUIREMENTS

The MTCA cleanup regulations provide that a cleanup action must comply with cleanup levels for identified COPCs, points of compliance, and applicable or relevant and appropriate requirements (ARARs) based on federal and state laws (WAC 173-340-710). The Site cleanup levels, points of compliance, and ARARs for the selected cleanup remedy are briefly summarized in the following sections.

3.1 Human Health and Environmental Concerns

Because Site groundwater is not a current or reasonably likely future source of drinking water, cleanup levels for Site soil need not be protective of groundwater as drinking water. Additionally, an empirical demonstration presented in the RI/FS (GeoEngineers et al. 2008) verified that, excluding petroleum hydrocarbon releases in certain parts of the shoreline at the Port Parcel 3 Seafarers' Memorial Park and the central parking lot at Port Parcel 2, existing chemical concentrations in Site soils are protective of groundwater and marine surface water receptors. Although current soil concentrations in most areas of the Site are protective of groundwater, within the identified petroleum hydrocarbon release areas further soil and/or free product remediation is nonetheless necessary to ensure protection of human health and the environment.

3.1.1 Future Land Use Considerations

Soil cleanup levels for unrestricted land use were developed in accordance with WAC 173-340-740. The Port Uplands and MJB North Areas are currently zoned Commercial Marine 1 (CM1), which provides for a mix of commercial, industrial, and recreational uses. Because the Site is not zoned for strictly industrial use, soil cleanup levels were developed based on unrestricted land use, including the more stringent MTCA Method B cleanup levels that assume ground floor residential land use (WAC 173 340 740[3]).

Development of Port Parcel 1 (which is currently vacant) will be implemented concurrent with or immediately following cleanup of this area of the Site. As discussed above, SHS built an office park on Port Parcel 2 in 2000. SHS subsequently subdivided Parcel 2 into four sub-lots, sold these lots to four entities, and since then additional buildings have been constructed on these

properties. Construction of Seafarers' Memorial Park on Port Parcel 3 began in 1995, and included grass, landscaped areas, and a community building, as well as asphalt-paved roads and parking areas. Parcel 3 and the adjacent Marine Area are part of Seafarers' Memorial Park. At present, there are no plans to modify the site use of Parcel 3.

MJB has made a preliminary determination that a water and water view-dependent mixed-use development, with a residential component, is a viable future development option for the MJB North Area. A mixed-use uplands development might include a hotel, retail shops, offices, restaurants, residences, and parking structures. Residence styles would likely consist of townhouses, townhouses over flats, and/or stacked flats. These residences would primarily be situated above the lower non-residential levels or above an in-ground or above-ground parking structure. The uplands development area would be surfaced with concrete, asphalt, or structures, with localized and controlled landscaped areas. MJB has also considered a marina as a development option. The conceptual plan for the marina includes slips for pleasure boats and float planes, docks, and, potentially, a floating breakwater. A 12-ft-wide (approximate) promenade, adjacent to a 25-ft-wide landscaped set-back along the top of the bank, is planned as the transition from the offshore to the upland areas. The marina would also include upland support facilities (e.g., parking).

3.1.2 Ecological Risk Considerations

Terrestrial ecological evaluations (TEEs) were performed for both the Port Uplands and MJB North Areas, and are presented in the RI/FS (GeoEngineers et al. 2008). For the purposes of the TEEs, chemicals that exceeded unrestricted land use soil screening concentrations in WAC 173-340-900 (Table 749-2) were identified as COPCs.

The wildlife exposure model in WAC 173-340-900 (Table 749-4), supplemented with recently-published USEPA soil-worm bioaccumulation factors (BAFs) for selected metals, was used to calculate soil concentrations of copper, nickel, and zinc that are protective of terrestrial wildlife at the Site (see Appendix E of the FS

report; GeoEngineers et al. 2008). The calculations were performed with specific reference to the MJB North Area, but the methods used and the values derived are also applicable to the Port Uplands Area. USEPA has not developed a BAF for nickel, so the value in WAC 173-340-900 (Table 749-5) was used. The lowest concentration calculated for three wildlife receptor groups (avian predator, mammalian predator, and mammalian herbivore) was selected as the protective soil concentration for wildlife.

The Port Uplands Area currently consists of paved surfaces, buildings, and controlled landscape areas over most of the property (Parcels 2 and 3). Parcel 1 will be developed as a marine skills center educational facility with or immediately following Site cleanup. These features are expected to be maintained into the future.

The MTCA regulation assumes that the biologically active soil zone extends to a depth of 6 ft BGS, and allows a conditional point of compliance to be established at 6 ft BGS for ecological-based soil cleanup levels, provided that environmental covenants are implemented to address potential excavation of deeper soil (WAC 173-340-7490[4][a]; GeoEngineers et al. 2008). At the Port Uplands Area, COPCs exceeding terrestrial ecological criteria in soil between 0 and 6 ft BGS west of the shoreline zone include arsenic, copper, lead, zinc, and dioxins/furans. At the MJB North Area, COPCs exceeding terrestrial ecological criteria in soil between 0 and 6 ft BGS west of the shoreline zone include arsenic, lead, and zinc.

The MJB North Area uplands currently provide low quality habitat for wildlife, and this condition is expected to continue under a future site development scenario. Much of this area is covered with a 1-ft-thick layer of quarry spalls (rock fragments) that contains little organic matter to support soil biota or surface vegetation. Plants are not considered to be at risk within the MJB North Area uplands. Currently, there is very little vegetation within this area, except for sparse blackberries, scotch broom, and grasses. Under a future development scenario, paved surfaces and buildings will cover most of the property except for controlled landscaped areas. Any future planting of vegetation within the area

where COPCs exceed MTCA soil screening concentrations for plant exposures (WAC 173-340-900, Table 749-3) would require adding a minimum of 1 ft of topsoil over the existing quarry spalls to promote grass and/or shallow rooted plant growth. The addition of clean soil would limit grass and/or shallow rooted plant exposure to existing metal concentrations in the soil.

In accordance with WAC 173-340-7493(3)(b), the Port and K-C will conduct more detailed TEE soil bioassays in selected locations at the Port Uplands Area where COPC concentrations exceed preliminary TEE levels listed in WAC 173-340-900, Table 749-2. Potential risks to plant life and soil biota in such target areas of the Site will be assessed directly using tests described in the *Early Seedling Growth Protocol for Soil Toxicity Screening* (Ecology Publication No. 96-324) and the *Earthworm Bioassay Protocol for Soil Toxicity Screening* (Ecology Publication No. 96-327), respectively. The results of the soil bioassay sampling will be integrated into the remedial design (RD), and may reduce the extent of soil remediation required in parts of the Port Uplands Area where 0 to 6 ft BGS soil concentrations exceed preliminary TEE chemical criteria.

3.1.3 Shoreline Stability Considerations

The shoreline at the Site is protected from northerly wind and waves by the Cap Sante Boat Haven and its breakwater. The wave action in the Site area is predominantly from the southeast and northeast, and is strong enough to result in a mixed sand/gravel/cobble intertidal substrate (Antrim et al. 2000). Empirical observations and hydrodynamic modeling of the Site have shown that since approximately 1962, storm-generated wave and current action have resulted in erosion and loss of up to approximately 75 feet of shoreline, particularly in parts of Port Parcel 3 (GeoEngineers et al. 2008). The shoreline at the Site has recently been temporarily reinforced with riprap armor to minimize this erosion, and protection of the shoreline has required routine maintenance by the Port and MJB. In February 2005, the Port completed a Bank Stabilization Interim Action along the Seafarers' Memorial Park shoreline under the Port's Consent Decree (Landau Associates 2005). Monitoring suggests that shoreline erosion has now ceased in the Interim Action area. However, a long-term shoreline remedy is

needed for the Site that appropriately balances: 1) control of bank erosion and associated releases to the sediment from contaminated nearshore soils; 2) future land uses; and 3) habitat restoration. Ecology's selected cleanup action alternative for the Site described in Section 4 provides a reasonable balance of these and other MTCA and SMS objectives, including consideration of public and stakeholder preferences.

Detailed pre-RD shoreline stability modeling performed for the RI/FS demonstrated that within the North Marine Area, offshore wave attenuation structures would dissipate wave energy along the shoreline by breaking incoming storm-generated waves and by preventing wave reflection from the existing Cap Sante Boat Haven breakwater thereby reducing and/or preventing shoreline erosion in the Port Uplands Area. Construction of wave attenuation structures in the North Marine Area would also allow for permanent placement at the sediment surface of finer-grained (sand and gravel) habitat-suitable materials along the Port Uplands Area shoreline. Based on this information and other MTCA evaluations, Ecology's selected remedy for the North Marine Area includes construction of a permanent offshore wave attenuation structure (see Section 4).

The detailed pre-RD modeling also showed that for the South Marine Area offshore of the MJB North Area, application of beach armor rock would provide protection of the shoreline in a manner consistent with planned land uses. The cap would consist of a rock armor layer, with the interstices of the rock filled with gravel. This shoreline remedy would also be consistent with future development of the MJB North Area as a marina, subject to separate permitting requirements (see Section 3.1.1). Additional discussion of preliminary shoreline stabilization designs is presented in Section 4.

While the shoreline stabilization outlined above would control bank erosion in a manner consistent with land use and habitat restoration objectives (GeoEngineers et al. 2008), Ecology has determined that additional remediation of contaminated nearshore soils is necessary to meet the MTCA threshold

requirements for a remedy that protects human health and the environment and is permanent to the maximum extent practicable. Thus, Ecology's selected remedy includes removal to the maximum extent practicable of those contaminated soils from 0 to 10 ft BGS that are located within the existing shoreline buffer zone (up to 75 ft inland of the mean higher high water [MHHW] line) without adversely affecting existing buildings. Remediation levels in this application include sediment protection requirements, as discussed in more detail in Section 3.4.

3.2 Indicator Hazardous Substances

Under MTCA, "indicator hazardous substances" means the subset of hazardous substances present at a Site for monitoring and analysis during any phase of remedial action for the purpose of characterizing the Site or establishing cleanup requirements for that Site. Consistent with WAC 173-340-703, when defining cleanup requirements at a Site that is contaminated with a relatively large number of COPCs, Ecology may eliminate from consideration those hazardous substances that contribute a small percentage of the overall threat to human health and the environment. The remaining COPCs can then serve as indicator hazardous substances for purposes of defining Site cleanup requirements.

As outlined in Section 2.5, the list of COPCs (hazardous, toxic, and/or deleterious substances) identified at the Site includes:

- Metals (antimony, arsenic, chromium, copper, lead, mercury, nickel, thallium and zinc)
- Wood debris (sediment wood percentage and TVS) and its toxic degradation byproducts
- Diesel- and motor oil-range petroleum hydrocarbons
- cPAHs
- PCBs
- Dioxins/furans
- 4-Methylphenol

Indicator hazardous substances selected for the Site include all of the above COPCs with the exception of dioxins/furans and 4-methylphenol. Based on information presented in the RI/FS report (GeoEngineers et al. 2008), dioxins/furans contribute only a small percentage of the overall threat to human health and the environment at the Site, and exceedances of preliminary cleanup levels for these substances generally co-occur with other COPCs (see Figures 3 to 10 and Ecology 2008). Moreover, because of the relatively long turnaround times (typically 6 to 8 weeks) for reliable chemical analyses of dioxin/furan concentrations, near real-time compliance monitoring of dioxins/furans is problematic and was determined to be impracticable during the course of the RA. Hence, indicator hazardous substance concentrations will be used to demonstrate real-time compliance with cleanup levels during the conduct of the RA. However, soil confirmation samples will be collected and submitted for dioxin/furan analyses. If these confirmation samples subsequently reveal exceedances of dioxin/furan cleanup levels (see Table 1), residual contamination areas will be managed on-site through the implementation of appropriate institutional controls.

Only one sediment sample collected at the Site (of 43 samples total) exceeded the SQS chemical criterion for 4-methylphenol, a common woody debris degradation product, and this COPC was not detected at levels of concern in Site soils or groundwater. Because sediment cleanup levels for wood debris (as measured by wood volume and TVS) were developed using site-specific biological assessments that include the effects of 4-methylphenol, a separate cleanup level for this COPC is not necessary. Accordingly, the final list of indicator hazardous substances at the Site excludes 4-methylphenol.

3.3 Cleanup Levels

Cleanup standards consist of 1) cleanup levels that are protective of human health and the environment; and 2) the point of compliance at which the cleanup levels must be met. Preliminary site-specific cleanup standards were developed in the RI/FS and detailed information regarding the derivation of cleanup levels can be found in the RI/FS report (GeoEngineers et al. 2008). Final media-specific cleanup levels and points of compliance are summarized below.

Site-specific cleanup levels for soil that are protective of human health and terrestrial ecological receptors, and cleanup levels for groundwater that are protective of marine surface water, were developed in accordance with MTCA requirements. As discussed in Section 3.1.2, further testing of soil will be performed in selected portions of the Port Uplands Area to refine the evaluation of risk to terrestrial ecological receptors using soil bioassays. The results of the soil bioassays will be integrated into the RD, and may reduce the extent of soil remediation required in parts of the Port Uplands Area where 0 to 6 ft BGS soil concentrations exceed preliminary TEE chemical criteria.

Because Site groundwater is not a current or reasonably likely future source of drinking water, cleanup levels for Site soil need not be protective of groundwater as drinking water. Additionally, an empirical demonstration was used in the RI/FS and showed that existing chemical concentrations in Site soil are protective of groundwater as marine surface water at the proposed conditional point of compliance for groundwater.

Cleanup levels for sediments that are protective of benthic infauna were developed in accordance with MTCA and SMS requirements. Remediation levels for shoreline areas of the Site where erosion of soils could lead to deposition of contaminants in the Marine Area are discussed in Section 3.4.

Media-specific cleanup levels are discussed in the sections below.

3.3.1 Soil

Soil cleanup levels for unrestricted land use were developed in accordance with WAC 173-340-740, conservatively assuming potential future ground floor residential land use. Soil cleanup levels will apply to the soil from 0 to 15 ft BGS. Under MTCA Method B, soil cleanup levels must be as stringent as:

- Concentrations established under applicable state and federal laws
- Concentrations protective of terrestrial ecological receptors
- Concentrations protective of direct human contact with soil
- Concentrations protective of groundwater

Each of these criteria was considered during the development of soil cleanup levels, as detailed in the RI/FS report (GeoEngineers et al. 2008). Cleanup levels used in this CAP for constituents detected in Site soil are presented in Table 1.

As discussed in Section 3.1.2, the Port and K-C will conduct more detailed TEE soil bioassays in selected locations at the Port Uplands Area where COPC concentrations exceed preliminary TEE levels. The soil bioassay results will be integrated into the RD.

3.3.2 Groundwater

As discussed in Section 2.6.2, human ingestion of hazardous substances in groundwater is not a potential exposure pathway because groundwater at the Site or potentially affected by the Site is not a current or reasonable future source of drinking water. Consequently, the Site groundwater qualifies as a non-potable water source.

Arsenic and/or petroleum constituents were detected at interior monitoring wells at concentrations above levels protective of marine surface water, though concentrations at or near the groundwater/surface water interface in the porewater discharge zone are protective of Site receptors. However, as a result of remedial excavation of contaminated soil as described in Section 4, other COPCs that are present in Site soils could potentially be mobilized and released to the groundwater. Thus, post-construction groundwater compliance monitoring described in Section 7 will include all indicator hazardous substances identified in Site soils (see Section 3.2). Site-specific groundwater cleanup levels for indicator hazardous substances are presented in Table 1.

3.3.3 Sediment

Sediment cleanup levels were developed by Ecology according to MTCA and SMS requirements. Two SMS criteria are promulgated by Ecology (WAC 173-204-320). These include the SQS, the concentration below which effects to benthos are unlikely, and the CSL, the concentration above which more than minor adverse biological effects may be expected. The SQS and CSL values have

been developed for a suite of hazardous substances. SQS criteria have been selected as sediment cleanup levels for the Site, although CSLs were used to develop appropriate shoreline soil remediation levels to ensure permanent sediment protection (see Section 3.4). A summary of the site-specific sediment cleanup levels for sediment indicator hazardous/deleterious substances is provided in Table 2.

There is no promulgated SMS criterion for wood debris in sediment. In fall 2007 and 2008, supplemental sediment investigations were performed in the Marine Area (GeoEngineers et al. 2008 and Appendix B to this CAP). The primary objective of these supplemental investigations was to conduct a suite of confirmatory biological tests on synoptic surface sediment samples collected from locations representing the range of wood debris content at the Site with the potential for deleterious effects. These data were then used to develop sediment cleanup levels for wood debris at the Site. Based on interpretation of the available biological data, surface sediment TVS levels greater than 12.2 percent (dry weight basis) and/or wood debris levels greater than 25 percent (by volume) were identified as having the potential for site-specific deleterious effects exceeding SQS biological criteria.

There is also no promulgated SMS criterion for diesel and motor oil-range hydrocarbons. Based on Ecology's review of sediment bioassay data available from other similar sites with relatively weathered hydrocarbons, the MTCA Method A soil cleanup level for diesel and motor oil-range hydrocarbons is also likely to be protective of sediment and aquatic life exposures (see Tables 1 and 2). During RD, the Port and K-C will conduct more detailed site-specific biological testing, including focused bioassay testing of Site soil elutriates using Microtox® 15-minute reduction in bioluminescence (Ecology 1995), to verify the protectiveness of the Table 2 sediment cleanup levels for Site hydrocarbons. If necessary, the Table 2 sediment cleanup levels for diesel and/or motor oil-range hydrocarbons will be modified during RD to conform to the bioassay data.

3.4 Shoreline Buffer Zone Soil Remediation Levels

A remediation level defines a concentration of a hazardous substance in a particular medium above which a particular cleanup action component must be used (WAC 173-340-200). In practice, a remediation level is a contaminant concentration that is above a cleanup level. When soil contamination is above the soil remediation level, different cleanup actions may be employed than for contamination between the remediation level and the cleanup level. For example, soil with contamination above a remediation level may be excavated whereas soil with contamination below the remediation level but above the cleanup level may be managed on site using other technologies.

As discussed in Section 3.1.3, while the shoreline stabilization and capping actions included as part of the selected Site remedy would control bank erosion in a manner consistent with land use and habitat restoration objectives, additional remediation of contaminated nearshore soils is necessary to meet MTCA threshold requirements for permanent protection of human health and the environment and protection of sediment quality. Nearshore soil remediation levels were developed for this element of the remedy and will be used as one of the performance monitoring standards when excavating nearshore contaminated soils from 0 to 10 ft BGS that are located within the existing shoreline buffer zone (up to 75 ft inland of MHHW). Nearshore soil remediation levels in this application consider CSL chemical criteria for sediment indicator hazardous substances (applicable to the entire 0 to 10 ft BGS depth interval to ensure that adjacent sediments will not recontaminate above SQS chemical criteria, considering sediment transport conditions at the Site under a hypothetical future shoreline erosion scenario). The other performance monitoring standard is based on MTCA Method B (human health and ecological) cleanup levels (applicable to the 0 to 6 ft BGS depth interval; see Figures 17 to 19). The more restrictive of the two performance monitoring standards applicable to each depth interval will be used. A summary of the site-specific nearshore soil remediation levels that are applicable to the 0 to 6 and 6 to 10 ft BGS depth intervals is provided in Table 2.

As discussed in Section 3.3.3, during RD, the Port and K-C will conduct more detailed site-specific bioassay testing of Site soil elutriates to verify the protectiveness of the Table 2 nearshore soil remediation levels for Site hydrocarbons. If necessary, the Table 2

remediation levels for diesel and/or motor oil-range hydrocarbons will be modified during RD to conform to the bioassay data.

3.5 Points of Compliance

Under MTCA, the point of compliance is the point or location on a site where the cleanup levels must be attained. This section describes the points of compliance for soil, groundwater, and sediment.

3.5.1 Soil

The standard point of compliance for the soil cleanup levels shown in Table 1 will be throughout the soil column from the ground surface to 15 ft BGS, in accordance with WAC 173-340-740(6)(d) and WAC 173-340-7490(4)(b). For potential terrestrial ecological exposures, MTCA regulations allow a conditional point of compliance to be established from 0 to 6 ft BGS (the biologically active zone according to MTCA default assumptions), provided that environmental covenants are used to address potential excavation of deeper soil [WAC 173-340-7490(4)(a)]. Accordingly, in areas of the Site where potential ecological exposures are a concern, and where appropriate environmental covenants can be implemented, a conditional point of compliance for soil concentrations protective of terrestrial ecological receptors will apply throughout the soil column from 0 to 6 ft BGS.

Subject to final engineering and RD analyses, there will likely be limited areas of the Site where attainment of soil cleanup levels within the 0 to 6 ft BGS conditional point of compliance is impracticable, such as immediately adjacent to or beneath existing Site buildings. In such localized areas, and consistent with WAC 173-340-740(6)(f), other engineering approaches such as capping the soil with asphalt or concrete barriers, or placement of an indicator layer and clean soil cap (similar to remedial actions previously implemented at Port Parcel 2 (ThermoRetec 2000) may provide the necessary environmental protection. Cleanup specifications for these localized areas will be developed during RD, as appropriate.

3.5.2 Groundwater

Because groundwater cleanup levels are based on protection of marine surface water and not protection of groundwater as drinking water, the conditional point of compliance for the groundwater cleanup levels is the point of groundwater discharge to the Cap Sante Waterway and Fidalgo Bay. This corresponds to the groundwater/surface water interface at the Port Uplands Area and the MJB North Area. At the Port Uplands and MJB North Areas, shoreline wells will be used to evaluate compliance.

3.5.3 Sediment

For marine sediments potentially affected by hazardous substances, the point of compliance for protection of the environment is surface sediments within the biologically active surface water habitat zone, represented by samples collected across the top 10 cm (i.e., 0 to 0.3 ft) below the mudline.

3.6 Applicable Regulatory Requirements

In addition to the cleanup standards developed through the MTCA and SMS process and presented in Section 3.1, other regulatory requirements must be considered in the selection and implementation of the cleanup action. MTCA and SMS require the cleanup standards to be "at least as stringent as all applicable state and federal laws" [WAC 173-340-700(6)(a)]. Besides establishing minimum requirements for cleanup standards, applicable state and federal laws may also impose certain technical and procedural requirements for performing cleanup actions. These requirements are described in WAC 173-340-710. Applicable state and federal laws are discussed below.

The cleanup action at the Site will be performed pursuant to MTCA under the terms of the Consent Decree between Ecology and the implementing Potentially Liable Parties (PLPs). Accordingly, the selected cleanup action meets the permit exemption provisions of MTCA, obviating the need to follow most procedural requirements of the various local and state regulations that would otherwise apply to the action. Ecology will determine the substantive provisions of state and local laws and regulations that are applicable to this project, following consultation with appropriate state and local regulators.

Similarly, the cleanup action also qualifies for a USACE Nationwide Permit 38 (NWP 38). Nevertheless, federal consultation under the Endangered Species Act, Section 401 Water Quality Certification, and other substantive requirements must still be met by the cleanup action. The USACE will separately be responsible for issuing approval of the project under NWP 38, following Endangered Species Act consultation with the federal Natural Resource Trustees, and also incorporating Ecology's 401 Water Quality Certification.

3.6.2 Solid and Hazardous Waste Management

The Washington Hazardous Waste Management Act and the implementing regulations, the Dangerous Waste Regulations (Chapter 173-303 WAC), would apply if dangerous wastes are generated during the cleanup action. There is no indication of listed wastes being generated or disposed of at the Site. Based on the analytical data generated during the RI/FS, only limited volumes of soil and/or sediment at the Site may be characterized as dangerous waste if excavated or dredged. The Dangerous Waste Regulations would be applicable only if post-removal sampling of excavated/dredged material (e.g., toxicity characteristic leaching procedure [TCLP] sampling, if required by the receiving landfill) indicated contaminant concentrations exceeding levels associated with dangerous waste characteristics or criteria. Related regulations include state and federal requirements for solid waste handling and disposal facilities (40 Code of Federal Regulations [CFR] 241, 257; Chapter 173-350 and -351 WAC) and land disposal restrictions (40 CFR 268; WAC 173-303-340).

3.6.3 Puget Sound Dredged Material Management Program

In Puget Sound, the open water disposal of sediments is managed under the DMMP. This program is administered jointly by the USACE, USEPA, the Washington Department of Natural Resources (WDNR), and Ecology. The DMMP developed the Puget Sound Dredge Disposal Analysis (PSDDA 2000) protocols, which include testing requirements to determine whether dredged sediments are appropriate for open water disposal. The DMMP has also designated disposal sites throughout Puget Sound. While initial characterization

data collected during the RI/FS (GeoEngineers et al. 2008) indicates that sediment woody debris to be dredged from the Site is likely suitable for confined openwater disposal (e.g., at the Port Gardiner DMMP site), if this option is pursued by the implementing parties, additional characterization work would be required during RD to complete the suitability determination and permitting process. Use of PSDDA facilities would need to comply with other DMMP requirements including material approval, disposal requirements, and payment of disposal site fees.

3.6.4 State Environmental Policy Act

The State Environmental Policy Act (SEPA) (Revised Code of Washington [RCW] 43.21C; WAC 197-11) and the SEPA procedures (WAC 173-802) are intended to ensure that state and local government officials consider environmental values when making decisions. The SEPA process begins when an application for a permit is submitted to an agency, or an agency proposes to take some official action such as implementing a MTCA CAP. Prior to taking any action on a proposal, agencies must follow specific procedures to ensure that appropriate consideration has been given to the environment. The severity of potential environmental impacts associated with a project determines whether an Environmental Impact Statement is required. A SEPA checklist will be required prior to initiating remedial construction activities. Because the Site cleanup action will be performed under a Consent Decree, SEPA and MTCA requirements will be coordinated, if possible. The Port is the lead SEPA agency for this action.

3.6.5 Shoreline Management Act

The Shoreline Management Act (RCW 90.58) and its implementing regulations establish requirements for substantial developments occurring within water areas of the state or within 200 feet of the shoreline. The City of Anacortes has set forth requirements based on local considerations such as shoreline use, economic development, public access, circulation, recreation, conservation, and historical and cultural features. Local shoreline management plans are adopted under state regulations, creating an enforceable state law. Because the Site

cleanup action will be performed under a Consent Decree, compliance with substantive requirements will be necessary, but a shoreline permit will not be required.

3.6.6 Washington Hydraulics Code

The Washington Hydraulics Code establishes regulations for the construction of any hydraulic project or the performance of any work that will use, divert, obstruct, or change the natural flow or bed of any of the salt or fresh water of the state. The code also creates a program requiring Hydraulic Project Approval (HPA) permits for any activities that could adversely affect fisheries and water resources. Timing restrictions and technical requirements under the hydraulics code are applicable to dredging and placement of cover sediments if necessary. The RI/FS was prepared using costs and durations that recognize potential fish closure periods, during which time dredging and any in-water work will not be permitted (GeoEngineers et al. 2008). Exact closure periods will be determined through agency consultation.

3.6.7 Water Quality Management

The Clean Water Act (CWA) is the primary federal law for protecting water quality from pollution. In addition to federal law, water quality is regulated by Ecology under the state water quality act, RCW 90.48. The CWA regulations provide requirements for the discharge of dredged or fill material to waters of the United States and are applicable to any in-water work. The CWA regulations also prescribe requirements for point source and non-point source discharges. Section 404 of the CWA requires permits from the USACE for discharges of dredged or fill material into waters of the United States, including wetlands. Requirements for all known, available, and reasonable technologies for treating wastewater prior to discharge to state waters are applicable to any dewatering of marine sediment prior to upland disposal. Section 401 of the CWA requires the state to certify that federal permits are consistent with state water quality standards are reflected in the Section 401 certification, the certification generally stands in the stead of a stand-alone determination by Ecology of state water quality provisions

applicable to the cleanup action. The substantive requirements of a certification determination are applicable. State and federal standards for marine waters will be applicable to discharges to surface water during sediment dredging, and return flows (if necessary) to surface waters from dewatering operations.

Construction activities that disturb 1 acre or more of land need to comply with the provisions of state construction stormwater regulations, and a stormwater permit will be required for the cleanup action. A stormwater pollution prevention plan or equivalent MTCA construction quality assurance project plan (CQAPP) will be prepared prior to activities that will disturb 1 acre or more of soil. The CQAPP will document planned procedures designed to prevent stormwater pollution by controlling erosion of exposed soil and by containing soil stockpiles and other materials that could contribute pollutants to stormwater. It is anticipated that a CQAPP will be prepared as part of the RD process, and supplemented as appropriate by the remedial contractor.

3.6.8 Other Potentially Applicable Regulatory Requirements

The following is a list of other potentially applicable regulations for the cleanup action:

- Air Emissions Applicable for site grading or excavation work that could generate dust. Controls would need to be in place during construction (e.g., wetting or covering exposed soils and stockpiles), as necessary, to meet the substantive restrictions on off-site transport of airborne particulates by the local agency, the Northwest Clean Air Agency.
- Archeological and Historical Preservation The Archeological and Historical Preservation Act (16 USCA 496a-1) would be applicable if any subject materials are discovered during RD or site grading and excavation/dredging activities. A cultural resources assessment will be performed during RD to determine whether cleanup activities could affect historical archaeological remains that might be located in on-site fill, or affect prehistoric archaeological remains that could be located beneath the fill in past upland and marine locations. This information will be included in the USACE NWP 38 permit application.

- Health and Safety Site cleanup-related construction activities will need
 to be performed in accordance with the requirements of the Washington
 Industrial Safety and Health Act (RCW 49.17) and the federal
 Occupational Safety and Health Act (29 CFR 1910, 1926). These
 applicable regulations include requirements that workers are to be
 protected from exposure to contaminants and that excavations are to be
 properly shored.
- Minimum Standards for Construction and Maintenance of Wells –
 Groundwater monitoring wells will need to be installed as a part of the
 post-construction water quality confirmation monitoring outlined in
 Section 7. The new wells will be constructed in accordance with the
 requirements of WAC 173-160 to further ensure protection of
 groundwater resources at the Site.

4.0 SELECTED SITE CLEANUP ACTION

The cleanup action selected by Ecology for the Site incorporates different actions targeted to different zones of the Site. The actions to be taken for each zone are interdependent. The following sections discuss the actions for each zone.

Many of the cleanup actions require soil excavation and backfilling to restore grade. All backfill soils must come from a source approved by Ecology and must have suitable geotechnical characteristics.

4.1 Port Uplands Interior Area Cleanup

Alternative PUA-4 has been selected as the cleanup action for the Port Uplands Area. This section describes the components of the cleanup action for those portions of the Port Uplands Area located landward of the shoreline buffer zone (Figure 13). Based on the evaluations presented in the RI/FS (GeoEngineers et al. 2008), PUA-4 has been selected as the remedy for this area of the Site, as described in more detail in Section 5 below. For the purpose of describing the cleanup action in this CAP, the Port Uplands Area has been divided into two separate areas that address different exposure characteristics. The shoreline area consists of a buffer zone that extends inland from the MHHW level to approximately 75 ft west (upland) of the MHHW mark. Remedial actions in the shoreline zone are described along with sediment cleanup actions in Section 4.3 below. The remainder of the Port Uplands Area inland of the shoreline area addresses soil cleanup based on different exposure pathways and remediation levels and will be addressed separately from the shoreline area. These two adjacent areas are shown in Figure 13 and are described in the following sections.

The cleanup action for the Port Uplands Area, outside of the shoreline buffer zone, achieves removal to the maximum extent practicable of contaminated soil in the vicinity of MW-110 and focused removal of remaining contaminated soil at depths designed to address primary exposure pathways. The proposed cleanup action includes the following components:

 Excavate contaminated soil and overburden from areas of the Port Uplands Area west of the shoreline buffer zone, including:

- Soil between 0 and 10 ft BGS at Port Parcel 1 containing metals at concentrations exceeding MTCA Method B cleanup levels
- Soil between 0 and 15 ft BGS in the vicinity of monitoring well MW-110 containing free product, diesel-range hydrocarbons, or motor oil-range hydrocarbons exceeding MTCA Method B cleanup levels
- o Soil between 0 and 6 ft BGS in remaining areas of the Port Uplands Area containing metals, diesel or motor oil-range hydrocarbons, or cPAHs at concentrations exceeding MTCA Method B cleanup levels
- Transport contaminated soil to appropriate disposal facility
- Backfill excavations with clean imported fill and restore original Site topography, features, and surfaces
- Develop environmental covenants to address remaining contaminated soil left in place below 6 ft BGS across the remainder of the Port Uplands Area
- Install a monitoring well network and monitor groundwater quarterly for at least one year

Prior to implementation, applicable regulatory requirements will be addressed as described in Section 3.6. The following sections provide further description of the components of the selected cleanup action for the Port Uplands Area.

4.1.1 Soil Removal

The anticipated areas and depths of soil removal activities associated with the selected cleanup action are depicted in Figure 13. Soil exceeding cleanup levels for metals, diesel and motor oil-range hydrocarbons, and cPAHs would be removed to varying degrees. Petroleum-contaminated soil adjacent to the Park Building would be excavated to the maximum practicable extent as generally shown on Figure 13. The existing Seafarers' Memorial Park building on Port Parcel 3 is expected to be preserved during remedial actions. It is anticipated that soils exceeding the cleanup level for diesel and motor oil-range petroleum hydrocarbons, if found to be located below the building, would be left in place.

The soil removal associated with this cleanup action is expected to be performed using commonly available excavation techniques. The construction methods

would be specified during the RD phase or by the selected cleanup contractor, but are likely to include the following:

- Excavation near buildings would utilize sheetpile walls or other
 equivalent shoring methods to protect the structural integrity of the
 buildings. Soil removal will be performed in a manner that will prevent
 demolition of the Park Building, potentially resulting in leaving
 inaccessible contaminated soil in place
- Excavations extending below 10 ft BGS would be completed using commonly available dewatering techniques to minimize the water content of the excavated materials to the extent possible
- The excavations would be completed in a manner that allows segregation and reuse of clean overburden soil, resulting in approximately 30 percent of excavated soil allowed to be reused as clean backfill
- Excavation of soil from areas on the upland side of the shoreline buffer zone as shown in Figure 13 results in approximately 21,600 cy excavated, including approximately 15,100 cy of contaminated soil and 6,500 cy of overburden soil assumed to be clean and suitable for on-site beneficial reuse.

4.1.2 Soil Disposal and Treatment

As noted above, it is assumed that the soil removal activities would be completed in a manner that allows segregation and reuse of clean overburden soil. The excavated soil would be characterized for disposal as required by MTCA and Washington State Dangerous Waste regulations and the selected disposal facility. Excavated contaminated soil is expected to fall into two categories: 1) non-dangerous waste suitable for disposal at a Subtitle D landfill; or 2) dangerous waste requiring either disposal at a Subtitle C (hazardous/dangerous waste) facility or treatment prior to disposal at a Subtitle D facility.

For soil to be categorized as non-dangerous waste and suitable for disposal at a Subtitle D landfill, it will be necessary to demonstrate that Site contaminants are not present at concentrations greater than ten times the Universal Treatment

Standards (UTS), as defined in 40 CFR 268.48. This requirement includes the results of TCLP testing for metals.

It is expected that some of the excavated soil would be precluded from disposal at a Subtitle D landfill as non-dangerous waste based on exceeding ten times the UTS for lead. Lead has been detected in soil at the Port Uplands Area at concentrations that would potentially result in failure of this rule based on the TCLP. However, soil samples collected during the September 2008 supplemental sampling described in Appendix A indicated undetectable concentrations of lead in leachate samples of soil with total lead concentrations up to 680 mg/kg. Based on the TCLP results from the supplemental sampling, it is expected that the volume of upland soil that will fail TCLP for lead may be a small percentage of the total contaminated soil volume. Soil that fails TCLP as a result of disposal characterization will be transported to a permitted facility for off-site treatment and/or disposal.

4.2 MJB North Interior Upland Area Cleanup

Alternative MJB-4 has been selected as the cleanup action for the MJB North Area. Similar to Section 4.1, for the purpose of describing the cleanup action in this CAP, the MJB North Area uplands have been divided into two separate areas that address different exposure characteristics. The shoreline area consists of a buffer zone that extends inland from the MHHW level to approximately 75 ft west (upland) of the MHHW mark. Remedial actions in the shoreline zone are described along with sediment remedial actions in Section 4.3 below. The remainder of the MJB upland area inland of the shoreline area addresses soil cleanup based on different exposure pathways.

Figure 14 presents the anticipated areas and depths of soil removal in Alternative MJB-4, which consists of the following elements for the upland area west of the 75 ft buffer zone:

Remove shallow soil (0 to 6 ft BGS) with contaminant concentrations that exceed
 MTCA Method B cleanup levels throughout the uplands area

- Characterize and dispose of excavated soil in accordance with applicable regulations
- Perform confirmation sampling
- Backfill excavations with clean fill
- Develop environmental covenants to address remaining contaminated soil left in place below 6 ft BGS across the remainder of the MJB North Area
- Install a monitoring well network and monitor groundwater quarterly for at least one year

The estimated MJB North Area interior (i.e., west of the 75 ft buffer zone) excavation volume is approximately 3,500 cy. Excavation within the MJB interior under Alternative MJB-4 is not intended to extend below 6 ft BGS. However, localized areas of deeper soil contamination containing indicator hazardous substance concentrations above the cleanup levels listed in Table 1 may remain on-site. The 6 ft excavation depth will limit worker restrictions and concurrently limit the potential for terrestrial biota exposure. Furthermore, site characterization data presented in the RI/FS report (GeoEngineers et al. 2008) demonstrated that groundwater at the shoreline wells complies with the groundwater cleanup levels, indicating that leaching of subsurface (greater than 6 ft BGS) soil contaminants to groundwater is not an exposure pathway of concern. Therefore, soils left in place under this alternative will not be a source of mobile contamination that would affect marine surface water or sediments. The upland and marine area remedies (see Section 4.3) will be completed concurrently to minimize the amount of excavation dewatering needed and ensure as complete a removal as practicable within the designated areas.

Prior to implementation, applicable regulatory requirements will be addressed as described in Section 3.6. A utility locate will be conducted, the concrete rails on the east side of the property (Figure 14) will be demolished, and the existing monitoring wells (MW-1 through MW-7) abandoned except where they can be protected during cleanup and redevelopment activities. Approximately four post-construction monitoring events will likely be sufficient to confirm that groundwater at the MJB North Area has not been affected by Site contaminants.

4.3 Shoreline Buffer Zone and Marine Area Cleanup

Marine Alternative M-1 has been selected as the cleanup action for the Marine Area of the Site. This alternative removes surface sediments that exceed SQS chemical and biological criteria in intertidal and subtidal areas, backfills all excavations with clean material, and provides for a protective cap throughout the intertidal area. Where the Marine Area connects to the shoreline buffer excavation area, a thickened cap edge will be created to provide additional armoring and prevent potential undermining of the cap from wave action along the shoreline. Erosion protection is provided by offshore wave attenuation structures along the Port property, and with armor rock along the MJB North Area property. Alternative M-1 also provides for mitigation of eelgrass disturbed by the remedial activities (as required by forthcoming permit actions), as well as potential on-site opportunities for additional eelgrass and/or macroalgae restoration.

The Marine Area cleanup action includes the following elements:

- Remove surficial debris in the intertidal area
- Remove the existing seasonal dock structure to allow for access
- Cut off or remove (as practicable) piles along the shoreline
- Dredge subtidal surface sediments exceeding SQS criteria
- Backfill subtidal excavations with clean sand to restore existing grades and manage anticipated dredge residuals within the excavation area
- Construct wave attenuation structures offshore of the Port Uplands Area and armored caps offshore of the MJB North Area to provide transitional slope cap protection (Figure 16)
- Excavate the shoreline transitional slope to facilitate cap placement while maintaining the approximate existing grades
- Transition shoreline buffer zone and Marine Area excavations as appropriate (Figures 17 to 19)
- Place a minimum of 2 ft of suitably-sized cap material (including a lower minimum 1-ft layer of quarry spalls) along the Port Uplands and MJB North Area shorelines that have been identified as potential erosional sources of localized contaminated sediment deposits
- Restore the existing seasonal dock structure.

The Marine Area dredge and excavation volume is approximately 30,200 cy, and the shoreline buffer excavation volume is approximately 10,600 cy (5,100 cy in the Port Uplands Area and 5,500 cy in the MJB North Area). Approximately 3,000 to 4,000 cy of this shoreline buffer excavation volume is anticipated to be clean soil potentially suitable for on-site beneficial reuse as backfill.

The Port and K-C will conduct supplemental soil sampling in parts of the shoreline buffer zone (including near PP-12 and PP-19), primarily to refine characterization of cPAH concentrations in soils between 0 and 6 ft BGS. These additional data may result in minor modifications to the shoreline buffer excavation boundary and volume in the MJB North Area. The focused soil sampling is currently scheduled for late January 2009, and the results will be integrated into RD.

4.3.1 Shoreline Buffer Zone Excavation

The intent of the shoreline buffer zone remediation is to ensure that the cleanup action protects human health and ecological receptors, prevents recontamination of marine sediments, and supports habitat development along the shoreline. Figure 15 presents the anticipated areas of shoreline soil and marine sediment removal associated with the selected site-wide cleanup action. This selected cleanup action removes contaminated nearshore soil (exceeding cleanup and/or remediation levels, as appropriate) to a depth of 10 feet BGS as practicable within the shoreline buffer zone; this is the soil considered most likely to impact marine sediments if the shoreline erodes. Deeper contaminated soil within the shoreline buffer zone will either be left in place below existing clean soil or will be covered by clean fill following excavation of overlying contaminated soil. Shoreline soil within the upper 6 ft that exceeds MTCA Method B soil cleanup levels, CSLs, or (forthcoming) site-specific petroleum remediation levels (see Figures 17 to 19) will be excavated for off-site disposal under this alternative. Similarly, shoreline soil within the 6 to 10 ft BGS interval that exceeds CSLs or petroleum remediation levels will be excavated for off-site disposal. Clean soil overlying deeper contaminated soil will be stockpiled, sampled, and re-used if it is determined to be chemically and geotechnically suitable.

4.3.2 Intertidal Area Shoreline Transitional Slope Cleanup

In the intertidal area, the selected cleanup action includes excavation of sediment, wood debris, and brick, along with removal or cutting of piles, to facilitate placement of a minimum 2-ft-thick cap to contain deeper wood debris deposits, and backfilling the area to retain the original bathymetry within the shoreline transitional slope (Figure 16). Removal and disposal methods are similar to those described for the subtidal area.

Where the intertidal area connects to the shoreline buffer zone excavation (as shown in Figures 17 and 19), the intertidal excavation will be performed to an elevation sufficient to provide for an armor layer to minimize the potential for erosion at the edge of the cap from breaking waves on the slope. While the wave attenuation and armoring schemes described in Section 4.3.5 are designed to prevent cap erosion, the armored intertidal cap section in the Port Marine Area will provide extra protection (e.g., quarry spalls will comprise the lower 1 ft of the cap). The 10 ft BGS excavation at the MHHW line will extend offshore at an approximate 1H:1V slope up to the bottom of the proposed offshore cap prior to backfilling and capping (see Figures 17 and 19). At the MHHW line, the intertidal excavation will connect directly to the 75 ft shoreline buffer zone excavation. Supplemental characterization of the Port Parcel 2 soils immediately east of the existing sheetpile wall is planned for January/February 2009. The results of this supplemental sampling will be integrated into RD to ensure that shoreline contamination that may exist in this area is appropriately addressed.

Adjacent to the northern and central Port Uplands Area, the transitional intertidal slope cap will be a minimum of 2 ft thick, consisting of a lower 1-ft layer of quarry spalls covered with a 1-ft layer of surficial sand and gravel mixture. The cap does not extend into subtidal areas (Figure 16). Along the contiguous northern portion of the MJB North Area south of the influence of the wave attenuation structure, a minimum 2-ft-thick armored cap will be placed, consisting of a thicker base layer of rock armor material with an overlying surface layer of sand and gravel. Detailed cap designs for various areas of the Site will be developed during RD. The current contours of the shoreline and

subsurface areas will be re-established by backfilling once the transitional slope excavation has been completed.

4.3.3 Subtidal Zone Dredging

The selected cleanup action includes dredging of surficial sediment and wood debris in the subtidal area that exceeds SQS criteria (Figure 15). A 2-ft thickness of sediment will be excavated using either land- and/or water-based equipment depending on work area and tidal conditions. After excavation, sediment will be handled using one of the following three options:

- Sediment will be offloaded and transported to an approved off-site location for beneficial reuse (e.g., for use in regional topsoil products)
- Sediment will be transported by barge and disposed of at a suitable openwater disposal site such as the Port Gardner non-dispersive DMMP disposal site after larger wood and debris greater than 2 ft in any dimension is removed
- If neither of the above options is determined to be practicable, sediment will be offloaded, transported, and disposed of at an approved upland landfill

During RD, sediment within the prospective dredge prism will be further characterized to verify its suitability for open-water disposal. Potential disposal and upland beneficial reuse opportunities for these materials will also be finalized during RD. Based on these RD evaluations, one or more of the three options above may be implemented.

The majority of the subtidal dredged areas will be backfilled to approximate the existing grade with clean materials of differing grain size, depending on stable grain sizes and habitat design specifications for specific locations within the Marine Area. Grain size and other engineering specifications for the backfill material will be determined during RD, following USEPA, USACE, and other relevant design guidance (e.g., Palermo et al. 1998) and in consultation with natural resource agencies for habitat considerations.

4.3.4 Erosion Protection

As described in the RI/FS report, the present wave environment in the Marine Area has resulted in significant erosion along the Port Uplands Area and MJB North Area shorelines. Erosion of soil from these shoreline areas is the likely source of down-drift sediment contamination observed just to the south, and erosion protection is thus a key required element of the cleanup action. Integrated shoreline erosion and source control objectives will be achieved under the cleanup action using offshore wave attenuation and cap armoring as described below.

Offshore Wave Attenuation

To protect the Port Uplands Area shoreline from future erosion, offshore wave attenuation structures will be constructed to provide permanent and effective wave attenuation as described in the RI/FS, at the locations shown in Figure 16. By reducing incoming wave energy, a permanently stable shoreline cap can be constructed using sand and gravel materials in lieu of larger armor substrate materials.

Numerical modeling of wave conditions in the Marine Area is described in the FS. The numerical modeling evaluated both armored cap and wave attenuator alternatives for shoreline protection. The modeling results showed that the wave attenuation structures will more effectively dissipate the wave energy along the Port Uplands Area shoreline by breaking incoming storm-generated waves and by preventing wave reflection from the existing Cap Sante Boat Haven breakwater. The wave attenuation structure will allow for permanent placement of sand and gravel materials along the shoreline.

The wave attenuation structure will be constructed using imported rock with crest elevations ranging up to +12 ft MLLW.

Armored Cap

Along the Port Upland Area, the lower portion of the transition slope cap will be comprised of a quarry spall armor layer with a minimum thickness of 1 ft. This

armor layer will provide long-term protection of the confined underlying sediments from direct wave-break action when exposed by tides. The armor layer will be covered by a minimum of 1 ft of sand and gravel material as shown on Figures 17 and 18.

Along the MJB North Area shoreline, the shoreline cap will be protected from erosion with a rock armor layer placed along the shoreline. Armored caps will be constructed to the extents shown in the Figure 16 plan view. The cap will include a minimum 2-ft-thick rock armor layer along with a 0.5-ft-thick top-dressing of sand and gravel that will be placed in the interstices of the rock.

4.3.5 Eelgrass and Macroalgae Restoration Opportunities

Disturbance of existing eelgrass beds is anticipated during construction of the remedy. Subtidal sediments in the dredge area will be excavated and replaced with clean sand and gravel in areas of existing eelgrass. Eelgrass beds disturbed by the cleanup action will be re-planted after backfilling and capping have been completed. There are potential opportunities for additional eelgrass and macroalgae restoration and placement of clean sand in areas located outside of the dredging limits within the protected area created by the wave attenuation structures, as shown in Figure 16.

4.4 Contamination Remaining On-Site Following Remedy

The selected cleanup action for the Port and MJB Uplands Areas is expected to contain subsurface soil and sediment in place at several locations across the Site with hazardous substance concentrations exceeding MTCA soil or SMS sediment cleanup levels listed in Tables 1 or 2 for the respective properties.

As described in Section 4.2.1 above, the cleanup strategy for the Port Uplands Area relies on removing contaminated soil to varying depths across the Site for the purpose of removing particular exposure pathways. These areas of residual contaminated soil will be documented following completion of the cleanup action and will continue to be addressed through the use of confirmation monitoring and environmental covenants implemented at the Site, as described in Sections 4.5 and 4.7 below. The anticipated

areas where contaminated soil will be contained in place at the Port Uplands Area include the following:

- Park Building Soil contaminated with petroleum hydrocarbons, lead, and cPAHs is expected to underlie the Park Building. If encountered, the contaminated soil will be left in place under the current Park Building and at a distance from the building established to ensure the structural integrity of the building rather than demolishing the building to achieve complete removal. The contaminated soil under the building is expected to extend to a depth of 15 ft BGS or greater, based on data collected from the perimeter of the building.
- Monitoring Well MW-102 Soil in the vicinity of monitoring well MW-102 on the east side of R Avenue has arsenic concentrations above cleanup levels. Soil in this location will be removed to the extent possible. However, it is suspected that the arsenic-contaminated soil may extend under the Anacortes Concepts, L.L.C building to the east as well as below R Avenue. The excavation of this location will be performed so as not to impact the structural integrity of the building or the roadway, resulting in the potential to leave contaminated soil in place.
- Northwest Educational Service Building The east end of the Northwest
 Educational Service Building was constructed over deep contaminated soil
 present near the southern end of the subsurface containment wall installed
 adjacent to the shoreline in Parcel 2. The excavation activity performed in the
 vicinity of this building is expected to leave contaminated soil in place
 underneath the building and at a distance from the building established to
 ensure the structural integrity of the structure.
- Other Areas Below 6 ft BGS The cleanup goal for the areas of the Port Uplands Area outside of the shoreline buffer zone, the deep hydrocarbon removal area adjacent to the Park Building, and the Parcel 1 arsenic removal area is to remove contaminated soil in the upper 6 ft that exceeds cleanup levels. There are areas of soil contamination below 6 ft BGS that will be left in place. These areas are either currently underneath a 6-ft column of clean soil, or will be below 6 ft of clean backfill following soil removal activities in the upper 6 ft.

In the MJB North Area, the estimated maximum depth of contamination is approximately 16 ft BGS along the shoreline, which is the maximum depth of the fill material. Because excavation for Alternative MJB-4 will not extend below 10 ft BGS, localized areas of deeper soil contamination will remain at the MJB North Area at concentrations exceeding soil cleanup levels, particularly within portions of the shoreline buffer zone.

In both the Port Uplands and MJB North Areas, post-excavation confirmation samples will be collected and will be used to verify the remaining contaminant mass at the Site following completion of remedial actions (see Section 4.5 below). The remaining contamination will be protected from erosion by up to 10 ft of clean shoreline backfill.

Section 4.9 below discusses environmental covenants required for the portions of the Port and MJB North Areas where complete removal of soil exceeding applicable cleanup levels (Table 1) will not be achieved.

4.5 Construction Performance Monitoring

Performance monitoring following excavation and/or dredging will initially include topographic or bathymetric surveys to verify that at least 90 percent of the excavation/dredge area has achieved the required cut elevations, with the caveat that "high-spots" above the required elevations (i.e., up to 10 percent of the area) are relatively isolated (i.e., non-contiguous), and not the result of intentional bias during implementation.

Once required excavation or dredging elevations have been verified as outlined above, performance monitoring will involve collecting soil or sediment samples from the base and sidewalls of the upland excavations to confirm that soil remediation and/or cleanup levels have been achieved and/or to document concentrations of contaminants remaining on site. Performance monitoring activities will include the following:

 Collect discrete grab samples from the final limits of the upland and sediment remedial excavations, with the sampling density appropriately tailored to the location and size of the excavation (detailed post-construction verification sampling plans will be developed during RD).

- The confirmatory soil and sediment samples will be submitted for analysis of
 indicator hazardous substances and wood debris indicators listed on Tables 1
 and 2, respectively to verify that the excavation actions are complete or to
 document remaining contaminant mass at the Site.
- Samples will be analyzed on a short turnaround basis to allow the results to be compared with soil remediation and/or soil/sediment cleanup levels shown in Tables 1 and 2 to evaluate whether the final limits of the remedial excavations have been achieved.
- Within certain Site areas including the shoreline buffer zone, the base of the
 excavations will be dictated by the design excavation depth rather than
 achievement of cleanup or remediation levels. In these areas, the excavation base
 samples will be used to document the extent of contamination remaining
 following completion of the excavation.

Performance samples exceeding the cleanup levels will be evaluated using Ecology's Statistical Guidance for Site Managers (Ecology 1992). For shallow upland excavations in the MJB North Area, the statistics will be completed using all of the post-excavation performance samples in the remaining upland area to obtain a large enough sample set to achieve statistical significance.

In addition to the remedial excavation performance samples described above, sampling of dewatering effluent may be performed if this water is discharged to the City of Anacortes sewer system. Samples of water that is pumped from the excavations by the construction dewatering system would be collected on a periodic basis as required by the City of Anacortes. The treated dewatering effluent sample results will be evaluated by the Port for compliance with the City's water quality standards for discharge to the sanitary sewer.

4.6 Post-Construction Confirmation Monitoring

The limited groundwater impacts identified at several interior monitoring wells within the Port Uplands Area and nearshore sediment impacts are directly associated with areas of soil contamination that will be addressed by the selected cleanup action. The soil removal included as part of the remedy is expected to result in a reduction of contaminant concentrations in groundwater, thereby obviating the need for active groundwater remediation. However, to verify that the proposed cleanup action is protective of groundwater, a network of monitoring wells will be installed at the Site and sampled for Site indicator hazardous substances. Similarly, to verify that shoreline protection, cap/backfill stability, substrate suitability, and habitat mitigation/restoration objectives are met, long-term sediment and habitat recovery monitoring will be performed. Detailed post-construction water and sediment quality confirmation monitoring plans will be developed during RD. The exact locations of monitoring wells and sediment sampling locations will be determined following the completion of remedial actions based on the final dimensions of the excavation areas.

Groundwater will be sampled on a quarterly basis at each of the 12 monitoring wells either retained or installed following remedy implementation for a minimum of four consecutive quarters. Groundwater samples will be analyzed for the full list of soil indicator hazardous substances (see Table 1), including dissolved metals, diesel- and motor oil-range hydrocarbons, PCBs, and PAHs. Following completion of four consecutive quarters of groundwater sampling, long-term groundwater monitoring will be implemented. The monitoring frequency, number and locations of monitoring wells, and analytical parameters will be modified by the Port and K-C as appropriate based on the results of the quarterly monitoring and in consultation with Ecology.

4.7 Contingency Actions

Post-excavation performance sampling will ensure that contaminated soils and sediments are removed, as practicable. Groundwater and sediment monitoring will ensure that deeper contaminated soils left in place do not pose a risk to the marine area via contaminant migration to groundwater and sediment/surface water. Investigations completed during the RI/FS demonstrated that groundwater at the shoreline wells complies with the proposed groundwater cleanup levels, indicating that leaching of soil contaminants to groundwater is not currently an exposure pathway of concern. However, the remedial efforts along the shoreline may create soil disturbances that mobilize deeper contaminants. If contaminants exceed the cleanup levels in groundwater and/or sediment samples after an initial four quarters of post-remedial action confirmation monitoring, semi-annual groundwater and/or sediment monitoring

will be conducted, as appropriate. If groundwater and/or sediment samples continue to exceed the cleanup levels without abating, additional actions will be considered. Similarly, if long-term monitoring data reveal that shoreline protection, cap/backfill stability, substrate suitability, and/or habitat mitigation/ restoration objectives have not been achieved, contingency actions will be developed and implemented as appropriate.

4.8 Future Site Use

The selected site-wide cleanup action is compatible with future expected land use for both the Port and MJB properties, and provides significant public access opportunities. The future expected land use of the MJB property is commercial/residential on the uplands, with possible marina use in the adjacent Marine Area. The selected cleanup action allows for this expected future development activity and provides opportunities for enhanced public access, including shoreline public access.

The selected cleanup action would also provide significantly enhanced public access to Fidalgo Bay at the Port Uplands Area. The Port Uplands Area is expected to continue to be used in its current configuration, with commercial uses on Parcels 1 and 2, and Seafarers' Memorial Park on Parcel 3. The selected cleanup action, including environmental covenants, is compatible with this continued pattern of land use. In addition, the cleanup action will provide opportunities for enhanced shoreline public access amenities as part of an integrated habitat/landscape architecture plan for the Port Uplands Area, and a new focus on small boat use at Seafarers' Memorial Park. The enhanced small boat use would be facilitated by the new beach and calm water area created by the wave attenuation structures, allowing safe launching/landing of small watercraft and an inviting public space for staging small boat excursions and events. Mitigation requirements determined during the forthcoming RD and permitting phase of this cleanup action may necessitate modifications of the project plans.

4.9 Environmental Covenants

The proposed cleanup action will leave soil exceeding MTCA Method B cleanup levels (Table 1) in place below 6 ft BGS in portions of the Port Uplands Area and MJB North Area, and potentially below existing structures such as the Park Building. While the contaminated soil below 6 ft BGS is deep enough to not pose current risks to human

health and terrestrial ecological receptors, future development within areas of the contaminated soil could potentially generate conditions requiring appropriate safe handling procedures, stormwater controls, and consideration of disposal options for the specific indicator hazardous substances and concentrations encountered.

Environmental covenants will be required for the portions of the Port and MJB North Areas where complete removal of soil exceeding applicable cleanup levels (Table 1) will not be achieved. The covenants will identify specific contaminated soil locations and depths that will require special management if disturbed, unless the soil contamination is removed at a later time. Soil management plans will be required that instruct property owners on Ecology's requirements for performing invasive work in areas of remaining contaminated soil. The environmental covenants will be recorded following completion of excavation activities described in this CAP.

4.10 Habitat Restoration Opportunities

Under the Puget Sound Initiative, MTCA cleanup actions are designed to coincidentally enhance and/or restore habitat. The selected site-wide cleanup action provides significant habitat restoration opportunities and will restore almost 14 acres of currently degraded and injured intertidal/subtidal habitat.

Existing riprap and degraded sediments in the intertidal parts of the Marine Area will be replaced with clean sand and gravel beaches that will provide higher quality habitat, particularly for forage fish. In the North Marine Area, degraded sediments within the transitional beach area protected by the wave attenuation structures will be replaced with clean sand and gravel to provide high quality substrate, and eelgrass will be planted in the areas protected by the wave attenuators. This area currently supports little to no eelgrass because of degraded sediment quality. Potential eelgrass and/or macroalgae restoration areas are depicted in Figure 16. Detailed habitat mitigation, enhancement, and restoration plans will be developed during RD.

5.0 ALTERNATIVES CONSIDERED AND BASIS FOR REMEDY SELECTION

A range of potential site-wide cleanup action alternatives were evaluated in the final RI/FS report (GeoEngineers et al. 2008). This section summarizes the cleanup technologies and alternatives considered, and the basis for selection of the site-wide remedy.

5.1 Cleanup Technologies

The RI/FS report presents a detailed screening evaluation of potentially applicable general response actions and remediation technologies. The screening evaluation was carried out for each of the environmental media (soil, groundwater, and sediment) requiring cleanup action evaluation. During the development of the RI/FS, cleanup action alternatives were developed by assembling the technologies that were carried forward from this screening evaluation.

5.2 Feasibility Study Alternatives

The RI/FS presents a detailed evaluation of a range of potential cleanup action alternatives for the Port Uplands Area (PUA-1 to -4), MJB North Area (MJB-1 to -4), and Marine Area (M-1 to -2), which are summarized in Tables 3, 6 and 9, respectively. The RI/FS also presents detailed evaluations of each alternative, which are summarized in Section 5.3 below.

5.3 MTCA Disproportionate Cost Analysis

The MTCA disproportionate cost analysis (DCA) is used to evaluate which of the alternatives that meet MTCA threshold requirements are permanent to the maximum extent practicable. This analysis compares the relative benefits and costs of cleanup alternatives in selecting the alternative whose incremental cost is not disproportionate to the incremental benefits. Seven criteria are used in the disproportionate cost analysis as specified in WAC 173-340-360(2) and (3):

- Protectiveness
- Permanence
- Cost
- Long-Term Effectiveness
- Management of Short-Term Risks

- Implementability
- Consideration of Public Concerns

The comparison of benefits relative to costs may be quantitative, but will often be qualitative. Costs are disproportionate to the benefits if the incremental costs of a more permanent alternative exceed the incremental degree of benefits achieved by a lower-cost alternative [WAC 173-340-360(3)(e)(i)]. When two or more alternatives are equal in benefits, Ecology shall select the less costly alternative [WAC 173-340-360(3)(e)(ii)(C)].

The comparison of benefits relative to costs may be quantitative or qualitative based on the availability of quantitative data, such as mass of contaminants removed, estimated areas that will be contained, and volume of contaminated soils remaining on the Site. However, the benefits for some of the categories will be qualitative. For this reason, Ecology's analysis of which alternative is permanent to the maximum extent practicable is largely qualitative. The MTCA regulation allows Ecology to use best professional judgment to assess benefits qualitatively, and use its discretion to favor or disfavor qualitative benefits and use that information in selecting a cleanup action [WAC 173-340-360 (3)(e)(ii)(C)]. In order to document Ecology's qualitative analysis for the Site, Ecology assigned weighing factors to each of the six non-cost benefits criteria. The weighting factors represent Ecology's opinion on the importance of each benefit criterion at the Site, relative to protection of human health and the environment. The factors weighed for each of the criteria are briefly discussed in the following section and presented in Tables 3 to 11.

Protectiveness

The overall protectiveness of a cleanup action alternative is evaluated based on several factors, including the extent to which human health and the environment are protected and the degree to which overall risk at a site is reduced [WAC 173-340-360(3)(f)(i)]. Both on-site and off-site reductions in risk resulting from implementing the alternative are considered. Protectiveness is determined by evaluating the degree of improvement in overall environmental quality. At this Site, Ecology believes a weighting factor of 30 percent is appropriate for protectiveness. This represents the greatest value of all categories and is necessary based on the overall importance of protection of human

health and the environment, especially in relation to Ecology's goal of restoring the health of Puget Sound.

Permanence

Under MTCA, the permanence of an alternative is evaluated based on the degree to which the remedy permanently reduces the toxicity, mobility, or mass of hazardous substances, including the effectiveness of the alternative in destroying hazardous substances, the reduction or elimination of hazardous substance releases and sources of releases, the degree of irreversibility of waste treatment processes, and the characteristics and quantity of treatment residuals generated [WAC 173-340-360(3)(f)(ii)]. Based on the importance of the restoration of Puget Sound, Ecology believes this factor to be second only to protectiveness in importance and used a weighting factor of 20 percent for this evaluation criterion.

Cost

The analysis of cleanup action alternative costs under MTCA includes consideration of all costs associated with implementing an alternative, including design, construction, long-term monitoring, and environmental covenants [WAC 173-340-360(3)(f)(iii)]. Costs are intended to be comparable among different alternatives to assist in the overall analysis of relative costs and benefits of the alternatives. Costs are compared against benefits to assess cost-effectiveness and practicability of the cleanup action alternatives. No weighting factor is applied to this quantitative category, as costs are compared against the numeric analysis.

Long-Term Effectiveness

Long-term effectiveness expresses the degree of certainty that the alternative will be successful in maintaining compliance with cleanup standards over the long-term [WAC 173-340-360(3)(f)(iv)]. The MTCA regulations contain a specific preference ranking for different types of technologies that is to be considered as part of the comparative analysis. The ranking places the highest preference on technologies such as reuse/recycling, treatment, immobilization/solidification, and disposal in an engineered, lined, and monitored facility. Lower preference rankings are applied to technologies such as on-site isolation/containment with attendant engineered controls, and

environmental covenants and monitoring. The regulations recognize that, in most cases, the selected cleanup remedy will combine multiple technologies. The MTCA preference ranking must be considered along with other site-specific factors in the evaluation of long-term effectiveness. Ecology considers a weighting for this factor of 20 percent to be appropriate at this Site.

Management of Short-term Risks

This criterion is a measure of the relative magnitude and complexity of actions required to maintain protection of human health and the environment during implementation of the cleanup action [WAC 173-340-360-(3)(f)(v)]. Cleanup actions carry short-term risks, such as potential mobilization of contaminants during construction, or safety risks typical of large construction projects. In-water dredging and excavation of contaminated soils along the shoreline carry a risk of temporary water quality degradation and potential sediment recontamination. Some short-term risks can be managed through the use of best management practices during the project design and construction, while other risks are inherent to certain project alternatives. A weighting factor of 10 percent is being used for this Site. This lower rating is based on the limited timeframe associated with the risks and the general ability to modify any alternative to reduce short-term risks during construction without significant effect on human health and the environment.

Implementability

Implementability is the ability to implement the selected remedy. It measures the overall relative difficulty and uncertainty of implementing the cleanup action. It includes technical factors such as the availability of proven technologies and experienced contractors to accomplish the cleanup work [WAC 173-340-360(3)(f)(vi)]. It also includes administrative factors associated with permitting and completing the cleanup. The weighting factor Ecology used for implementability is 10 percent. Implementability is less associated with the primary goal of the cleanup action, protection of human health and the environment, and therefore has a lower weighting factor. In addition, the issues associated with the implementability of a remedy are often duplicated in the remedy costs. Engineering design considerations are often of primary

importance in this category and often refined during the development of the engineering design report.

Consideration of Public Concerns

The public involvement process under MTCA is used to identify potential public concerns regarding cleanup action alternatives [WAC 173-340-360(3)(f)(vi)]. The extent to which an alternative addresses those concerns is considered as part of the remedy selection process. This includes concerns raised by individuals, community groups, local governments, tribes, federal and state agencies, and other organizations that may have an interest in or knowledge of the Site. A weighting factor of 10 percent is being used for the evaluation of this category. The public concerns voiced during the public involvement process can also be included in the other categories identified above such as protectiveness and long-term effectiveness. Public concerns that can be incorporated into alternative categories are more appropriately considered in the scoring of those other categories. In particular, the public concerns for this Site would generally be associated with environmental concerns and performance of the cleanup action, which are addressed under other criteria such as protectiveness and permanence.

5.4 Evaluation and Comparison of Port Uplands Area Alternatives

This section provides a comparative analysis of the remedial alternatives considered for the Port Uplands Areas. The MTCA evaluation criteria discussed in Section 5.3 were used to evaluate the each remedial alternative, with the alternatives then ranked based on their expected performance under each criterion. Table 3 presents the details about the Port Uplands Areas remedial alternatives. The details of evaluation and their results are presented in Tables 4 and 5, respectively.

Threshold Requirements

Based on the review of final RI/FS report (GeoEngineers et al. 2008), Ecology determined that Port Alternative PUA-3 does not meet threshold requirements under MTCA for cleanup actions [WAC 173-340-360(2)], because this alternative does not address soil contamination present between 6 and 10 ft BGS along the shoreline. Thus, this alternative is not considered for the DCA.

The other three remaining alternatives (PUA-1, PUA-2, and PUA-4) meet MTCA threshold requirements, because they address the contamination present at the Site to varying degrees, including the contamination present between 6 to 10 ft BGS along the shoreline and at other locations on the Site. The details of depths, quantity of excavation, and costs are presented in Table 3.

5.4.1 Comparison of Port Uplands Areas Alternatives by Criteria

The evaluation of disproportionate cost is based on a comparative analysis of costs against the remaining six MTCA evaluation criteria. Relative rankings of each alternative for these six criteria are summarized in Table 4. These rankings are briefly discussed below:

Protectiveness

Alternatives PUA-2 and PUA-4 are less protective than Alternative PUA-1. Alternative PUA-1 is the most protective because it removes all contaminated soils to the maximum extent practicable. Alternatives PUA-2 and PUA-4 receive the same ranking, since these remove contaminated soils to varying depths, resulting in leaving some subsurface contamination, and both rely on environmental covenants to prevent exposure to the contaminated soils left in place. Alternative PUA-1 ranks the highest for the overall protectiveness, and both Alternatives PUA-2 and PUA-4 rank 80 percent in overall protectiveness compared with PUA-1.

Permanence

Alternative PUA-1 achieves the highest level of performance relative to other two alternatives, since it includes the removal of soil contamination to the maximum extent practicable. However, Alternatives PUA-2 and PUA-4 are also considered to be permanent because they will be implemented in conjunction with the inwater remedy that includes shoreline protection. Alternatives PUA-2 and PUA-4 rank at 80 percent for permanence compared with Alternative PUA-1.

Long-Term Effectiveness

Alternative PUA-1 achieves a higher degree of long-term effectiveness than the other two alternatives as a result of the greater amount of contaminated removed under that alternative, and is assigned a ranking of 5. Alternatives PUA-2 and PUA-4 are assigned rankings of 4 and 3 respectively, since lesser amounts of contaminated soils will be removed under those alternatives compared to PUA-1. The ranking order for long-term effectiveness from highest to lowest is PUA-1, PUA-2, and PUA-4. Alternatives PUA-2 and PUA-4 rank 80 percent and 60 percent in long-term effectiveness compared with PUA-1, respectively.

Management of Short Term-Risks

Alternative PUA-1 receives the lowest ranking for this category, since a greater amount of contaminated soils will be removed under this alternative along the shoreline, near the Park Building relative to the other two alternatives. In addition, excavation of deep soils along the shoreline may require extensive shoring and dewatering. Alternatives PUA-2 and PUA-4 received a higher score for the management of short-term risks, relative to Alternative PUA-1, as a result of the reduced extent of soil excavation.

Implementability

The lowest score for implementability was assigned to Alternative PUA-1, which received a score of 2. This is as a result of the extensive excavation, including the deep excavation along the shoreline, which may require extensive shoring and dewatering. Both Alternatives PUA-2 and PUA-4 received a higher score of 3 for implementability relative to PUA-1 because of the reduced extent and difficulty of soil removal associated with these alternatives.

Consideration of Public Comments

Public concerns were identified based on the public comments received on the draft RI/FS report. No public concerns were expressed concerning Alternative PUA-1, since it is the most protective alternative. Alternative PUA-1 was therefore assigned a score of 5 for consideration of public concerns. Alternatives PUA-2 and PUA-4 leave some residual soils contamination below 10 ft BGS, so

they were assigned a score of 4 for this category. The rankings provided in Table 4 are intended to reflect on balance, how well the alternatives address the cross-section of public comments received. Alternative PUA-4 is relatively responsive to public concerns that have been raised and receives a moderately high score of 4. This alternative makes significant use of removal of contamination between 6 to 10 ft BGS that is present along the shoreline and in the vicinity of the Park Building. Favorable public comments were received for Alternative PUA-4 with respect to its balancing of cost and environmental benefits.

5.4.2 Reasonable Restoration Time Frame

The restoration time pertains to the time required to meet cleanup levels. The restoration time for all three of the Port Uplands Area alternatives is in the order of 2 to 3 years. This includes project design, permitting, construction, and closure activities. Alternatives PUA-2 and PUA-4 would leave some residual contamination in place requiring long-term monitoring and consequently could extend the duration of time for monitoring to confirm that cleanup levels are being maintained.

5.4.3 Overall Comparison of Remedy Costs and Benefits

Tables 3 and 4 and Figure 20 summarize costs and remedy benefits for each Port Uplands Area alternative. The estimated costs of the alternatives range from \$9.1 to \$18.3 million. The RI/FS report (GeoEngineers et al. 2008) presents detailed cost estimates for the Port Uplands Area alternatives. These costs are expressed in 2008 dollars without adjustment to future cost inflation and without present value discount of future costs. The probable remedy costs are expected to vary with a range of +50 percent to -30 percent.

The overall benefits associated with each alternative using a composite ranking are presented in Table 5 and Figure 20. The calculated benefits using the categorical weighting factors are in Table 4. The calculated benefits integrate the rankings for each evaluation criterion discussed above, multiplied by the weighting within that category and summed to reach the benefits total. As per MTCA requirements, the relative benefits and cost of each alternative are

compared to the Alternative PUA-1, since PUA-1 represents the most permanent remedial alternative evaluated. It therefore represents the benchmark for comparison of the other remedies' incremental benefits and costs.

Alternative PUA-1 receives an overall benefit score of 4.4, because it uses the greatest extent of contaminated soils excavation and off-site disposal. The remedy is considered to provide high benefit rankings under protectiveness, performance and long-term effectiveness. However, this alternative has a low ranking for short-term risks management and implementability. The calculated score of 4.4 for PUA-1 is slightly higher than that for Alternative PUA-2. However, Alternative PUA-1 is substantially more costly than PUA-2 (60 percent more). Because the cost of Alternative PUA-1 is substantially higher (\$6.8 million) than that of Alternative PUA-2, whereas the benefit level is only slightly greater (0.6), the incremental cost of Alternative PUA-1 is disproportionate to the incremental environmental benefits gained.

Alternative PUA-2 achieves a calculated benefit score of 3.8. This alternative has a high score for protectiveness, performance, and long-term effectiveness, although its scores are generally lower than those for Alternative PUA-1. It has a medium ranking for the management of short-term risk, implementability, and consideration of public concerns. This alternative is consistent with the Port's future land use plans.

Alternative PUA-4 receives an overall benefit score of 3.7. This alternative has a relatively high score for protectiveness, performance, and long-term effectiveness, although generally lower than Alternative PUA-1. It receives a moderate score for the management of short-term risk and implementability. However, it received favorable public comments with respect to balancing the cost to the overall environmental benefits. This alternative is also designed to accommodate for the Port's future land use plans.

The calculated ranking of Alternative PUA-4 (3.7) is slightly lower than that of Alternative PUA-1 (4.4). Rankings for Alternatives PUA-2 and PUA-4 were

nearly identical (3.8 and 3.7 respectively); however, both ranked lower for protectiveness, permanence, and long-term effectiveness relative to Alternative PUA-1, since PUA-2 and PUA-4 do not remove all the contaminated soils to the maximum extent practicable. Estimated costs of Alternatives PUA-2 and PUA-4 are \$11.5 million and \$9.1 million respectively. Though the overall score for these two alternatives is almost the same, Alternative PUA-4 provides a relatively high level of benefits as measured against MTCA criteria, because of the higher cost associated with Alternative PUA-2.

Figure 20 presents a graphical comparison between remedy costs and benefits for each of the alternatives. Remedy benefits are plotted in gray using the calculated rankings from Table 5 and hatched portion of the graph represents the costs of alternatives. The relative costs and benefits are graphically represented by a line.

A substantial increase in costs among Alternatives PUA-4, PUA-2, and PUA-1 is apparent from this graph. Since the increase in cost is not accompanied by a corresponding increase in remedy benefits, Alternative PUA-1 is impractical (incremental costs are disproportionate to the incremental environmental benefits gained) and lower cost alternatives should be considered.

The overall benefit score for Alternatives PUA-2 and PUA-4 are almost the same (3.8 and 3.7 respectively). However, Alternative PUA-2 costs \$2.4 million (20 percent) more than Alternative PUA-4. Hence, Alternative PUA-4 has a greater degree of overall environmental benefits per unit of incremental cost than Alternative PUA-2. In addition, this alternative is considered permanent to the maximum extent practicable under MTCA.

5.4.4 Conclusions

Based on the above DCA evaluation per MTCA requirements, Alternative PUA-4 is identified as the preferred alternative for the Port Uplands Area. This alternative uses a high-performance technology (removal) and provides a high level of calculated ranking, achieving the highest degree of environmental benefits for the unit of incremental cost while still remaining practical.

The high costs of contaminated soil removal under this alternative are appropriately targeted at the soils that:

- Have the highest contamination levels (i.e., source areas)
- Can be removed safely without an excessive level of short-term risk
- Consider the community concerns raised during the public comment period

Alternative PUA-4 is permanent to the maximum extent practicable and hence identified as the preferred alternative for the Port Uplands Area.

5.5 Evaluation and Comparison of MJB North Area Upland Alternatives

Comparative analyses of remedial alternatives considered for the MJB North Area are discussed in this section. The MTCA evaluation criteria as discussed in Section 5.3 were used to evaluate the each remedial alternative and compared to each other relative to their expected performances under each criterion. Table 6 presents the details on each of the MJB North Area remedial alternatives. The details of the evaluation and their results are presented in Tables 7 and 8, respectively.

Threshold Requirements

Based on review of the RI/FS report, Ecology determined that Alternatives MJB-1 and MJB-4 meet the MTCA threshold requirements for cleanup actions. However, Ecology determined that Alternatives MJB-2 and MJB-3 do not meet MTCA threshold requirements because these alternatives would leave a significant amount of contaminated soil exceeding cleanup levels in place between 6 and 10 ft BGS along the shoreline. Consequently, Alternatives MJB-2 and MJB-3 are not carried through the DCA.

Alternatives MJB-1 and MJB-4 are evaluated and compared with respect to MTCA criteria described in Section 5.3. Table 7 presents a summary of criteria used in the evaluation of above two alternatives.

5.5.1 Comparison of MJB North Area Alternatives by Criteria

The evaluation of disproportionate cost is based on a comparative analysis of costs against six criteria. Relative rankings of each alternative for these six criteria are summarized in Table 7. These rankings are briefly discussed below:

Protectiveness

Alternative MJB-1 removes the soil contamination to the maximum extent practicable, and consequently receives the highest score in this category. Alternative MJB-4 receives a lower ranking than Alternative MJB-1, since some of the localized deeper soil contamination will be left in place under MJB-4. However, Alternative MJB-4 reduces terrestrial ecological risks within the remaining uplands area by reducing constituent concentrations to protective levels. Alternative MJB-4 ranks 80 percent in performance compared to Alternative MJB-1.

Permanence

Alternative MJB-1 is rated highest for this criterion because it reduces the contaminant mass, toxicity, and volume to the greatest extent by the removal of all contaminated soils present along the shoreline. Alternative MJB-4 removes less contaminant volume than Alternative MJB-1, and thus receives a lower (medium) score than Alternative MJB-1. Contaminated soils removed under both alternatives would be placed in an off-site landfill. Alternative MJB-4 ranks 80 percent in performance compared to Alternative MJB-1.

Long-Term Effectiveness

Both alternatives would produce wastes requiring off-site management. Alternative MJB-1 receives the highest score, since it is expected to excavate all the contaminated soils along the shoreline and uplands. Alternative MJB-4 would leave some deeper contaminated soils in place, and as a result is ranked lower than Alternative MJB-1. Alternative MJB-4 ranks 80 percent in long-term effectiveness compared to Alternative MJB-1.

Management of Short-Term Risks

Alternative MJB-1 ranked lowest for this criterion, since the greatest volume of soil disturbed/excavated including below the water table may require shoring and dewatering. Alternative MJB-4 ranked higher than Alternative MJB-1 as a result of the reduced amount of soil disturbance/excavation and consequently reduced risk during the excavation and in transportation.

Technical and Administrative Implementability

Both alternatives can be implemented. Alternative MJB-4 would be easier to implement than Alternative MJB-1, as MJB-1 would require excavation of soil below the water table, which may require shoring and dewatering. Consequently, Alternative MJB-4 is ranked higher than Alternative MJB-1 for this criterion.

Consideration of Public Comments

Alternative MJB-1 provides the most complete removal and therefore has high level of public acceptance, as a result scored higher than Alternative MJB-4. Alternative MJB-4 leaves some contaminated soils in place. Both alternatives are acceptable to the public. However, based on the public comments received, Alternative MJB-4 was favored because of balancing the cost and environmental benefits.

5.5.2 Reasonable Restoration Time Frame

Both Alternatives MJB-1 and MJB-4 require approximately 2 to 3 years for implementation, and therefore have the same initial restoration timeframe. Since Alternative MJB-4 would leave some contaminated soil in place, and would require long-term monitoring and environmental covenants to meet cleanup goals, Alternative MJB-4 may require a longer restoration time.

5.5.3 Overall Comparison of Remedy Costs and Benefits

Alternative MJB-1 removes a greater volume of contaminated soils than Alternative MJB-4. However, both alternatives are protective of human health and the environment because both will meet the final soil cleanup levels at the

standard point of compliance or at a conditional point of compliance through the excavation of contaminated soils along the shoreline and in other upland areas of the Site.

The MJB North Area alternative costs and benefits are summarized in Tables 6 and 7 and Figure 21. The estimated costs of Alternatives MJB-1 and MJB-4 are \$7.0 million and \$4.8 million respectively. The RI/FS report presents detailed cost estimates for the MJB North Area alternatives. These costs are expressed in 2008 dollars without adjustment to future cost inflation and without present value discount of future costs. The probable remedy costs are expected to vary with a range of +50 percent to -30 percent.

Alternative MJB-1 represents the most permanent remedial alternative evaluated, owing to excavation of all contaminated soils exceeding soil cleanup levels at the standard point of compliance required by MTCA (soil between 0 and 15 ft BGS), whereas Alternative MJB-4 is capable of meeting soil cleanup levels at a conditional point of compliance corresponding the upper 10 ft BGS in soil. Hence, Alternative MJB-1 is considered as the bench mark against which the relationship between incremental remedy benefits and incremental costs are evaluated.

Alternative MJB-4 receives an overall score of 3.8. This alternative has a relatively high score for protectiveness and performance, and a medium score for long-term effectiveness, lower than Alternative MJB-1. It received a moderate score for the management of short-term risks and implementability, which are the same or higher than Alternative MJB-1. Nonetheless, this alternative received favorable comments from the public with respect to balancing the cost and environmental benefits. This alternative is also consistent with the future land use plans.

The overall calculated ranking of Alternatives MJB-1 and MJB-4 are 4.5 and 3.8, respectively. Alternative costs are \$7.0 million and \$4.8 million respectively. While the overall ranking of Alternative MJB-1 (4.5) is slightly higher than that of

Alternative MJB-4 (3.8), Alternative MJB-1 costs \$2.2 million more than the Alternative MJB-4. Thus, Alternative MJB-1 costs approximately 70 percent more, but overall environmental benefit gained is relatively small (0.7).

Figure 21 presents a graphical comparison between remedy costs and benefits for each of MJB North Area alternatives. Remedy benefits are plotted in gray using the calculated rankings from Table 8 and the hatched portion of the graph represents the costs of alternatives. The relative costs and benefits are graphically represented by a line.

The increase in costs between Alternative MJB-1 and MJB-4 is apparent in Figure 21. However, both alternatives provide appropriate human health and environmental protection, permanent solution to the maximum extent practicable, and reasonable restoration time. Hence, the increase in cost (\$2.2 million) of Alternative MJB-1 is not accompanied by a corresponding increase in remedy benefits (0.7). Therefore, as per MTCA, Alternative MJB-1 is impractical and lower cost alternatives should be considered. Ecology considers the incremental cost of Alternative MJB-1 to be substantial and disproportionate to the negligible degree of environmental benefits achieved by this alternative. Thus, Alternative MJB-4 has a greater degree of overall benefits achieved for incremental unit cost than Alternative MJB-1. This alternative is also considered permanent to the maximum extent practicable under MTCA.

5.5.4 Conclusions

Based on the above DCA evaluation per MTCA requirements, Alternative MJB-4 is identified as the preferred remedial alternative for the MJB North Area. This alternative uses high performance technologies and provides a high level of calculated ranking, achieving the highest environmental benefits to the unit incremental cost while remaining practical.

The high costs of contaminated soil removal under this alternative are appropriately targeted at the soils that:

Have the highest contaminants levels (i.e., source areas)

- Have medium metal concentrations for soil homogenization
- Can be removed and/or homogenized safely without an excessive level of short-term risk
- Considers the community concerns raised during the public comment

Alternative MJB-4 is permanent to the maximum extent practicable and hence is identified as the preferred alternative for the MJB North Area.

5.6 Evaluation and Comparison of Marine Area Alternatives

Comparative analyses of remedial alternatives considered for the site-wide Marine Area are discussed in this section. The MTCA evaluation criteria as discussed in Section 5.4 were used to evaluate the each remedial alternative and compared to each other relative to their expected performances under each criterion. Table 9 presents the details on each of the Marine Area remedial alternatives. The details of the evaluation and their results are presented in Tables 10 and 11, respectively.

Threshold Requirements

Based on review of the RI/FS report, Ecology determined that both Alternatives M-1 and M-2 meet the MTCA threshold requirements for cleanup actions. Alternatives M-1 and M-2 are evaluated and compared with respect to MTCA criteria described in Section 5.4. Table 10 presents a summary of criteria used in the evaluation of above two alternatives.

5.6.1 Comparison of Marine Area Alternatives by Criteria

The evaluation of disproportionate cost is based on a comparative analysis of costs against six criteria. Relative rankings of each alternative for these six criteria are summarized in Table 10. These rankings are briefly discussed below:

Protectiveness

Alternatives M-1 and M-2 are both protective and provide risk reduction because contamination is removed from the aquatic area. Alternative M-1 ranks higher than Alternative M-2 because it removes a greater volume of impacted sediment (Table 9).

Permanence

Neither alternative achieves permanent destruction of metals or organic constituents (including wood debris). Where upland disposal is used, considerations about long-term management of sediments in the landfill are the same as those described for the upland soils. However, both alternatives achieve a permanent risk reduction in the aquatic environment by removing impacted sediments. Alternative M-1 achieves marginally greater permanence because it removes sediments above the SQS; however, the greater increment of permanence is achieved at additional cost (see below). For both alternatives, the unavoidable generation of dredge residuals requires the placement of a 0.5-ft-thick sand layer to ensure a clean post-dredge surface and achieve cleanup standards in a reasonable timeframe.

The wave attenuation structures offshore of the Port property will allow for placement of permanent caps consisting of finer-grained, habitat-enhancing materials. These caps will not require long-term maintenance and will allow for a stable environment for aquatic habitat to develop.

Long-Term Effectiveness

Alternative M-1 is considered marginally more effective than Alternative M-2 in the long term because it removes a greater volume of potentially harmful sediment from the aquatic environment. However, due to the greater volume of dredging in Alternative M-1, there may be an increased potential for dredge residuals compared to Alternative M-2. In both alternatives, residuals would be managed using a post-dredge cover of clean material.

Management of Short-Term Risks

Alternative M-1 entails a greater volume of dredging and post-dredge cover placement. The construction duration is longer. During construction, there would be a greater potential for short-term water quality impacts associated with dredging, backfilling, capping, and cover placement. In comparison, Alternative M-2 requires a lower volume of dredging, backfill, capping, and cover materials.

Thus, Alternative M-2 ranks marginally higher than Alternative M-1 for management of short-term risks.

Implementability

Both alternatives are technically possible to implement relative to complexity, administrative/regulatory requirements, size, access, and integration with existing operations. Alternative M-1 entails removing a greater volume of material from the dredge area, and consequently requires management of more excavated material for disposal or beneficial reuse. Without considering beneficial reuse, both Alternatives M-1 and M-2 are considered to rank equally for technical and administrative implementability. If a beneficial reuse option for wood debris material were to become available and determined to be practicable during RD, Alternative M-1 would rank lower for implementability because as the dredge volume increases, a greater area of upland space would be needed for staging beneficial reuse activities.

5.6.2 Reasonable Restoration Time Frame

Both Alternatives M-1 and M-2 require approximately two to three years for implementation, and therefore have the same initial restoration timeframe. Since Alternative M-2 would leave some contaminated sediment in place, and would require long-term monitoring and environmental covenants to meet cleanup goals, Alternative M-2 may require a longer restoration time.

5.6.3 Overall Comparison of Remedy Costs and Benefits

Alternative M-1 removes a greater volume of contaminated sediments than Alternative M-2. However, both alternatives are protective of human health and the environment because both will meet the final sediment cleanup levels at the standard point of compliance at the Site.

The Marine Area alternative costs and benefits are summarized in Tables 9 and 10. The estimated costs of Alternatives M-1 and M-2 are \$7.1 million and \$5.8 million, respectively. The RI/FS report presents detailed cost estimates for the Marine Area alternatives. These costs are expressed in 2008 dollars without

adjustment to future cost inflation and without present value discount of future costs. The probable remedy costs are expected to vary with a range of +50 percent to -30 percent.

Alternative M-1 has the highest cost, while Alternative M-2 is approximately 20 percent lower. Therefore, Alternative M-2 ranks higher for cost. However, overall environmental benefit score associated with Alternative M-1 (4.7) is also approximately 20 percent higher than Alternative M-2 (3.9; Table 11).

5.6.4 Conclusions

Based on the above DCA evaluation per MTCA requirements, Alternative M-1 is identified as the preferred remedial alternative for the Marine Area. This alternative uses high performance technologies and provides a high level of calculated ranking, achieving the highest environmental benefits that are proportionate to the unit incremental costs while remaining practical.

6.0 IMPLEMENTATION OF THE CLEANUP ACTION

Preliminary RD of the selected cleanup action described in this CAP was initiated in November 2008 under Agreed Order DE-08TCPHQ-6208 between Ecology and the Port. Preliminary RD completed to date has included development of concept-level designs sufficient to complete necessary applications, including a Joint Aquatic Resources Permit Application (JARPA) for project permitting, a cultural resources plan, and a habitat mitigation plan. The permitting process is ongoing. Remedial actions are currently targeted to begin in summer 2009, subject to the permitting schedule.

EXHIBIT C of the Consent Decree contains an outline of the schedule to complete RD and implementation activities. The Consent Decree will be entered in Skagit County Superior Court, and will become effective once entered.

Consistent with Chapter 70.105D RCW, as implemented by Chapter 173-340 WAC (MTCA Cleanup Regulation), Ecology has determined that the selected Site cleanup action described in Section 4 of this CAP is protective of human health and the environment, will attain federal and state requirements that are applicable or relevant and appropriate, complies with cleanup standards, and provides for compliance monitoring. The selected cleanup action satisfies the preference expressed in WAC 173-340-360 for the use of permanent solutions to the maximum extent practicable, and provides for a reasonable restoration time frame.

7.0 COMPLIANCE MONITORING

Compliance monitoring and contingency responses (as needed) will be implemented in accordance with WAC 173-340-410, Compliance Monitoring Requirements. Detailed requirements will be described in the Construction Quality Assurance Plan (CQAP) and Operations, Maintenance, and Monitoring Plan (OMMP) to be prepared as a part of RD. The objective of these plans is to confirm that cleanup standards have been achieved, and also to confirm the long-term effectiveness of cleanup actions at the Site. The plans will contain discussions on duration and frequency of monitoring, the trigger for contingency response actions, and the rationale for terminating monitoring. The three types of compliance monitoring to be conducted include:

- Protection Monitoring to confirm that human health and the environment are adequately protected during the construction period of the cleanup action
- Performance Monitoring to confirm that the cleanup action has attained cleanup standards and other performance standards
- Confirmation Monitoring to confirm the long-term effectiveness of the cleanup action once performance standards have been attained

Cleanup levels and associated points of compliance for the cleanup action are described above in Section 3.

7.1 Monitoring Objectives and Rationale

The cleanup action incorporates monitoring to determine whether cleanup standards have been achieved during and after the cleanup action. Three broad categories of compliance monitoring will be undertaken at the Site as follows:

• Water Quality (Protection and Confirmation Monitoring) – During the cleanup action, construction controls and protection monitoring will be implemented as practicable to ensure surface water quality protection within the Site area. As discussed in Section 4.6, following completion of remedial actions, groundwater will be sampled on a quarterly basis at retained or newly constructed groundwater monitoring wells for a minimum of four consecutive quarters following completion of remedial actions. After four consecutive quarters of confirmation groundwater sampling, the subsequent sampling frequency will be determined by the Port, MJB, and K-C in consultation with Ecology. If the

- groundwater samples continue to exceed the groundwater cleanup levels after two years without abating, additional actions will be considered. Detailed monitoring and contingency response requirements will be described in the Construction Quality Assurance Plan (CQAP) and Operations, Maintenance, and Monitoring Plan (OMMP) to be prepared as a part of RD.
- Physical Limits and Integrity (Performance and Confirmation Monitoring) As discussed in Section 4.5, topographic and bathymetric performance monitoring will be conducted during the cleanup action to guide the limits of construction activities. Following completion of construction, physical confirmation monitoring of upland and intertidal sediment cap surfaces will be performed to verify that caps are not substantially eroded over time by natural and/or anthropogenic forces. During these confirmation monitoring events, sediment cap thickness will be assessed and compared with the minimum required thickness determined during RD to ensure integrity of the caps to protect human health and the environment (Palermo et al. 1998). Again, detailed monitoring and contingency response requirements will be described in the CQAP and OMMP to be prepared as a part of RD.
- discussed in Section 4.5, once required excavation or dredging elevations have been verified, performance monitoring will involve collecting soil or sediment samples from the base and/or sidewalls of excavations to confirm that cleanup and remediation levels have been achieved and to document concentrations of contaminants remaining on site. If individual samples exceed cleanup or remediation levels (e.g., in sidewalls of upland excavations or at the base of offshore dredge prisms), a statistical analysis of the data will be performed to assess the extent and degree of exceedance(s) (Ecology 1992). Following Ecology approval, response actions will be implemented as appropriate, including:
 - No further action (i.e., cleanup determined to be successful within a given grid area)
 - Additional wood debris indicator and/or chemical sampling to further characterize residual contamination within and/or adjacent to the excavation and dredge areas

- Placement of a clean sand cover as necessary to address identified sediment residuals
- Placement of a confining cap layer or backfill to achieve isolation of underlying contaminants
- Supplemental excavation or dredging as practicable, followed by additional post-construction performance sampling, as appropriate
- Following completion of construction, confirmation monitoring of surface sediments within the shoreline cap area will be conducted. Chemical monitoring will be performed to verify that these areas achieve and maintain SQS chemical and woody debris criteria (Table 2). Again, detailed monitoring and contingency response requirements will be described in the CQAP and OMMP to be prepared as a part of RD.

All selected remedial actions to be implemented within the Site use demonstrated technologies with at least a 15 to 20-year record of successful performance. Additional research is not necessary to demonstrate the effectiveness of the technologies. Accordingly, long-term monitoring is appropriately focused toward routine maintenance objectives and verification that the cleanup action is achieving its intended goals.

8.0 FIVE-YEAR REVIEW

Because the cleanup action described in Section 4 will result in hazardous substances remaining at the Site at concentrations exceeding cleanup levels (e.g., beneath the Park Building and below 6 ft BGS in other areas of the Site), and because environmental covenants are included as part of the remedy, Ecology will review the selected cleanup action described in this CAP every 5 years to ensure protection of human health and the environment. Consistent with the requirements of WAC 173-340-420, the 5-year review shall include the following:

- A review of the title of the real property subject to the environmental covenant to verify that the covenant is properly recorded
- A review of available monitoring data to verify the effectiveness of completed cleanup actions, including engineered caps and institutional controls, in limiting exposure to hazardous substances remaining at the Site
- A review of new scientific information for individual hazardous substances or mixtures
 present at the Site
- A review of new applicable state and federal laws for hazardous substances present at the Site
- A review of current and projected future land and resource uses at the Site
- A review of the availability and practicability of more permanent remedies
- A review of the availability of improved analytical techniques to evaluate compliance with cleanup levels.

Ecology will publish a notice of all periodic reviews in the Site Register and will provide an opportunity for review and comment by the potentially liable persons and the public. If Ecology determines that substantial changes in the cleanup action are necessary to protect human health and the environment at the Site, a revised CAP will be prepared and provided for public review and comment in accordance with WAC 173-340-380 and 173-340-600.

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