

An Assessment of the PCB and Dioxin Background in Washington Freshwater Fish, with Recommendations for Prioritizing 303(d) Listings



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An Assessment of the PCB and Dioxin Background in Washington Freshwater Fish, with Recommendations for Prioritizing 303(d) Listings

by Art Johnson, Keith Seiders, and Dale Norton

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Waterbody Numbers: See Appendix A

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Abstract

Section 303(d) of the federal Clean Water Act requires states to prepare a list every two years of waterbodies that do not meet water quality standards. A Total Maximum Daily Load (TMDL) must be developed for every waterbody and pollutant on the list. A TMDL includes a written, quantitative assessment of water quality problems and pollutant sources that cause the problems. The TMDL determines the amount of a pollutant that can be discharged to the waterbody and allocates that pollutant load among sources.

Washington State has approximately 90 waterbodies 303(d) listed for not meeting human health criteria for polychlorinated biphenyls (PCBs) and dioxin (2,3,7,8-TCDD) in edible fish tissue samples from rivers, lakes, and streams. In order to prioritize the state's TMDL resources to address human health concerns, the Washington State Department of Ecology (Ecology) determined that better information was needed on what constitutes background levels of PCBs and dioxin in local fish populations.

In response to this need, Ecology collected 52 fish fillet samples from 24 lakes, rivers, and streams across Washington considered representative of background conditions. The samples were analyzed for PCBs, polychlorinated dibenzodioxins, polychlorinated dibenzofurans, and percent lipids. The results are used to recommend approaches for prioritizing 303(d) listings for these chemicals, with the intent of accelerating cleanup actions across the state.

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Introduction

303(d) Listings for PCBs and Dioxin

Section 303(d) of the federal Clean Water Act requires states to prepare a list every two years of waterbodies that do not meet water quality standards. The Act requires that a Total Maximum Daily Load (TMDL) be developed for every waterbody and pollutant on the list. A TMDL includes a written, quantitative assessment of water quality problems and pollutant sources that cause the problems. The TMDL determines the amount of a pollutant that can be discharged to the waterbody and allocates that pollutant load among sources.

Washington's current (2008) 303(d) list has 113 Category 5 listings for PCBs and 48 Category 5 listings for dioxin (2,3,7,8-tetrachlorodibenzo-*p*-dioxin or TCDD) not meeting (exceeding) U.S. Environmental Protection Agency (EPA) human health criteria for fish consumption. The listings cover 59 waterbodies for PCBs and 32 waterbodies for dioxin (Figures 1 and 2, Appendix B and C). Approximately one-third of the listings are for waterbodies where there are no obvious local sources of these chemicals. A number of the listings are based on fish tissue concentrations that are not substantially above criteria.

There are an additional 47 Category 2 (Waters of Concern) listings (32 waterbodies) for dioxin toxicity equivalents (TCDD TEQs) exceeding human health criteria (Appendix D). TEQs are a measure of the combined toxicity of a mixture of polychlorinated dibenzodioxins (PCDDs) and related compounds called polychlorinated dibenzofurans (PCDFs), based on the toxicity of TCDD, the most toxic of these compounds. Here again, sources of contamination are not always readily apparent.

Category 5 listings require a TMDL. Category 2 listings are for waters where there is some evidence of a problem, but not enough data to qualify for Category 5 and require a TMDL.

Studies have shown there can be a significant underlying PCB, PCDD, and PCDF background in aquatic and terrestrial environments unrelated to local sources (e.g., Meijer et al., 2003; Wania and Mackay, 1993; Gillian and Wania, 2005). The question that arises in connection with 303(d) listings is what level of contamination indicates the presence of significant local sources which may be amenable to cleanup actions? Without better information on what constitutes background levels of these chemicals in local fish populations, it is difficult to prioritize the state's TMDL resources to address human health concerns.



Figure 1. Locations of Waterbodies with Washington State Category 5 Listings for PCBs in Freshwater Fish.



Figure 2. Locations of Waterbodies with Washington State Category 5 Listings for Dioxin (2,3,7,8-TCDD) in Freshwater Fish.

Study Objectives and Overview

The objective of this project is to characterize PCB, TCDD, and TCDD TEQ levels in edible fish tissue from background lakes, rivers, and streams in Washington. The results are used in conjunction with existing data and other information (e.g., human health-based water quality criteria and screening values) to recommend approaches for prioritizing 303(d) listings for these chemicals, with the intent of accelerating cleanup actions across the state.

The Background Study analyzed 52 fish tissue samples from 24 waterbodies in four regions of Washington: western Washington, West Slope of the Cascades, East Slope of the Cascades, and eastern Washington. From one to six species were analyzed from each waterbody, depending on availability. The fish were collected during the summer and fall of 2007 and 2008.

Composite fillets, typically from four to five individuals of each species, were analyzed for PCBs, PCDDs, PCDFs, and percent lipids. High resolution gas chromatography and mass spectrometry (HR-GC/MS) methods were used to achieve low detection limits. A subset of samples was analyzed for PCB Aroclor-equivalents by gas chromatography/electron capture detection (GC/ECD).

The Washington State Department of Ecology (Ecology) Environmental Assessment Program (EA Program) conducted the study, following a Quality Assurance Project Plan (Johnson, 2008).

PCBs, Dioxins, and Furans

PCBs, dioxins, and furans are classed as persistent, bioaccumulative, and toxic substances <u>www.ecy.wa.gov/programs/swfa/pbt</u>. Their chemical structure is shown in Figure 3.

PCBs are a class of 209 discrete chemical compounds containing from 1 to 10 chlorines. Commercial PCBs were manufactured as mixtures with varying chlorine content. In the United States, PCBs were primarily sold under the trade name Aroclors. PCBs were used in closed industrial systems such as electrical transformers and capacitors, plasticizers, lubricants, and hydraulic fluids, as well as in inks and sealers for gaskets and furnaces. Manufacture and use of PCBs was banned by EPA in the 1970s and 1980s due to ecological concerns.

Polychlorinated dioxins and furans are unintended byproducts found in association with certain industrial sites, waste incinerators, and combustion, especially of chlorinated material. Pulp and paper mills that used chlorine in their bleaching process were a major historical source of dioxins and furans in the Pacific Northwest. This practice was discontinued in the 1990s. Nationwide, reductions in dioxin and furan emissions have occurred from a combination of regulatory activities, improved emission controls, voluntary actions on the part of industry, and the closing of a number of facilities.

PCBs, dioxins, and furans enter waterbodies through direct discharge, runoff, and atmospheric deposition. Atmospheric deposition occurs because these compounds are sufficiently volatile to evaporate and then deposit in cooler regions.



Figure 3. Chemical Structure of PCBs, Dioxin, and Furans.

PCB, dioxin, and furan levels can be elevated in the absence of local sources solely due to atmospheric deposition. Meijer et al. (2003), for example, showed that >80% of the soil burden of PCBs in the northern hemisphere was derived from atmospheric as opposed to local sources. Contamination of mountain lake and polar food webs with PCBs, dioxins, and other organochlorines has been attributed to atmospheric sources (Wania and Mackay, 1993; Blais et al., 1998; Gillian and Wania, 2005).

303(d) Human Health Criteria

Washington's human health criteria were issued to the state in 1992 in what is commonly called the EPA National Toxics Rule (NTR; 40 CFR Part 131; Federal Register Vol. 57, No. 246; Vol. 63, No. 63 – Revision of Polychlorinated Biphenyls (PCB) Criteria). Washington was one of 14 states issued criteria in the federal promulgation. The criteria apply to approximately 91 different chemicals.

EPA's human health criteria minimize or specify the potential risk of adverse human effects due to substances in ambient water. The criteria are generally assumed to be set at levels that will avert negative effects to the majority of the public, as well as special subpopulations that,

because of high water or fish-intake rates or biological sensitivities, have an increased risk of receiving a dose that will elicit adverse effects. If the criteria are met, the concentrations of contaminants should remain below the levels that would generally raise public health concerns and result in fish advisories.

The criteria are based on a set of exposure assumptions as follows: a chemical-specific laboratory-derived bioconcentration factor, 6.5 g/day fish consumption, 2 L/day water consumption, 70 kg body weight, 70 year duration of exposure, and 70 year lifespan.

The toxicity values used to calculate the criteria for both non-carcinogens and carcinogens are from the EPA Integrated Risk Information System (IRIS) database. Criteria for carcinogens such as PCBs and dioxins are based on a one-in-one-million (10^{-6}) additional cancer risk from exposure to a specific contaminant in fish and shellfish. Criteria for non-carcinogens are calculated so that non-cancer effects should not occur (criteria are set so that toxicity thresholds are not exceeded)

Ecology's 303(d) listing criteria for PCBs and dioxin in edible fish tissue are shown in Table 1. The criteria are derived from EPA bioconcentration factors and human health water column criteria established for fish consumption under the NTR. In essence, the 303(d) fish tissue criteria are the human health water quality criteria expressed in tissue form.

| Chemical | Water Quality Criteria | X | Bioconcentration Factor | x | Unit Conversion Factor | = | Fish Tissue Criteria |
|--------------------------|---------------------------|---|----------------------------|---|---------------------------|---|-------------------------|
| Total PCBs | 0.17 ng/L† | | 31,200 L/Kg | | 0.001 | | 5.3 ug/Kg** |
| Dioxin (2,3,7,8-TCDD) | 0.013 pg/L†† | | 5,000 L/Kg | | 0.001 | | 0.07 ng/Kg† |

Table 1. Human Health Criteria Used to 303(d) List Waterbodies for PCBs and Dioxin in Edible Fish Tissue* (wet weight).

 $*10^{-6}$ health risk @ 6.5 grams per day.

†parts per trillion.

**parts per billion.

††parts per quadrillion.

Ecology (2006) sets out the policies for assessing 303(d) listed waters (www.ecy.wa.gov/programs/wq/303d/wqp01-11-ch1Final2006.pdf). A waterbody segment is placed in Category 5 if either the mean of three single fish samples with the highest concentration or one composite sample made up of at least five fish exceeds the applicable criteria for total PCBs or dioxin. Only detected PCB congeners or detected Aroclor-equivalents are used in calculating total PCBs.

The water quality standards do not include criteria for dioxin TEQs. Applying dioxin TEQs to the dioxin criterion confers greater protection than provided by using the TCDD concentrations alone. Waterbodies that exceed the dioxin criterion based on TEQs are placed in Category 2 (Waters of Concern).

WDOH Screening Values

The Washington State Department of Health (WDOH), Office of Environmental Health, Safety, and Toxicology (OEHST), uses a different set of criteria called *screening values* to determine if chemical contaminants in edible fish tissue warrant further assessment, possibly leading to a fish consumption advisory. When screening values are exceeded, OEHST compares the concentrations with known background levels, concentrations in other foods, reductions from cleaning and cooking techniques, and known benefits from fish consumption. A range of fish consumption rates are considered, from recreational anglers to Native American subsistence consumers.

Because WDOH uses a non-cancer approach, the NTR target cancer risk is often exceeded without prompting an advisory. Another important consideration for WDOH is that advisories must be fashioned so that they can be communicated effectively to the public.

Table 2. WDOH Screening Values for PCBs and Dioxin in Edible Fish Tissue: Eight Meals per Month.* (wet weight)

| | Fish Tissue Screening Values | | | | |
|------------|------------------------------|----------------|----------------|----------------|--|
| Chemical | Non- | Carcinogens | | | |
| | carcinogens | $(RL=10^{-4})$ | $(RL=10^{-5})$ | $(RL=10^{-6})$ | |
| Total PCBs | 23 | 59 | 6 | 0.6 | |
| (ug/Kg) | | • • | ÷ | | |
| TCDD TEQs | | 0.75 | 0.075 | 0.008 | |
| (ng/Kg) | | 0.75 | 0.075 | 0.008 | |

*60 grams per day.

RL = risk level.

OEHST was contacted about using the WDOH screening values to help prioritize the 303(d) listings for PCBs and dioxin. For PCBs they recommended using the 23 ug/Kg screening value for non-cancer effects (Table 2). The Table 2 values are for a consumption rate of eight meals per month at 60 grams per day. The consumption rate behind Washington State's human health criteria is 6.5 grams per day.

Because PCBs often exceed 23 ug/Kg in both freshwater and marine fish species, OEHST typically views 100 ug/Kg total PCBs as a cutoff above which there is a strong possibility that a fish advisory will be needed. Again, the final decision on issuing an advisory depends on consumption rates and favored fish species of the population at risk, as well as mitigating factors described above.

WDOH uses TCDD TEQs in its screening values. The screening values for TCDD TEQs are extremely low (Table 2) and, thus, are exceeded almost anytime TCDD is detected. Unlike PCBs, WDOH has not yet established a level of concern below which TCDD is considered to represent an insignificant health risk. WDOH will re-evaluate both the dioxin screening level and level of concern once EPA completes their comprehensive reassessment of dioxin exposure and human health effects that is expected to come out sometime in 2010 (Dave McBride, personal communication).

http://cfpub.epa.gov/ncea/cfm/recordisplay.cfm?deid=55265&CFID=8435202&CFTOKEN=223 03812&jsessionid=4a30f63c0215f4705a991c5180233233802e.

EPA Screening Values

EPA has also developed screening values for carcinogenic and non-carcinogenic substances to help prioritize areas that may present risks to humans from fish consumption (EPA, 2000). The EPA values are for guidance only and are not regulatory thresholds. The approach in their development was the same as for the NTR/303(d) criteria except (1) a cancer risk level of 10^{-5} was used rather than 10^{-6} and (2) consumption rates were set at 17.5 grams/day for recreational fishers and 142 grams/day for subsistence fishers. Table 3 shows the EPA screening values for PCBs and dioxins.

| | Recreation | al Fishers+ | Subsistence Fishers** | | |
|-----------------------|---------------------|-------------|-----------------------|-------------|--|
| Chemical | Non- carcinogens | Carcinogens | Non- carcinogens | Carcinogens | |
| Total PCBs (ug/Kg) | 80 | 20 | 9.83 | 2.5 | |
| TCDD TEQs (ng/Kg) | | 0.26 | | 0.032 | |

 Table 3. EPA Screening Values for PCBs and Dioxins in Edible Fish Tissue.* (wet weight)

*10⁻⁵ health risk.

+17.5 grams per day.

**142 grams per day.

The 303(d) criteria are compared to WDOH and EPA screening values in Table 4.

| | D' 1 | Consumption | Criteria/ Screening Value | | | | |
|--------------------------|-----------------|-------------------|------------------------------|-------------------|--|--|--|
| Туре | Level | Rate (gms/day) | PCBs | TCDD/TCDD TEQs | | | |
| | | | (ug/Kg) | (ng/Kg) | | | |
| Carcinogens | Carcinogens | | | | | | |
| 303(d) Listing Criteria* | 10-6 | 6.5 | 5.3 | 0.07 | | | |
| WDOH Saraaning Values* | 10-6 | 60 | 0.6 | 0.008 | | | |
| wDOH Screening values | 10-5 | 60 | 6.0 | 0.075 | | | |
| EDA Saraaning Valuag** | 10-5 | 17.5 | 20 | 0.26 | | | |
| EFA Screening values. | 10-5 | 142 | 2.5 | 0.032 | | | |
| Non-carcinogens | Non-carcinogens | | | | | | |
| WDOH Screening Values† | 10-5 | 60 | 23 | | | | |
| EDA Sereening Volues** | 10-5 | 17.5 | 80 | | | | |
| EFA Screening values. | 10-5 | 142 | 9.8 | | | | |

Table 4. Comparison of Human Health Edible Fish Tissue Criteria and Screening Values.

*EPA National Toxics Rule.

†8 meals/month. **EPA (2000).

Study Design

Selection of Background Waterbodies

For purposes of this study, the term "background" denotes a waterbody where the only known or likely PCB or dioxin source of significance is the atmosphere. These are lakes, rivers, and streams where no reason was found to suspect significant past use or sources of PCBs, dioxins, or furans in the watershed.

Background sites were selected by examining Washington State maps and GIS coverages showing population density, agricultural land use, public lands, annual precipitation, and wind direction. This exercise identified areas that have a low probability of local sources of contamination.

Fisheries biologists and resource managers for the Washington Department of Fish and Wildlife (WDFW), National Park Service, U.S. Forest Service, and Ecology were asked to identify potential background lakes and rivers within these areas, using the following criteria:

- Elevation under approximately 3,000 ft.
- Watershed relatively undisturbed or logging only.
- At least two nonplanted fish species of catchable size.
- Good accessibility.

Based on the mapping exercise and waterbody recommendations, four lakes and two rivers were selected for sampling in each of four regions: Western Washington, West Slope of the Cascades, East Slope of the Cascades, and eastern Washington. The location of these regions relative to the Pacific Ocean air mass and atmospheric sources in urban Puget Sound, and the barrier to these influences caused by the Cascade Mountains were viewed as having the potential to result in substantially different rates of chemical deposition. There are large differences in precipitation across the study area.

An effort was made to distribute the sampling effort along a north-south gradient within each of the four regions. The waterbodies selected included a mix of impoundments and natural waterbodies of various sizes, as is the case with the 303(d) list. The sampling effort was weighted toward lakes because of (1) the low diversity of fish species in most rivers that might qualify as background and (2) the greater ability of fish to move into and out of rivers as opposed to lakes. Lakes and impoundments also dominate the 303(d) list. It was not possible to locate good background sites within the major Eastern Washington agricultural basins of the Yakima, Columbia, Palouse, and Walla Walla Rivers.

The appropriateness of the waterbodies selected as background sites was checked against Ecology's Facility Site Identification System (<u>http://ecyapps3/facilitysite/</u>). Facility Site identifies sites known to Ecology as having an active or potential impact on the environment. Table 5 lists the 24 waterbodies where fish samples were ultimately collected for the study; their locations are shown in Figure 4.

| Region and Waterbody Name | Location | Lake Elevation (ft) | Lake Area (acres) | River Drainage Area (mi ²) | | | |
|------------------------------|------------------------------------|---------------------------|-------------------------|---|--|--|--|
| Western Washington | | | | | | | |
| Mountain Lake | Moran State Park, Orcas Island | 914 | 198 | | | | |
| Ozette Lake | Olympic NP | 29 | 7,787 | | | | |
| Quinault River | Quinault Reservation/Olympic NP | | | 264 | | | |
| Cushman Lake | Olympic NF | 731 | 4,003 | | | | |
| North River | Chehalis River drainage | | | 219 | | | |
| Loomis Lake | Loomis Lake State Park, Long Beach | 25 | 200 | | | | |
| West Slope of the Case | cades | | | | | | |
| Baker Lake | N. Cascade NP/Baker-Snoqualmie NF | 724 | 3,616 | | | | |
| Sauk River | Baker-Snoqualmie NF | | | 714 | | | |
| Spada Lake | Baker-Snoqualmie NF | 1,435 | 1,870 | | | | |
| Chester Morse Lake | Baker-Snoqualmie NF | 1,555 | 1,682 | | | | |
| Panamaker Creek | Gifford Pinchot NF | | | ~10 | | | |
| Merrill Lake | Gifford Pinchot NF | 1,541 | 344 | | | | |
| East Slope of the Case | ades | | | | | | |
| Patterson Lake | Okanogan NF | 2,740 | 18 | | | | |
| Entiat River | Wenatchee NF | | | 419 | | | |
| Fish Lake | Wenatchee NF | 2,257 | 513 | | | | |
| Cle Elum Lake | Wenatchee NF | 2,224 | 4,810 | | | | |
| Bumping Lake | Wenatchee NF | 3,426 | 1,310 | | | | |
| Klickitat River | Klickitat County | | | 1,297 | | | |
| Eastern Washington | | | | | | | |
| Kettle River | Colville NF | | | 4,070 | | | |
| Sullivan Lake | Colville NF | 2,546 | 1,290 | | | | |
| South Twin Lake | Colville Indian Reservation | 2,572 | 977 | | | | |
| Omak Lake | Colville Indian Reservation | 954 | 3,244 | | | | |
| Badger Lake | South of Turnbull NWR | 2,178 | 244 | | | | |
| Pataha Creek | Umatilla NF | | | 200 | | | |

 Table 5. Waterbodies Where Fish Samples Were Collected.

NP – National Park.

NF – National Forest.

NWR – National Wildlife Reserve.





| Western Washington |
|--------------------|
| 1 = Mountain Lake |

- 2 = Ozette Lake
- 3 = Quinault River
- 4 = Cushman Lake
- 5 =North River
- 6 = Loomis Lake
- 8 = Sauk River 9 = Spada Lake 10 = Chester Morse Lake 11 = Merrill Lake

West Slope Cascades

7 = Baker Lake

12 = Panamaker Creek

East Slope Cascades 13 = Patterson Lake 14 = Entiat River 15 = Fish Lake 16 = Cle Elum Lake 17 = Bumping Lake 18 = Klickitat River Eastern Washington 19 = Sullivan Lake 20 = Kettle River 21 = Omak Lake 22 = South Twin Lake 23 = Badger Lake 24 = Pataha Creek

While pristine, high mountain lakes obviously qualify as background, they were not included in this study because of enhanced atmospheric deposition due to colder temperatures and greater amounts of precipitation (Blais et al., 1998; Gillian and Wania, 2005). High lakes have the additional drawbacks of difficult access and low fish diversity.

The U.S. Geological Survey (USGS) has analyzed PCBs in fish from 14 Washington mountain lakes over 3,000 ft. in elevation (Moran et al., 2007). Of the 19 samples in the study, 78% were nondetect for PCBs because a relatively insensitive GC/ECD method was used (50 ug/Kg reporting limit). However, USGS estimated concentrations of 17 - 20 ug/Kg total PCBs in approximately 20% of the samples, higher levels than a number of the waterbodies on the 303(d) list. This finding supports the approach of using lower elevation sites as a background reference for 303(d) listed waters.

Some historical data were available on PCBs and dioxin in fish from several of the lakes and rivers selected for the 2007-08 Background Study. The results confirmed low concentrations in Mountain, Ozette, Chester Morse, and Patterson Lakes, and the Entiat River (0.9 to <10 ug/Kg total PCBs, <0.1 ng/Kg TCDD; Seiders et al., 2007, EPA, 2005).

Fish Samples

The Background Study focused on the larger fish species more likely to be consumed and on which most of the contaminant data and 303(d) listings for Washington exist. The species of primary interest were rainbow trout (*Oncorhynchus mykiss*), cutthroat trout (*Oncorhynchus clarki*), brook trout (*Salvelinus fontinalis*), kokanee (*Oncorhynchus nerka*), mountain whitefish (*Prosopium williamsoni*), brown trout (*Salmo trutta*), yellow perch (*Perca flavescens*), and largemouth bass (*Micropterus salmoides*). A few samples were also obtained of largescale suckers (*Catostomus macrocheilus*), pike minnow (*Ptychocheilus oregonensis*), and peamouth (*Mylocheilus caurinus*).

Based on past experience, it was anticipated that from one to three species could be collected from each waterbody. An effort was made to collect at least one predator and one bottom feeder from each site, as recommended by EPA (2000). Planted fish were avoided to the extent possible, unless planted as fry.

Fish were collected in 2007 and 2008, primarily during the late summer and fall. The majority of the 303(d) listings are based on studies conducted around this time of year. Lipid content of most fish species is highest at this time, and PCBs and dioxins are primarily associated with lipids.

Fillets were analyzed for all fish samples. Each sample consisted of a composite of pooled tissues, typically from four to five individual fish. Composite samples provide a more cost efficient estimate of mean contaminant concentrations than single fish. At least one composite sample was analyzed from each waterbody. To provide an indication of the potential effect fish size might have on the results, two size classes were analyzed for several waterbodies. Length, weight, age, and sex were determined for each fish used in the composites.

Chemical Analysis

All samples were analyzed for PCBs (209 compounds), PCDDs/PCDFs (17 compounds), and percent lipids. HR-GC/MS methods were used to give detection limits down to 0.04 ug/Kg for PCB congeners and down to 0.01 ng/Kg for PCDDs/PCDFs. Lipids were analyzed for potential use in normalizing concentrations between samples. PCB Aroclor-equivalents were analyzed in a subset of samples for comparison with the congener data.

Sampling Procedures

Fish Collection

Fish sampling followed the EA Program SOP (Sandvik, 2006a). Fish were collected by electroshocking, beach seines, gill net, or hook and line. Only legal size fish were taken. For species with no size limits, only those large enough to reasonably be retained for consumption were taken.

Fish selected for analysis were killed by a blow to the head. Each fish was given a unique identifying number and its length and weight recorded. The fish were individually wrapped in aluminum foil, put in plastic bags, and placed on ice for transport to Ecology headquarters, where the samples were frozen pending preparation of the tissue samples.

Tissue Preparation

Tissue samples were prepared follow the EA Program SOP (Sandvik, 2006b). Techniques to minimize potential for contamination were used. People preparing the samples wore non-talc nitrile gloves and worked on heavy duty aluminum foil or a polyethylene cutting board. The gloves and foil were changed between samples, and the cutting board was cleaned between samples as described below.

The fish were thawed to remove the foil wrapper and rinsed with tap water, then deionized water, to remove any adhering debris. The entire fillet from one or both sides of each fish was removed with stainless steel knives and homogenized in a Kitchen-Aid blender. The fillets were scaled and analyzed skin-on. The sex of each fish was recorded, and hard structures were saved for age determination (scales, otoliths, opercles, dorsal, and/or pectoral spines as appropriate for each species). Age was determined by WDFW or a private laboratory.

Two to seven individual fish (19 small fish in one case) were used for each composite sample. To the extent possible, the length of the smallest fish in a composite was no less than 75% of the length of the largest fish. The composites were prepared using equal weights from each fish. The pooled tissues were homogenized to uniform color and consistency, using three passes through the blender. The homogenates were placed in glass jars with Teflon lid liners, cleaned to EPA (1990) QA/QC specifications.

Cleaning of resecting instruments, cutting boards, and blender parts was done by washing in tap water with Liquinox detergent, followed by sequential rinses with tap water, de-ionized water, and pesticide-grade acetone. The items were then air dried on aluminum foil in a fume hood before use.

The fish tissue samples were refrozen for shipment with chain-of-custody records to the Ecology Manchester Laboratory. The samples were stored frozen at Manchester until shipped to the contract laboratory. Excess tissue was retained for all samples and stored frozen at Ecology Headquarters. The holding time for frozen tissue samples being analyzed for PCBs and PCDDs/PCDFs is up to one year (Methods 1668A and 1613B).

Laboratory Methods

Table 6 shows the laboratory methods used in the study. PCB and PCDD/PCDF congeners were analyzed by accredited laboratories, Pace Analytical Services of Minneapolis, MN, and Pacific Rim Laboratories Inc. of Surrey, BC, through a contract with Manchester Laboratory. Manchester analyzed PCB Aroclor-equivalents and percent lipids. Manchester reviewed and verified all contract laboratory results.

At Ecology's request, the PCDD/PCDF results were reported down to the limit of detection, with values qualified as estimates if they were between the detection limit and quantitation limit. Because the lipid data from contract labs have been highly variable in the past, all samples were analyzed for percent lipids by Manchester to provide a more consistent data set.

| Amolessia | Mathad | Reporting Limit* | Laboratory | | |
|-----------------------------|-------------------------------------|------------------|------------|-------------|--|
| Analysis | Method | (wet weight) | 2007 | 2008 | |
| PCB congeners | High Resolution. GC/MS EPA 1668A | 0.04 - 1.2 ug/Kg | Pace | Pacific Rim | |
| PCB Aroclor- equivalents | GC/ECD EPA 8082 | 0.98 - 2.2 ug/Kg | Manchester | Manchester | |
| PCDDs/PCDFs | High Resolution GC/MS EPA 1613B | 0.01 - 1.0 ng/Kg | Pace | Pacific Rim | |
| Percent lipids | Gravimetric MEL SOP 700009 | 0.1% | Manchester | Manchester | |

Table 6. Laboratory Procedures.

*Varies with congener.

MEL – Manchester Environmental Laboratory.

Data Quality

Data Review and Verification

Manchester Laboratory reviewed the chemical data for this project. For results generated by Manchester, final review was performed by the unit supervisor or an analyst experienced with the method. Chemists on the Manchester staff performed the review for analytical work subcontracted to commercial laboratories. Quality assurance and quality control at Manchester are described in their Laboratory Users Manual.

Manchester prepared written case narratives assessing the quality of all data collected. These reviews include a description of analytical methods and an assessment of holding times, initial and continuing calibration and degradation checks, method blanks, surrogate and labeled compound recoveries, matrix spike recoveries, laboratory control samples, and laboratory duplicates. The reviews and the complete Manchester data reports are available from the author on request.

A Quality Assurance Project Plan (Johnson, 2008) established measurement quality objectives (MQOs) for accuracy, bias, and reporting limits. To determine if MQOs were met, the project lead compared results on field and laboratory quality control samples to the MQOs. To evaluate whether the reporting limit targets were met, the results were examined for non-detects and to determine if any values exceeded the lowest concentration of interest. Based on these assessments and a review of the laboratory data packages and Manchester's data verification reports, the data were either accepted, accepted with appropriate qualifications, or rejected and re-analysis or re-sampling considered.

The data contained in this report are useable as qualified. Qualified data are flagged as follows:

- U = The analyte was not detected at or above the reported sample quantitation limit.
- UJ = The analyte was not detected at or above the reported sample quantitation limit (PCB congeners) or detection limit (PCDDs/PCDFs). However, the reported quantitation and detection limits are approximate and may or may not represent the actual limits necessary to accurately measure the analyte in the sample.
- J = The analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sample.
- N = The analysis indicates the presence of an analyte for which there is presumptive evidence to make a tentative identification.
- NJ = The analysis indicates the presence of an analyte that has been tentatively identified, and the associated numerical value represents its approximate concentration.

Precision

The field variability inherent in the chemical residues accumulated by fish was minimized by using composite samples. Estimates of analytical precision were obtained by analyzing laboratory duplicates (splits). The results are summarized in terms of relative percent difference (RPD) in Appendix F. RPD is the difference between duplicates as a percent of the mean value.

RPDs for PCB congeners were 26% or better, except 66% for a single low-level sample. The duplicates analyzed for Aroclor-equivalents agreed within 13% or better. TCDD was unambiguously detected in one sample only, with an RPD of 24%. RPDs for TCDD TEQs ranged from 6-73%; the higher RPDs were again associated with the lower level samples. The lipid duplicates were in good agreement, 0-16% RPD.

Results¹ and Discussion

Samples and Species Analyzed

A total of 52 fish fillet samples were analyzed for the 2007-08 PCB and Dioxin Background Study (Table 7). These samples were obtained from 24 waterbodies, six in each of four regions of the state (see Figure 4). Similar numbers of samples were collected from each region (13-16). Two species were analyzed per waterbody, on average. Each sample typically consisted of composite muscle tissue from four or five individual fish.

| 1 / | - |
|------------------------------------|-------|
| Samples Analyzed | 52 |
| Waterbodies Sampled | 24 |
| Waterbodies per Region | 6 |
| Samples per Region | 13-16 |
| Fish Species Sampled | 11 |
| Fish Species per Waterbody | 1-6 |
| Average Fish Species per Waterbody | 2 |

Table 7. Number of Samples, Waterbodies, and Species Analyzed.

Salmonids were encountered more frequently than spiny rayed fishes (35 versus 19 samples; Table 8). Of the 11 species collected for the study, cutthroat, rainbow trout, mountain whitefish, and largescale suckers were most common at 11-17% of samples. Brook trout, largemouth bass, kokanee, and northern pike minnow constituted 6-9% of samples. One brown trout and one peamouth were also analyzed. Cutthroat, mountain whitefish, and largescale suckers were collected from all four regions.

Table 8. Fish Species Analyzed.

| Spacios | Number | Percent | |
|----------------------|------------|------------|--|
| species | of Samples | of Samples | |
| Cutthroat Trout | 9 | 17% | |
| Rainbow Trout | 9 | 17% | |
| Mountain Whitefish | 7 | 13% | |
| Largescale Sucker | 6 | 11% | |
| Brook Trout | 5 | 9% | |
| Largemouth Bass | 5 | 9% | |
| Yellow Perch | 4 | 8% | |
| Kokanee | 3 | 6% | |
| Northern Pike Minnow | 3 | 6% | |
| Brown Trout | 1 | 2% | |
| Peamouth | 1 | 2% | |

¹ See Appendix G for the complete data.

PCB Concentrations

Congeners

Results from the PCB congener analysis of the fish fillet samples are summarized as total PCBs in Table 9. Total PCBs are the sum of all detected congeners, including estimated values (J qualified) and tentative identifications (N and NJ qualified). Table 10 has a statistical summary of the data. Units of ug/Kg are equivalent to parts per billion. All data in this report are on a wet weight basis.

| Waterbody | Species Total P | | PCBs | | | |
|---------------------|------------------------------|------|------|--|--|--|
| Western Washington | | | | | | |
| Mountain Lake | Kokanee | 11 | | | | |
| | Largemouth Bass | 0.26 | J | | | |
| Oratta Laka | Northern Pike Minnow (large) | 1.4 | J | | | |
| Ozette Lake | Northern Pike Minnow (small) | 1.3 | J | | | |
| | Yellow Perch | 0.25 | J | | | |
| Quinault River | Cutthroat Trout | 0.45 | | | | |
| | Kokanee | 1.1 | J | | | |
| Cushman Lake | Largescale Sucker | 0.60 | J | | | |
| | Mountain Whitefish | 1.3 | J | | | |
| North River | Cutthroat Trout | 1.0 | | | | |
| | Largemouth Bass | 0.35 | | | | |
| Loomis Lake | Yellow Perch (large) | 0.04 | NJ | | | |
| | Yellow Perch (small) | 0.04 | NJ | | | |
| West Slope Cascades | | | | | | |
| | Cutthroat Trout | 1.2 | J | | | |
| Baker Lake | Largescale Sucker | 0.51 | J | | | |
| | Rainbow Trout | 0.91 | J | | | |
| Sauk River | Mountain Whitefish | 2.0 | J | | | |
| | Cutthroat Trout (large) | | J | | | |
| Spada Lake | Cutthroat Trout (small) | 1.2 | J | | | |
| | Rainbow Trout | 1.7 | J | | | |
| Chester Morse Lake | Rainbow Trout | 1.4 | J | | | |
| Merrill Lake | Cutthroat Trout | 4.4 | | | | |
| Panamaker Creek | Kokanee | | | | | |
| East Slope Cascades | | | | | | |
| D.4 L.1 | Rainbow Trout | 3.4 | | | | |
| Patterson Lake | Yellow Perch | 0.23 | | | | |

Table 9. Total PCB Concentrations in Fish Fillets from 24 Background Waterbodies Across Washington State (ug/Kg, wet weight, congener data).

| Waterbody | Species | Total PCBs | | | |
|--------------------|----------------------------|------------|----|--|--|
| | Brook Trout | 0.29 | | | |
| Entiat River | Rainbow Trout | 0.53 | | | |
| | Brown Trout | 88 | | | |
| | Largemouth Bass | 25 | | | |
| F' 1 T 1 | Largescale Sucker (large) | 3.5 | | | |
| Fish Lake | Largescale Sucker (small) | 2.4 | | | |
| | Northern Pike Minnow | 38 | | | |
| | Rainbow Trout | 5.4 | | | |
| | Largescale Sucker | 6.6 | | | |
| Cle Elum Lake | Mountain Whitefish (large) | 5.1 | | | |
| | Mountain Whitefish (small) | 1.7 | | | |
| Dumning Laka | Brook Trout | 7.3 | | | |
| bumping Lake | Cutthroat Trout | 2.6 | | | |
| Klickitat River | Mountain Whitefish | 6.0 | | | |
| Eastern Washington | | | | | |
| | Cutthroat Trout | 3.5 | J | | |
| Sullivan Lake | Mountain Whitefish (large) | 3.4 | J | | |
| | Mountain Whitefish (small) | 2.0 | J | | |
| Vattla Divan | Largescale Sucker | 0.24 | UJ | | |
| Kettle Kiver | Rainbow Trout | 0.31 | J | | |
| Ormaly Lalva | Cutthroat Trout | 3.8 | | | |
| Отак Lake | Peamouth | 3.0 | | | |
| | Brook Trout (large) | 1.2 | | | |
| South Twin Lake | Brook Trout (small) | 0.83 | | | |
| | Largemouth Bass | 0.25 | NJ | | |
| Dadgar Laka | Largemouth Bass | 1.4 | | | |
| Dauger Lake | Rainbow Trout | 1.0 | | | |
| Pataha Creek | Brook Trout | 0.71 | | | |

U = Not detected at or above reported result.

J = Result is an estimated value.

NJ = There is evidence the analyte is present. The associated numerical result is an estimate.

Table 10. Summary of Results for Total PCB Congeners (ug/Kg, wet weight).

| Number of Samples | 52 |
|---------------------|------|
| Detection Frequency | 98% |
| Minimum | 0.04 |
| Maximum | 88 |
| Mean | 4.9 |
| Median | 1.4 |
| 90th Percentile | 6.5 |

PCBs were detected in all but one fish tissue sample. Approximately 130 individual congeners were reported overall, with about 60 of these detected in at least half the samples. Congeners found in more the 90% of the samples included: PCB-52/69, -70, -74, -85, -86/97/117, -87/115, -93/95/98/102, -101, -105, -128, -138, -146, -153, -163/164, -180, and -182/187.²

Total PCB concentrations ranged from 0.04 to 88 ug/Kg, wet weight. The mean and median were 4.9 and 1.4 ug/Kg, respectively. The 90^{th} percentile value was 6.5 ug/Kg. The 90^{th} percentile is the value exceeded by only 10% of the samples.

Figure 5 shows the PCB data plotted on a regional basis. The highest total PCB concentrations were observed in waterbodies on the east slope of the Cascades. Regional differences were statistically significant for PCBs (Kruskal-Wallis test, p < 0.05). The gradual increase in PCB levels from Western Washington to the East Slope Cascades may reflect an elevation effect, previously discussed (see Study Design).



Figure 5. Box-and-Whisker Plot of the PCB Congener Data by Region.

² Some compounds could not be distinguished due to coelution.

Aroclor-Equivalents

The 303(d) fish tissue listings for PCBs are either based on an analysis of individual congeners or Aroclor-equivalents. Congeners are usually analyzed by high-resolution GC/MS methods that are more costly, but more sensitive and thus give lower detection limits than the GC/ECD method typically employed for Aroclor mixtures. Most of the historical fish tissue data for Washington State are from an Aroclor analysis.

Fish tissues from the Background Study were analyzed in conjunction with other Ecology fish samples from the Washington State Toxics Monitoring Program (WSTMP), discussed later in this report. Forty samples from the Background Study and the WSTMP were analyzed for PCBs by both high resolution GC/MS and GC/ECD methods. The results are compared in Figure 6.



Figure 6. Results of Split Fish Fillet Samples Analyzed for PCB Congeners and Aroclor-Equivalents.

There was a good correlation between congeners and Aroclor-equivalents $(R^2 = 0.81)^3$. Samples having approximately 10 ug/Kg or greater total PCBs showed better agreement that samples with lower concentrations approaching the 1 - 2 ug/Kg reporting limit for Aroclors. Overall, 60% of the Aroclor results had the higher total PCB concentration. In four instances, the Aroclor result qualified for a 303(d) listing, while the congener result did not (Appendix G).

Weathering of PCBs in environmental samples, mixing of different Aroclors, and bioaccumulation complicate quantitation by the Aroclor method. In addition, the Aroclor method does not allow for recovery correction as the congener method does. These factors can lead to errors in assessing the total PCB concentration. Most studies point to a congener analysis

 $^{^{3}}$ R², the coefficient of determination, represents the proportion of common variation in two variables, i.e., the strength or magnitude of the relationship. A value of 1.0 represents a perfect positive correlation.

as being more accurate (e.g., Conner et al., 2005; Sather et al., 2001; Bernhard and Petron, 2001; Schwartz and Stalling 1987).

PCB Toxicity Equivalents (TEQs)

PCB compounds with chlorine atoms located at the ends of the molecule (3/3' to 5/5' in Figure 3) are structurally similar to TCDD and are more toxic than PCBs with chlorines at the inner positions (2/2' and 6/6'). Human and mammalian Toxicity Equivalency Factors (TEFs) have been established for 12 of these compounds, based on their toxicity relative to TCDD (Van den Berg et al., 2006; Appendix H). The TEFs can be used to calculate the TCDD TEQ of a PCB mixture in fish tissue or other types of environmental samples. Ecology does not use PCB TEQs for 303(d) listing purposes (Connor et al., 2005). PCB TEQs can, however, be used to refine the toxicity estimate. This is a potentially useful means of ranking 303(d) listed waterbodies for TMDLs.

An example of how waterbody ranking can differ when done according to total PCB congeners and PCB TEQs is shown in Table 11 for the background lakes and rivers sampled in the Puget Sound basin for the present study. The waterbodies in question move up or down in rank order by up to three places when prioritized by TEQs.

| Ranked | by Total PCBs | | Ranked by PCB TEQs | | Change | |
|-----------------------|-----------------------|---------------|-----------------------|-----------------------|------------|--------------|
| Waterbody | Species | Total PCBs | Waterbody | Species | PCB TEQ | in Rank |
| Mountain Lake | Kokanee | 11 | Mountain Lake | Kokanee | 0.00034 | none |
| Spada Lake | Cutthroat – large | 2.4 | Sauk River | Mountain whitefish | 0.0000040 | 3rd to 2nd |
| Sauk River | Mountain whitefish | 2.0 | Chester Morse Lake | Rainbow trout | 0.0000030 | 5th to 3rd |
| Spada Lake | Rainbow trout | 1.7 | Spada Lake | Cutthroat – large | 0.0000025 | 2nd to 4th |
| Chester Morse Lake | Rainbow trout | 1.4 | Spada Lake | Rainbow trout | 0.0000022 | 4th to 5th |
| Cushman Lake | Mountain whitefish | 1.3 | Spada Lake | Cutthroat – small | 0.0000020 | 7th to 6th |
| Spada Lake | Cutthroat – small | 1.2 | Cushman Lake | Mountain whitefish | 0.0000019 | 5th to 7th |
| Cushman Lake | Kokanee | 1.1 | Cushman Lake | Kokanee | 0.0000019 | none |
| Baker Lake | Rainbow trout | 0.91 | Baker Lake | Rainbow trout | 0.0000016 | none |
| Cushman Lake | Sucker | 0.60 | Baker Lake | Sucker | 0 | none |
| Baker Lake | Sucker | 0.51 | Cushman Lake | Sucker | 0 | tie for 10th |

Table 11. Puget Sound Background Waterbodies Ranked by Total PCBs versus PCB TEQs (ug/Kg, wet weight).

Statewide, the congeners contributing the most to the PCB TEQ in the background samples were PCB-118, -105, and -126 (Figure 7). These compounds were detected in 79%, 62%, and 8% of the samples, respectively.



Figure 7. Percent Contribution of Dioxin-Like PCBs to TEQ in Background Fish Fillet Samples.

In the remainder of this report, the term *total PCBs* refers to the sum of individual PCB congeners and not total PCB TEQs, unless stated otherwise.

Dioxin and Furan Concentrations

Table 12 has the results for TCDD (dioxin) and TCDD TEQs in the background fish samples. In this table, detections are highlighted in bold to better identify where these compounds were detected. A statistical summary follows in Table 13. Units of ng/Kg are equivalent to parts per trillion.

TCDD was detected in only 27% of the fish tissue samples. Most of the detections (10 out of 14) were in western Washington. Overall, concentrations ranged from less than 0.008 to 0.12 ng/Kg. The mean and median were non-detect at 0.03 ng/Kg, and the 90th percentile was 0.041 ng/Kg.

Most waterbodies had one or more PCDDs/PCDFs detected (Figure 8). Thus, the detection frequency for TCDD TEQs was high (94%). TEQ concentrations ranged from 0.007 to 0.65 ng/Kg, with a mean of 0.091 ng/Kg, a median of 0.051 ng/Kg, and a 90th percentile of 0.18 ng/Kg. 1,2,3,7,8-PeCDD, 1,2,3,4,7,8-HxCDF, 2,3,7,8-TCDD, and 2,3,7,8-TCDF were the primary contributors to the TEQ (Figure 9).

| Waterbody | Species | TCDD | | TCDD TEQ | | |
|-----------------------|------------------------------|-------|----|-------------|---|--|
| Western Washington | | | | | | |
| Mountain Lake | Kokanee | 0.12 | | 0.65 | | |
| Largemouth Bass | | 0.019 | UJ | 0.028 | J | |
| 0 | Northern Pike Minnow (large) | 0.033 | UJ | 0.010 | J | |
| Ozette Lake | Northern Pike Minnow (small) | 0.026 | UJ | 0.012 | J | |
| | Yellow Perch | 0.014 | UJ | 0.016 | J | |
| Quinault River | Cutthroat Trout | 0.03 | U | 0.071 | J | |
| | Kokanee | 0.016 | NJ | 0.14 | J | |
| Cushman Lake | Largescale Sucker | 0.008 | UJ | 0.051 | J | |
| | Mountain Whitefish | 0.014 | NJ | 0.056 | J | |
| North River | Cutthroat Trout | 0.032 | NJ | 0.032 | J | |
| | Largemouth Bass | 0.03 | U | 0.042 | | |
| Loomis Lake | Yellow Perch (large) | 0.03 | U | 0.033 | | |
| | Yellow Perch (small) | 0.031 | NJ | 0.091 | J | |
| West Slope of the Ca | scades | | | | | |
| | Cutthroat Trout | 0.031 | UJ | 0.092 | J | |
| Baker Lake | Largescale Sucker | 0.011 | NJ | 0.035 | J | |
| | Rainbow Trout | 0.063 | NJ | 0.35 | J | |
| Sauk River | Mountain Whitefish | 0.060 | NJ | 0.18 | J | |
| | Cutthroat Trout (large) | 0.045 | UJ | 0.15 | J | |
| Spada Lake | Cutthroat Trout (small) | 0.027 | NJ | 0.12 | J | |
| | Rainbow Trout | 0.039 | UJ | 0.017 | J | |
| Chester Morse Lake | Rainbow Trout | 0.024 | UJ | 0.070 | J | |
| Merrill Lake | Cutthroat Trout | 0.03 | U | 0.050 | U | |
| Panamaker Creek | Kokanee | 0.03 | NJ | 0.45 | J | |
| East Slope of the Cas | cades | | | | | |
| D.() I.1 | Rainbow Trout | 0.03 | U | 0.031 | | |
| Patterson Lake | Yellow Perch | 0.03 | U | 0.10 | J | |
| | Brook Trout | 0.03 | U | 0.049 | J | |
| Entiat River | Rainbow Trout | 0.059 | NJ | 0.090 | J | |
| | Brown Trout | 0.03 | U | 0.062 | J | |
| | Largemouth Bass | 0.03 | U | 0.048 | | |
| | Largescale Sucker (large) | 0.03 | U | 0.13 | J | |
| rish Lake | Largescale Sucker (small) | 0.03 | U | 0.050 | | |
| | Northern Pike Minnow | 0.03 | U | 0.025 | J | |
| | Rainbow Trout | 0.03 | U | 0.026 | | |

Table 12. TCDD and TCDD TEQ Concentrations in Fish Fillets from 24 Background Waterbodies Across Washington State (ng/Kg, wet weight).

| Waterbody | Species | TCDD | | TCDD TEQ | | | |
|----------------------------|----------------------------|-------|----|-------------|---|--|--|
| | Largescale Sucker | 0.03 | U | 0.15 | J | | |
| Cle Elum Lake | Mountain Whitefish (large) | 0.03 | U | 0.021 | J | | |
| | Mountain Whitefish (small) | 0.03 | U | 0.095 | J | | |
| Dumning Lalza | Brook Trout | 0.03 | U | 0.025 | J | | |
| bumping Lake | Cutthroat Trout | 0.03 | U | 0.014 | | | |
| Klickitat River | Mountain Whitefish | 0.041 | | 0.19 | | | |
| Eastern Washington | Eastern Washington | | | | | | |
| | Cutthroat Trout | 0.017 | NJ | 0.12 | J | | |
| Sullivan Lake | Mountain Whitefish (large) | 0.041 | NJ | 0.19 | J | | |
| | Mountain Whitefish (small) | 0.039 | UJ | 0.011 | J | | |
| Vattla Divar | Largescale Sucker | 0.04 | UJ | 0.008 | J | | |
| Kettle Kivel | Rainbow Trout | 0.03 | UJ | 0.16 | J | | |
| Omaly Laka Cutthroat Trout | | 0.03 | U | 0.013 | J | | |
| Olliak Lake | Peamouth | 0.03 | U | 0.077 | J | | |
| | Brook Trout (large) | 0.03 | U | 0.075 | | | |
| South Twin Lake | Brook Trout (small) | 0.03 | U | 0.035 | J | | |
| | Largemouth Bass | 0.03 | U | 0.007 | | | |
| Radger Lake | Largemouth Bass | 0.03 | U | 0.050 | U | | |
| Dauger Lake | Rainbow Trout | 0.03 | U | 0.050 | U | | |
| Pataha Creek | Brook Trout | 0.03 | U | 0.056 | J | | |

Detections are highlighted in **bold.** U = Not detected at or above reported result.

J = Result is an estimated value.

NJ = There is evidence the analyte is present. The associated numerical result is an estimate.

| Table 13. Summary of Results for TCDD and TCDD | TEQs (ng/Kg, wet weight). |
|--|---------------------------|
|--|---------------------------|

| | TCDD | | TCDD TEQ |
|---------------------|-------|----|-------------|
| Number of Samples | 52 | | 52 |
| Detection Frequency | 27% | | 94% |
| Minimum | 0.008 | UJ | 0.007 |
| Maximum | 0.12 | | 0.65 |
| Mean | 0.03 | U | 0.091 |
| Median | 0.03 | U | 0.051 |
| 90th Percentile | 0.041 | | 0.18 |

U = Not detected at or above reported result.

J = Result is an estimated value.



Figure 8. Detection Frequency of PCDDs and PCDFs in Background Fish Fillet Samples.



Figure 9. Percent Contribution of PCDD/PCDF Congeners to the TCDD TEQ in Background Fish Fillet Samples.
There were no significant regional differences in either TCDD or TCDD TEQ concentrations (Figure 10).



Figure 10. Box-and-Whisker Plots of the TCDD and TCDD TEQ Data by Region (non-detects plotted at the detection limit).

Background versus Non-background

The 303(d) list is not representative of PCB or dioxin levels in Washington State because the list is biased toward contaminated sites and includes some older studies. A more recent and balanced assessment is provided by Ecology's WSTMP (Seiders and Deligeannis, 2009). WSTMP is a screening-level effort which targets waterbodies across Washington. Results are used primarily to identify areas of concern for follow-up actions. The bulk of the newer fish tissue listings comes from this program.

Results from the Background Study were compared to statewide results from the WSTMP to assess the extent to which the sites selected as background were in fact "cleaner" than those having potential for contamination. Fifty-seven of the waterbodies analyzed for PCB congeners or PCDDs/PCDFs in fish fillets by WSTMP between 2001 and 2008 were identified as non-background, using the same approach as in the present 2007-08 study.

The background and non-background total PCB congener data are compared in Figure 11. Table 14 summarizes the two data sets.



Figure 11. Total PCBs in Fish Fillets from Background (present study) and Non-Background (selected WSTMP sites) Waterbodies Across Washington State.

| | Background Sites 2007-08 | Non-Background Sites 2001-08 |
|---------------------|--------------------------------|---------------------------------|
| Number of Samples | 52 | 84 |
| Detection Frequency | 98% | 74% |
| Minimum | 0.044 | 0.63 |
| Maximum | 88 | 1632 |
| Mean | 4.9 | 58 |
| Median | 1.4 | 10 |
| 90th Percentile | 6.5 | 90 |

Table 14. Summary of Total PCB Data on Fish Fillets from Background and Non-Background Waterbodies Across Washington State (ug/Kg, wet weight; congener analysis).

Waterbodies sampled in the Background Study clearly have much lower PCB concentrations than non-background waterbodies surveyed through the WSTMP. The background mean, median, and 90th percentile total PCB values are an order of magnitude below non-background. Detection frequency is somewhat lower for non-background due to differences in reporting limits.

The TCDD and TCDD TEQ data are similarly compared in Figure 12 and Table 15.

Roughly a third of the non-background samples had low to non-detectable TCDD concentrations similar to background. TCDD concentrations in the remaining samples were elevated over background to a greater extent that was evident for PCBs. The distribution of the two data sets suggests that significant TCDD sources are less widespread in Washington than PCB sources.

TCDD TEQs in background and non-background waterbodies differ to a much lesser extent than either PCBs or TCDD. Mean and median TCDD TEQs for background and non-background are within a factor of 2 or less. This relatively modest elevation in non-background waterbodies suggests there is a significant background contribution to the TCDD TEQ at many locations in Washington.



Figure 12. TCDD and TCDD TEQs in Fish Fillets from Background and Non-Background Waterbodies Across Washington State.

| | | r | ГCDD | TCDD TEQ | | | |
|---------------------|---------------------|----|-------------------------|---------------------|-------------------------|--|--|
| | Background Sites | | Non-Background Sites | Background Sites | Non-Background Sites | | |
| Number of Samples | 52 | | 109 | 52 | 109 | | |
| Detection Frequency | 27% | | 96% | 94% | 96% | | |
| Minimum | 0.0075 | UJ | 0.0005 | 0.0066 | 0.015 | | |
| Maximum | 0.12 | | 12 | 0.65 | 1.9 | | |
| Mean | 0.033 | U | 0.51 | 0.091 | 0.17 | | |
| Median | 0.03 | U | 0.20 | 0.051 | 0.066 | | |
| 90th Percentile | 0.041 | | 0.71 | 0.18 | 0.55 | | |

Table 15. Summary of TCDD and TCDD TEQ Data on Fish Fillets from Background and Non-Background Waterbodies Across Washington State (ng/Kg, wet weight).

U = Not detected at or above reported result.

J = Result is an estimated value.

Biological Parameters

Lipid Content

Concentrations of lipophilic contaminants such as PCBs and dioxins are frequently corrected for variation in lipid content of the tissues analyzed. The lipid normalized data may reveal patterns not apparent on a wet weight basis. Lipid normalizing, however, is only appropriate when the chemical concentrations are directly proportional to lipids.

The total PCB congeners and TCDD TEQ data from the Background Study are plotted against percent lipids in Figures 13 and 14. The correlation between PCBs and percent lipids was poor $(R^2 = 0.26)$. There was no apparent correlation between TCDD TEQs and lipids $(R^2 = 0.03)$. TCDD was not detected frequently enough to assess the correlation.

Lipid content, therefore, was not a major factor influencing either total PCB congeners or TCDD TEQ levels in the background data set overall. Herbert and Keenleyside (1995), Stow et al. (1997), Exponent (2003), and others have concluded that lipid normalization may be justified when bioconcentration (uptake from water) is the primary mechanism of accumulation, but is not warranted under many situations for biomagnifying chemicals such as PCBs or dioxins. PCB and dioxin concentrations increase with trophic level due to uptake from food and long retention time in tissue.

Lipid content is probably related to site-specific factors including but not limited to temperature, trophic status, and food availability. Much better correlations were evident for individual lakes. For example, R^2 values for total PCBs versus percent lipids were 0.76 - 0.77 in Fish Lake and Ozette Lake, where the largest variety of species were collected for the background study.



Figure 13. Total PCBs versus Lipid Content in Fish Fillets from Background Waterbodies.



Figure 14. TCDD TEQs versus Lipid Content in Fish Fillets from Background Waterbodies.

Fish Size

Larger, older fish at the tops of food chains tend to have higher concentrations of bioaccumulative chemicals (e.g., DeVault et al., 1989).

As part of this study, two size classes of fish were analyzed for six species to provide estimates of the influence of fish size on PCB and dioxin concentrations. The samples came from seven waterbodies (see Tables 9 and 12).

Figure 15 shows how total PCB and TCDD TEQ concentrations varied by fish length. Size effects could not be assessed for TCDD because concentrations were at or below detection limits in these samples.



Figure 15. Total PCB congeners and TCDD TEQs in Large versus Small Fish of Six Species. (*Individual sample results from Tables 9 and 12*)

Total PCBs were clearly affected by fish size. PCB levels were substantially higher (by factors of 1.5 to 3) in five of the large fish samples out of the seven sample pairs analyzed and marginally higher in one additional pair of samples. One low-level sample had similar PCB concentrations in both large and small fish. A consistent size affect was not readily apparent for TCDD TEQs.

The Background Study focused on the larger fish more likely to be retained and eaten by anglers. The same is true of most efforts that have resulted in 303(d) listings. If there is a size bias in the background data, would tend to favor higher concentrations.

Species Differences

Contaminant residues in fish vary from species to species due to the many factors that affect bioaccumulation. This study was not designed to evaluate species differences in PCB or dioxin levels. Species were generally analyzed in accordance with their abundance and ease of collection.

Salmonids were encountered more frequently than spiny rayed fishes – 35 versus 19 samples – and thus dominate the sample set. Out of the 10 samples with the highest total PCB concentrations, seven were salmonids. Spiny rays had eight out of the 10 lowest total PCB concentrations. Of the 13 samples where TCDD was detected, 11 were salmonids. Similarly, salmonids tended to dominate TCDD TEQs, with eight out of the 10 highest concentrations.

Chemical Patterns

The PCB and PCDD/PCDF background data were examined for spatial patterns that might indicate local versus regional sources.

Penta-, hexa-, and hepta-chlorobiphenyls were the dominant PCB homologues in the background fish tissue samples. Trichlorobiphenyls were substantially more abundant in several northwest Washington waterbodies (Ozette, Baker, Sauk, and Cushman) and in one northeastern Washington lake (Sullivan) (Figure 16). Other than the suggestion of a clustering in northwest Washington, it is unclear what these locations might have in common that would cause this result. Detection limit differences between 2007 and 2008, coupled with a number of coeluting congeners in the 2007 data, obscured spatial patterns that might have existed among specific PCB congeners.



Figure 16. Relative Abundance of Trichlorobiphenyls in Fish Tissue Samples from Background Waterbodies (mean TriCB concentration as percent of total PCBs).

None of the PCDD or PCDF congeners were detected consistently enough statewide to analyze spatial patterns in the data. 1,2,3,4,7,8-HxCDF, 1,2,3,7,8-PeCDD, 1,2,3,4,6,7,8-HpCDD, and 2,3,7,8-TCDF were most frequently found, but were only reported in about half (44-56%) of the samples. TCDD was detected in about one-third (29%) of the samples (see Figure 8). On the basis of detection frequency alone, Western Washington and the West Slope of the Cascades, with eight out of the 10 samples detected, appeared to be more impacted by TCDD sources than Eastern Washington. Unlike PCB congeners, detection limits for PCDDs/PCDFs were comparable between sampling years

Comparison with Human Health Criteria

The total PCB concentrations measured in the Background Study are compared to the human health criterion used for 303(d) listing in Figure 17.



Figure 17. Total PCBs in Background Fish Fillet Samples Compared to 303(d) Listing Criterion (ug/Kg, wet weight; congener analysis).

Most samples (85%) met the 5.3 ug/Kg criterion for total PCBs. Criteria exceedances were observed for PCBs in several waterbodies on the East Slope of the Cascades. These included Fish Lake (25-88 ug/Kg), Bumping Lake (7.3 ug/Kg), Cle Elum Lake (6.6 ug/Kg), and the Klickitat River (6.0 ug/Kg). Mountain Lake on Orcas Island in the San Juans also exceeded the criterion at 11 ug/Kg. Mountain Lake is currently 303(d) listed based on a total PCB concentration of 10 ug/Kg measured in a sample of kokanee fillets collected in 2004 (Seiders et al., 2007).

Plots of the TCDD and TCDD TEQ data are in Figure 18. Only Mountain Lake (0.12 ng/Kg) exceeded (did not meet) the 0.07 ng/Kg human health criterion for TCDD. One sample each from Baker Lake, Sauk River, and the Entiat River approached this criterion, 0.059 – 0.063 ng/Kg.

Forty percent of the samples exceeded human health criteria when viewed in terms of TCDD TEQs. Mountain Lake (0.65 ng/Kg), Panamaker Creek (0.45 ng/Kg), and Baker Lake (0.35 ng/Kg) had substantially higher concentrations than other waterbodies. Mountain Lake, Ozette Lake, and the Quinault River are currently in Category 2 for TCDD TEQs exceeding the TCDD human health criterion.



Figure 18. TCDD and TCDD TEQs in Background Fish Fillet Samples Compared to 303(d) Listing Criteria (ng/Kg, wet weight).

Table 16 summarizes the criteria exceedances observed, showing the number of waterbodies exceeding in each of the four areas where background samples were collected.

| Chemical | Human Health | Percent of Samples Exceeding | Range of Concentrations | Region and Number of Waterbodies Where Exceedances Occurred (out of 6 each) | | | | |
|------------|-----------------|------------------------------------|----------------------------|--|---------------|---------------|-----------------------|--|
| | Criterion | Criterion (n=52) | Criterion | Western Washington | West Slope | East Slope | Eastern Washington | |
| Total PCBs | 5.3 ug/Kg | 15% | 5.4 - 88 | 1* | | 4 | | |
| TCDD | 0.07 ng/Kg | 2% | 0.12 | 1* | | | | |
| TCDD TEQ | 0.07 ng/Kg | 40% | 0.071-0.65 | 4 | 5 | 5 | 4 | |

Table 16. Summary of Human Health Criteria Exceedances Observed in Fish Fillets from Background Waterbodies.

*Mountain Lake, Orcas Island.

Four of the five exceedances for total PCBs were in lakes or rivers on the East Slope of the Cascades. The only western Washington waterbody exceeding the PCB criterion was Mountain Lake, as noted previously.

Other high elevation waterbodies on the east slope of the Cascades have also exceeded the PCB criterion in recent fish tissue studies. Total PCB levels in Keechelus Lake and Kachess Lake, the two large Yakima River reservoirs adjacent to the Cle Elum Lake reservoir, exceeded the criterion in 16 out of 19 fillet samples analyzed in 2006, with concentrations reaching 17 ug/Kg (Johnson et al., 2009 draft).

A sample of kokanee fillets analyzed in 2004 from Cascade Lake, west of Mountain Lake on Orcas Island, had a low concentration of total PCBs, 1.1 ug/Kg (Seiders et al., 2007). Cascade Lake is at a lower elevation than Mountain Lake (346 ft. versus 914 ft.). The difference in PCB concentrations could be partly due to an elevation effect.

Mountain Lake was also the only waterbody in the Background Study that exceeded the TCDD criterion. The Cascade Lake kokanee sample was not analyzed for PCDDs/PCDFs.

Criteria exceedances were common and more equitably distributed statewide for TCDD TEQs. Four to five waterbodies exceeded the criteria in each of the four regions.

The human health exceedances in background versus non-background waterbodies from the WSTMP are compared in Table 17.

| | Human | Percent Exceedi | of Samples | Percent of Waterbodies Exceeding Criterion | | |
|------------|---------------------|---------------------|--------------------------|---|--------------------------|--|
| Chemical | Health Criterion | Background Sites | Non-Background Sites* | Background Sites | Non-Background Sites* | |
| | | (n = 52) | (n =84 - 109) | (n=24) | (N = 46 - 57) | |
| Total PCBs | 5.3 ug/Kg | 15% | 73% | 21% | 81% | |
| TCDD | 0.07 ng/Kg | 2% | 30% | 4% | 38% | |
| TCDD TEQ | 0.07 ng/Kg | 40% | 73% | 75% | 77% | |

Table 17. Comparison of Human Health Criteria Exceedances Observed in Background and Non-Background Waterbodies.

*Data from the Washington State Toxics Monitoring Program (see text).

Conclusions

The fish tissue data generated through the 2007-08 PCB and Dioxin Background Study are summarized in Table 18.

Table 18. Summary Statistics on Background Concentrations of PCBs, TCDD, and TCDD TEQs Measured in Fish Fillets from 24 Washington Lakes, Rivers, and Streams During 2007-08 (N=52, wet weight).

| Chemical | Detection Frequency | Minimum | Maximum | Mean | Median | 90th Percentile |
|--------------------|------------------------|----------|---------|--------|--------|--------------------|
| Total PCBs (ug/Kg) | 98% | 0.04 | 88 | 4.9 | 1.4 | 6.5 |
| TCDD (ng/Kg) | 27% | 0.008 UJ | 0.12 | 0.03 U | 0.03 U | 0.041 |
| TCDD TEQ (ng/Kg) | 94% | 0.007 | 0.65 | 0.091 | 0.051 | 0.18 |

U = Not detected at or above reported result.

J = Result is an estimated value.

The following overall conclusions are indicated for PCBs and dioxins in Washington State freshwater fish:

- In non-background waterbodies, it is common for PCBs and TCDD to be higher than (exceed) Washington State human health criteria for edible fish tissue. Exceedances of the TCDD criteria are also common when viewed in terms of TCDD TEQs.
- Marginal to occasionally substantial (factor of 2 or more) exceedances of the human health criterion can be expected for PCBs in fish from background waterbodies. In the present study, the highest concentrations were encountered in waterbodies on the east slope of the Cascades, possibly due to an elevation effect.
- The TCDD human health criterion is approached but infrequently exceeded in fish from background waterbodies. TCDD sources to Washington waterbodies appear to be less widespread than PCB sources.
- Exceedances of the TCDD human health criterion when calculated using TCDD TEQs are common in background waterbodies. There appears to be a significant background contribution to the TCDD TEQ in Washington freshwater fish.

Recommendations

- Recommendations for prioritizing Washington State's 303(d) listings for PCBs and dioxins in edible freshwater fish tissue are provided in Table 19. Decision points in the prioritization scheme are based on: (1) fish tissue concentrations measured in the 2007-08 Background Study; (2) Washington State human health criteria; and (3) WDOH and EPA screening values. Situations where subsistence or other fishers are consuming certain fish species at especially high rates may justify allocating TMDL resources to waterbodies with lower levels of contamination.
- 2. Caution should be used in applying the results of this work outside the confines of 303(d). The use of statewide values may inappropriately bias outcomes, for example in MTCA actions. In view of regional variability, case-by-case decision making based on additional data gathering may be appropriate.
- 3. Other recommendations that follow from the Background Study include:
 - All Category 5 (the 303(d) List) and Category 2 (Waters of Concern) listings for TCDD, dioxin, and TCDD TEQs should be reviewed and a consistent rule applied for Category 5 and 2 before prioritizing sites for further action.
 - In evaluating the 303(d) listings, be cognizant of the effect that waterbody elevation, fish size, and fish species can have on the chemical residues in fish.
 - Consider using PCB TEQs to help prioritize 303(d) PCB listings for TMDLs or other follow-up work.
 - Where practical, include a high resolution GC/MS analysis for PCB congeners in fish tissue studies when the data will be used for 303(d) listing. This is especially important when marginal exceedances of criteria are anticipated.
 - Conduct verification sampling before taking action on listings for PCBs, TCDD/dioxin, or TCDD TEQs unless based on a large sample size. A range of size classes and fish species should be analyzed.
- 4. The Background Study identified several waterbodies that qualify for Category 5 or Category 2 listing:
 - Fish Lake, Bumping Lake, Cle Elum Lake, and the Klickitat River should be placed in Category 5 of the 303(d) list for not meeting (exceeding) the PCB human health criterion in edible fish tissue. (Mountain Lake is currently listed, see Appendix B.)
 - Mountain Lake should be placed in Category 5 of the 303(d) list for exceeding the TCDD human health criterion in edible fish tissue.
 - Cushman Lake, Loomis Lake, Baker Lake, Sauk River, Spada Lake, Patterson Lake, Entiat River, Fish Lake, Cle Elum Lake, Klickitat River, Sullivan Lake, Kettle River, Omak Lake, and South Twin Lake qualify for Category 2 listing due to the TCDD TEQ exceeding the TCDD human health criterion in edible fish tissue.

| Chemical/ Concentration | Recommendation | Rationale |
|----------------------------|--|--|
| Total PCBs (u | g/Kg) [303(d) listing crite | rion is 5.3 ug/Kg for Category 5: TMDL required] |
| < 10 | No further action | Fish from most background waterbodies have less than 10 ug/Kg. EPA screening level for non-carcinogenic effects for subsistence fishers is 9.8 ug/Kg. |
| 10 - 20 | Low priority for TMDL | Concentrations sometimes encountered in background waterbodies. EPA screening level for recreational fishers for carcinogenic effects is 20 ug/Kg. |
| 20 - 100 | Medium priority for TMDL | Concentrations sometimes encountered in background waterbodies. WDOH non-carcinogen screening level is 23 ug/Kg. Concentrations above 53 ug/Kg exceed Washington human health criteria by a factor of 10 (10⁻⁵ risk level). EPA screening level for recreational fishers for non-carcinogenic effects is 80 ug/Kg. |
| >100 | High priority for TMDL | Exceeds maximum concentration encountered in background waterbodies. WDOH considers >100 ug/Kg as indicating strong possibility of need for fish consumption advisory. |
| TCDD (ng/Kg |) [303(d) listing criterion i | s 0.07 ng/Kg for Category 5: TMDL required] |
| >0.07 - 0.30 | Low priority for TMDL | Concentrations exceeding 0.07 ng/Kg potentially encountered in background waterbodies. EPA screening level for recreational fishers for carcinogenic effects is 0.26 ng/Kg. |
| 0.30 - 0.70 | Medium priority for TMDL | • Exceeds EPA screening level for recreational fishers for carcinogenic effects. |
| >0.70 | High priority for TMDL | • Concentrations above 0.70 ng/Kg exceed Washington human health criteria by a factor of 10 (10 ⁻⁵ risk level). |
| TCDD TEQs (| ng/Kg) [Waters of Conce | rn, listing criterion is 0.07 ng/Kg for Category 2: TMDL not required] |
| <0.20* | No further action | • Background waterbodies often approach 0.20 ng/Kg. |
| 0.20 - 0.70* | Low priority for follow-up action | • Within range of concentrations encountered in background waterbodies. |
| >0.70* | Take follow-up actions (e.g., verification sampling, source identification) | Exceeds maximum concentration encountered in background waterbodies. Concentrations above 0.70 ng/Kg exceed Washington human health TCDD criteria by a factor of 10 (10⁻⁵ risk level). |

Table 19. Recommendations for Prioritizing 303(d) *Freshwater* Fish Tissue Listings for PCBs and Dioxins.

* Assumes TCDD contribution is <0.07 ng/Kg; if not, then qualifies for Category 5.

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Appendices

Appendix A. Waterbody Numbers for 2007-08 Background Study of Lakes, Rivers, and Streams.

Table A-1. Waterbody Numbers for the PCB and Dioxin Background Study.

| Region and Waterbody Name | Waterbody Number | | | |
|------------------------------|---------------------|--|--|--|
| Western Washington | | | | |
| Mountain Lake | WA-02-9060 | | | |
| Ozette Lake | WA-20-9040 | | | |
| Quinault River | WA-21-2020 | | | |
| Cushman Lake | WA-16-9010 | | | |
| North River | WA-24-1010 | | | |
| Loomis Lake | WA-24-9040 | | | |
| West Slope of the Case | cades | | | |
| Baker Lake | WA-04-9010 | | | |
| Sauk River | WA-04-1080 | | | |
| Spada Lake | WA-07-9710 | | | |
| Chester Morse Lake | WA-08-9060 | | | |
| Panamaker Creek | | | | |
| Merrill Lake | WA-27-9020 | | | |
| East Slope of the Case | ades | | | |
| Patterson Lake | WA-13-9120 | | | |
| Entiat River | WA-46-1020 | | | |
| Fish Lake | WA-56-9020 | | | |
| Cle Elum Lake | WA-39-9010 | | | |
| Bumping Lake | WA-38-9010 | | | |
| Klickitat River | WA-30-1010 | | | |
| Eastern Washington | | | | |
| Kettle River | WA-60-1010 | | | |
| Sullivan Lake | WA-62-9190 | | | |
| South Twin Lake | WA-58-9040 | | | |
| Omak Lake | WA-49-9250 | | | |
| Badger Lake | WA-34-9060 | | | |
| Pataha Creek | WA-35-2013 | | | |

Appendix B. 2008 303(d) Category 5 Listings for PCBs in Freshwater Edible Fish Tissue.

| Listing Detail | WRIA | Waterbody Name | Map Link |
|-------------------|------|--|--------------|
| <u>7022</u> | 13 | Ward Lake | <u>7022</u> |
| 7023 | 18 | Elwha River | 7023 |
| <u>7350</u> | 37 | Yakima River | <u>7350</u> |
| <u>8201</u> | 57 | Spokane River | <u>8201</u> |
| <u>8202</u> | 57 | Spokane River | <u>8202</u> |
| <u>8207</u> | 57 | Spokane River | <u>8207</u> |
| <u>8741</u> | 23 | Chehalis River | <u>8741</u> |
| <u>8746</u> | 24 | Columbia River | <u>8746</u> |
| <u>8755</u> | 33 | Snake River | <u>8755</u> |
| <u>8765</u> | 25 | Columbia River | <u>8765</u> |
| <u>8772</u> | 25 | Columbia River | <u>8772</u> |
| <u>8773</u> | 25 | Columbia River | <u>8773</u> |
| <u>8783</u> | 27 | Columbia River | <u>8783</u> |
| <u>8863</u> | 37 | Yakima River | <u>8863</u> |
| <u>8864</u> | 37 | Yakima River | <u>8864</u> |
| <u>8953</u> | 41 | Lower Crab Creek | <u>8953</u> |
| <u>9015</u> | 54 | Spokane Lake | <u>9015</u> |
| <u>9021</u> | 54 | Spokane Lake | <u>9021</u> |
| <u>9027</u> | 54 | Spokane River | <u>9027</u> |
| <u>9033</u> | 54 | Spokane River | <u>9033</u> |
| <u>9051</u> | 55 | Little Spokane River | <u>9051</u> |
| <u>14025</u> | 1 | Whatcom Lake | <u>14025</u> |
| <u>14036</u> | 3 | Skagit River | <u>14036</u> |
| <u>14090</u> | 9 | Duwamish River | <u>14090</u> |
| <u>14261</u> | 37 | Yakima River | <u>14261</u> |
| <u>14299</u> | 45 | Wenatchee River | <u>14299</u> |
| <u>14385</u> | 54 | Spokane River | <u>14385</u> |
| <u>14397</u> | 57 | Spokane River | <u>14397</u> |
| <u>14398</u> | 57 | Spokane River | <u>14398</u> |
| <u>14400</u> | 54 | Spokane River | <u>14400</u> |
| <u>14402</u> | 57 | Spokane River | <u>14402</u> |
| <u>14407</u> | 61 | Columbia River (Franklin D. Roosevelt Lake) | <u>14407</u> |
| <u>17164</u> | 26 | Cowlitz River | <u>17164</u> |
| 17299 | 1 | Padden Lake | <u>17299</u> |

Table B-1. Category 5 303(d) Listings for Total PCBs in Edible Fish Tissue.

| Listing Detail | WRIA | Waterbody Name | Map Link |
|-------------------|------|--------------------------------|--------------|
| <u>17366</u> | 3 | Samish (West Arm) Lake | <u>17366</u> |
| <u>17383</u> | 8 | Green Lake | <u>17383</u> |
| 17431 | 13 | McIntosh Lake | <u>17431</u> |
| <u>17484</u> | 57 | Liberty Lake | <u>17484</u> |
| 18801 | 31 | Columbia River (Lake Umatilla) | <u>18801</u> |
| 18802 | 31 | Columbia River (Lake Umatilla) | <u>18802</u> |
| <u>19120</u> | 35 | Snake River | <u>19120</u> |
| <u>19121</u> | 35 | Snake River | <u>19121</u> |
| <u>19390</u> | 36 | Columbia River | <u>19390</u> |
| <u>19391</u> | 36 | Columbia River | <u>19391</u> |
| <u>19393</u> | 36 | Columbia River | <u>19393</u> |
| <u>20045</u> | 37 | Yakima River | <u>20045</u> |
| <u>20047</u> | 37 | Yakima River | <u>20047</u> |
| <u>20219</u> | 39 | Yakima River | <u>20219</u> |
| <u>20306</u> | 45 | Icicle Creek | <u>20306</u> |
| <u>36440</u> | 54 | Spokane Lake | <u>36440</u> |
| <u>36441</u> | 54 | Spokane Lake | <u>36441</u> |
| <u>42169</u> | 12 | American Lake | <u>42169</u> |
| <u>42170</u> | 15 | Kitsap Lake | <u>42170</u> |
| <u>42171</u> | 41 | Moses Lake | <u>42171</u> |
| <u>42172</u> | 28 | Vancouver Lake | <u>42172</u> |
| <u>42173</u> | 43 | Medical Lake, West | <u>42173</u> |
| <u>42269</u> | 42 | Banks Lake | <u>42269</u> |
| 42386 | 34 | Sprague Lake | <u>42386</u> |
| 42412 | 34 | Silver Lake | <u>42412</u> |
| <u>43146</u> | 39 | Keechelus Lake | <u>43146</u> |
| 43215 | 53 | Buffalo Lake | <u>43215</u> |
| 43251 | 7 | Calligan Lake | <u>43251</u> |
| 43383 | 62 | Pend Oreille River | <u>43383</u> |
| <u>43385</u> | 36 | Scooteney Reservoir | <u>43385</u> |
| 43465 | 28 | Lacamas Lake | <u>43465</u> |
| <u>43482</u> | 8 | Washington Lake | <u>43482</u> |
| <u>52646</u> | 8 | Ballinger Lake | <u>52646</u> |
| <u>52647</u> | 62 | Bead Lake | <u>52647</u> |
| <u>52648</u> | 23 | Black Lake | <u>52648</u> |
| <u>52651</u> | 22 | Chehalis River | <u>52651</u> |
| <u>52653</u> | 25 | Columbia River | <u>52653</u> |
| <u>52654</u> | 44 | Columbia River | <u>52654</u> |
| <u>52655</u> | 45 | Columbia River | <u>52655</u> |

| Listing Detail | WRIA | Waterbody Name | Map Link |
|-------------------|------|-------------------------------------|--------------|
| <u>52656</u> | 47 | Columbia River (Lake Entiat) | <u>52656</u> |
| <u>52657</u> | 47 | Columbia River (Lake Entiat) | <u>52657</u> |
| <u>52658</u> | 41 | Columbia River (Priest Rapids Lake) | <u>52658</u> |
| <u>52659</u> | 26 | Cowlitz River | <u>52659</u> |
| <u>52662</u> | 15 | Haven Lake | <u>52662</u> |
| <u>52663</u> | 17 | Leland Lake | <u>52663</u> |
| <u>52666</u> | 13 | Long Lake | <u>52666</u> |
| <u>52667</u> | 59 | Loon Lake | <u>52667</u> |
| <u>52668</u> | 14 | Mason Lake | <u>52668</u> |
| <u>52669</u> | 26 | Mayfield Lake | <u>52669</u> |
| <u>52670</u> | 9 | Meridian Lake | <u>52670</u> |
| <u>52671</u> | 27 | Merwin Lake | <u>52671</u> |
| <u>52673</u> | 2 | Mountain Lake | <u>52673</u> |
| <u>52674</u> | 57 | Newman Lake | <u>52674</u> |
| <u>52675</u> | 29 | Northwestern Lake | <u>52675</u> |
| 52676 | 13 | Offutt Lake | <u>52676</u> |
| 52683 | 62 | Pend Oreille River | 52683 |
| 52684 | 41 | Potholes Reservoir | 52684 |
| 52685 | 21 | Queets River | 52685 |
| 52686 | 21 | Quinault River | 52686 |
| 52689 | 25 | Sacajawea Lake | 52689 |
| <u>52690</u> | 8 | Sammamish Lake | <u>52690</u> |
| <u>52691</u> | 9 | Sawyer Lake | <u>52691</u> |
| <u>52692</u> | 26 | Silver Lake | <u>52692</u> |
| <u>52693</u> | 8 | Silver Lake | <u>52693</u> |
| <u>52695</u> | 33 | Snake River | <u>52695</u> |
| <u>52696</u> | 35 | Snake River | <u>52696</u> |
| <u>52697</u> | 35 | Snake River | <u>52697</u> |
| <u>52698</u> | 33 | Snake River (Sacajawea Lake) | <u>52698</u> |
| <u>52699</u> | 7 | Snohomish River | <u>52699</u> |
| <u>52701</u> | 14 | Summit Lake | <u>52701</u> |
| <u>52703</u> | 8 | Washington Lake | <u>52703</u> |
| <u>52704</u> | 8 | Washington Lake | <u>52704</u> |
| <u>52705</u> | 8 | Washington Lake | <u>52705</u> |
| <u>52707</u> | 45 | Wenatchee River | <u>52707</u> |
| <u>52708</u> | 1 | Whatcom Lake | <u>52708</u> |
| <u>52833</u> | 38 | Cowiche Creek | <u>52833</u> |
| <u>52935</u> | 62 | Pend Oreille River | <u>52935</u> |
| <u>52947</u> | 45 | Wenatchee River | <u>52947</u> |
| 53209 | 28 | Lake River | 53209 |

Appendix C. 2008 303(d) Category 5 Listings for Dioxin (2,3,7,8-TCDD) in Freshwater Edible Fish Tissue.

| Listing Detail | WRIA | Waterbody Name | Map Link |
|----------------|------|------------------------------|--------------|
| <u>42381</u> | 43 | Medical Lake, West | <u>42381</u> |
| <u>42405</u> | 34 | Sprague Lake | <u>42405</u> |
| <u>42410</u> | 54 | Spokane Lake | <u>42410</u> |
| <u>42411</u> | 54 | Spokane River | <u>42411</u> |
| <u>42431</u> | 34 | Silver Lake | <u>42431</u> |
| <u>42434</u> | 41 | Moses Lake | <u>42434</u> |
| <u>42440</u> | 42 | Banks Lake | <u>42440</u> |
| <u>42443</u> | 12 | American Lake | <u>42443</u> |
| <u>51543</u> | 8 | Ballinger Lake | <u>51543</u> |
| <u>51544</u> | 62 | Bead Lake | <u>51544</u> |
| <u>51549</u> | 25 | Columbia River | <u>51549</u> |
| <u>51552</u> | 26 | Cowlitz River | <u>51552</u> |
| <u>51554</u> | 17 | Leland Lake | <u>51554</u> |
| <u>51556</u> | 13 | Long Lake | <u>51556</u> |
| <u>51560</u> | 9 | Meridian Lake | <u>51560</u> |
| <u>51562</u> | 48 | Methow River | <u>51562</u> |
| <u>51572</u> | 41 | Potholes Reservoir | <u>51572</u> |
| <u>51577</u> | 9 | Sawyer Lake | <u>51577</u> |
| <u>51578</u> | 26 | Silver Lake | <u>51578</u> |
| <u>51581</u> | 33 | Snake River | <u>51581</u> |
| <u>51582</u> | 35 | Snake River | <u>51582</u> |
| <u>51583</u> | 33 | Snake River (Sacajawea Lake) | <u>51583</u> |
| <u>51584</u> | 7 | Snohomish River | <u>51584</u> |
| <u>51586</u> | 54 | Spokane River | <u>51586</u> |
| <u>51587</u> | 57 | Spokane River | <u>51587</u> |
| <u>51588</u> | 41 | Stan Coffin Lake | <u>51588</u> |
| <u>51591</u> | 8 | Washington Lake | <u>51591</u> |
| <u>51592</u> | 8 | Washington Lake | <u>51592</u> |
| <u>51593</u> | 8 | Washington Lake | <u>51593</u> |
| <u>51595</u> | 1 | Whatcom Lake | <u>51595</u> |
| <u>53204</u> | 28 | Vancouver Lake | <u>53204</u> |
| <u>53206</u> | 28 | Lake River | <u>53206</u> |
| <u>34887</u> | 37 | Yakima River | <u>34887</u> |
| <u>34889</u> | 39 | Yakima River | <u>34889</u> |
| <u>34904</u> | 36 | Columbia River | <u>34904</u> |

| $T_{-1} = 0 = 1$ | Cata a mar 5 202 | (1) | T: | 227 | | D : | E 131.1 . E | т: |
|------------------|------------------|-----|----------------|-------|---------|---------|---------------|---------|
| Table C-1. | Category 5 303 | (a) |) Listings for | 2,3,1 | /,8-ICD | $D \ln$ | i Edible Fish | Tissue. |

| Listing Detail | WRIA | Waterbody Name | Map Link | |
|----------------|------|-------------------|--------------|--|
| <u>34905</u> | 37 | Yakima River | <u>34905</u> | |
| <u>34913</u> | 37 | Yakima River | <u>34913</u> | |
| <u>34926</u> | 36 | Columbia River | <u>34926</u> | |
| <u>34927</u> | 35 | Snake River | <u>34927</u> | |
| <u>34942</u> | 35 | Snake River | <u>34942</u> | |
| <u>36354</u> | 23 | Chehalis River | <u>36354</u> | |
| <u>36355</u> | 23 | Dillenbaugh Creek | <u>36355</u> | |
| <u>43061</u> | 47 | Chelan Lake | <u>43061</u> | |
| <u>43094</u> | 7 | Dorothy Lake | <u>43094</u> | |
| <u>43128</u> | 39 | Keechelus Lake | <u>43128</u> | |
| <u>43162</u> | 6 | Lone Lake | <u>43162</u> | |
| <u>43197</u> | 53 | Buffalo Lake | <u>43197</u> | |
| <u>43233</u> | 7 | Calligan Lake | <u>43233</u> | |

Appendix D. 2008 Category 2 Listings for TCDD TEQs in Freshwater Edible Fish Tissue.

| Listing Detail | WRIA | Waterbody Name | Map Link |
|-------------------|------|-------------------------------------|--------------|
| 42437 | 15 | Kitsap Lake | 42437 |
| 51596 | 8 | Ballinger Lake | 51596 |
| 51597 | 62 | Bead Lake | 51597 |
| 51598 | 22 | Chehalis River | 51598 |
| 51599 | 22 | Chehalis River | 51599 |
| <u>51600</u> | 25 | Columbia River | <u>51600</u> |
| <u>51601</u> | 44 | Columbia River | <u>51601</u> |
| <u>51602</u> | 45 | Columbia River | <u>51602</u> |
| <u>51603</u> | 47 | Columbia River (Lake Entiat) | <u>51603</u> |
| <u>51604</u> | 41 | Columbia River (Priest Rapids Lake) | <u>51604</u> |
| <u>51605</u> | 26 | Cowlitz River | <u>51605</u> |
| <u>51606</u> | 15 | Haven Lake | <u>51606</u> |
| <u>51607</u> | 17 | Leland Lake | <u>51607</u> |
| <u>51609</u> | 13 | Long Lake | <u>51609</u> |
| <u>51610</u> | 59 | Loon Lake | <u>51610</u> |
| <u>51613</u> | 9 | Meridian Lake | <u>51613</u> |
| <u>51614</u> | 27 | Merwin Lake | <u>51614</u> |
| <u>51615</u> | 48 | Methow River | <u>51615</u> |
| <u>51616</u> | 2 | Mountain Lake | <u>51616</u> |
| <u>51618</u> | 29 | Northwestern Lake | <u>51618</u> |
| <u>51620</u> | 20 | Ozette Lake | <u>51620</u> |
| <u>51622</u> | 34 | Palouse River | <u>51622</u> |
| <u>51623</u> | 34 | Palouse River | <u>51623</u> |
| <u>51624</u> | 34 | Palouse River, S.F. | <u>51624</u> |
| <u>51625</u> | 41 | Potholes Reservoir | <u>51625</u> |
| <u>51626</u> | 21 | Queets River | <u>51626</u> |
| <u>51627</u> | 21 | Quinault River | <u>51627</u> |
| <u>51629</u> | 8 | Sammamish Lake | <u>51629</u> |
| <u>51630</u> | 9 | Sawyer Lake | <u>51630</u> |
| <u>51631</u> | 26 | Silver Lake | <u>51631</u> |
| <u>51632</u> | 3 | Skagit River | <u>51632</u> |
| 51633 | 33 | Snake River | <u>51633</u> |
| <u>51634</u> | 35 | Snake River | <u>51634</u> |
| <u>51635</u> | 35 | Snake River | <u>51635</u> |
| 51636 | 33 | Snake River (Sacajawea Lake) | <u>51636</u> |

Table D-1. Category 2 Listings for TCDD TEQs in Edible Fish Tissue.

| Listing Detail | WRIA | Waterbody Name | Map Link |
|-------------------|------|------------------|--------------|
| <u>51637</u> | 7 | Snohomish River | <u>51637</u> |
| <u>51638</u> | 54 | Spokane River | <u>51638</u> |
| <u>51639</u> | 57 | Spokane River | <u>51639</u> |
| <u>51640</u> | 57 | Spokane River | <u>51640</u> |
| <u>51641</u> | 41 | Stan Coffin Lake | <u>51641</u> |
| <u>51642</u> | 14 | Summit Lake | <u>51642</u> |
| <u>51643</u> | 13 | Ward Lake | <u>51643</u> |
| <u>51644</u> | 8 | Washington Lake | <u>51644</u> |
| <u>51645</u> | 8 | Washington Lake | <u>51645</u> |
| <u>51646</u> | 8 | Washington Lake | <u>51646</u> |
| <u>51647</u> | 45 | Wenatchee River | <u>51647</u> |
| <u>51648</u> | 1 | Whatcom Lake | <u>51648</u> |

Appendix E. Latitude and Longitude of Waterbodies Sampled for the 2007-08 PCB and Dioxin Background Study.

| Table E-1. | Latitude and Longitude of Waterbodies Sampled for the 2007-08 PCB and Dioxin |
|------------|--|
| Backgroun | d Study. |

| Region and Waterbody Name | Location | Latitude | Longitude |
|------------------------------|-----------------------------------|----------|-----------|
| Western Washington | | | |
| Mountain Lake | Moran State Park, Orcas Island | 48.66 | 122.82 |
| Ozette Lake | Olympic NP | 48.10 | 124.64 |
| Quinault River | Quinault Reservation/Olympic NP | 47.54 | 123.74 |
| Cushman Lake | Olympic NF | 47.47 | 123.25 |
| North River | Chehalis River drainage | 46.83 | 123.59 |
| Loomis Lake | Loomis Lake SP, Long Beach | 46.44 | 124.04 |
| West Slope of the Cascad | les | | |
| Baker Lake | N. Cascade NP/Baker-Snoqualmie NF | 48.72 | 121.66 |
| Sauk River | Baker-Snoqualmie NF | 48.10 | 121.39 |
| Spada Lake | Baker-Snoqualmie NF | 47.97 | 121.65 |
| Chester Morse Lake | Baker-Snoqualmie NF | 47.39 | 121.70 |
| Panamaker Creek | Gifford Pinchot NF | 46.06 | 122.29 |
| Merrill Lake | Gifford Pinchot NF | 46.09 | 122.33 |
| East Slope of the Cascad | les | · | |
| Patterson Lake | Okanogan NF | 48.46 | 120.24 |
| Entiat River | Wenatchee NF | 48.00 | 120.61 |
| Fish Lake | Wenatchee NF | 47.83 | 120.70 |
| Cle Elum Lake | Wenatchee NF | 47.29 | 121.11 |
| Bumping Lake | Wenatchee NF | 46.85 | 121.32 |
| Klickitat River | Klickitat County | 45.71 | 121.26 |
| Eastern Washington | | | |
| Kettle River | Colville NF | 48.89 | 118.60 |
| Sullivan Lake | Colville NF | 48.80 | 117.29 |
| South Twin Lake | Colville Indian Reservation | 48.27 | 118.39 |
| Omak Lake | Colville Indian Reservation | 48.28 | 119.40 |
| Badger Lake | South of Turnbull NWR | 47.35 | 117.63 |
| Pataha Creek | Umatilla NF | 46.26 | 117.53 |

NAD 83.

NP – National Park; NF – National Forest; NWR – National Wildlife Refuge.

Appendix F. Results of Duplicate Samples

| Chemical/ Sample No. | Field ID | Subsample #1 | Subsample #2 | RPD |
|-------------------------|-------------------|-----------------|-----------------|-----|
| Total PCBs: C | ongeners (ug/Kg) | | | |
| 081201107 | FISHNPM | 38.2 | 38.5 | 1% |
| 081201118 | OMAKCTT | 3.84 | 3.86 | 1% |
| 081201121 | KLICKMWF | 6.03 | 5.06 | 17% |
| 081201262 | PANAMKOK | 3.01 | 3.17 | 5% |
| 07494503 | OZETTNPM | 1.38 | 1.79 | 26% |
| 07494537 | SPADACTT | 2.38 | 1.20 | 66% |
| Total PCBs: A | roclor-equivalent | s (ug/Kg) | | |
| 0812011-13 | SNOQMWF* | 14.4 J | 14.2 J | 1% |
| 0812011-27 | STEVEKOK* | 9.8 J | 9.6 J | 2% |
| 07494508 | SAMLNPM* | 13.8 J | 15.3 J | 10% |
| 07494519 | SULBNT* | 28.4 J | 32.2 J | 13% |
| TCDD (ng/Kg) | | | | |
| 081201107 | FISHNPM | 0.03 U | 0.03 U | |
| 081201118 | OMAKCTT | 0.03 U | 0.03 U | |
| 081201121 | KLICKMWF | 0.041 | 0.052 | 24% |
| 081201262 | PANAMKOK | 0.03 NJ | 0.03 UJ | |
| 07494503 | OZETNPML | .033 UJ | 0.039 UJ | |
| 07494537 | SPADCTTL | 0.045 UJ | 0.024 UJ | |
| TCDD TEQs (1 | ng/Kg) | | | |
| 081201107 | FISHNPM | 0.0250 J | 0.50 U | |
| 081201118 | OMAKCTT | 0.0130 J | 0.0167 J | 25% |
| 081201121 | KLICKMWF | 0.187 | 0.156 | 18% |
| 081201262 | PANAMKOK | 0.448 J | 0.422 | 6% |
| 07494503 | OZETNPML | 0.010 J | 0.018 J | 58% |
| 07494537 | SPADCTTL | 0.145 J | 0.067 J | 73% |
| Percent Lipids | | | | |
| 081201107 | FISHNPM | 3.43 | 2.91 | 16% |
| 081201118 | OMAKCTT | 4.66 | 4.58 | 2% |
| 081201262 | PANAMKOK | 0.28 | 0.28 | 0% |

Table F-1. Results of Duplicate (Split) Sample Analyses.

U = Not detected at or above reported result.

J = Result is an estimated value.

NJ = There is evidence the analyte is present. The associated numerical result is an estimate.

*WSTMP samples analyzed in conjunction with samples from present 2007-08 study.
Appendix G. Chemical and Biological Data for the 2007-08 PCB and Dioxin Background Study.

Table G-1. Chemical Data for the PCB and Dioxin Background Study.

(See Ecology's Environmental Information Management system for individual congener data, <u>www.ecy.wa.gov/eim/</u>.)

| Site | Species Code | Sample Number | Date Collected | Total PCB Congeners | (ug/Kg) | Total PCB Aroclors | (ug/Kg) | 2378 TCDD (ng/Kg) | | 2378 TCDD TEQ | (ng/Kg) | Lipids (%) |
|-----------------------|-----------------|------------------|-------------------|------------------------|---------|-----------------------|---------|----------------------|----|------------------|---------|---------------|
| Badger Lake | LMB | 081201123 | 11/18/08 | 1.38 | | 1.3 | | 0.03 | U | 0.050 | U | 0.46 |
| Badger Lake | RBT | 081201122 | 11/18/08 | 1.01 | | 1.9 | U | 0.03 | U | 0.050 | U | 0.63 |
| Baker Lake | CTT | 07494512 | 10/9/07 | 1.16 | J | 0.98 | U | 0.031 | UJ | 0.092 | J | 1.26 |
| Baker Lake | LSS | 07494513 | 10/9/07 | 0.51 | J | na | | 0.011 | NJ | 0.035 | J | 0.21 |
| Baker Lake | RBT | 07494511 | 10/9/07 | 0.91 | J | 0.98 | U | 0.063 | NJ | 0.35 | J | 1.48 |
| Bumping Lake | BKT | 081201117 | 10/16/08 | 7.27 | | na | | 0.03 | U | 0.025 | J | 0.75 |
| Bumping Lake | CTT | 081201115 | 10/16/08 | 2.61 | | 1.9 | J | 0.03 | U | 0.014 | | 3.15 |
| Chester Morse Lake | RBT | 07494545 | 11/5/07 | 1.36 | J | 1.2 | J | 0.024 | UJ | 0.070 | J | 1.12 |
| Cle Elum Lake | LSS | 081201260 | 7/17/08 | 6.55 | | na | | 0.03 | U | 0.15 | J | 4.42 |
| Cle Elum Lake | MWF | 081201102 | 10/30/08 | 5.14 | | 6.0 | J | 0.03 | U | 0.021 | J | 2.32 |
| Cle Elum Lake | MWF | 081201250 | 10/30/08 | 1.65 | | na | | 0.03 | U | 0.095 | J | 3.19 |
| Panamaker Cr. | KOK | 081201262 | 11/24/08 | 3.01 | | na | | 0.03 | NJ | 0.45 | J | 0.28 |
| Cushman Lake | KOK | 07494515 | 10/4/07 | 1.14 | J | 2.0 | U | 0.016 | NJ | 0.14 | J | 1.21 |
| Cushman Lake | LSS | 07494520 | 9/27/07 | 0.60 | J | na | | 0.0075 | UJ | 0.051 | J | 0.27 |
| Cushman Lake | MWF | 07494516 | 9/27/07 | 1.30 | J | 1.6 | J | 0.014 | NJ | 0.056 | J | 2.31 |
| Entiat River | BKT | 081201259 | 7/28/08 | 0.29 | | na | | 0.03 | U | 0.049 | J | na |
| Entiat River | RBT | 081201255 | 7/28/08 | 0.53 | | na | | 0.059 | NJ | 0.090 | J | 2.49 |
| Fish Lake | BNT | 081201109 | 7/16/08 | 87.7 | | 98 | J | 0.03 | U | 0.062 | J | 6.53 |
| Fish Lake | LMB | 081201106 | 7/16/08 | 24.9 | | 24 | J | 0.03 | U | 0.048 | | 1.94 |
| Fish Lake | LSS | 081201258 | 7/16/08 | 3.54 | | na | | 0.03 | U | 0.13 | J | 2.87 |
| Fish Lake | LSS | 081201257 | 7/16/08 | 2.38 | | na | | 0.03 | U | 0.050 | | 2.25 |
| Fish Lake | NPM | 081201107 | 7/16/08 | 38.2 | | 42 | J | 0.03 | U | 0.025 | J | 3.43 |
| Fish Lake | RBT | 081201108 | 7/16/08 | 5.41 | | 2.1 | J | 0.03 | U | 0.026 | | 0.18 |
| Kettle River | LSS | 07494525 | 10/10/07 | 0.24 | UJ | na | | 0.04 | UJ | 0.008 | J | 0.38 |
| Kettle River | RBT | 07494526 | 10/10/07 | 0.31 | J | 0.99 | U | 0.028 | UJ | 0.16 | J | 1.23 |
| Klickitat River | MWF | 081201121 | 12/8/08 | 6.03 | | 8.8 | J | 0.041 | | 0.19 | | 5.89 |

| Site | Species Code | Sample Number | Date Collected | Total PCB Congeners | (ug/Kg) | Total PCB Aroclors | (ug/Kg) | 2378 TCDD (ng/Kg) | | 2378 TCDD TEQ | (ng/Kg) | Lipids (%) |
|----------------|-----------------|------------------|-------------------|------------------------|---------|-----------------------|---------|----------------------|----|------------------|---------|---------------|
| Loomis Lake | LMB | 081201254 | 11/5/08 | 0.35 | | na | | 0.03 | U | 0.042 | | 0.6 |
| Loomis Lake | YP | 081201252 | 11/5/08 | 0.044 | Ν | na | | 0.03 | U | 0.033 | | 0.26 |
| Loomis Lake | YP | 081201253 | 11/5/08 | 0.044 | NJ | na | | 0.031 | NJ | 0.091 | J | 0.39 |
| Merrill Lake | CTT | 081201133 | 6/24/08 | 4.44 | | 1.3 | J | 0.03 | U | 0.050 | U | 1.58 |
| Mountain Lake | KOK | 081201261 | 11/12/08 | 11.4 | | na | | 0.12 | | 0.65 | | 0.49 |
| North River | CTT | 081201132 | 8/29/08 | 1.04 | | 2.2 | UJ | 0.032 | NJ | 0.032 | J | 1.83 |
| Omak Lake | CTT | 081201118 | 7/1/08 | 3.84 | | 2.2 | UJ | 0.03 | U | 0.013 | J | 4.66 |
| Omak Lake | PEA | 081201120 | 7/1/08 | 3.02 | | 1.3 | J | 0.03 | U | 0.077 | J | 0.45 |
| Ozette Lake | LMB | 07494506 | 9/12/07 | 0.26 | J | na | | 0.019 | UJ | 0.028 | J | 0.49 |
| Ozette Lake | NPM | 07494503 | 9/12/07 | 1.38 | J | 1.7 | J | 0.033 | UJ | 0.010 | J | 0.81 |
| Ozette Lake | NPM | 07494504 | 9/12/07 | 1.28 | J | na | | 0.026 | UJ | 0.012 | J | 0.7 |
| Ozette Lake | YP | 07494505 | 9/12/07 | 0.25 | J | na | | 0.014 | UJ | 0.016 | J | 0.15 |
| Pataha Creek | BKT | 081201114 | 10/15/08 | 0.71 | | 2.2 | UJ | 0.03 | U | 0.056 | J | 0.59 |
| Patterson Lake | RBT | 081201263 | 12/23/08 | 3.43 | | na | | 0.03 | U | 0.031 | | 0.33 |
| Patterson Lake | YP | 081201251 | 12/23/08 | 0.23 | | na | | 0.03 | U | 0.103 | J | 0.98 |
| Quinault River | CTT | 081201101 | 8/28/08 | 0.45 | | 1.1 | U | 0.03 | U | 0.071 | J | 0.41 |
| Sauk River | MWF | 07494514 | 10/4/07 | 1.97 | J | 3.7 | J | 0.06 | NJ | 0.18 | J | 3.81 |
| South Twin L. | BKT | 081201103 | 6/25/08 | 1.17 | | 1.1 | U | 0.03 | U | 0.075 | | 2.24 |
| South Twin L. | BKT | 081201256 | 6/25/08 | 0.83 | | na | | 0.03 | U | 0.035 | J | 1.55 |
| South Twin L. | LMB | 081201104 | 6/25/08 | 0.25 | NJ | 1.7 | UJ | 0.03 | U | 0.007 | | 0.64 |
| South Twin L. | RBT | 081201105 | 6/25/08 | na | | 1.1 | U | na | | na | | 1.25 |
| Spada Lake | CTT | 07494537 | 10/17/07 | 2.38 | J | 4.1 | J | 0.045 | UJ | 0.15 | J | 0.39 |
| Spada Lake | CTT | 07494538 | 10/17/07 | 1.21 | J | na | | 0.027 | NJ | 0.12 | J | 1.04 |
| Spada Lake | RBT | 07494539 | 10/17/07 | 1.70 | J | na | | 0.039 | UJ | 0.017 | J | 0.76 |
| Sullivan Lake | CTT | 07494521 | 9/17/07 | 3.53 | J | 6.5 | J | 0.017 | NJ | 0.12 | J | 3.57 |
| Sullivan Lake | MWF | 07494522 | 9/17/07 | 3.37 | J | 4.9 | J | 0.041 | NJ | 0.19 | J | 2.91 |
| Sullivan Lake | MWF | 07494523 | 9/17/07 | 1.99 | J | na | | 0.039 | UJ | 0.011 | J | 1.62 |

| Table G-2. | Biological Data on | Fish Samples | Collected fo | or the 2007-08 | PCB and D | ioxin |
|------------|--------------------|--------------|--------------|----------------|-----------|-------|
| Background | d Study. | - | | | | |

| Site | Species Code | Mean Total Length (mm) | Mean Weight (g) | Mean Age (years) | # Fish in Composite |
|--------------------|-----------------|---------------------------|--------------------|---------------------|------------------------|
| Badger Lake | LMB | 244 | 189 | 3.0 | 5 |
| Badger Lake | RBT | 300 | 279 | 1.0 | 5 |
| Baker Lake | CTT | 235 | 121 | 2.2 | 5 |
| Baker Lake | LSS | 366 | 411 | na | 5 |
| Baker Lake | RBT | 215 | 103 | 1.0 | 5 |
| Bumping Lake | BKT | 237 | 110 | 2.7 | 3 |
| Bumping Lake | CTT | 256 | 151 | 1.0 | 5 |
| Chester Morse Lake | RBT | 412 | 603 | 3.6 | 5 |
| Cle Elum Lake | LSS | 495 | 1355 | 15.2 | 5 |
| Cle Elum Lake | MWF | 353 | 428 | 5.4 | 5 |
| Cle Elum Lake | MWF | 258 | 149 | 2.0 | 5 |
| Panamaker Creek | KOK | 352 | 353 | 2.0 | 4 |
| Cushman Lake | KOK | 254 | 146 | 3.0 | 5 |
| Cushman Lake | LSS | 295 | 262 | na | 5 |
| Cushman Lake | MWF | 276 | 181 | 6.0 | 5 |
| Entiat River | BKT | 143 | 75 | 3.5 | 2 |
| Entiat River | RBT | 200 | 85 | 4.0 | 5 |
| Fish Lake | BNT | 487 | 1430 | 4.4 | 5 |
| Fish Lake | LMB | 425 | 1371 | 7.4 | 5 |
| Fish Lake | LSS | 409 | 804 | 7.8 | 5 |
| Fish Lake | LSS | 370 | 571 | 5.8 | 5 |
| Fish Lake | NPM | 404 | 620 | 6.8 | 5 |
| Fish Lake | RBT | 313 | 241 | 1.0 | 5 |
| Kettle River | LSS | 262 | 165 | na | 5 |
| Kettle River | RBT | 283 | 193 | 2.0 | 5 |
| Klickitat River | MWF | 413 | 724 | 5.0 | 5 |
| Loomis Lake | LMB | 198 | 110 | 1.0 | 4 |
| Loomis Lake | YP | 221 | 155 | 1.8 | 4 |
| Loomis Lake | YP | 170 | 68 | 1.0 | 4 |
| Merrill Lake | CTT | 307 | 274 | 3.0 | 5 |
| Mountain Lake | KOK | 270 | 160 | 2.6 | 5 |
| North River | CTT | 219 | 102 | 2.0 | 5 |
| Omak Lake | CTT | 385 | 516 | 2.2 | 5 |
| Omak Lake | PEA | 233 | 116 | 9.0 | 7 |

| Site | Species Code | Mean Total Length (mm) | Mean Weight (g) | Mean Age (years) | # Fish in Composite |
|-----------------|-----------------|---------------------------|--------------------|---------------------|------------------------|
| Ozette Lake | LMB | 377 | 768 | 3.8 | 5 |
| Ozette Lake | NPM | 377 | 433 | 9.0 | 5 |
| Ozette Lake | NPM | 318 | 267 | 6.8 | 5 |
| Ozette Lake | YP | 251 | 199 | 4.0 | 5 |
| Pataha Creek | BKT | 174 | 51 | 3.6 | 19 |
| Patterson Lake | RBT | 263 | 185 | 1.0 | 5 |
| Patterson Lake | YP | 174 | 72 | 2.0 | 5 |
| Quinault River | CTT | 276 | 175 | 3.3 | 4 |
| Sauk River | MWF | 255 | 173 | 2.5 | 4 |
| South Twin Lake | BKT | 297 | 308 | 1.2 | 5 |
| South Twin Lake | BKT | 243 | 156 | 1.0 | 3 |
| South Twin Lake | LMB | 230 | 193 | 4.0 | 5 |
| South Twin Lake | RBT | 302 | 321 | 1.0 | 5 |
| Spada Lake | CTT | 291 | 173 | 3.4 | 5 |
| Spada Lake | CTT | 229 | 102 | 2.4 | 5 |
| Spada Lake | RBT | 241 | 113 | 2.0 | 4 |
| Sullivan Lake | CTT | 300 | 267 | 2.4 | 5 |
| Sullivan Lake | MWF | 265 | 164 | 2.6 | 5 |
| Sullivan Lake | MWF | 247 | 125 | 2.2 | 5 |

Species Codes: BKT = Brook trout, BNT = Brown trout, CTT = Cutthroat trout, KOK = Kokanee salmon, LMB = Largemouth bass, LSS = Largescale sucker, MWF = Mountain whitefish, NPM = Northern pikeminnow, PEA = Peamouth, RBT = Rainbow trout, YP = Yellow Perch

Appendix H. Toxicity Equivalency Factors for PCBs, Dioxins, and Furans.

Table H-1. Human and Mammalian Toxic Equivalency Factors (TEFs) for PCBs, Dioxins, and Furans.

| Compound | WHO (2005) TEF* | | | | | | |
|-------------------------------|--------------------|--|--|--|--|--|--|
| Non-ortho substituted PCBs | | | | | | | |
| PCB-77 | 0.0001 | | | | | | |
| PCB-81 | 0.0003 | | | | | | |
| PCB-126 | 0.1 | | | | | | |
| PCB-169 | 0.03 | | | | | | |
| Mono-ortho substituted | d PCBs | | | | | | |
| PCB-105 | 0.00003 | | | | | | |
| PCB-114 | 0.00003 | | | | | | |
| PCB-118 | 0.00003 | | | | | | |
| PCB-123 | 0.00003 | | | | | | |
| PCB-156 | 0.00003 | | | | | | |
| PCB-157 | 0.00003 | | | | | | |
| PCB-167 | 0.00003 | | | | | | |
| PCB-189 | 0.00003 | | | | | | |
| Chlorinated Dibenzo-p-dioxins | | | | | | | |
| 2,3,7,8-TCDD | 1 | | | | | | |
| 1,2,3,7,8-PeCDD | 1 | | | | | | |
| 1,2,3,4,7,8-HxCDD | 0.1 | | | | | | |
| 1,2,3,6,7,8-HxCDD | 0.1 | | | | | | |
| 1,2,3,7,8,9-HxCDD | 0.1 | | | | | | |
| 1,2,3,4,6,7,8-HpCDD | 0.01 | | | | | | |
| OCDD | 0.0003 | | | | | | |
| Chlorinated Dibenzofu | rans | | | | | | |
| 2,3,7,8-TCDF | 0.1 | | | | | | |
| 1,2,3,7,8-PeCDF | 0.03 | | | | | | |
| 2,3,4,7,8-PeCDF | 0.3 | | | | | | |
| 1,2,3,4,6,7,8-HpCDF | 0.1 | | | | | | |
| 1,2,3,4,7,8,9-HpCDF | 0.1 | | | | | | |
| 1,2,3,4,7,8-HxCDF | 0.1 | | | | | | |
| 1,2,3,6,7,8-HxCDF | 0.1 | | | | | | |
| 1,2,3,7,8,9-HxCDF | 0.01 | | | | | | |
| 2,3,4,6,7,8-HxCDF | 0.01 | | | | | | |
| OCDF | 0.0003 | | | | | | |

Van den Berg et al., 2006.

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Appendix I. Glossary, Acronyms, and Abbreviations

Glossary

303(d) list: Section 303(d) of the federal Clean Water Act requires Washington State to periodically prepare a list of all surface waters in the state for which beneficial uses of the water – such as for drinking, recreation, aquatic habitat, and industrial use – are impaired by pollutants. These are water quality limited estuaries, lakes, and streams that fall short of state surface water quality standards, and are not expected to improve within the next two years.

90th percentile: A statistical number obtained from a distribution of a data set, above which 10% of the data exists and below which 90% of the data exists.

Background: For this study, denotes a waterbody where the only known or likely PCB or dioxin source of significance is the atmosphere.

Clean Water Act: A federal act passed in 1972 that contains provisions to restore and maintain the quality of the nation's waters. Section 303(d) of the Clean Water Act establishes the TMDL program.

Dioxin: Polychlorinated dibenzo-p-dioxins.

Furan: Polychlorinated dibenzofuran.

Exceeded Criteria: Did not meet criteria.

Salmonid: Any fish that belong to the family *Salmonidae*. Basically, any species of salmon, trout, or char. <u>www.fws.gov/le/ImpExp/FactSheetSalmonids.htm</u>.

Spiny Rayed Fish: Fish with fins supported by both spines and soft rays, such as bass and yellow perch, as opposed to salmonids and other soft rayed fish.

Total Maximum Daily Load (TMDL): A distribution of a substance in a waterbody designed to protect it from exceeding water quality standards. A TMDL is equal to the sum of all of the following: (1) individual wasteload allocations for point sources, (2) the load allocations for nonpoint sources, (3) the contribution of natural sources, and (4) a Margin of Safety to allow for uncertainty in the wasteload determination. A reserve for future growth is also generally provided.

Acronyms and Abbreviations

Following are acronyms and abbreviations used frequently in this report.

| BC | British Columbia, Canada |
|------------|--|
| Ecology | Washington State Department of Ecology |
| EA | Environmental Assessment |
| EIM | Environmental Information Management database |
| EPA | U.S. Environmental Protection Agency |
| GIS | Geographic Information System |
| Manchester | Ecology Manchester Environmental Laboratory |
| MTCA | Model Toxics Control Act |
| NTR | National Toxics Rule |
| PBT | persistent, bioaccumulative, and toxic substance |
| PCB | polychlorinated biphenyls |
| PCDD/Fs | polychlorinated dibenzo-p-dioxins and -furans |
| QA/QC | Quality Assurance/Quality Control |
| RPD | Relative percent difference |
| SOP | Standard operating procedures |
| TCDD | tetrachlorodibenzo-p-dioxin |
| TEF | Toxicity Equivalent Factor |
| TEQ | Toxic Equivalent |
| TMDL | Total Maximum Daily Load (water cleanup plan) |
| USGS | U.S. Geological Survey |
| WAC | Washington Administrative Code |
| WDFW | Washington Department of Fish and Wildlife |
| WDOH | Washington Department of Health |
| WRIA | Water Resources Inventory Area |
| WSTMP | Washington State Toxics Monitoring Program |

Units of Measurement

| a unit of mass |
|---|
| kilograms, a unit of mass equal to 1,000 grams. |
| millimeter, one thousandth of a meter |
| nanograms per kilogram (parts per trillion) |
| nanograms per liter (parts per trillion) |
| picograms per liter (parts per quadrillion) |
| micrograms per kilogram (parts per billion) |
| |