# Bellingham Bay Regional Background Sediment Characterization 

Bellingham, WA

Final Data Evaluation and Summary Report

February 27, 2015
Publication no. 15-09-044

## Publication and Contact Information

This report is available on the Department of Ecology's website at www.ecy.wa.gov/biblio/1509044.html

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# Final Data Evaluation and Summary Report 

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## List of Acronyms

AOI
ARI
Axys
CAP
CD
cPAH carcinogenic polycyclic aromatic hydrocarbon
CSL Cleanup Screening Level
CSO combined sewer overflow
DGPS differential global positioning system
DQO data quality objectives
Ecology Washington State Department of Ecology
EMPC estimated maximum potential concentration
EPA U.S. Environmental Protection Agency
KM Kaplan-Meier
MDL method detection limit
MLLW mean lower low water
NAD83 1983 North American Datum
PCB polychlorinated biphenyl
PeCDD pentachlorodibenzodioxin
PQL practical quantitation limit
PSEP Puget Sound Estuary Program
PSRM Puget Sound reference material
QA/QC Quality Assurance/Quality Control
RI/FS remedial investigation/feasibility study
RPD relative percent difference
RSD relative standard deviation
SAP Sampling and Analysis Plan
SCO sediment cleanup objective
SCUM II Sediment Cleanup User's Manual II
SDL sample specific detection limit
SIM select ion monitoring
SMS Sediment Management Standards
SVOC semi-volatile organic chemical
TBT tributyltin
TCDD tetrachlorodibenzodioxin
TEC toxic equivalent concentration
TEF toxic equivalent factor
TEQ toxic equivalent quotient
TOC total organic carbon

UCL upper confidence limit
UTL upper tolerance limit
WAC Washington Administrative Code
WHO World Health Organization

### 1.0 Introduction

In early 2013, Ecology revised its Sediment Management Standards (SMS) to establish a new framework for identification and cleanup of contaminated sediment sites (Washington Administrative Code [WAC] 173-204). A key component of this framework was the concept of regional background sediment concentrations, which could potentially serve as the Cleanup Screening Level (CSL) for sediment sites. During the advisory group process for the rule revisions, it was recommended that Ecology be responsible for establishing regional background sediment concentrations for the state.

Bellingham Bay is the location of a multi-organization, comprehensive cleanup initiative known as the Bellingham Bay Demonstration Pilot. Under this initiative, Ecology is addressing 12 cleanup sites, including 8 sediment sites. Controlling pollution sources is also a component of the initiative. Regional background concentrations are needed to support these cleanup and source control activities.

The current study was prepared in accordance with the SMS and Sediment Cleanup User’s Manual (SCUM II; Ecology 2013a \& 2015). Sediment sampling, analytical procedures, and methods corresponded to the SMS, SCUM II, and WAC 173-340-830.

### 1.1. Regional Background Definition

For a number of bioaccumulative chemicals, risk-based values protective of human health and upper trophic levels fall below natural background concentrations, as defined in the SMS (WAC 173-204-505). Sediments are a sink for chemicals from potentially hundreds of sources, including a mix of permitted and unpermitted stormwater, atmospheric deposition, and historical releases from industrial activities. In urban embayments with developed shorelines, chemical concentrations in sediment are frequently higher than natural background concentrations.

The 2013 SMS rule revisions retained the two-tiered framework used to establish sediment cleanup levels, but now incorporates natural background (as the potential Sediment Cleanup Objective (SCO)) and a new term and concept, regional background as the potential CSL. The SMS rule includes a definition for regional background (WAC 173-204-505(16)) and parameters for establishing regional background (WAC 173-204-560(5)):
"Regional Background" means the concentration of a contaminant within a department defined geographic area that is primarily attributable to diffuse sources, such as atmospheric deposition or storm water, not attributable to a specific source or release.

The SMS is intended to provide flexibility to establish regional background on a case-by-case basis and does not prescribe specifically how regional background should be established. The approach and methods contained in the Bellingham Bay Regional Background Sampling and Analysis Plans (SAP; Ecology 2014a) were developed by Ecology to establish regional background concentrations for the following analytes: lead, carcinogenic polycyclic hydrocarbons (cPAHs), dioxins/furans, and polychlorinated biphenyls (PCBs). This study serves as one example of how regional background concentrations can be established in a particular Ecology-defined geographic area.

Ecology's approach to establishing regional background has evolved over time through working on initial bays and after receiving comments from stakeholders and tribes, as described below.

### 1.2. Stakeholder Discussions

Ecology received a number of comments from stakeholders and tribes related to the regional background sampling conducted at Port Gardner and North Olympic Peninsula (Ecology 2013b, Ecology 2013c, and Ecology 2014b). Many stakeholders requested that for future regional background characterizations, they would like to work with Ecology before SAPs were drafted and submitted for public comment. In response, Ecology engaged stakeholders earlier in the process for the initial discussions regarding establishing regional background for Elliott Bay and/or the Lower Duwamish Waterway. A similar process was followed for Bellingham Bay, with Ecology holding a stakeholder review meeting in August 2014 to discuss the draft SAP (Ecology 2014a).

Based on these collective comments and discussions, Ecology determined that some modifications to the original sampling designs used to establish regional background were appropriate. The concepts described below were included in the Bellingham Bay Regional Background SAP and incorporated into the final Sediment Cleanup Users Manual II (SCUM II; Ecology 2013a \& 2015).

The following modifications were incorporated into the Bellingham Bay regional background study design:

- Rationale and Conceptual Bay Model. This included a discussion of the selected analytes, existing information used to develop the sampling area of interest, and the rationale for the selected sampling methods. These choices were based on a conceptual bay model for Bellingham Bay and key features of Bellingham Bay. These included known sites and sources, existing chemistry data, existing modeling information, hydrodynamic information, bathymetry, etc.
- Determining Areas of Primary Influence. The area in which sediment samples were collected was consistent with the SMS definition of regional background (WAC 173-204$505(16))$. This entailed sampling closer to the shoreline, sources, and sites, while
remaining outside areas of direct influence. Bay-specific information was used, where available, to determine areas directly associated with depositional zones of outfalls or other point sources and areas directly affected by sites.
- Differentiating from Natural Background. Existing data were examined to identify areas that were within the range of natural background concentrations (as defined in SCUM II; Ecology 2013a \& 2015). These areas were excluded from sampling and calculation of regional background.
- Differing Areas of Interest for Different Analytes. Different analytes were found to be elevated above natural background in different areas of the bay. In Bellingham Bay, it was determined that cPAHs were elevated over a larger area than other chemicals. Therefore, a larger area of interest (AOI) was used for sampling regional background concentrations of cPAHs.


### 1.3. Bellingham Bay Area of Interest

Regional background concentrations for Bellingham Bay were determined based on data from samples collected within the AOI (Figure 1). The AOI was defined as a hydraulically connected marine environment that excluded areas directly influenced by known cleanup sites, potential sources (e.g., active or historical outfalls), dredged disposal sites, and areas more representative of natural background (as defined in WAC 173-204-505) such as the Nooksack River delta (Figure 1). Specifically, the AOI boundaries were determined as described below:

1. Cleanup sites along the shoreline were excluded based on clearly elevated levels of the bioaccumulative chemicals of concern. Bellingham Bay has been the focus of a number of sediment cleanup projects in the inner harbor and in other locations, along with a comprehensive waterfront redevelopment plan currently underway. Together, these processes have eliminated or reduced many contaminant sources (particularly industrial sources) to the bay, allowing natural recovery to begin. However, a number of sites have remaining contamination that may directly influence sediment concentrations in Bellingham Bay. The following cleanup sites were excluded from the AOI and a $75 \mathrm{~m}(\sim 250 \mathrm{ft})$ exclusion buffer was added beyond the cleanup site boundaries (Figure 1). The cleanup sites are described below (from north to south):

- Oeser Company. This is a wood treating company is located adjacent to Little Squalicum Park, which includes Little Squalicum Creek. Little Squalicum Creek discharges to the bay just north of Squalicum Creek. Historical wood treating practices resulted in the contamination of company property, as well as the park and creek. The site has been designated as a federal Superfund site. Contaminants at the site include pentachlorophenol, dioxins/furans, copper, zinc, and polycyclic aromatic hydrocarbons (PAHs). A Record of Decision was finalized in 2003 and site cleanup activities were
undertaken in 2005 - 2009. In 2010, EPA determined that additional cleanup activities were required in Little Squalicum Park and these were conducted in 2010-2011.
- Eldridge Municipal Landfill. This is a former City of Bellingham municipal landfill located in Little Squalicum Park, next to Little Squalicum Creek. Contaminants at the site included PAHs, phthalates, pentachlorophenol, and metals (cadmium, copper, lead, mercury, and zinc). In 2011, over 4,000 tons of debris and soil were excavated and disposed of in Roosevelt Landfill. A remedial investigation/feasibility study (RI/FS) and a consent decree/cleanup action plan (CD/CAP) are in the process of being completed.
- Weldcraft Steel \& Marine. This site is located in Outer Squalicum Harbor and has been used for boat repair, maintenance, and fabrication. Contaminants in sediments from past practices include metals, tributyltin (TBT), gasoline, and diesel. An interim cleanup action to remove contaminated sediment was completed in 2006. Contamination remains in upland areas of the site and an RI/FS has been completed.
- Marine Services Northwest. This site is located in Inner Squalicum Harbor. Contaminated sediments are located adjacent to a historical boat maintenance lift. Contaminants include metals, TBT, and PAHs. This is a quiescent and enclosed area. In 2001, a draft RI/FS was developed by the Port under the Voluntary Cleanup Program (VCP) prior to Ecology's decision to not allow sediment cleanups under VCP.
- I\&J Waterway. This site consists of contaminated sediments in and adjacent to the federally authorized navigation lane. Historical industrial operations along the waterfront include a lumber mill, rock-crushing plant, and frozen food processing. A seafood processing plant is currently in operation. Contaminants identified in sediments include phthalates, phenols, PAHs, dioxins/furans, nickel, and mercury. An RI/FS is in the process of being completed.
- Central Waterfront. This site is located between I\&J and Whatcom Waterways and consists of 55 acres of the waterfront with various historical industrial uses, including concrete manufacturing, lumber mill operations, boat repair, two bulk fuel terminal, rock processing, and a landfill. The site is adjacent to the former Georgia-Pacific treatment lagoon. Contaminants found in upland areas of this site include petroleum mixtures, PAHs, and a variety of metals. In 2013, an interim cleanup action was completed which included removal of over 1000 tons of petroleum-contaminated soil, removal of creosoted pilings, and beach restoration. The RI/FS for the rest of the site is expected to be completed in 2015.
- Holly St. Landfill. This historical City of Bellingham municipal waste landfill is bisected by Whatcom Creek as it enters Whatcom Waterway. Contaminants at the site include refuse along the shoreline, as well as copper and zinc releases to the waterway. Cleanup was completed in 2005.
- Whatcom Waterway. This sediment site is over 200 acres and includes the waterway itself, as well as the adjacent former Georgia-Pacific treatment lagoon. Contamination consists mainly of mercury and phenols from the former Georgia-Pacific pulp mill
operations. Untreated wastewater was discharged to the bay until 1979, when the Georgia-Pacific wastewater treatment lagoon and deepwater outfall (with a diffuser) was built. Prior to construction of the Post Point Wastewater Treatment Plant and outfall in 1974, primary treated wastewater was also discharged from the Whatcom Creek Waterway Wastewater Treatment Plant into the Whatcom Waterway. Contaminants in sediments include wood waste, phthalates, phenols, mercury, PAHs, various metals, and dioxins/furans. In 2001, an interim action resulted in the placement of dredged material over the approximately 6 -acre Log Pond area of the site. The first phase of the final cleanup action is expected to begin in 2015 and will include dredging and capping in the inner portion of the waterway, additional capping in the Log Pond, and removal of structures and pilings.
- Georgia-Pacific West. This is the upland property where the main operations area of the former Georgia-Pacific mill was located on the south shoreline of Whatcom Waterway. Contaminants on the 74-acre site include PAHs, metals, dioxins/furans, various petroleum mixtures, and volatile organic compounds. The RI was completed in 2013 and interim actions to remove petroleum-contaminated and mercury-contaminated soil were conducted in 2011 and 2013. The site has two separate and distinct areas of contamination and has been divided accordingly. An FS and CD/CAP for the northern Pulp \& Tissue Mill Remedial Action Unit was completed in 2014. An FS and CD/CAP amendment for the southern Chlor-Alkali Remedial Action Unit are expected to be completed in 2015.
- R.G. Haley. This site is located south of the Georgia-Pacific west site and north of the Cornwall Avenue Landfill site. Former wood treating operations, as well as lumber, coal, and wharf facilities have resulted in soil, groundwater, and sediment contamination. Contaminants include wood waste, diesel plumes, pentachlorophenol, dioxins/furans, and PAHs. An emergency action was taken in 2001to contain oil releases to Bellingham Bay, including building a sheet-pile wall, installing oil recovery wells, and some sediment removal. In 2013, a sand/clay layer was placed on a portion of the shoreline to address an oil seep. A draft RI/FS has been completed.
- Cornwall Ave. Landfill. This site is south and adjacent the R.G. Haley site and was originally used as a sawmill and for wood storage. It was later used as a City of Bellingham municipal waste landfill, followed by log storage after closure of the landfill. Erosion of the landfill into the bay has occurred over the years. Sediment contaminants include wood waste, metals, phthalates, polychlorinated biphenyls (PCBs), PAHs, phenols, diesel, and heavy oils. An interim action was conducted in 2011/2012 to cap much of the upland solid waste with stabilized dredged sediments and an impermeable liner. An RI/FS was completed in 2013 and a CD/CAP completed in 2014.
- South State St. Manufactured Gas Plant. This site is located along the southeast shoreline of Bellingham Bay at the north end of Boulevard Park. A former manufactured gas plant operated at this location until the late 1940s. Contaminants in sediments include
various petroleum products, PAHs, and volatile hydrocarbons. An RI has been conducted and the FS is underway.
- Harris Avenue Shipyard. This site is located at the southern end of the study area along the shoreline in Fairhaven. Sediments at the site have been contaminated by former shipbuilding and maintenance operations and contain metals, PCBs, phthalates, PAHs, and possibly dioxins/furans. An RI/FS is in the process of being completed.

2. In addition to the cleanup sites listed above, the following potential point and nonpoint sources of contamination were identified and excluded from the AOI (Figure 1):

- Post Point Wastewater Treatment Plant Outfall. This outfall is at the southern boundary of the study area and discharges offshore at approximately 100 ft depth. The treatment plant has treated municipal sewage, stormwater, and industrial wastewater, and was upgraded to secondary treatment in 1993. Due to high organic loading, high ammonia and sulfides have been measured in sediments resulting in intermittent sediment bioassay failures. The depositional zone of this outfall is within the established exclusion zone of the AOI.
- Former Georgia-Pacific Deepwater Outfall. This outfall extends 8,000 feet in a southwesterly direction from the former Georgia-Pacific treatment lagoon and includes a 2000 -ft-long diffuser section discharging into 55 feet of water. Since the mill operations stopped in 2007, substantial natural recovery has occurred in this area. A 75 meter exclusion buffer was added around the diffuser.
- Urban Watersheds. As noted above, several creeks carry stormwater discharges and historically received industrial discharges and can represent historical and/or current sources of contaminants to Bellingham Bay. Whatcom Creek discharges into Whatcom Waterway and would be included in that site. Little Squalicum Creek and Padden Creek could also have areas of elevated contaminants near their mouths, but recent data is not available to determine their status. The established exclusion buffers were sufficient to account for any direct influence from these potential sources.
- CSOs. After the Post Point Treatment Plant was built in 1974, combined sewer overflow (CSO) events occurred at four locations, including at the Post Point Treatment Plant, at the C Street Interceptor (former Whatcom Waterway Treatment Plant outfall), at the lower Cornwall pump station, and at the Oak Street pump station. CSO reduction programs since then have minimized CSO events so that they only occur at the C Street Interceptor in Whatcom Waterway. The established sampling exclusion buffers were sufficient to account for any direct influence from these potential sources.
- Starr Rock. In 1969, 130,000 cubic yards of sediment was dredged from Whatcom Waterway for maintenance and disposed of near Starr Rock, a natural rocky formation south of the Cornwall Avenue Landfill site. These sediments were likely contaminated with chemicals from wastewater discharged from the Georgia-Pacific pulp mill into Whatcom Waterway. Known contaminants include mercury, dioxins/furans, and various
semi-volatile organic chemicals (SVOCs). An additional 75 meter exclusion buffer was added around this potential source.
- Early Sediment Disposal Sites. The I\&J Waterway was dredged in 1966 and the sediments were disposed of in an area about $1 / 4$ mile west of the terminus of the GeorgiaPacific deepwater outfall. The Squalicum Creek Waterway was also dredged in 1963 and these sediments were disposed of just east of the mouth of Little Squalicum Creek. Between 1979 and 1983, several dredging projects disposed of sediments in a disposal site northeast of the I\&J site. Many of these sites may have contained contaminated sediments and several were likely impacted by mercury from the deepwater GeorgiaPacific outfall. However, all have been subject to decades of natural recovery as well. The established sampling exclusion buffers were sufficient to account for any direct influence from these potential sources.
- DMMP Deepwater Disposal Site. A deepwater non-dispersive disposal site has been established offshore of Post Point in the deeper areas of the bay. This disposal site has received mainly clean dredged material that is not necessarily representative of surrounding sediments. However, the last disposal event at this site was in 1998. This area was not included in the AOI.

3. Areas that represented natural background concentrations or that contained natural features that would hinder sampling were also excluded (Figure 1):

- Areas where existing data were below natural background concentrations, as defined by the Bold Plus data set (SCUM II; Ecology 2013a \& 2015), were excluded from the regional background AOIs. Boundaries were drawn approximately half-way between existing data points above and below the natural background 90/90 upper tolerance limit (UTL). A larger area of the bay was within natural background for dioxins/furans. Therefore, the regional background AOI for cPAHs, PCBs, and lead was larger than that for dioxins/furans (Figure 1).
- Shallow areas $<6 \mathrm{ft}$ mean lower low water (MLLW; $\sim 2 \mathrm{~m}$ ) in depth were excluded for logistical sampling reasons. These include all areas within the Nooksack River delta and south along the shoreline, some areas within the waterways, and very narrow areas along the southeast shoreline.


### 2.0 Sampling Methods and Analysis

The analytes selected for the calculation of regional background included lead, cPAHs, dioxin/furan congeners, and PCB congeners. Arsenic and cadmium were not included because they did not appear elevated in the bay outside of clearly contaminated areas within sites. Mercury was not included for two reasons: 1) current concentrations throughout the bay represent legacy contamination primarily from a single historic source, and 2) concentrations of mercury in the bay are rapidly recovering due to natural processes.
Lead was not evaluated as part of the conceptual bay model, but was discussed as a potential analyte at the August 2014 Bellingham Bay workshop (Section 1.2). Lead was added to the analyte list as a result of this discussion.

Two AOIs were established within Bellingham Bay based on distributions of existing data for cPAHs and dioxin/furan congeners. Concentrations of cPAHs were higher than natural background throughout a larger area than those of dioxin/furan congeners. As a result, the AOI for cPAHs was larger than that of dioxin/furan congeners (Figure 1). Due to a lack of existing bay-wide data, lead and PCB congeners were sampled in the larger cPAH AOI. A priori precision calculations based on existing data for cPAHs and dioxin/furan congeners from within the Bellingham Bay AOIs indicated that a minimum of eight dioxin samples and 22 cPAH samples would be expected to achieve the precision target for the data set (Ecology 2014a). Based on these estimates, no secondary samples were collected. The final sample counts for each AOI were as follows:

- 23 samples were collected from the larger AOI and analyzed for dioxin/furan congeners, cPAHs, PCB congeners, lead, total organic carbon (TOC), total solids, and grain size.
- 7 samples were collected from the dioxin/furan exclusion area and were analyzed for all of the above mentioned analytes except dioxin/furan congeners.

Sediment sampling was conducted from September 9 through September 11, 2014. The target sampling stations were selected within each AOI using a spatially balanced random sampling procedure such that no samples would be collected within 250 meters of each other.

### 2.1. Station Positioning and Navigation

The R/V Kittiwake was used for the surface sediment grabs. A differential global positioning system (DGPS) was used aboard the R/V Kittiwake for station positioning. The sampling station target coordinates were provided in advance and programmed into the R/V Kittiwake's navigation system. Upon sampling device deployment, the actual position was recorded once the device reached the seafloor and the winch cable was in a vertical position. Latitude and longitude station coordinates were recorded in degrees decimal minutes using the 1983 North American Datum (NAD83). Water depths were measured using the winch meter wheel and verified by the
ship's fathometer. Table 1 provides the actual coordinates, water depths, and distance between the target and actual stations. Figure 2 shows the actual sample stations.

Two attempts were made at the target coordinates for station BB-19. A section of $2 \times 4$ lumber was caught in the grab on the first attempt. The second attempt came up empty. The target station was very near shore, and the bottom appeared to be a hard substrate (i.e. gravel and/or cobble). After consulting with Ecology, the station was moved 84 meters southwest (Table 1).

### 2.2. Surface Sediment Grabs

Surface sediment grabs were collected at 30 stations. All samples were collected using a stainless steel double van Veen grab sampler ( $0.2 \mathrm{~m}^{2}$ ). Sampling followed the step wise procedure outlined in the SAP (Ecology 2014a). Notes related to sampling activities are presented in the sediment grab logbook in Appendix A. A brief summary of field sampling methods is provided below.

Established deployment and recovery procedures for the grab sampling gear, described by the Puget Sound Estuary Program (PSEP), were followed to ensure recovery of the best possible samples and minimize risks to personnel and equipment (PSEP 1997). Once a grab sample was retrieved, the overlying water was carefully siphoned off one side of the sampler. If the sample was judged to be acceptable according to PSEP specifications, the penetration depth was measured with a decontaminated stainless steel ruler, and sample quality, color, odor, and texture were described in the sample log. Scanned copies of the surface sediment grab logbook are presented in Appendix A.

The target depth for surface sediment collection was 12 cm . The penetration depth for seven samples was less than the target. However, penetration in all seven of these samples exceeded 10 cm.

Percent fines were determined at each station by rinsing 40 ml of sediment through a $63.5 \mu \mathrm{~m}$ sieve until the water was clear. Percent fines are equal to 40 minus the volume of remaining sediment divided by 40 . The amount of sediment retained on the sieve was recorded in the surface sediment grab logbook (Appendix A).

### 2.3. Sample Storage, Delivery, and Chain of Custody

After filling the jars with homogenized aliquots of sediment, all samples were labeled and the lids were wrapped with electrical tape to seal the jars and prevent leakage. Each label was marked with a jar tag number for tracking purposes. Sample identification and jar tag numbers were recorded in the sample container logbook (Appendix B).

After labeling, all samples were stored in insulated coolers and preserved by cooling to a temperature of $4^{\circ} \mathrm{C}$.

Samples were delivered to Analytical Resources Incorporated (ARI; Tukwila, WA) and shipped to Axys Analytical (Sidney, BC). Sediment samples collected for archival were also submitted to ARI. All archive samples were frozen at $-18^{\circ} \mathrm{C}$. The Chain of Custody forms for all samples are presented in Appendix C.

### 2.3.1.Laboratory analysis

Samples were submitted to laboratories subcontracted by NewFields to conduct the chemical analysis. Axys analyzed the samples for dioxin/furan and PCB congeners. ARI analyzed samples for the sediment conventionals (TOC, total solids, and grain size), lead, and cPAHs. Table 2 presents a list of all samples collected as part of this sampling effort and includes the relevant analytical methods.

Additional samples collected for quality assurance/quality control (QA/QC) purposes are listed in Table 2. Field duplicates and triplicates were collected at stations BB-20 and BB-24. Rinsate blanks and equipment rinsate samples for metals and cPAHs were also collected. Further details relating to chemical analysis can be found in the SAP (Ecology 2014a).

Because of expected low concentrations, the data quality objectives (DQOs) used in this study were more stringent than those required under most sediment characterizations. As a result, the target practical quantitation limits (PQLs) for analysis were lower than most standard methods could provide. The PQLs for the analytes are listed in Table 3. This table includes the PQLs for the dioxin-like PCB congeners. The PQLs for the non-listed PCB congeners were all $0.4 \mathrm{ng} / \mathrm{kg}$. The PQLs for the conventional parameters and the full list of PCB congeners can be found in the SAP (Ecology 2014a).

All non-detect sample results for cPAHs were reported to the method detection limit (MDL) and detected results less than the target PQL were "J" qualified. All non-detect results for metals were reported at the PQL. Metals data are not qualified below the PQL. Non-detect results for dioxin/furan and PCB congeners were reported at the sample-specific detection limit (SDL). All detected congener results less than the PQL were "J" qualified.

Laboratories do not provide PQL values for toxic equivalent quotient (TEQ) concentrations. Instead, these values were calculated for cPAHs , dioxin/furan congeners, and PCB congeners using the toxicity equivalency factors (TEF) from SCUM II (Ecology 2013a \& 2015) for determining TEQ values and the individual compound or congener-specific PQLs in Table 3. The Ecology guidance for determining TEQs uses the dioxin/furan TEF values updated by the World Health Organization (WHO) in 2005 (Van den Berg et al. 2006). The resulting PQL for cPAHs was $0.76 \mu \mathrm{~g} \mathrm{TEQ} / \mathrm{kg}$. The PQLs for dioxin/furan and PCB congeners were 2.3 and 0.052 ng TEQ/kg, respectively.

### 3.0 Data Validation

A QA2 (EPA Stage 3/4) chemistry data review was conducted by EcoChem, Inc. (Seattle, WA) who examined the complete analytical process from calculation of instrument and method detection limits, PQLs, final dilution volumes, sample size, and wet-to-dry ratios to quantification of calibration compounds and all analytes detected in blanks and environmental samples (PTI 1989a; PTI 1989b; USEPA 2009). The intent of the independent data validation was to ensure that the data are defensible and usable for regulatory purposes. This section provides a brief summary of the data validation reports. Two validation reports were completed. The first was for the majority of the chemistry samples, while the second was for two PCB congener samples that needed re-extraction and analysis by Axys. Both full validation reports are provided in Appendix E and available as electronic files upon request.

When necessary, EcoChem applied the following data qualifiers to the chemical results:

- U: The analyte was analyzed but not detected above the reported sample quantitation limit.
- UJ: The analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitation necessary to accurately and precisely measure the analyte in the sample.
- J: The analyte was positively identified. The associated numerical value is the approximate concentration of the analyte in the sample. " J " qualifiers were assigned by the laboratories for results less than the PQL and greater than the MDL, or by EcoChem for results that failed to meet study specific QA/QC criteria.
- DNR: Do not report. A more appropriate result is reported from another analysis or dilution.

The use of the DNR qualifier was limited to selecting the appropriate results for $2,3,7,8-\mathrm{TCDF}$, as results were reported for analysis on two separate columns. The remainder of the data was usable. Reason codes for applying the "U," "UJ," and "J" qualifiers and the definition for these codes are given in the validation reports (Appendix E).

Several qualifiers given by Axys were reclassified by EcoChem. Axys gave a "B" qualifier to all results where the analyte was detected in the method blank. EcoChem established an action level of five times the blank concentration. If a sample result was above that, the " B " qualifier was removed. If the result was below the action level, the result was qualified as not-detected "U." The laboratory assigned " K " qualifiers to dioxin/furan and PCB congener data. This qualifier implied that a peak was detected, but did not meet identification criteria. These data were considered estimated maximum possible concentrations (EMPC). All EMPC results were given a "U" qualifier by EcoChem, but remained at the reported concentration which represented an
elevated PQL for that congener.
Project specific QA/QC measures were employed during sample collection and analysis to ensure the precision, accuracy, and reproducibility of the results. This included field QA/QC samples such as equipment rinsates, rinsate blanks, and field duplicates and triplicates. Laboratory measures included the analysis of standard reference materials (SRM).

The equipment rinsate blank and decontamination water rinsate provided a quality control check on the potential for cross contamination by measuring the effectiveness of the sampling and processing decontamination procedures. Rinsate samples were collected and analyzed for lead and PAH. No PAH were detected in the rinsate samples. Lead was detected in each of the rinsates, but no sample results were within five times the blank concentrations.

Field duplicates and triplicates were collected at the same time as the original samples using identical sampling techniques. Duplicate/triplicates were used to determine the precision of the sample collection process and determine the representativeness of the sample. Table 2 lists the specific duplicates and triplicates collected for this study.

The relative percent difference (RPD) was used to evaluate duplicate samples, while the relative standard deviation (RSD) was used to evaluate triplicates. In general, if the RPD or RSD was greater than 50 percent, the affected results of the duplicate/triplicate sample were " J " qualified. If a result was already "U" qualified, the affected results were requalified "UJ." For the duplicate BB-24-S/D, the RPD for 22 PCB congeners and 3 homologs was greater than the control limits and the results were qualified. For duplicate BB-20-S/D, congeners PCB-80 and PCB-128 were qualified for an elevated RPD. Precision was acceptable for all other analytes.

Overall, the high precision of the field duplicates and triplicates indicated that the study results were representative of the sediment they were collected from, which is important for reducing variability in the data set.

The recently developed Puget Sound reference material (PSRM) was submitted for analysis for the analysis of dioxin and PCB congeners. The published acceptance criteria for this PSRM are $\pm 50$ percent of the mean.
(http://www.nws.usace.army.mil/Missions/CivilWorks/Dredging/SRM.aspx).
The results for $1,2,3,7,8,9-\mathrm{HxCDF}$ were less than the lower control limit, while the results for PCB-120 were greater than the higher control limit. No results were qualified based on these outliers as the reference material is still undergoing evaluation and is not yet certified.

PCB congeners in samples BB-22 and BB-30 were analyzed in a separate batch due to issues with the initial extraction impacting the mono-chlorinated biphenyls. As a result, no laboratory quality control data was available for PCB-01, PCB-02, and PCB-03. Results for these congeners in both samples were "J" qualified. The validation for these samples was completed as part of a separate report (Appendix E).

### 4.0 Data Results

A summary of the results from the laboratory analysis is provided in this section. The results are presented in terms of general usability by listing the number of undetected and qualified results for each analyte (Figure 3). The results of the conventionals analyses (grain size distribution and TOC) are presented in Figure 4. The spatial distributions of the measured analytes throughout the AOI are presented in Figures 5 through 8. Complete data results are presented in Appendix F. Laboratory data packages are available electronically as Appendix D.

### 4.1. Calculation of Toxicity Equivalents

Calculation of the TEQ when many of the congener concentrations within a sample are reported below the detection limits can be problematic. A common approach is to substitute $0,1 / 2$, or 1 times the detection limit in place of a non-detected concentration. A more robust method for calculating total TEQs when non-detect values are present is the Kaplan-Meier (KM) approach, which is a statistical method for estimating a sum or mean when part of the population is censored (Helsel 2010, 2012). The methods for addressing non-detects, including KM, are discussed in greater detail in the SCUM II guidance, Chapter 6 (Ecology 2013a \& 2015).

KM TEQs were calculated separately for the PCB congeners, dioxin/furan congeners, and cPAHs from each sample. The KM means reported for the TEQ data were calculated using R version 3.1.1 (R Core Team 2014) using the cenfit function from the NADA package (Lee 2013).The KM sum was calculated and the number and distribution of censored values was evaluated. The following rules were applied to the final KM TEQs:

- If the number of non-detect congeners within a sample exceeded 50 percent, the KM TEQ value was qualified as a less-than value (L qualified), followed by the number of censored congeners (see data tables in Appendix F). For example, if 12 of the 17 dioxin/furan congeners were undetected, the detection frequency was $29 \%$ and the KM TEQ would be calculated and qualified with L12.
- If the lowest detection limit for a non-detect was lower than all detected values, the positive bias in the KM estimate was adjusted downwards using Efron's bias correction (Klein and Moeschberger 2003). This method treats the lowest ranked value as detected even if it was reported as a non-detected data point.
- If the highest detection limit is greater than the highest detected value, the highest nondetect value provides no meaningful information and is ignored in the KM estimation of the mean. The highest toxic equivalent concentration (TEC) value is always treated as uncensored in the KM TEQ calculation, and the TEQ is qualified with an L if the original value was censored. All L-qualified TEQ values were treated as non-detects in the distributional assessments and when calculating summary statistics across samples.

Calculated KM TEQs are presented in the data tables in Appendix F along with the traditional 0, $1 / 2$, and 1 detection limit substitutions. A brief comparison was made of the results from these four estimates of total TEQ.

The mean and $90^{\text {th }}$ percentiles were calculated for each method. For cPAHs the mean and $90^{\text {th }}$ percentiles differed by less than $0.1 \mu \mathrm{~g}$ TEQ $/ \mathrm{kg}$ regardless of the method used, indicating nondetects had little influence on concentrations. The mean dioxin/furan TEQ concentrations were all within $0.06 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$, while the $90^{\text {th }}$ percentiles were consistent at $13.2 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$.

Larger differences were noted for PCB TEQ concentrations. PCB TEQ means ranged from 0.09 $\mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$ for the 0 DL substitution to $0.13 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$ for the 1 DL substitution. The $90^{\text {th }}$ percentiles ranged from 0.21 to $0.25 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$. The greatest difference between the methods was seen in the maximum concentrations for PCB TEQ. The maximum concentrations for the 0 and $1 / 2$ DL substitutions were 0.28 and 0.29 ng TEQ $/ \mathrm{kg}$, respectively. The maximum concentrations for the 1 DL substitution and the KM estimate were 0.43 and $0.38 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$, respectively. Sample BB-12 had the maximum PCB TEQ concentration which was driven by PCB-126. This congener has the highest TEF value of the dioxin-like PCBs, and was a nondetect at an elevated detection limit due to the presence of wood waste in this sample. The use of this elevated DL at the reported concentration is likely an overestimate of the true concentration. This sample was excluded in the calculation of the summary statistics in Section 4.3.3 and is discussed as an outlier in Section 5.3.

Given the small differences between the methods, the more statistically robust KM TEQ values are used in statistical summaries and analysis for the remainder of this report when discussing total TEQ concentrations.

### 4.2. Summary of Qualified Results

The DQOs for this study necessitated PQLs that were lower than those typically used in sediment investigations, as the intent of this regional background study is to obtain as few nondetects and as many unqualified results as possible. Too many non-detects could create a skewed distribution that would not meet the project requirements for precision (Section 5.2), while too many data qualified as estimated for a given analyte could result in an unreliable regional background concentration or one that is below the project-specific PQLs summarized in Table 3.

The number of qualified (both non-detect and estimated) results for each analyte are shown in Figure 3. Non-detect results are represented by dark blue and included all data given a qualifier flag of "U" or "UJ." Estimated values were given a qualifier flag of " $J$ " and are represented by a medium blue color. A "J" qualifier indicates the result was considered an estimate either because the value was less than the PQL and greater than the MDL, or the EcoChem data validation indicated QA/QC issues. The light blue color indicates sample results that were not qualified. The total sample counts in Figure 3 include the field duplicates and have not been screened for outliers.

A total of 32 samples and 2 duplicates were analyzed for lead. None of these results were qualified.

Dibenz( $\mathrm{a}, \mathrm{h}$ )anthracene was the only PAH with non-detect results. Twenty six out of 32 results were non-detects, with an additional three " J " qualified results. The remaining cPAH compounds were detected in all samples. However, results from twelve samples were "J" qualified for each cPAH compound. The total cPAH TEQ concentration was above the calculated PQL of $0.76 \mu \mathrm{~g}$ TEQ/kg in all samples (Figure 3).

There were also relatively few non-detects for the dioxin/furan congeners. 2,3,7,8-TCDD and $1,2,3,7,8-\mathrm{PeCDD}$ have the greatest impact on total TEQ (TEF of 1). These two congeners alone comprised nearly 25 percent of the total TEQ on average, but were not detected in one sample each. The hexa-, hepta-, and octachlorinated dioxin congeners were detected with the greatest frequency. Two samples had a total TEQ less than the PQL of 2.3 ng TEQ/kg (Figure 3).

Of the dioxin-like PCBs, PCB-169 was not detected in any of the samples and PCB-81 was not detected in 15 samples. The remaining congeners were detected in at least 75 percent of the samples. PCB-126 has the greatest impact on total TEQ with a TEF of 0.1 and was a non-detect in six samples. The total PCB TEQ was less than the PQL of 0.052 ng TEQ/kg in seven samples (Figure 3).

Overall, the data quality for calculation of regional background is high, as most of the analytes were detected and without qualifiers in more than two-thirds of the samples analyzed.

### 4.3. Summary and Spatial Distribution of Results

This section provides an initial evaluation of the sample results prior to the more in-depth statistical evaluation of Section 5.0. The spatial distributions of each analyte are presented in Figures 4 through 8 . Summary statistics including the minimum, median, average, and maximum concentrations for each analyte are presented in Table 4 for the Bellingham Bay data set. Table 4 also includes the Pearson correlation coefficient ( $r$-value) and its significance level ( $p$-value) for correlations of each analyte to TOC.

Field duplicates and triplicates were averaged prior to mapping the spatial distributions and calculating the summary statistics in Table 4. Only detected concentrations were averaged for a given station. If all concentrations were non-detects, the maximum detection limit was used. The TEQ values presented in this section were calculated using the KM method described in Section 4.1.

Potential outliers were excluded from the calculation of the summary statistics. The excluded samples are listed in the following text and in the footnotes of Table 4. A more detailed evaluation of these samples as outliers is provided in Section 5.3.

### 4.3.1.Conventional parameters

Conventional parameters analyzed for this study included grain size, TOC, and total solids. Figure 4 presents combined results for the grain size distribution and percent TOC for all stations. In this figure, percent fines are represented by the color gradient and percent TOC is represented by the size of the circle.

There was little variation in percent fines throughout the AOI. With the exception of BB-12, all samples were between 84 and 99.6 percent (Figure 4). While measured fines in BB-12 were reported at 55.2 percent, this is likely an underestimate. This sample contained a large amount of fine, woody debris that skewed the final determination of fines low.

The large amount of wood waste in sample BB-12 was apparent from its high concentration of TOC and was indicative of a direct anthropogenic influence. TOC in this sample was 20.8 percent, more than five times higher than the next highest concentration in Bellingham Bay, and more than four times higher than the maximum observed in the Bold survey. Excluding BB-12, TOC in Bellingham Bay ranged from 1.0 to 4.0 percent. Most of the analytes were correlated to TOC. No correlation analyses were conducted against percent fines due to the limited range of fines in Bellingham Bay.

### 4.3.2. Metals

All 30 samples were analyzed for lead (Figure 5). The lowest concentrations were found in the northwest corner of the AOI, where the highest natural recovery was observed. The highest concentration of $53 \mathrm{mg} / \mathrm{kg}$ was present at station BB-28. This concentration was a statistical outlier, and because the station was located near a historical source (Section 5.3), it was excluded from the summary statistics as an outlier. BB-12 was also excluded due to elevated TOC and its implication of direct anthropogenic influence. In the remaining 28 samples, lead concentrations ranged from 8.4 to $16 \mathrm{mg} / \mathrm{kg}$ (Table 4). The correlation of lead to TOC had an $r=0.818$ and was statistically significant ( $p<0.0001$ ).

### 4.3.3. Organics

Sample BB-12 is shown in Figures 6, 7, and 8, but was excluded from the calculation of summary statistics for all analytes (Table 4) due to the presence of wood waste and high TOC (Section 5.3). PCB congener and cPAH TEQs were removed for station BB-28 due to elevated concentrations of these analytes that potentially originated from the historical source; the dioxin/furan concentration at this station was retained as it was not likely impacted by the historical source (Section 5.3).

The cPAH TEQ concentration was highest at station BB-28 (2,050 $\mu \mathrm{g} / \mathrm{kg}$, Figure 6). Excluding this sample, the next highest concentration was $94 \mu \mathrm{~g} / \mathrm{kg}$. The median cPAH TEQ concentration was $20 \mu \mathrm{~g} / \mathrm{kg}$. cPAH TEQs were significantly correlated to TOC with $r=0.749(p<0.0001$, Table 4).

The spatial distribution of dioxin/furan TEQ concentrations is shown in Figure 7. Concentrations less than 4.0 ng TEQ/kg were assumed to be representative of natural background. The dioxin/furan AOI was smaller than that of the other analytes due to low dioxin/furan concentrations in the existing data in the northwest corner of the AOI. Samples BB-14 and BB27 near the boundary with the natural background area both had concentrations near 2.0 ng TEQ/kg (Figure 7). These stations appear to be representative of natural background and were excluded from the regional background data set. In the remaining 20 samples, dioxin/furan TEQ concentrations ranged from 4.2 to $14 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$, with a median of $9.9 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$. Dioxin/furan TEQ concentrations were not significantly correlated to TOC ( $p>0.05$, Table 4).

The PCB congener TEQ is based on the toxicity of dioxin/furan congeners. However, the TEFs for PCBs are lower than those of dioxin/furan congeners, resulting in lower TEQs. The PCB congener TEQ was highest at station $\mathrm{BB}-12$ and was due to a particularly influential nondetected congener (Section 4.1). Station BB-28, located near a historical source, had the next highest PCB TEQ concentration ( 0.28 ng TEQ $/ \mathrm{kg}$ ). Both stations BB-12 and BB-28 (Figure 8) were excluded from the summary of results for PCB congener TEQ concentrations (Section 5.3). PCB TEQ results for the remaining 28 samples ranged from 0.028 to $0.21 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$ with a median of 0.08 ng TEQ $/ \mathrm{kg}$. PCB TEQ concentrations had a statistically significant correlation to TOC ( $p=0.005$, Table 4).

### 4.3.4.Chemical distribution summary

Overall, the physical and chemical distributions shown on Figures 4 through 8 indicate the following patterns and similarities:

- Consistent concentrations of percent fines at all stations, with the exception of station BB-12.
- Wood waste and elevated TOC at station BB-12.
- Direct influence on concentrations of lead, cPAHs, and PCBs from a historical source at station BB-28.
- Low concentrations in the northwest corner of the AOI due to the influence of the Nooksack River.
- Strong or moderate correlations of most chemicals with TOC.

When stations BB-12 and BB-28 are excluded, these chemical distributions suggest that:

- The AOI did not contain areas directly affected by sites or sources.
- Variations in the data were primarily correlated with natural characteristics of the sediments and processes within the bay.

These features confirm that the overall data set is appropriate for calculation of regional background concentrations. A more formal outlier analysis is presented in Section 5.3.

### 5.0 Data Analysis

This section describes the approach used to evaluate the data for Bellingham Bay with the objective of determining regional background concentrations for those analytes that were elevated above natural background.

### 5.1. Natural Background for Bellingham Bay

The natural background data set suitable for use in Puget Sound is defined in SCUM II (Ecology 2013a \& 2015). Comparison to this natural background data set was important for establishing the western and southern boundaries of the AOIs in Bellingham Bay, determining which analytes were elevated above natural background, and evaluating potential outliers.

Ecology has determined that data from the OSV Bold survey (DMMP 2009) plus select data sets from reference areas (Bold Plus) are appropriate for use as natural background for sites throughout Puget Sound. Bold Plus consists of the 70 samples collected as part of the OSV Bold survey and analyzed for a full suite of analytes plus additional samples from reference areas. The Bold Plus data set was used for comparison with the Bellingham Bay data set (Section 5.3) to identify which analytes were present at concentrations above natural background. Information on the full suite of Bold Plus data can be found in SCUM II, Chapter 10 (Ecology 2013a \& 2015).

### 5.2. Data Sufficiency

The first step in ensuring that the data set was sufficient to calculate regional background values was to evaluate the precision of the mean, expressed as the width of the 95 percent upper confidence limit (UCL) of the mean and divided by the mean. For normally distributed data precision is calculated as:

$$
\text { Precision }=\frac{t_{0.05(1), d f} S / \sqrt{n}}{\bar{X}}
$$

where:
$\bar{X}=$ the arithmetic mean of the $n$ baseline samples
$t_{0.05(1), d f}=$ the 1 - tailed critical value from the $t$-distribution, for $d f$ degrees of freedom and $\alpha$ of 0.05 .
$d f=$ the degrees of freedom associated with the sample standard deviation $(\mathrm{S})$. This is $n-1$, where $n$ is the number of observations used to estimate the variance.

$$
S=\text { standard deviation of the sample }=\sqrt{\frac{\sum_{i=1}^{n}\left(X_{i}-\bar{x}\right)^{2}}{(n-1)}}
$$

Precision of the mean expressed in this way is a common frame of reference for quantifying uncertainty in the population estimates that are necessary for the calculation of the background threshold value. For non-normal data, precision is calculated using the most appropriate estimate of the 95 UCL on the mean (e.g., the gamma formula and critical values for data that follow the gamma distribution, or a non-parametric bootstrap method for data that do not follow a discernible distribution).

A precision value of 25 percent was selected as a rough guideline for regional background data sets. After excluding samples based on the outlier analysis (Section 5.3), the precision for lead, dioxin/furan TEQ, and PCB congener TEQ were all less than 25 percent (Table 5). The precision for cPAH was higher at 30 percent, but deemed sufficient for calculation of regional background (Table 5).

### 5.3. Outlier Analysis

Ecology has formulated a weight of evidence approach to identify and evaluate potential outliers and determine whether they should be excluded from the calculation of regional background. The recommended steps for this approach are as follows:

- A statistical analysis was conducted to identify potential outliers. This analysis included a variety of techniques such as Q-Q plots, box plots, and univariate outlier tests appropriate to the distribution. This analysis is summarized in Appendix G.
- The bay-specific distribution was compared to the Bold Plus natural background distribution, both visually for the entire distribution and with respect to the calculated 90/90 UTLs for the bay-specific and natural background distributions.
- If the bay-specific distribution for an analyte was within the natural background distribution, the analyte and any potential outliers associated with it were not evaluated further. Alternatively, if the bay-specific distribution for an analyte appeared to exceed natural background, any potential outliers within that distribution were evaluated further.
- If a station was elevated for any analyte and determined to be directly influenced by a current or historical source, the analyte(s) from that station believed to be associated with the source was excluded from the calculation of regional background.
- If a station was elevated for any analyte, but not directly impacted by a current or historical source, other factors that may explain the elevated value(s) were considered. Factors may include gradients or patterns in the data set for that analyte (or lack thereof), correlations with natural geologic factors (grain size or TOC), sediment transport processes, etc. Correlations to TOC are explored in Appendix G.
- If deemed necessary, the 90/90 UTL of the data set was calculated with and without any elevated values and/or statistically identified outliers. If the resulting 90/90 UTL calculated values were within the range of analytical variability and were not substantially different from one another, Ecology may have decided to retain the elevated
concentrations in the calculation of regional background. The decisions from the previous steps deemed this unnecessary for the Bellingham Bay data set.
- If available, Ecology may also choose to analyze additional supplemental samples if it would be expected to further clarify the upper tail of the bay-specific distribution. This was deemed unnecessary for the Bellingham Bay data set.

A summary of the results for each analyte is presented in Tables 4 and 5 and in Figure 9. A full description of the potential outlier investigations is presented in Appendix G. The key findings are as follows:

- Sample BB-28 was identified as a statistical outlier for lead and cPAHs, and also had somewhat elevated levels of PCB TEQs. Further review of potential sites and sources identified a former refueling station in the vicinity of BB-28 that was a likely source for all three analytes. Therefore, data for these analytes at this station were excluded from the calculation of regional background.
- Sample BB-12 was identified as a statistical outlier due to elevated TOC. The TOC was linked to wood waste that was observed in this sample during the field investigation, suggesting a direct anthropogenic influence. The wood waste in the sample also caused elevated detection limits for some analytes, particularly PCB-126, which skewed the TEQ calculation (Section 4.1). This elevated detection limit resulted in PCB TEQ being identified as a statistical outlier at this station. Because of the anthropogenic influence and the analytical interference caused by the wood waste, data for all analytes at this station were excluded from the calculation of regional background.
- No additional statistical outliers were found. However, dioxin/furan TEQ concentrations at stations BB-14 and BB-27 had recovered to levels below natural background. Given that these stations were located adjacent to the dioxin/furan AOI boundary (Figure 7), these stations were considered to be part of that natural background area and were excluded from the calculation of regional background.


### 5.4. Calculation of Bellingham Bay 90/90 UTLs and Regional Background Values

Table 6 presents the 90/90 UTL values for Bellingham Bay alongside the Bold Plus 90/90 UTL natural background concentrations. All 90/90 UTL values were calculated in ProUCL 5.0 (USEPA 2013).

The Bellingham Bay 90/90 UTLs for lead and PCB TEQ were consistent with those of natural background. The Bellingham Bay 90/90 UTL for lead was $16 \mathrm{mg} / \mathrm{kg}$, compared to $21 \mathrm{mg} / \mathrm{kg}$ for the Bold Plus natural background. The Bellingham Bay 90/90 UTL for PCB TEQ was 0.21 ng $T E Q / k g$, equal to that for natural background.

The Bellingham Bay 90/90 UTL for dioxin/furan TEQ was 15 ng TEQ/kg, nearly four times greater than the natural background of 4 ng TEQ/kg. The Bellingham Bay cPAH 90/90 UTL of $86 \mu \mathrm{~g} \mathrm{TEQ} / \mathrm{kg}$ was more than five times greater than the natural background of $16 \mu \mathrm{~g} \mathrm{TEQ} / \mathrm{kg}$. The following conclusions regarding regional background can be drawn from these results:

- The Bellingham Bay and natural background 90/90 UTLs for lead and PCB TEQ were within the range of analytical variability and considered the same. Therefore, regional background values cannot be calculated for these analytes and will default to natural background.
- The Bellingham Bay 90/90 UTL for dioxin/furan TEQ is above natural background by nearly a factor of 4 . A regional background concentration of $15 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$ has been calculated.
- The Bellingham Bay 90/90 UTL for cPAH TEQ is above natural background by nearly a factor of 5. A regional background concentration of $86 \mu \mathrm{~g} \mathrm{TEQ} / \mathrm{kg}$ has been calculated.


### 6.0 References

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## Tables

Table 1. Sampling station depths, actual coordinates, distance from target, and percent fines for the Bellingham Bay Regional Background data set.

| Station <br> ID | Mudline <br> Depth (m <br> MLLW) | Northing <br> (SPN <br> NAD83) | Easting <br> (SPN <br> NAD83) | Latitude <br> (NAD83) | Longitude <br> (NAD83) | Distance <br> from <br> Target (m) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BB-01 | -9.7 | 641884.8 | 1234001.8 | 48.747430 | -122.517903 | 0.6 |
| BB-02 | -5.2 | 644345.9 | 1234823.4 | 48.754223 | -122.514722 | 0.2 |
| BB-03 | -6 | 641881.6 | 1237271.8 | 48.747617 | -122.504353 | 4 |
| BB-04 | -7.1 | 642707.1 | 1235645.5 | 48.749782 | -122.511167 | 0.6 |
| BB-05 | -9.1 | 642708.4 | 1233998.9 | 48.749687 | -122.517990 | 1.6 |
| BB-06 | -4.5 | 643526.6 | 1237283.8 | 48.752125 | -122.504452 | 0.3 |
| BB-07 | -4.1 | 644343.1 | 1235647.8 | 48.754265 | -122.511305 | 1.5 |
| BB-08 | -16.1 | 634503.1 | 1234822.8 | 48.727252 | -122.513833 | 0.3 |
| BB-09 | -12.5 | 636964.0 | 1236466.6 | 48.734093 | -122.507247 | 0.8 |
| BB-10 | -14.4 | 635325.1 | 1235644.7 | 48.729553 | -122.510503 | 0.4 |
| BB-11 | -11.4 | 639430.3 | 1234822.2 | 48.740753 | -122.514282 | 1.6 |
| BB-12 | -3.3 | 642706.4 | 1238922.3 | 48.749975 | -122.497588 | 0.8 |
| BB-13 | -19.5 | 636145.9 | 1233181.7 | 48.731655 | -122.520780 | 0.7 |
| BB-14 | -10.8 | 641066.1 | 1234000.0 | 48.745187 | -122.517837 | 1.1 |
| BB-15 | -22.2 | 634502.6 | 1233179.9 | 48.727152 | -122.520638 | 1.2 |
| BB-16 | -9.7 | 639421.9 | 1236461.7 | 48.740828 | -122.507488 | 1.2 |
| BB-17 | -13.5 | 637785.3 | 1235646.4 | 48.736295 | -122.510718 | 0.8 |
| BB-18 | -9.7 | 632857.0 | 1235642.8 | 48.722790 | -122.510288 | 2.1 |
| BB-19 | -9 | 636033.9 | 1237031.7 | 48.731578 | -122.504822 | 84 |
| BB-20 | -7.8 | 640241.8 | 1237273.6 | 48.743123 | -122.504198 | 3.5 |
| BB-21 | -16.7 | 637788.7 | 1233183.5 | 48.736157 | -122.520922 | 1.3 |
| BB-22 | -10.1 | 633683.6 | 1235642.7 | 48.725055 | -122.510363 | 0.4 |
| BB-23 | -6.9 | 639423.8 | 1238924.5 | 48.740980 | -122.497285 | 0.4 |
| BB-24 | -16.2 | 636145.8 | 1234825.7 | 48.731753 | -122.513970 | 0.8 |
| BB-25 | -6 | 641068.4 | 1238927.0 | 48.745487 | -122.497422 | 1.1 |
| BB-26 | -9.9 | 639426.5 | 1238099.5 | 48.740938 | -122.500703 | 1.6 |
| BB-27 | -9 | 641065.5 | 1235645.5 | 48.745283 | -122.511018 | 0.5 |
| BB-28 | -5.3 | 634504.4 | 1236467.0 | 48.727353 | -122.507023 | 0.9 |
| BB-29 | -9.8 | 638603.6 | 1237284.3 | 48.738635 | -122.504007 | 0.4 |
| BB-30 | -14.4 | 637786.9 | 1234821.3 | 48.736250 | -122.514137 | 1 |

## Notes:

MLLW - mean lower low water
SPN NAD - state plane north North American datum

Table 2. Collected sediment samples, target analytes, and analytical methods.

| Sampling <br> Location | Grain Size | TOC | Total Solids | Lead | cPAH | Dioxin/Furan Congeners | PCB <br> Congeners | Archive |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Method | PSEP | PSEP | PSEP | EPA 200.8 | LL SIM 8270 | EPA 1613B | EPA 1668A |  |
| BB-01-S | X | X | X | X | X | - | X | A |
| BB-02-S | X | X | X | X | X | - | X | A |
| BB-03-S | X | X | X | X | X | - | X | A |
| BB-04-S | X | X | X | X | X | - | X | A |
| BB-05-S | X | X | X | X | X | - | X | A |
| BB-06-S | X | X | X | X | X | - | X | A |
| BB-07-S | X | X | X | X | X | - | X | A |
| BB-08-S | X | X | X | X | X | X | X | A |
| BB-09-S | X | X | X | X | X | X | X | A |
| BB-10-S | X | X | X | X | X | X | X | A |
| BB-11-S | X | X | X | X | X | X | X | A |
| BB-12-S | X | X | X | X | X | X | X | A |
| BB-13-S | X | X | X | X | X | X | X | A |
| BB-14-S | X | X | X | X | X | X | X | A |
| BB-15-S | X | X | X | X | X | X | X | A |
| BB-16-S | X | X | X | X | X | X | X | A |
| BB-17-S | X | X | X | X | X | X | X | A |
| BB-18-S | X | X | X | X | X | X | X | A |
| BB-19-S | X | X | X | X | X | X | X | A |
| BB-20-S | X | X | X | X | X | X | X | A |
| BB-20-D | X | X | X | X | X | X | X | - |
| BB-20-T | X | X | X | - | - | - | - | - |
| BB-21-S | X | X | X | X | X | X | X | A |
| BB-22-S | X | X | X | X | X | X | X | A |
| BB-23-S | X | X | X | X | X | X | X | A |
| BB-24-S | X | X | X | X | X | X | X | A |
| BB-24-D | X | X | X | X | X | X | X | - |
| BB-24-T | X | X | X | - | - | - | - | - |
| BB-25-S | X | X | X | X | X | X | X | A |
| BB-26-S | X | X | X | X | X | X | X | A |
| BB-27-S | X | X | X | X | X | X | X | A |
| BB-28-S | X | X | X | X | X | X | X | A |
| BB-29-S | X | X | X | X | X | X | X | A |
| BB-30-S | X | X | X | X | X | X | X | A |

## Notes:

A - archive $\quad \mathrm{cPAH}$ - carcinogenic polycyclic aromatic hydrocarbons $\quad \mathrm{PCB}$ - polychlorinated biphenyl D - duplicate T - triplicate TOC - total organic carbon EPA - Environmental Protection Agency PSEP - Puget Sound Estuary Program

Table 3. Target analytes, methods, and practical quantitation limits.

| Analyte | Preparation Method | Analytical Method | PQL |
| :--- | :---: | :---: | :---: |
| Metals (mg/kg DW) | EPA 3050B/3051 | EPA 200.8 | 0.1 |
| Lead |  |  |  |

Carcinogenic PAHs ( $\mu \mathrm{g} / \mathrm{kg}$ DW)

| Benzo(a)pyrene | EPA 3546 | EPA 8270 SIM LL | 0.5 |
| :--- | :---: | :---: | ---: |
| Benz(a)anthracene | EPA 3546 | EPA 8270 SIM LL | 0.5 |
| Benzo(b)fluoranthene | EPA 3546 | EPA 8270 SIM LL | 0.5 |
| Benzo(k)fluoranthene | EPA 3546 | EPA 8270 SIM LL | 0.5 |
| Chrysene | EPA 3546 | EPA 8270 SIM LL | 0.5 |
| Indeno(1,2,3-cd)pyrene | EPA 3546 | EPA 8270 SIM LL | 0.5 |
| Dibenz(a,h)anthracene | EPA 3546 | EPA 8270 SIM LL | 0.5 |
| cPAH TEQ ${ }^{\text {a }}$ | -- | -- | 0.76 |


| PCB Congeners (ng/kg DW) |  |  |  |
| :---: | :---: | :---: | :---: |
| PCB 77 | EPA 1668A | EPA 1668 | 0.4 |
| PCB 81 | EPA 1668A | EPA 1668 | 0.4 |
| PCB 105 | EPA 1668A | EPA 1668 | 0.4 |
| PCB 114 | EPA 1668A | EPA 1668 | 0.4 |
| PCB 118 | EPA 1668A | EPA 1668 | 0.4 |
| PCB 123 | EPA 1668A | EPA 1668 | 0.4 |
| PCB 126 | EPA 1668A | EPA 1668 | 0.4 |
| PCB 156 | EPA 1668A | EPA 1668 | 0.8 |
| PCB 157 | EPA 1668A | EPA 1668 |  |
| PCB 167 | EPA 1668A | EPA 1668 | 0.4 |
| PCB 169 | EPA 1668A | EPA 1668 | 0.4 |
| PCB 189 | EPA 1668A | EPA 1668 | 0.4 |
| PCB Congener TEQ ${ }^{\text {a }}$ | -- | -- | 0.052 |

Dioxin/Furan Congeners (ng/kg DW)

| 2,3,7,8-TCDD | EPA 1613B/3540C | EPA 1613B | 0.2 |
| :---: | :---: | :---: | :---: |
| 1,2,3,7,8-PeCDD | EPA 1613B/3540C | EPA 1613B | 1 |
| 1,2,3,4,7,8-HxCDD | EPA 1613B/3540C | EPA 1613B | 1 |
| 1,2,3,6,7,8-HxCDD | EPA 1613B/3540C | EPA 1613B | 1 |
| 1,2,3,7,8,9-HxCDD | EPA 1613B/3540C | EPA 1613B | 1 |
| 1,2,3,4,6,7,8-HpCDD | EPA 1613B/3540C | EPA 1613B | 1 |
| OCDD | EPA 1613B/3540C | EPA 1613B | 2 |
| 2,3,7,8-TCDF | EPA 1613B/3540C | EPA 1613B | 0.2 |
| 1,2,3,7,8-PeCDF | EPA 1613B/3540C | EPA 1613B | 1 |
| 2,3,4,7,8-PeCDF | EPA 1613B/3540C | EPA 1613B | 1 |
| 1,2,3,4,7,8-HxCDF | EPA 1613B/3540C | EPA 1613B | 1 |
| 1,2,3,6,7,8-HxCDF | EPA 1613B/3540C | EPA 1613B | 1 |
| 1,2,3,7,8,9-HxCDF | EPA 1613B/3540C | EPA 1613B | 1 |


| Analyte | Preparation Method | Analytical Method | PQL |
| :--- | :---: | :---: | :---: |
| $2,3,4,6,7,8-\mathrm{HxCDF}$ | EPA $1613 \mathrm{~B} / 3540 \mathrm{C}$ | EPA 1613B |  |
| $1,2,3,4,6,7,8-\mathrm{HpCDF}$ | EPA $1613 \mathrm{~B} / 3540 \mathrm{C}$ | EPA 1613B | 1 |
| $1,2,3,4,7,8,9-\mathrm{HpCDF}$ | EPA $1613 \mathrm{~B} / 3540 \mathrm{C}$ | EPA 1613 B | 1 |
| OCDF | EPA $1613 \mathrm{~B} / 3540 \mathrm{C}$ | EPA 1613 B | 1 |
| Dioxin/Furan TEQ $^{\text {a }}$ | -- | -- | 2 |

Notes:
DW - dry weight TEQ - toxicity equivalent quotient PQL - practical quantitation limit
PAH - polycyclic aromatic hydrocarbon $\quad \mathrm{PCB}$ - polychlorinated biphenyl
SIM - selected ion monitoring
Rose highlighting indicates the project specific PQL
a. TEQ values were calculated by multiplying the PQL by the appropriate TEF.

Table 4. Summary statistics and correlation to total organic carbon (TOC) for target analytes.

| Location ID | Lead | $\text { cPAH TEQ }{ }^{\text {a }}$ | Dx/F TEQ ${ }^{\text {a }}$ | PCB TEQ ${ }^{\text {a }}$ |
| :---: | :---: | :---: | :---: | :---: |
| Units | mg/kg | Hg TEQ/kg | ng TEQ/kg | ng TEQ/kg |
|  |  |  |  |  |
| Sample Size | $28^{\text {b }}$ | $28^{\text {c }}$ | $20^{\text {d }}$ | $28^{\text {e }}$ |
| Minimum | 8.4 | 7.4 | 4.2 | 0.028 |
| Average | 12 | 30 | 9.3 | 0.098 |
| Median | 11 | 21 | 9.9 | 0.086 |
| Maximum | 16 | 94 | 14 | 0.25 |
|  |  |  |  |  |
| r-value | 0.818 | 0.749 | 0.288 | 0.517 |
| p-value | $<0.0001$ | $<0.0001$ | 0.218 | 0.005 |

Notes:
cPAH - carcinogenic polycyclic aromatic hydrocarbon $\quad \mathrm{Dx} / \mathrm{F}$ - dioxin/furan congener PCB - polychlorinated biphenyl
a. Toxicity equivalency quotient - calculated as described in Section 4.1.
b. Excludes BB-12 due to elevated TOC and BB-28 due to an elevated lead concentration of $53 \mathrm{mg} / \mathrm{kg}$ (Section 5.3).
c. Excludes BB-12 due to elevated TOC and BB-28 due to an elevated cPAH concentration of $2,050 \mu \mathrm{~g} / \mathrm{kg}$ (Section 5.3).
d. Excludes BB-12 due to elevated TOC and BB-14 and BB-27 due to concentrations consistent with natural background (Section 5.3).
e. Excludes BB-12 due to elevated TOC and BB-28 due to an elevated PCB TEQ concentration of $0.3 \mathrm{ng} \mathrm{TEQ} / \mathrm{kg}$ (Section 5.3).

Table 5. Statistical summary of the Bellingham Bay data set.

| Parameter | Units | $\mathbf{N}$ | \% <br> Detect | CV | Precision $^{\mathbf{a}}$ | Distribution $^{\mathbf{b}, \mathbf{c}}$ | Samples $_{\text {Excluded }^{\mathbf{d}}}$ <br> Lead mg/kg |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| cPAH | $\mu \mathrm{g} \mathrm{TEQ/kg}$ | 28 | $100 \%$ | 0.19 | $6 \%$ | Nonparametric | $\mathrm{BB}-12, \mathrm{BB}-28$ |
| Dioxins/Furans | ng TEQ $/ \mathrm{kg}$ | 20 | $100 \%$ | 0.80 | $30 \%$ | Nonparametric | $\mathrm{BB}-12, \mathrm{BB}-28$ |
| PCB congener | ng TEQ/kg | 28 | $96 \%$ | 0.60 | $23 \%$ | Gamma | BB-12, BB-28 |

Notes:
cPAH - carcinogenic polycyclic aromatic hydrocarbon $\quad \mathrm{PCB}$ - polychlorinated biphenyl $\quad \mathrm{N}$ - sample size $\quad \mathrm{CV}$ - coefficient of variance a. The precision column shows the half-width of the $95 \% \mathrm{UCL}$ on the mean relative to the mean (e.g., for a normal distribution, $\mathrm{t} \times$ (std.dev./sqrt(n))/mean); the target was precision $\leq 25 \%$.
b. Outlier tests included Rosner's (for normal distributions, $n \geq 25$ ), or Tukey's rule of $2 \times I Q R$ from the median (non-parametric). No samples were identified as outliers after excluding samples identified in the last column.
c. Best fit distribution as determined by the goodness-of-fit tests in ProUCL and the highest correlation coefficient for the probability plots (detected concentrations only).
d. Samples were removed from the data set prior to the statistical evaluations summarized here (see Section 5.3 for further details regarding exclusion of these samples).

Table 6. Calculated 90/90 upper tolerance limits (UTLs) for the Bellingham Bay and Bold Plus natural background data sets.

| Analyte | Units | Bellingham Bay |  | Bold Plus |
| :---: | :---: | :---: | :---: | :---: |
|  |  | N | 90/90 UTL $^{\text {a }}$ | 90/90 UTL ${ }^{\text {a }}$ |
| Lead | $\mathrm{mg} / \mathrm{kg}$ | 28 | 16 | 21 |
| cPAH TEQ | $\mu \mathrm{g} \mathrm{TEQ/kg}$ | 28 | 86 | 16 |
| PCB TEQ | ng TEQ/kg | 28 | 0.21 | 0.20 |
| Dioxin/Furan TEQ | ng TEQ/kg | 20 | 15 | 3.6 |

Notes:
cPAH - carcinogenic polycyclic aromatic hydrocarbon $\quad \mathrm{PCB}$ - polychlorinated biphenyl $\quad \mathrm{TEQ}$ - toxicity equivalency quotient $\quad \mathrm{N}-$ sample size
a. All values rounded to two significant figures.

Figures




Figure 3. Summary of undetected and qualified results.


Figure 4. Total organic carbon and percent fines throughout the Bellingham Bay area of interest.

Washington State Plane North Coordinate System NAD 83 Datum
Date: $1 / 9 / 2015$ Date: 1/9/2015




Figure 7. Dioxin/furan congener TEQ concentrations (KM) throughout the Bellingham Bay area of interest.

Washington State Plane North Coordinate System NAD 83 Datum
Date: $1 / 9 / 2015$ Date: 1/9/2015



Figure 8. Polychlorinated biphenyl (PCB) congener TEQ concentrations (KM) throughout the Bellingham

Washington State Plane North Coordinate System NAD 83 Datum Date: 1/9/2015


Figure 9. Boxplots showing the distribution of analyte concentrations in Bellingham Bay sediments and Puget Sound wide natural background (Bold Plus).
Note: The 1st, 2nd, and 3rd quartiles are estimated using Kaplan-Meier for censored data; horizontal lines indicate the level of the highest detection limit. Note the break in the scale of the $y$-axis to accommodate the single extreme cPAH value in Bellingham Bay.

## Appendices A-E. Available Upon Request

## Appendix F. Data Tables

Table F-1. Summary of Bellingham Bay Sediment Conventionals, Metals, and Carcinogenic Polycyclic Aromatic Hydrocarbons.

| Location ID | BB-01-S | Q | BB-02-S | Q | BB-03-S | Q | BB-04-S | Q | BB-05-S | Q | BB-06-S | Q | BB-07-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/11/2014 | 834 | 9/10/2014 | 1135 | 9/10/2014 | 1512 | 9/10/2014 | 1438 | 9/10/2014 | 1410 | 9/10/2014 | 1115 | 9/10/2014 | 1356 |
| Conventionals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Organic Carbon | 1.22 |  | 1.3 |  | 1.5 |  | 0.984 |  | 1.11 |  | 1.53 |  | 0.967 |  |
| Total Solids | 40.73 |  | 54.68 |  | 45.98 |  | 42.02 |  | 44.47 |  | 49.29 |  | 55.83 |  |
| Particle/Grain Size, Phi Scale <-1 | 0.1 | U | 0.1 | U | 0.2 |  | 0.1 | U | 0.1 | U | 0.2 |  | 0.1 |  |
| Particle/Grain Size, Phi Scale -1 to 0 | 0.4 |  | 0.2 |  | 0.1 |  | 0.6 |  | 0.1 |  | 0.1 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 0 to 1 | 0.5 |  | 0.2 |  | 0.2 |  | 0.6 |  | 0.2 |  | 0.2 |  | 0.4 |  |
| Particle/Grain Size, Phi Scale 1 to 2 | 0.5 |  | 0.2 |  | 0.3 |  | 0.4 |  | 0.1 |  | 0.2 |  | 0.3 |  |
| Particle/Grain Size, Phi Scale 2 to 3 | 0.3 |  | 0.1 |  | 0.3 |  | 0.3 |  | 0.1 |  | 0.1 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 3 to 4 | 0.5 |  | 1.2 |  | 0.5 |  | 1.1 |  | 0.3 |  | 0.1 |  | 0.6 |  |
| Particle/Grain Size, Phi Scale 4 to 5 | 4.7 |  | 13.9 |  | 6.4 |  | 5 |  | 5.4 |  | 5.8 |  | 11.6 |  |
| Particle/Grain Size, Phi Scale 5 to 6 | 31.3 |  | 26.4 |  | 18.8 |  | 22.4 |  | 19.8 |  | 17.4 |  | 27.9 |  |
| Particle/Grain Size, Phi Scale 6 to 7 | 24.2 |  | 20.5 |  | 21.9 |  | 23.3 |  | 27.2 |  | 23.3 |  | 21.7 |  |
| Particle/Grain Size, Phi Scale 7 to 8 | 8.9 |  | 11.8 |  | 15.5 |  | 14.3 |  | 13.1 |  | 16.2 |  | 11.7 |  |
| Particle/Grain Size, Phi Scale 8 to 9 | 6.7 |  | 7.2 |  | 10.6 |  | 9 |  | 9.2 |  | 10.5 |  | 7.4 |  |
| Particle/Grain Size, Phi Scale 9 to 10 | 7.4 |  | 5.7 |  | 8.2 |  | 7.9 |  | 8.3 |  | 7.3 |  | 5.3 |  |
| Particle/Grain Size, Phi Scale >10 | 14.4 |  | 12.5 |  | 17.1 |  | 15.2 |  | 16.2 |  | 18.5 |  | 12.6 |  |
| Particle/Grain Size, Fines (Silt/Clay) | 97.7 |  | 98.1 |  | 98.5 |  | 97 |  | 99.2 |  | 99 |  | 98.2 |  |
| Metals (mg/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lead | 10.1 |  | 8.4 |  | 10.6 |  | 9 |  | 9.6 |  | 10.9 |  | 8.4 |  |
| carcinogenic PAH (ug/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Benzo(a)anthracene | 9.14 |  | 6.37 | J | 25.7 |  | 14.1 |  | 10.8 |  | 13.9 | J | 15.4 |  |
| Chrysene | 21.6 |  | 14.4 | J | 43 |  | 27.6 |  | 26.6 |  | 23.7 | J | 51 |  |
| Benzo(b)fluoranthene | 15.2 |  | 10.2 | J | 28.9 |  | 19.2 |  | 17.3 |  | 16.6 | J | 21.5 |  |
| Benzo(k)fluoranthene | 4.29 |  | 3.17 | J | 10.3 |  | 6.24 |  | 5.45 |  | 6.15 | J | 7.49 |  |
| Benzo(a)pyrene | 7.65 |  | 4.87 | J | 19 |  | 9.64 |  | 8.79 |  | 11.1 | J | 10.9 |  |
| Indeno(1,2,3-cd)pyrene | 6.17 |  | 3.67 | J | 14.8 |  | 8.1 |  | 7.48 |  | 8.77 | J | 8.61 |  |
| Dibenz(a,h)anthracene | 2.41 | U | 2.42 | UJ | 4.95 | U | 4.96 | U | 4.82 | U | 4.86 | UJ | 4.87 | U |
| cPAH TEQ (0 DL) | 11.3 |  | 7.36 |  | 27.4 |  | 14.7 |  | 13.2 |  | 15.9 |  | 16.7 |  |
| cPAH TEQ (1/2 DL) | 11.4 |  | 7.40 |  | 27.5 |  | 14.8 |  | 13.2 |  | 16.0 |  | 16.8 |  |
| cPAH TEQ (1 DL) | 11.4 |  | 7.44 |  | 27.6 |  | 14.9 |  | 13.3 |  | 16.1 |  | 16.9 |  |
| cPAH TEQ (KM) | 11.4 |  | 7.44 |  | 27.6 |  | 14.9 |  | 13.3 |  | 16.1 |  | 16.9 |  |

U-the analytie was analyzed for, but was not detected above the reported sample quantitation limit
$j$-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sample
UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximate.
L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detects
KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicates

Table F-1. Summary of Bellingham Bay Sediment Conventionals, Metals, and Carcinogenic Polycyclic Aromatic Hydrocarbons (continued).

| Location ID | BB-08-S | Q | BB-09-S | Q | BB-10-S | Q | BB-11-S | Q | BB-12-S | Q | BB-13-S | Q | BB-14-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 933 | 9/9/2014 | 1540 | 9/9/2014 | 1147 | 9/11/2014 | 1024 | 9/10/2014 | 1330 | 9/9/2014 | 1423 | 9/11/2014 | 853 |
| Conventionals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Organic Carbon | 1.72 | J | 1.12 | J | 1.87 | J | 1.28 |  | 20.8 |  | 1.11 | J | 1.4 |  |
| Total Solids | 31.45 |  | 34.47 |  | 37.96 |  | 33.33 |  | 40.37 |  | 32.05 |  | 38.4 |  |
| Particle/Grain Size, Phi Scale <-1 | 0.1 |  | 0.1 |  | 0.2 |  | 0.1 |  | 7.4 |  | 0.1 | U | 0.1 | U |
| Particle/Grain Size, Phi Scale -1 to 0 | 3.4 |  | 0.1 |  | 0.1 | U | 0.2 |  | 4.6 |  | 0.1 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 0 to 1 | 2.1 |  | 0.2 |  | 0.1 |  | 0.4 |  | 9.2 |  | 0.1 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 1 to 2 | 0.9 |  | 0.2 |  | 0.1 |  | 0.4 |  | 14.9 |  | 0.1 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 2 to 3 | 0.6 |  | 0.2 |  | 0.1 |  | 0.2 |  | 4.5 |  | 0.1 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 3 to 4 | 0.8 |  | 0.2 |  | 0.1 |  | 0.3 |  | 4.2 |  | 0.1 |  | 0.4 |  |
| Particle/Grain Size, Phi Scale 4 to 5 | 2.5 |  | 4.1 |  | 2.7 |  | 3.6 |  | 3 |  | 2.4 |  | 5.1 |  |
| Particle/Grain Size, Phi Scale 5 to 6 | 9.7 |  | 15.4 |  | 10.5 |  | 52.2 |  | 11.4 |  | 6.6 |  | 29.8 |  |
| Particle/Grain Size, Phi Scale 6 to 7 | 17.9 |  | 22.9 |  | 18.1 |  | 14 |  | 11.7 |  | 17.6 |  | 27.6 |  |
| Particle/Grain Size, Phi Scale 7 to 8 | 16.4 |  | 14.1 |  | 18.6 |  | 6.9 |  | 7.2 |  | 20.2 |  | 9.1 |  |
| Particle/Grain Size, Phi Scale 8 to 9 | 14.8 |  | 12.8 |  | 15 |  | 4.2 |  | 5.7 |  | 16.2 |  | 6.5 |  |
| Particle/Grain Size, Phi Scale 9 to 10 | 9.5 |  | 9.9 |  | 12.1 |  | 4.9 |  | 5.2 |  | 12.7 |  | 7.1 |  |
| Particle/Grain Size, Phi Scale $>10$ | 21.2 |  | 19.9 |  | 22.5 |  | 12.6 |  | 10.9 |  | 23.9 |  | 13.7 |  |
| Particle/Grain Size, Fines (Silt/Clay) | 92 |  | 99.1 |  | 99.4 |  | 98.3 |  | 55.2 |  | 99.6 |  | 98.8 |  |
| Metals (mg/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lead | 12.5 |  | 10.2 |  | 12 |  | 10.6 |  | 13.2 |  | 11.2 |  | 10.5 |  |
| carcinogenic PAH (ug/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Benzo(a)anthracene | 15 | J | 15.6 |  | 26.9 | J | 15.5 |  | 31.9 | J | 8.34 |  | 11.8 |  |
| Chrysene | 31.1 | J | 27.6 |  | 41.4 | J | 28.5 |  | 45.6 | J | 14.9 |  | 26 |  |
| Benzo(b)fluoranthene | 16.2 | J | 17.3 |  | 29.2 | J | 21.2 |  | 31.1 | J | 10.2 |  | 17.7 |  |
| Benzo(k)fluoranthene | 7.25 | J | 7.5 |  | 12.1 | J | 6.9 |  | 16.4 | J | 3.81 |  | 5.19 |  |
| Benzo(a)pyrene | 15.5 | J | 14.9 |  | 28 | J | 13.4 |  | 27.2 | J | 8.83 |  | 9.79 |  |
| Indeno(1,2,3-cd)pyrene | 9.39 | J | 11.8 |  | 20.5 | J | 12.5 |  | 19.1 | J | 6.65 |  | 8.57 |  |
| Dibenz(a,h)anthracene | 4.85 | UJ | 4.9 | U | 9.7 | UJ | 4.94 | U | 4.86 | UJ | 2.42 | U | 4.87 | U |
| cPAH TEQ (0 DL) | 20.6 |  | 20.4 |  | 37.3 |  | 19.3 |  | 37.5 |  | 11.9 |  | 14.4 |  |
| cPAH TEQ (1/2 DL) | 20.7 |  | 20.5 |  | 37.5 |  | 19.4 |  | 37.6 |  | 11.9 |  | 14.5 |  |
| cPAH TEQ (1 DL) | 20.8 |  | 20.6 |  | 37.6 |  | 19.5 |  | 37.7 |  | 12.0 |  | 14.5 |  |
| CPAH TEQ (KM) | 20.8 |  | 20.6 |  | 37.6 |  | 19.5 |  | 37.7 |  | 12.0 |  | 14.5 |  |

U-the analytie was analyzed for, but was not detected above the reported sample quantitation limit
$J$-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sample
UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximate.
L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detect
KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicates

Table F-1. Summary of Bellingham Bay Sediment Conventionals, Metals, and Carcinogenic Polycyclic Aromatic Hydrocarbons (continued).

| Location ID | BB-15-S | Q | BB-16-S | Q | BB-17-S | Q | BB-18-S | Q | BB-19-S | Q | BB-20-S | Q | BB-20-D | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 915 | 9/11/2014 | 1044 | 9/9/2014 | 1524 | 9/9/2014 | 1042 | 9/9/2014 | 1326 | 9/10/2014 | 1530 | 9/10/2014 | 1530 |
| Conventionals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Organic Carbon | 1.87 | J | 1.41 |  | 1.27 | J | 3.3 | J | 2.87 | J | 1.53 |  | 1.39 |  |
| Total Solids | 27.6 |  | 42.49 |  | 37.95 |  | 33.8 |  | 35.65 |  | 44.51 |  | 44.59 |  |
| Particle/Grain Size, Phi Scale <-1 | 0.3 |  | 3.9 |  | 0.1 | U | 0.2 |  | 0.1 | U | 0.1 | U | 0.2 |  |
| Particle/Grain Size, Phi Scale -1 to 0 | 1.2 |  | 0.2 |  | 0.1 |  | 3.5 |  | 0.5 |  | 0.1 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 0 to 1 | 1.2 |  | 0.3 |  | 0.1 |  | 3 |  | 1.3 |  | 0.2 |  | 0.1 |  |
| Particle/Grain Size, Phi Scale 1 to 2 | 1 |  | 0.3 |  | 0.2 |  | 1.3 |  | 1 |  | 0.2 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 2 to 3 | 1 |  | 0.3 |  | 0.1 |  | 1 |  | 0.7 |  | 0.1 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 3 to 4 | 0.7 |  | 0.5 |  | 0.1 |  | 1.4 |  | 0.8 |  | 0.3 |  | 0.3 |  |
| Particle/Grain Size, Phi Scale 4 to 5 | 6.4 |  | 4.2 |  | 3.2 |  | 2.8 |  | 2.5 |  | 6 |  | 3.2 |  |
| Particle/Grain Size, Phi Scale 5 to 6 | 1.8 |  | 19.4 |  | 11.7 |  | 10.2 |  | 15.3 |  | 13.6 |  | 15 |  |
| Particle/Grain Size, Phi Scale 6 to 7 | 17.9 |  | 17.8 |  | 19.5 |  | 19.7 |  | 18.7 |  | 20.8 |  | 20.8 |  |
| Particle/Grain Size, Phi Scale 7 to 8 | 17.7 |  | 15.8 |  | 17.5 |  | 15.1 |  | 14.3 |  | 16.8 |  | 17.6 |  |
| Particle/Grain Size, Phi Scale 8 to 9 | 15.5 |  | 10.6 |  | 14.6 |  | 12.3 |  | 13 |  | 12.3 |  | 12.3 |  |
| Particle/Grain Size, Phi Scale 9 to 10 | 11.6 |  | 9.5 |  | 11.3 |  | 9.8 |  | 10.5 |  | 10 |  | 9.5 |  |
| Particle/Grain Size, Phi Scale >10 | 23.7 |  | 17 |  | 21.5 |  | 19.8 |  | 21.4 |  | 19.7 |  | 20.6 |  |
| Particle/Grain Size, Fines (Silt/Clay) | 94.6 |  | 94.4 |  | 99.3 |  | 89.6 |  | 95.7 |  | 99.2 |  | 98.9 |  |
| Metals (mg/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lead | 11.6 |  | 11.2 |  | 11.3 |  | 15.7 |  | 15.4 |  | 11 |  | 11.1 |  |
| carcinogenic PAH (ug/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Benzo(a)anthracene | 8.3 | J | 20.7 |  | 15.2 |  | 61.5 |  | 41.1 | J | 20.2 | J | 22.3 |  |
| Chrysene | 13.6 | J | 36.8 |  | 26.4 |  | 86.4 |  | 61.6 | J | 31.1 | J | 39.8 |  |
| Benzo(b)fluoranthene | 9.77 | J | 23.4 |  | 16.2 |  | 56.6 |  | 38.4 | J | 22.8 | J | 27.3 |  |
| Benzo(k)fluoranthene | 4.12 | J | 9.83 |  | 8.32 |  | 32.5 |  | 19.3 | J | 8.23 | J | 9.77 |  |
| Benzo(a)pyrene | 9.4 | J | 17.5 |  | 13.6 |  | 65.2 |  | 41.5 | J | 14.2 | J | 17.9 |  |
| Indeno(1,2,3-cd)pyrene | 7.21 | J | 15.4 |  | 10 |  | 42.7 |  | 26.6 | J | 11.1 | J | 14.4 |  |
| Dibenz(a,h)anthracene | 2.46 | UJ | 4.87 | U | 4.85 | U | 9.22 |  | 5.1 | J | 4.84 | UJ | 4.9 | U |
| CPAH TEQ (0 DL) | 12.5 |  | 24.8 |  | 18.8 |  | 86.3 |  | 55.2 |  | 20.7 |  | 25.7 |  |
| cPAH TEQ (1/2 DL) | 12.5 |  | 24.9 |  | 18.9 |  | 86.3 |  | 55.2 |  | 20.8 |  | 25.8 |  |
| cPAH TEQ (1 DL) | 12.6 |  | 25.0 |  | 19.0 |  | 86.3 |  | 55.2 |  | 20.9 |  | 25.9 |  |
| CPAH TEQ (KM) | 12.6 |  | 25.0 |  | 19.0 |  | 86.3 |  | 55.2 |  | 23.4 |  | -- |  |

U-the analytie was analyzed for, but was not detected above the reported sample quantitation limit
j-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sample
UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximate.
L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detect
KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicates

Table F-1. Summary of Bellingham Bay Sediment Conventionals, Metals, and Carcinogenic Polycyclic Aromatic Hydrocarbons (continued).

| Location ID | BB-20-T | Q | BB-21-S | Q | BB-22-S | Q | BB-23-S | Q | BB-24-S | Q | BB-24-D | Q | BB-24-T | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/10/2014 | 1530 | 9/9/2014 | 1453 | 9/9/2014 | 1026 | 9/10/2014 | 942 | 9/9/2014 | 1353 | 9/9/2014 | 1353 | 9/9/2014 | 1353 |
| Conventionals |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Organic Carbon | 1.45 |  | 1.86 | J | 2.09 | J | 1.88 |  | 1.62 | J | 1.45 | J | 1.48 | J |
| Total Solids | 43.93 |  | 36.95 |  | 40.07 |  | 40.87 |  | 36.96 |  | 37.08 |  | 36.88 |  |
| Particle/Grain Size, Phi Scale <-1 | 0.1 | U | 0.1 | U | 0.5 |  | 3.4 |  | 2.3 |  | 0.9 |  | 0.1 |  |
| Particle/Grain Size, Phi Scale -1 to 0 | 0.1 | U | 0.1 | U | 1.8 |  | 0.8 |  | 0.1 | U | 0.1 | U | 0.1 | U |
| Particle/Grain Size, Phi Scale 0 to 1 | 0.1 |  | 0.1 |  | 2.4 |  | 0.7 |  | 0.1 |  | 0.2 |  | 0.1 |  |
| Particle/Grain Size, Phi Scale 1 to 2 | 0.1 |  | 0.1 |  | 1.8 |  | 0.5 |  | 0.1 |  | 0.1 |  | 0.1 |  |
| Particle/Grain Size, Phi Scale 2 to 3 | 0.1 |  | 0.1 | U | 2.2 |  | 0.4 |  | 0.2 |  | 0.1 |  | 0.1 |  |
| Particle/Grain Size, Phi Scale 3 to 4 | 0.2 |  | 0.1 |  | 1.7 |  | 0.5 |  | 0.1 |  | 0.1 |  | 0.1 |  |
| Particle/Grain Size, Phi Scale 4 to 5 | 4.2 |  | 0.8 |  | 2.8 |  | 2.5 |  | 5.8 |  | 3.4 |  | 1.6 |  |
| Particle/Grain Size, Phi Scale 5 to 6 | 15.3 |  | 8.4 |  | 11.1 |  | 12.3 |  | 9.6 |  | 8.8 |  | 8.8 |  |
| Particle/Grain Size, Phi Scale 6 to 7 | 21.3 |  | 18.7 |  | 16.9 |  | 19.5 |  | 16.9 |  | 17.6 |  | 18 |  |
| Particle/Grain Size, Phi Scale 7 to 8 | 16.3 |  | 19.7 |  | 14.7 |  | 16.7 |  | 17.9 |  | 18.5 |  | 18.2 |  |
| Particle/Grain Size, Phi Scale 8 to 9 | 11.9 |  | 16.1 |  | 12.9 |  | 12.7 |  | 14.2 |  | 15.4 |  | 16.8 |  |
| Particle/Grain Size, Phi Scale 9 to 10 | 10.6 |  | 12.5 |  | 9.7 |  | 10.4 |  | 10.9 |  | 11.4 |  | 12.4 |  |
| Particle/Grain Size, Phi Scale >10 | 19.9 |  | 23.3 |  | 21.6 |  | 19.8 |  | 21.8 |  | 23.4 |  | 24 |  |
| Particle/Grain Size, Fines (Silt/Clay) | 99.5 |  | 99.6 |  | 89.6 |  | 93.8 |  | 97.2 |  | 98.6 |  | 99.5 |  |
| Metals (mg/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Lead | -- |  | 10.7 |  | 15.5 |  | 16 |  | 11.3 |  | 11 |  | -- |  |
| carcinogenic PAH (ug/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Benzo(a)anthracene | -- |  | 8.68 |  | 61.9 | J | 60.1 | J | 16.9 | J | 19.3 |  | -- |  |
| Chrysene | -- |  | 15.5 |  | 87.5 | J | 112 | J | 26.6 | J | 33.8 |  | -- |  |
| Benzo(b)fluoranthene | -- |  | 11.1 |  | 63.6 | J | 43.3 | J | 17.9 | J | 20.5 |  | -- |  |
| Benzo(k)fluoranthene | -- |  | 3.9 |  | 34.8 | J | 21.6 | J | 8.53 | J | 9.81 |  | -- |  |
| Benzo(a)pyrene | -- |  | 8.49 |  | 70.7 | J | 38.7 | J | 17.4 | J | 19.6 |  | -- |  |
| Indeno(1,2,3-cd)pyrene | -- |  | 7.59 |  | 51.4 | J | 23.9 | J | 14.2 | J | 15.6 |  | -- |  |
| Dibenz(a,h)anthracene | -- |  | 2.45 | U | 9.59 | J | 5.57 | J | 4.8 | UJ | 4.81 | U | -- |  |
| cPAH TEQ (0 DL) | -- |  | 11.8 |  | 93.7 |  | 55.3 |  | 23.4 |  | 26.5 |  | -- |  |
| cPAH TEQ (1/2 DL) | -- |  | 11.8 |  | 93.7 |  | 55.3 |  | 23.5 |  | 26.5 |  | -- |  |
| cPAH TEQ (1 DL) | -- |  | 11.9 |  | 93.7 |  | 55.3 |  | 23.6 |  | 26.6 |  | -- |  |
| CPAH TEQ (KM) | -- |  | 11.9 |  | 93.7 |  | 55.3 |  | 25.1 | Avg | -- |  | -- |  |

U-the analytie was analyzed for, but was not detected above the reported sample quantitation limit
J-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sample
UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximate.
L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detect
KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicates

Table F-1. Summary of Bellingham Bay Sediment Conventionals, Metals, and Carcinogenic Polycyclic Aromatic Hydrocarbons (continued).

| Location ID | BB-25-S | Q | BB-26-S | Q | BB-27-S | Q | BB-28-S | Q | BB-29-S | Q | BB-30-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/11/2014 | 950 | 9/10/2014 | 1012 | 9/11/2014 | 923 | 9/9/2014 | 1129 | 9/11/2014 | 1103 | 9/9/2014 | 1507 |
| Conventionals |  |  |  |  |  |  |  |  |  |  |  |  |
| Total Organic Carbon | 2.13 |  | 2.09 |  | 1.6 |  | 3.68 | J | 1.51 |  | 1.59 | J |
| Total Solids | 42.7 |  | 38.15 |  | 45 |  | 36.73 |  | 37.11 |  | 35.61 |  |
| Particle/Grain Size, Phi Scale <-1 | 0.1 | U | 2.8 |  | 0.3 |  | 1 |  | 0.1 | U | 0.1 |  |
| Particle/Grain Size, Phi Scale -1 to 0 | 0.5 |  | 0.2 |  | 0.1 |  | 2.5 |  | 0.1 |  | 0.1 | U |
| Particle/Grain Size, Phi Scale 0 to 1 | 0.7 |  | 0.4 |  | 0.2 |  | 3.1 |  | 0.2 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 1 to 2 | 0.5 |  | 0.4 |  | 0.2 |  | 2.7 |  | 0.3 |  | 0.3 |  |
| Particle/Grain Size, Phi Scale 2 to 3 | 0.6 |  | 0.4 |  | 0.2 |  | 2.7 |  | 0.3 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 3 to 4 | 1.1 |  | 0.7 |  | 0.4 |  | 3.8 |  | 0.4 |  | 0.2 |  |
| Particle/Grain Size, Phi Scale 4 to 5 | 3.9 |  | 3.5 |  | 5.1 |  | 3.7 |  | 2.9 |  | 2.3 |  |
| Particle/Grain Size, Phi Scale 5 to 6 | 17.4 |  | 14.7 |  | 15.2 |  | 11 |  | 32.2 |  | 12.4 |  |
| Particle/Grain Size, Phi Scale 6 to 7 | 20.1 |  | 22.4 |  | 21.9 |  | 17.4 |  | 20 |  | 24.7 |  |
| Particle/Grain Size, Phi Scale 7 to 8 | 15 |  | 14.5 |  | 16.9 |  | 11.9 |  | 12.1 |  | 15.6 |  |
| Particle/Grain Size, Phi Scale 8 to 9 | 12 |  | 11.1 |  | 11 |  | 12.7 |  | 7.2 |  | 12.9 |  |
| Particle/Grain Size, Phi Scale 9 to 10 | 9.1 |  | 9.9 |  | 9.7 |  | 8.3 |  | 7.9 |  | 10.7 |  |
| Particle/Grain Size, Phi Scale >10 | 19.2 |  | 18.8 |  | 19 |  | 19.3 |  | 16.4 |  | 20.4 |  |
| Particle/Grain Size, Fines (Silt/Clay) | 96.6 |  | 95 |  | 98.7 |  | 84.2 |  | 98.8 |  | 99 |  |
| Metals (mg/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |
| Lead | 15.2 |  | 12.8 |  | 9.8 |  | 53 |  | 12.3 |  | 11.1 |  |
| carcinogenic PAH (ug/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |
| Benzo(a)anthracene | 82.5 |  | 34.3 | J | 14.7 |  | 824 |  | 28.6 |  | 15.9 |  |
| Chrysene | 107 |  | 52.5 | J | 29.5 |  | 1430 |  | 43.9 |  | 31.6 |  |
| Benzo(b)fluoranthene | 77.5 |  | 33.3 | J | 20.7 |  | 1310 |  | 30.2 |  | 16.7 |  |
| Benzo(k)fluoranthene | 37.7 |  | 14.1 | J | 6.32 |  | 678 |  | 13.9 |  | 8.36 |  |
| Benzo(a)pyrene | 60.8 |  | 27 | J | 11.5 |  | 1620 |  | 23.5 |  | 14.4 |  |
| Indeno(1,2,3-cd)pyrene | 43.1 |  | 20 | J | 11.1 |  | 1070 |  | 21.1 |  | 10.4 |  |
| Dibenz(a,h)anthracene | 8.19 |  | 4.96 | UJ | 4.9 | U | 233 |  | 4.93 | U | 4.87 | U |
| cPAH TEQ (0 DL) | 86.8 |  | 37.7 |  | 17.1 |  | 2050 |  | 33.3 |  | 19.9 |  |
| cPAH TEQ (1/2 DL) | 86.8 |  | 37.8 |  | 17.2 |  | 2050 |  | 33.4 |  | 19.9 |  |
| cPAH TEQ (1 DL) | 86.8 |  | 37.9 |  | 17.3 |  | 2050 |  | 33.5 |  | 20.0 |  |
| cPAH TEQ (KM) | 86.8 |  | 37.9 |  | 17.3 |  | 2050 |  | 33.5 |  | 20.0 |  |

U-the analytie was analyzed for, but was not detected above the reported sample quantitation limit
$J$-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sample
UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximate.
L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detects
KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicates

Table F-2. Summary of Bellingham Bay Dioxin/Furan Congener Data.
 ers (ng/kg DW)

| 2,3,7,8-TCDD | 1.1 | 0.866 | 1.07 |  | 0.334 |  | 0.501 | 0.944 |  | 0.156 | J | 0.623 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1,2,3,7,8-PECDD | 2.94 | 2.13 | 2.28 | J | 0.958 | J | 1.22 | 1.51 |  | 0.473 | J | 1.01 | U |
| 1,2,3,4,7,8-HXCDD | 12.2 | 8.5 | 9.59 |  | 3.17 |  | 2.99 | 4.97 |  | 0.947 | J | 4.48 |  |
| 1,2,3,6,7,8-HXCDD | 17.6 | 12.4 | 13 |  | 5.67 |  | 10.9 | 8.1 |  | 2.57 |  | 7.35 |  |
| 1,2,3,7,8,9-HXCDD | 16.1 | 11.4 | 9.94 |  | 5.01 |  | 6.34 | 7.64 |  | 2.21 |  | 7.13 |  |
| 1,2,3,4,6,7,8-HPCDD | 188 | 160 | 153 |  | 74.8 |  | 214 | 90.9 |  | 41.7 |  | 86.8 |  |
| OCDD | 892 | 924 | 770 |  | 461 |  | 1850 | 507 |  | 327 |  | 503 |  |
| 2,3,7,8-TCDF | 16.3 | 13.8 | 17.6 |  | 4.86 |  | 5.98 | 10.5 |  | 1.83 |  | 6.62 |  |
| 1,2,3,7,8-PECDF | 1.11 | 0.933 | 0.875 | J | 0.348 | J | 1.18 | 0.75 | U | 0.195 | U | 0.665 | J |
| 2,3,4,7,8-PECDF | 1.42 | 1.22 | 1.14 |  | 0.437 | J | 1.29 | 0.854 | J | 0.374 | J | 0.867 | J |
| 1,2,3,4,7,8-HXCDF | 2.06 | 1.52 | 1.51 |  | 0.967 | J | 3.32 | 1.13 |  | 0.598 | J | 1.27 |  |
| 1,2,3,6,7,8-HXCDF | 1.31 | 1.06 | 1.12 |  | 0.615 | J | 1.28 | 0.645 | J | 0.38 | U | 0.778 | J |
| 1,2,3,7,8,9-HXCDF | 0.153 | 0.132 | 0.077 | J | 0.054 | U | 0.188 | 0.122 | J | 0.053 | U | 0.162 | J |
| 2,3,4,6,7,8-HXCDF | 0.962 | 0.699 | 0.817 | J | 0.433 | J | 0.881 | 0.589 | J | 0.264 | J | 0.649 | J |
| 1,2,3,4,6,7,8-HPCDF | 21.4 | 17.5 | 16.7 |  | 10.7 |  | 26.9 | 10.9 |  | 6.6 |  | 12.8 |  |
| 1,2,3,4,7,8,9-HPCDF | 1.39 | 1.22 | 1.04 |  | 0.657 | J | 1.61 | 0.672 | J | 0.458 | U | 0.874 | J |
| OCDF | 59.3 | 61.8 | 49.1 |  | 29.9 |  | 70 | 33.9 |  | 19.6 |  | 34.9 |  |
| Dx/F TEQ (0 DL) | 13.6 | 10.4 | 11.0 |  | 4.51 |  | 8.31 | 7.27 |  | 2.17 |  | 4.91 |  |
| Dx/F TEQ (1/2 DL) | 13.6 | 10.4 | 11.0 |  | 4.52 |  | 8.32 | 7.28 |  | 2.20 |  | 5.42 |  |
| Dx/F TEQ (1 DL) | 13.6 | 10.4 | 11.0 |  | 4.52 |  | 8.33 | 7.29 |  | 2.22 |  | 5.92 |  |
| Dx/F TEQ (KM) | 13.6 | 10.4 | 11.0 |  | 4.52 |  | 8.33 | 7.28 |  | 2.19 |  | 5.92 | L |

U-the analytie was analyzed for, but was not detected above the reported sample quantitation limit
$J$-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sample
UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximate
L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detects
KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicates

Table F-2. Summary of Bellingham Bay Dioxin/Furan Congener Data (continued).

| Location ID | BB-16-S | Q | BB-17-S | Q | BB-18-S | Q | BB-19-S | Q | BB-20-S | Q | BB-20-D | Q | BB-21-S | Q | BB-22-S | Q | BB-23-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/11/2014 | 1044 | 9/9/2014 | 1524 | 9/9/2014 | 1042 | 9/9/2014 | 1326 | 9/10/2014 | 1530 | 9/10/2014 | 1530 | 9/9/2014 | 1453 | 9/9/2014 | 1026 | 9/10/2014 | 942 |
| Dioxin/Furan Congeners (ng/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2,3,7,8-TCDD | 0.317 |  | 0.549 |  | 0.565 |  | 0.582 |  | 0.268 |  | 0.253 |  | 1.17 |  | 0.458 |  | 0.711 |  |
| 1,2,3,7,8-PECDD | 0.969 | J | 1.55 | J | 1.62 |  | 1.44 |  | 0.937 | J | 0.572 | J | 2.34 |  | 2.24 |  | 2.06 | J |
| 1,2,3,4,7,8-HXCDD | 3.33 |  | 6.67 |  | 5.41 |  | 5 |  | 2.62 |  | 1.92 |  | 8.58 |  | 7.18 |  | 6.66 |  |
| 1,2,3,6,7,8-HXCDD | 6.23 |  | 10.2 |  | 13.2 |  | 9.02 |  | 5.79 |  | 4.47 |  | 13.5 |  | 13.1 |  | 16.6 |  |
| 1,2,3,7,8,9-HXCDD | 5.62 |  | 8.97 |  | 9.73 |  | 6.5 |  | 4.76 |  | 3.37 |  | 13.1 |  | 10.7 |  | 11.8 |  |
| 1,2,3,4,6,7,8-HPCDD | 93.4 |  | 130 |  | 278 |  | 174 |  | 106 |  | 83.4 |  | 135 |  | 190 |  | 355 |  |
| OCDD | 669 |  | 752 |  | 2190 |  | 1310 |  | 779 |  | 657 |  | 564 |  | 1220 |  | 3120 |  |
| 2,3,7,8-TCDF | 4.24 |  | 9.41 |  | 7.54 |  | 8.37 |  | 3.29 |  | 2.54 |  | 14.1 |  | 13.1 |  | 7.11 |  |
| 1,2,3,7,8-PECDF | 0.467 | J | 0.58 | J | 0.84 | J | 0.789 | J | 0.434 | U | 0.374 | J | 0.851 | J | 1.15 |  | 1.18 |  |
| 2,3,4,7,8-PECDF | 0.526 | J | 0.875 | J | 1.02 |  | 0.955 | J | 0.624 | J | 0.424 | J | 1.22 |  | 1.19 |  | 1.45 |  |
| 1,2,3,4,7,8-HXCDF | 1.23 |  | 1.42 |  | 3.03 |  | 1.93 |  | 1.47 |  | 1.22 |  | 1.48 |  | 2.57 |  | 4.97 |  |
| 1,2,3,6,7,8-HXCDF | 0.632 | J | 0.838 | J | 1.48 |  | 0.992 |  | 0.798 | U | 0.565 | J | 1.02 |  | 1.48 |  | 2.03 |  |
| 1,2,3,7,8,9-HXCDF | 0.079 | U | 0.089 | J | 0.184 | U | 0.099 | J | 0.093 | U | 0.102 | U | 0.096 | J | 0.135 | J | 0.25 | J |
| 2,3,4,6,7,8-HXCDF | 0.451 | J | 0.632 | J | 1.22 |  | 0.801 | J | 0.522 | J | 0.463 | J | 0.632 | J | 1.05 |  | 1.35 |  |
| 1,2,3,4,6,7,8-HPCDF | 12.5 |  | 14.9 |  | 54.2 |  | 25.2 |  | 15.9 |  | 12.8 |  | 12.2 |  | 31.6 |  | 61 |  |
| 1,2,3,4,7,8,9-HPCDF | 0.824 | J | 1.13 |  | 2.98 |  | 2.06 |  | 1.08 |  | 0.82 | J | 0.806 | J | 1.99 |  | 4.01 |  |
| OCDF | 39.1 |  | 42.6 |  | 158 |  | 90.8 |  | 43.7 |  | 37.7 |  | 36.2 |  | 82.7 |  | 214 |  |
| Dx/F TEQ (0 DL) | 4.91 |  | 7.90 |  | 10.7 |  | 8.04 |  | 4.71 |  | 3.60 |  | 10.8 |  | 10.6 |  | 13.5 |  |
| Dx/F TEQ (1/2 DL) | 4.91 |  | 7.90 |  | 10.7 |  | 8.04 |  | 4.77 |  | 3.60 |  | 10.8 |  | 10.6 |  | 13.5 |  |
| Dx/F TEQ (1 DL) | 4.92 |  | 7.90 |  | 10.8 |  | 8.04 |  | 4.82 |  | 3.61 |  | 10.8 |  | 10.6 |  | 13.5 |  |
| Dx/F TEQ (KM) | 4.92 |  | 7.90 |  | 10.8 |  | 8.04 |  | 4.18 | Avg | -- |  | 10.8 |  | 10.6 |  | 13.5 |  |

U-the analytie was analyzed for, but was not detected above the reported sample quantitation limit
$J$-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sample
UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximate
L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detects
KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicates

Table F-2. Summary of Bellingham Bay Dioxin/Furan Congener Data (continued).

| Location ID | BB-24-S | Q | BB-24-D | Q | BB-25-S | Q | BB-26-S | Q | BB-27-S | Q | BB-28-S | Q | BB-29-S | Q | BB-30-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1353 | 9/9/2014 | 1353 | 9/11/2014 | 950 | 9/10/2014 | 1012 | 9/11/2014 | 923 | 9/9/2014 | 1129 | 9/11/2014 | 1103 | 9/9/2014 | 1507 |
| Dioxin/Furan Congeners (ng/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2,3,7,8-TCDD | 1.6 |  | 1.5 |  | 0.423 |  | 0.332 |  | 0.112 | U | 0.591 |  | 0.378 |  | 0.867 |  |
| 1,2,3,7,8-PECDD | 2.75 | J | 3.12 |  | 1.62 |  | 1.05 |  | 0.476 | J | 2.22 |  | 1.32 |  | 1.92 |  |
| 1,2,3,4,7,8-HXCDD | 12.4 |  | 13.2 |  | 4.04 |  | 3.33 |  | 1.58 |  | 5.38 |  | 4.69 |  | 8.81 |  |
| 1,2,3,6,7,8-HXCDD | 18.3 |  | 17.9 |  | 14.3 |  | 8.22 |  | 2.82 |  | 12.6 |  | 8.88 |  | 12.7 |  |
| 1,2,3,7,8,9-HXCDD | 16.6 |  | 18.1 |  | 9.08 |  | 6.43 |  | 2.61 |  | 11.1 |  | 7.95 |  | 12.3 |  |
| 1,2,3,4,6,7,8-HPCDD | 172 |  | 175 |  | 329 |  | 166 |  | 43.2 |  | 210 |  | 132 |  | 150 |  |
| OCDD | 645 |  | 638 |  | 2760 |  | 1510 |  | 313 |  | 1610 |  | 996 |  | 706 |  |
| 2,3,7,8-TCDF | 21.5 |  | 21.6 |  | 4.33 |  | 4.57 |  | 1.63 |  | 7.34 |  | 5.91 |  | 11.4 |  |
| 1,2,3,7,8-PECDF | 1.14 |  | 1.02 |  | 1.54 |  | 0.64 | J | 0.272 | U | 1.1 |  | 0.545 | J | 0.705 | J |
| 2,3,4,7,8-PECDF | 1.44 |  | 1.39 |  | 1.74 |  | 0.733 | J | 0.289 | J | 1.31 |  | 0.659 | J | 1.05 |  |
| 1,2,3,4,7,8-HXCDF | 1.68 |  | 1.49 |  | 8.4 |  | 2.22 |  | 0.669 | U | 2.78 |  | 1.58 |  | 1.44 |  |
| 1,2,3,6,7,8-HXCDF | 1.37 |  | 1.24 |  | 2.63 |  | 0.971 | J | 0.37 | U | 1.51 |  | 0.809 | J | 0.995 |  |
| 1,2,3,7,8,9-HXCDF | 0.117 | J | 0.096 | J | 0.391 | U | 0.142 | U | 0.13 | U | 0.126 | J | 0.117 | U | 0.102 | J |
| 2,3,4,6,7,8-HXCDF | 0.762 | J | 0.811 | J | 1.47 |  | 0.688 | J | 0.316 | U | 1.27 |  | 0.595 | J | 0.715 | J |
| 1,2,3,4,6,7,8-HPCDF | 13.6 |  | 13.7 |  | 63.6 |  | 24.7 |  | 6.64 |  | 46.2 |  | 18.2 |  | 13.4 |  |
| 1,2,3,4,7,8,9-HPCDF | 0.933 | J | 0.913 | J | 4.62 |  | 1.91 |  | 0.477 | U | 2.83 |  | 1.25 |  | 1.08 |  |
| OCDF | 38.4 |  | 44.1 |  | 165 |  | 96.5 |  | 20.4 |  | 164 |  | 62.3 |  | 46.3 |  |
| Dx/F TEQ (0 DL) | 14.2 |  | 14.6 |  | 11.9 |  | 6.67 |  | 2.03 |  | 10.6 |  | 6.79 |  | 9.84 |  |
| Dx/F TEQ (1/2 DL) | 14.2 |  | 14.6 |  | 11.9 |  | 6.68 |  | 2.16 |  | 10.6 |  | 6.79 |  | 9.84 |  |
| Dx/F TEQ (1 DL) | 14.2 |  | 14.6 |  | 11.9 |  | 6.69 |  | 2.30 |  | 10.6 |  | 6.80 |  | 9.84 |  |
| Dx/F TEQ (KM) | 14.4 | Avg | -- |  | 11.9 |  | 6.69 |  | 2.10 |  | 10.6 |  | 6.80 |  | 9.84 |  |

U-the analytie was analyzed for, but was not detected above the reported sample quantitation limit
$J$-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sample
$J$-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximate,
L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detects
KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicates

Table F-3. Summary of Bellingham Bay Polychlorinated Biphenyl (PCB) Congener Data.

| Location ID | BB-01-S | Q | BB-02-S | Q | BB-03-S | Q | BB-04-S | Q | BB-05-S | Q | BB-06-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/11/2014 | 834 | 9/10/2014 | 1135 | 9/10/2014 | 1512 | 9/10/2014 | 1438 | 9/10/2014 | 1410 | 9/10/2014 | 1115 |
| PCB Congeners (ng/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |
| PCB-001 | 4.22 |  | 2.79 |  | 23.7 |  | 4.46 |  | 3.33 | J | 17.5 |  |
| PCB-002 | 6.9 |  | 5.63 |  | 11.1 |  | 3.72 |  | 4.44 |  | 11.3 |  |
| PCB-003 | 6.23 |  | 5.4 |  | 25.2 |  | 5.56 |  | 5.21 |  | 24.2 |  |
| PCB-004 | 3.74 |  | 2.59 | UJ | 11.2 |  | 3.1 |  | 2.4 | J | 8.78 |  |
| PCB-005 | 0.473 | U | 1.45 | UJ | 0.876 |  | 0.155 | U | 0.833 | U | 2.08 | U |
| PCB-006 | 4.13 |  | 3.66 | UJ | 14.4 |  | 3.81 |  | 3.46 | U | 14.8 |  |
| PCB-007 | 0.815 | U | 1.34 | UJ | 3.05 |  | 0.68 |  | 0.767 | U | 2.66 | J |
| PCB-008 | 15.4 |  | 11.5 | J | 48.6 |  | 12 |  | 12.7 |  | 62.6 |  |
| PCB-009 | 0.676 | J | 1.29 | UJ | 2.66 |  | 0.608 |  | 0.742 | U | 1.84 | U |
| PCB-010 | 0.405 | U | 1.28 | UJ | 0.54 | U | 0.131 | U | 0.735 | U | 1.82 | U |
| PCB-011 | 20.7 |  | 15.8 | J | 26 |  | 12 |  | 15.8 |  | 27 |  |
| PCB-012/013 | 5.13 | U | 5.18 | UJ | 12.7 |  | 3.54 |  | 3.18 | U | 10.8 |  |
| PCB-014 | 0.443 | U | 1.36 | UJ | 0.562 |  | 0.103 | U | 0.78 | U | 1.99 | U |
| PCB-015 | 16.3 |  | 13.3 |  | 52.2 |  | 12.2 |  | 12 |  | 65 |  |
| PCB-016 | 4.86 |  | 3.98 |  | 13.2 |  | 3.29 |  | 3.92 |  | 18.6 |  |
| PCB-017 | 7.14 |  | 5.46 |  | 24.1 |  | 6.2 |  | 6.68 |  | 33.5 |  |
| PCB-018/030 | 12.6 |  | 9.94 |  | 37.4 |  | 9.67 |  | 11.1 |  | 47.9 |  |
| PCB-019 | 1.26 | U | 1.39 | U | 4 |  | 1.1 |  | 1.15 |  | 4.7 |  |
| PCB-020/028 | 43.8 |  | 39.5 |  | 143 |  | 33.8 |  | 37.2 |  | 206 |  |
| PCB-021/033 | 15.5 |  | 13 |  | 49.7 |  | 11.9 |  | 12.7 |  | 83.1 |  |
| PCB-022 | 12 |  | 10.7 |  | 40.4 |  | 9.89 |  | 10.4 |  | 59.5 |  |
| PCB-023 | 0.191 | U | 0.208 | U | 0.176 | J | 0.0497 | U | 0.0551 | U | 0.243 | J |
| PCB-024 | 0.2 | U | 0.177 | U | 0.607 |  | 0.178 | J | 0.188 | J | 0.862 |  |
| PCB-025 | 4.37 |  | 3.14 |  | 14.2 |  | 3.64 |  | 3.9 |  | 19.1 |  |
| PCB-026/029 | 9.24 |  | 7.32 |  | 29.6 |  | 7.37 |  | 8.22 |  | 37.7 |  |
| PCB-027 | 1.41 | J | 1.08 | J | 4.63 |  | 1.23 |  | 1.31 |  | 5.58 |  |
| PCB-031 | 30.7 | J | 27.3 |  | 94.6 |  | 24.2 |  | 27 |  | 141 |  |
| PCB-032 | 4.89 |  | 3.84 |  | 16.4 |  | 4.14 |  | 4.45 |  | 25.9 |  |
| PCB-034 | 0.185 | U | 0.21 | U | 0.773 |  | 0.175 | J | 0.198 | J | 1.01 |  |
| PCB-035 | 1.38 | U | 1.3 | J | 4.35 |  | 1.08 |  | 1.26 |  | 6.33 |  |
| PCB-036 | 0.487 | U | 0.354 | J | 0.757 |  | 0.308 | J | 0.451 |  | 0.934 |  |
| PCB-037 | 10.9 |  | 10.1 |  | 37.1 |  | 8.32 |  | 9.19 |  | 50 |  |
| PCB-038 | 0.191 | U | 0.223 | U | 0.538 | U | 0.176 | J | 0.212 | J | 0.638 |  |
| PCB-039 | 0.235 | U | 0.26 | U | 0.851 |  | 0.226 | J | 0.23 | J | 1.29 |  |
| PCB-040/041/071 | 14 |  | 12.2 |  | 44 |  | 10.8 |  | 11.8 |  | 62.9 |  |
| PCB-042 | 6.82 |  | 6.28 |  | 20.4 |  | 4.9 |  | 5.44 |  | 28.6 |  |
| PCB-043 | 0.895 | J | 0.875 | J | 2.73 |  | 0.741 |  | 0.84 |  | 4.01 |  |
| PCB-044/047/065 | 27.7 |  | 22.6 |  | 81.1 |  | 20.8 |  | 22.8 |  | 124 |  |
| PCB-045/051 | 3.67 |  | 2.96 |  | 11 |  | 2.9 |  | 3.13 |  | 15.3 |  |
| PCB-046 | 1.13 | J | 0.813 | J | 3.25 |  | 0.819 |  | 0.932 |  | 4.6 |  |
| PCB-048 | 4.66 |  | 3.96 |  | 14.9 |  | 3.37 |  | 3.62 |  | 21.7 |  |
| PCB-049/069 | 20.4 |  | 17.5 |  | 63.3 |  | 15.7 |  | 17.5 |  | 86.2 |  |
| PCB-050/053 | 3.17 |  | 2.48 |  | 10.5 |  | 2.82 |  | 2.99 |  | 13.8 |  |
| PCB-052 | 35.8 |  | 29.3 |  | 102 |  | 28.3 |  | 29.8 |  | 186 |  |
| PCB-054 | 0.11 | U | 0.0748 | U | 0.212 | U | 0.078 | U | 0.059 | U | 0.201 | J |
| PCB-055 | 0.239 | U | 0.346 | U | 2 |  | 0.604 |  | 0.502 |  | 4.2 |  |
| PCB-056 | 15.3 |  | 15.1 |  | 43.7 |  | 10.1 |  | 11 |  | 70.6 |  |
| PCB-057 | 0.214 | U | 0.324 | U | 0.574 |  | 0.166 | J | 0.172 | J | 1 |  |
| PCB-058 | 0.224 | U | 0.336 | U | 0.258 | U | 0.0932 | U | 0.134 | U | 0.306 | U |
| PCB-059/062/075 | 2.52 |  | 2.47 |  | 8.27 |  | 1.93 |  | 2.16 |  | 11.3 |  |
| PCB-060 | 8.26 |  | 9.27 |  | 24.9 |  | 5.86 |  | 6.36 |  | 40.3 |  |
| PCB-061/070/074/076 | 63.4 |  | 61.5 |  | 182 |  | 44.7 |  | 49.2 |  | 334 |  |
| PCB-063 | 1.31 | J | 1.35 | U | 4.15 |  | 0.953 |  | 1.09 |  | 6.85 |  |
| PCB-064 | 11.3 |  | 9.74 |  | 33.4 |  | 8.89 |  | 9.31 |  | 50.8 |  |
| PCB-066 | 33.9 |  | 32.4 |  | 95.3 |  | 22.4 |  | 25.2 |  | 156 |  |
| PCB-067 | 1.16 | J | 0.958 | J | 3.97 |  | 0.799 |  | 0.879 |  | 5.98 |  |
| PCB-068 | 0.427 | J | 0.45 | J | 1.04 |  | 0.323 | J | 0.342 | J | 1.28 |  |
| PCB-072 | 0.551 | U | 0.579 | J | 1.61 |  | 0.353 | J | 0.433 |  | 2.19 |  |
| PCB-073 | 0.0489 | U | 0.174 | U | 0.054 | U | 0.048 | U | 0.0482 | U | 0.0505 | U |


| Location ID | BB-01-S | Q | BB-02-S | Q | BB-03-S | Q | BB-04-S | Q | BB-05-S | Q | BB-06-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/11/2014 | 834 | 9/10/2014 | 1135 | 9/10/2014 | 1512 | 9/10/2014 | 1438 | 9/10/2014 | 1410 | 9/10/2014 | 1115 |
| PCB-077 | 3.9 |  | 3.24 |  | 11.7 |  | 2.87 |  | 3.24 |  | 16.4 |  |
| PCB-078 | 0.25 | U | 0.353 | U | 0.255 | U | 0.092 | U | 0.133 | U | 0.33 | U |
| PCB-079 | 0.749 | J | 0.673 | J | 1.81 |  | 0.493 |  | 0.474 |  | 4.98 |  |
| PCB-080 | 0.216 | U | 0.334 | U | 0.789 |  | 0.0853 | U | 0.123 | U | 0.281 | U |
| PCB-081 | 0.231 | U | 0.313 | U | 0.647 |  | 0.157 | U | 0.137 | J | 0.754 |  |
| PCB-082 | 5.28 |  | 5.38 |  | 15.8 |  | 4.83 | J | 5.49 |  | 36.5 |  |
| PCB-083/099 | 30.5 |  | 33.8 |  | 96.1 |  | 26.6 |  | 30.4 |  | 162 |  |
| PCB-084 | 10.5 |  | 10.8 |  | 0.129 | U | 10.2 |  | 10.6 |  | 60.2 |  |
| PCB-085/116/117 | 9.88 |  | 9.55 |  | 28.9 |  | 7.55 |  | 8.32 |  | 49.5 |  |
| PCB-087/097/108/119/125 | 31.8 |  | 30.7 |  | 100 |  | 28.3 |  | 29.9 |  | 176 |  |
| PCB-088/091 | 6.5 |  | 6.55 |  | 20.5 |  | 6.02 |  | 6.43 |  | 32.5 |  |
| PCB-089 | 0.552 | U | 0.363 | U | 1.35 |  | 0.439 | U | 0.379 | J | 2.28 |  |
| PCB-090/101/113 | 42.2 |  | 51.4 |  | 141 |  | 39.8 |  | 44.1 |  | 256 |  |
| PCB-092 | 7.33 |  | 9.4 |  | 26.4 |  | 7.79 |  | 8.4 |  | 44.9 |  |
| PCB-093/095/098/100/102 | 32.6 |  | 31.4 |  | 98.2 |  | 28.9 |  | 30.3 |  | 171 |  |
| PCB-094 | 0.292 | J | 0.35 | U | 0.123 | U | 0.213 | U | 0.244 | U | 0.947 |  |
| PCB-096 | 0.354 | J | 0.233 | J | 0.815 | U | 0.239 | J | 0.197 | J | 1.02 |  |
| PCB-103 | 0.586 | U | 0.479 | U | 0.095 | U | 0.558 |  | 0.62 |  | 2.43 |  |
| PCB-104 | 0.0489 | U | 0.0521 | U | 0.0574 | U | 0.048 | U | 0.0482 | U | 0.0505 | U |
| PCB-105 | 20.3 |  | 19.2 |  | 62.7 |  | 15.4 |  | 16.7 |  | 128 |  |
| PCB-106 | 0.2 | U | 0.484 | U | 0.261 | U | 0.131 | U | 0.0733 | U | 0.673 | U |
| PCB-107/124 | 1.74 | U | 1.88 | J | 6.63 |  | 1.78 |  | 1.69 |  | 12.9 |  |
| PCB-109 | 4.09 |  | 3.99 |  | 13.3 |  | 3.15 |  | 3.53 |  | 27.1 |  |
| PCB-110/115 | 55.3 |  | 52.4 |  | 162 |  | 41.6 |  | 44.9 |  | 300 |  |
| PCB-111 | 0.0945 | U | 0.265 | U | 0.146 | U | 0.08 | J | 0.0659 | U | 0.168 | U |
| PCB-112 | 0.0867 | U | 0.255 | U | 0.0756 | U | 0.0524 | U | 0.0628 | U | 0.0614 | U |
| PCB-114 | 0.969 | U | 0.811 | U | 3.39 |  | 0.873 |  | 0.994 |  | 7.72 |  |
| PCB-118 | 50.5 |  | 47.4 |  | 148 |  | 35.7 |  | 39.1 |  | 294 |  |
| PCB-120 | 0.239 | U | 0.31 | U | 0.806 |  | 0.204 | J | 0.269 | J | 0.888 |  |
| PCB-121 | 0.0928 | U | 0.262 | U | 0.0827 | U | 0.0573 | U | 0.0686 | U | 0.0622 | U |
| PCB-122 | 0.831 | J | 0.56 | U | 2 |  | 0.444 |  | 0.492 |  | 4.46 |  |
| PCB-123 | 1.28 | U | 1.33 | J | 2.92 |  | 0.883 |  | 0.987 |  | 5.79 |  |
| PCB-126 | 0.398 | J | 0.562 | U | 0.816 |  | 0.259 | U | 0.265 | J | 0.954 |  |
| PCB-127 | 0.243 | U | 0.566 | U | 0.269 | U | 0.135 | U | 0.0757 | U | 0.678 | U |
| PCB-128/166 | 12.1 |  | 9.7 |  | 34.2 |  | 0.203 | U | 0.262 | U | 1.27 | U |
| PCB-129/138/160/163 | 73.4 |  | 55.6 |  | 185 |  | 55.2 |  | 60.6 |  | 384 |  |
| PCB-130 | 4.67 |  | 3.71 |  | 12.2 |  | 3.32 | J | 3.37 | J | 22.4 |  |
| PCB-131 | 0.75 | U | 0.534 | U | 2.46 | J | 0.585 | U | 0.281 | U | 3.91 | J |
| PCB-132 | 20.9 |  | 13.6 |  | 50.5 |  | 14.9 |  | 13.5 |  | 109 |  |
| PCB-133 | 1.11 | U | 0.872 | J | 2.79 | J | 0.84 | U | 0.845 | J | 4.22 | U |
| PCB-134/143 | 0.175 | U | 1.98 | J | 8.94 |  | 2.13 | J | 2.21 | U | 15.7 |  |
| PCB-135/151/154 | 17.7 |  | 13.7 |  | 43.5 |  | 11 |  | 13.7 |  | 96.8 |  |
| PCB-136 | 4.97 |  | 4.49 |  | 15.2 |  | 4.05 | J | 4.6 |  | 30.5 |  |
| PCB-137 | 3.4 |  | 2.78 | U | 10.1 |  | 2.51 | U | 2.73 | U | 19.6 |  |
| PCB-139/140 | 1.24 | J | 0.905 | U | 2.96 | J | 1.1 | J | 1.02 | J | 6.03 | J |
| PCB-141 | 8.3 |  | 6.71 |  | 25.5 |  | 7.01 |  | 7.15 |  | 58.2 |  |
| PCB-142 | 0.187 | U | 0.286 | U | 0.647 | U | 0.225 | U | 0.283 | U | 1.31 | U |
| PCB-144 | 2.33 | U | 1.76 | J | 5.9 | U | 1.76 | J | 1.68 | U | 13.7 |  |
| PCB-145 | 0.063 | U | 0.0497 | U | 0.076 | U | 0.106 | U | 0.0482 | U | 0.0505 | U |
| PCB-146 | 9.84 |  | 9.68 |  | 25.8 |  | 6.79 |  | 6.7 |  | 48.9 |  |
| PCB-147/149 | 44.9 |  | 32.5 |  | 103 |  | 32.3 |  | 34.4 |  | 242 |  |
| PCB-148 | 0.151 | UJ | 0.0497 | U | 0.307 | U | 0.048 | U | 0.147 | U | 0.519 | U |
| PCB-150 | 0.0489 | U | 0.116 | U | 0.0461 | U | 0.048 | U | 0.0482 | U | 0.361 | U |
| PCB-152 | 0.0489 | U | 0.0497 | U | 0.0461 | U | 0.112 | U | 0.066 | U | 0.177 | U |
| PCB-153/168 | 56.7 |  | 51.8 |  | 139 |  | 40.5 |  | 42.2 |  | 317 |  |
| PCB-155 | 0.084 | U | 0.095 | U | 0.0823 | U | 0.048 | U | 0.0482 | U | 0.085 | J |
| PCB-156/157 | 6.22 |  | 7.48 |  | 21.1 |  | 5.23 |  | 5.58 |  | 43.3 |  |
| PCB-158 | 5.95 |  | 4.82 |  | 16.1 |  | 4.82 | J | 4.18 |  | 35.3 |  |
| PCB-159 | 0.53 | U | 0.229 | U | 1.79 | J | 0.368 | U | 0.563 | U | 4.89 | J |
| PCB-161 | 0.132 | U | 0.2 | U | 0.483 | U | 0.159 | U | 0.197 | U | 0.981 | U |
| PCB-162 | 0.24 | J | 0.233 | U | 0.549 | U | 0.388 | J | 0.231 | U | 1.11 | U |


| Location ID | BB-01-S | Q | BB-02-S | Q | BB-03-S | Q | BB-04-S | Q | BB-05-S | Q | BB-06-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/11/2014 | 834 | 9/10/2014 | 1135 | 9/10/2014 | 1512 | 9/10/2014 | 1438 | 9/10/2014 | 1410 | 9/10/2014 | 1115 |
| PCB-164 | 3.93 |  | 3.76 |  | 12.2 |  | 3.09 | J | 3.37 | U | 22.1 |  |
| PCB-165 | 0.147 | U | 0.23 | U | 0.55 | U | 0.177 | U | 0.227 | U | 1.12 | U |
| PCB-167 | 2.43 |  | 2.3 | U | 7.29 |  | 1.83 |  | 2.01 |  | 14.4 |  |
| PCB-169 | 0.162 | U | 0.282 | U | 0.553 | U | 0.218 | U | 0.229 | U | 1.15 | U |
| PCB-170 | 11.4 |  | 11.6 |  | 34.3 |  | 9.37 |  | 9.35 |  | 84 |  |
| PCB-171/173 | 4.34 |  | 3.64 | U | 12.3 |  | 3.31 |  | 3.75 |  | 26.8 |  |
| PCB-172 | 2.5 | U | 2.78 |  | 6.94 |  | 1.83 |  | 1.93 |  | 17.7 |  |
| PCB-174 | 13 |  | 10.3 |  | 35.7 |  | 10.3 |  | 10.8 |  | 95.7 |  |
| PCB-175 | 0.666 | J | 0.727 | U | 1.98 |  | 0.534 |  | 0.588 |  | 4.4 |  |
| PCB-176 | 2.1 |  | 1.36 | J | 5.49 |  | 1.44 |  | 1.65 |  | 12.2 |  |
| PCB-177 | 8.83 |  | 7.38 | U | 23.8 |  | 7.33 |  | 8.09 |  | 53.3 |  |
| PCB-178 | 3.94 | U | 3.26 | U | 9.76 |  | 2.81 |  | 3.1 |  | 21.4 |  |
| PCB-179 | 6.24 |  | 4.67 | U | 16.7 |  | 4.86 |  | 5.24 |  | 43 |  |
| PCB-180/193 | 24.1 |  | 23.6 |  | 71.4 |  | 19.3 |  | 20.1 |  | 229 |  |
| PCB-181 | 0.206 | U | 0.0497 | U | 0.561 |  | 0.0821 | U | 0.163 | U | 0.771 |  |
| PCB-182 | 0.205 | U | 0.176 | U | 0.384 | U | 0.115 | J | 0.113 | U | 0.505 | U |
| PCB-183/185 | 10.5 |  | 9.66 |  | 29.1 |  | 8.46 |  | 8.59 |  | 79.9 |  |
| PCB-184 | 0.075 | U | 0.124 | U | 0.123 | U | 0.0562 | U | 0.084 | U | 0.117 | U |
| PCB-186 | 0.0489 | U | 0.0497 | U | 0.0839 | U | 0.0666 | U | 0.0504 | U | 0.0649 | U |
| PCB-187 | 22.4 |  | 22.3 |  | 57.4 |  | 16.3 |  | 18.2 |  | 146 |  |
| PCB-188 | 0.14 | J | 0.123 | U | 0.188 | U | 0.077 | J | 0.082 | J | 0.197 | U |
| PCB-189 | 0.452 | J | 0.666 | U | 1.61 |  | 0.446 |  | 0.46 |  | 3.55 |  |
| PCB-190 | 2.22 |  | 1.94 | U | 6.89 |  | 1.9 | J | 2.21 | J | 16.9 |  |
| PCB-191 | 0.565 | J | 0.435 | U | 1.33 |  | 0.446 |  | 0.45 |  | 3.94 |  |
| PCB-192 | 0.0489 | U | 0.0497 | U | 0.0813 | U | 0.0645 | U | 0.0488 | U | 0.0822 | U |
| PCB-194 | 7.41 | U | 5.83 |  | 22.6 |  | 6.38 |  | 6.44 |  | 101 |  |
| PCB-195 | 2.61 |  | 2.01 |  | 7.74 |  | 1.97 |  | 1.96 |  | 36.1 |  |
| PCB-196 | 4.04 |  | 3.27 |  | 12.1 |  | 3.55 |  | 3.64 |  | 49.7 |  |
| PCB-197/199 | 1.48 | U | 1.18 | U | 4.3 |  | 1.11 |  | 1.16 |  | 17.2 |  |
| PCB-198/201 | 13 |  | 11.5 |  | 30 |  | 8.74 | U | 10 |  | 115 |  |
| PCB-200 | 1.49 | U | 1.43 | U | 3.72 |  | 0.951 |  | 0.983 |  | 14.5 |  |
| PCB-202 | 3.05 |  | 2.84 | U | 7.04 |  | 2.2 |  | 2.44 |  | 23.8 |  |
| PCB-203 | 6.31 |  | 6.12 |  | 17.9 |  | 5.33 |  | 5.8 |  | 73.6 |  |
| PCB-204 | 0.0489 | U | 0.0497 | U | 0.0461 | U | 0.048 | U | 0.0482 | U | 0.0505 | U |
| PCB-205 | 0.404 | U | 0.516 | U | 0.987 |  | 0.31 | J | 0.301 | J | 4.07 |  |
| PCB-206 | 8.22 |  | 8.44 |  | 17.6 |  | 6.18 |  | 7.13 |  | 40.8 |  |
| PCB-207 | 1.04 | J | 1.22 | J | 5.64 |  | 1.05 | J | 0.941 | J | 6.78 |  |
| PCB-208 | 3.4 |  | 3.09 |  | 10.4 |  | 2.73 |  | 3.03 |  | 16.3 |  |
| PCB-209 | 7.47 |  | 7.18 | U | 132 |  | 7.97 |  | 7.5 |  | 25.8 |  |
| Total PCBs* | 1240 |  | 1090 |  | 3750 |  | 965 |  | 1030 |  | 6870 |  |
| PCB TEQ (0 DL) | 0.0426 |  | 0.00259 |  | 0.0904 |  | 0.0021 |  | 0.0288 |  | 0.112 |  |
| PCB TEQ (1/2 DL) | 0.0451 |  | 0.0350 |  | 0.0987 |  | 0.0183 |  | 0.0323 |  | 0.129 |  |
| PCB TEQ (1 DL) | 0.0476 |  | 0.0675 |  | 0.107 |  | 0.0346 |  | 0.0357 |  | 0.147 |  |
| PCB TEQ (KM) | 0.0429 |  | 0.0592 | L6 | 0.0913 |  | 0.0283 | L | 0.0291 |  | 0.114 |  |

*total PCBs represents the sum of all detected congeners
U-the analytie was analyzed for, but was not detected above the reported sample quantitation limi
$J$-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sampl UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximat $\epsilon$ L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detect KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicate

Table F-3. Summary of Bellingham Bay Polychlorinated Biphenyl (PCB) Congener Data (continued).

| Location ID | BB-07-S | Q | BB-08-S | Q | BB-09-S | Q | BB-10-S | Q | BB-11-S | Q | BB-12-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/10/2014 | 1356 | 9/9/2014 | 933 | 9/9/2014 | 1540 | 9/9/2014 | 1147 | 9/11/2014 | 1024 | 9/10/2014 | 1330 |
| PCB Congeners (ng/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |
| PCB-001 | 4.5 |  | 54.5 |  | 51.4 |  | 46.9 |  | 15.9 |  | 55.6 |  |
| PCB-002 | 7.31 |  | 32.7 |  | 23.4 |  | 22.8 |  | 12.7 |  | 12.2 |  |
| PCB-003 | 5.9 |  | 57.7 |  | 53.1 |  | 52.7 |  | 18 |  | 82.6 |  |
| PCB-004 | 3.14 |  | 16.8 |  | 15.5 |  | 13.1 |  | 7.74 |  | 31.1 |  |
| PCB-005 | 1.5 | U | 2.58 | U | 1.73 | U | 1.44 |  | 1.26 | U | 3.14 | U |
| PCB-006 | 3.87 | U | 16.8 |  | 18.4 |  | 15.9 |  | 7.6 |  | 32.5 |  |
| PCB-007 | 1.38 | U | 5.06 |  | 4.56 |  | 4.88 |  | 1.15 | U | 6.87 |  |
| PCB-008 | 13.3 |  | 89 |  | 74.3 |  | 73.5 |  | 30.9 |  | 193 |  |
| PCB-009 | 1.33 | U | 3.56 | J | 2.89 | J | 2.79 |  | 1.11 | U | 6.97 |  |
| PCB-010 | 1.32 | U | 2.15 | U | 1.51 | U | 0.61 | U | 1.08 | U | 1.64 | U |
| PCB-011 | 14.6 | U | 48.6 |  | 36.2 |  | 35.9 |  | 28.2 |  | 24.2 |  |
| PCB-012/013 | 4.19 | U | 19 |  | 20.7 | U | 15.7 |  | 11.6 | U | 23.6 |  |
| PCB-014 | 1.4 | U | 2.37 | U | 1.63 | U | 1.56 | U | 1.18 | U | 1.8 | U |
| PCB-015 | 13.8 |  | 107 |  | 93.7 |  | 88.5 |  | 38.6 |  | 170 |  |
| PCB-016 | 4.32 |  | 14.9 |  | 12.4 |  | 11.9 |  | 6.83 |  | 80.3 |  |
| PCB-017 | 7.39 |  | 25.4 |  | 25.8 |  | 20.4 |  | 13 |  | 119 |  |
| PCB-018/030 | 12.4 |  | 45.9 |  | 52.9 |  | 33.5 |  | 22.2 |  | 222 |  |
| PCB-019 | 1.27 |  | 4.37 |  | 4.64 |  | 3.71 |  | 2.15 |  | 21.9 |  |
| PCB-020/028 | 46.6 |  | 174 |  | 152 |  | 129 |  | 78.2 |  | 602 |  |
| PCB-021/033 | 17.1 |  | 52.2 |  | 52.8 |  | 44 |  | 24.8 |  | 285 |  |
| PCB-022 | 13.6 |  | 43 |  | 39.5 |  | 32.4 |  | 20.1 |  | 210 |  |
| PCB-023 | 0.0694 | U | 0.41 | J | 0.386 | U | 0.4 |  | 0.312 | U | 0.583 | J |
| PCB-024 | 0.152 | J | 0.542 | U | 0.614 |  | 0.592 | U | 0.272 | J | 1.72 | U |
| PCB-025 | 4.82 |  | 18.7 |  | 20.1 |  | 15 |  | 8.33 |  | 43.2 |  |
| PCB-026/029 | 9.92 |  | 36.4 |  | 40.4 |  | 28.2 |  | 17.6 |  | 98 |  |
| PCB-027 | 1.47 |  | 5.21 | U | 6.42 |  | 4.54 |  | 2.76 |  | 15.2 |  |
| PCB-031 | 31.2 |  | 136 |  | 126 |  | 91 |  | 58.2 |  | 472 |  |
| PCB-032 | 5.99 |  | 22.2 |  | 20.4 |  | 15.1 |  | 9.08 |  | 90.3 |  |
| PCB-034 | 0.242 | J | 0.78 | J | 0.894 | J | 0.703 |  | 0.44 |  | 2.43 | J |
| PCB-035 | 1.7 |  | 6.73 |  | 5.86 |  | 5.16 |  | 2.87 |  | 10.9 |  |
| PCB-036 | 0.388 | J | 2.4 |  | 1.67 |  | 1.64 |  | 1.1 |  | 9.98 |  |
| PCB-037 | 11.8 |  | 45.6 |  | 35.5 |  | 31.6 |  | 21.1 |  | 137 |  |
| PCB-038 | 0.359 | J | 1.12 |  | 0.807 |  | 0.728 | U | 0.468 |  | 0.512 | U |
| PCB-039 | 0.308 | J | 1.4 |  | 1.11 |  | 0.936 |  | 0.534 |  | 3.68 | J |
| PCB-040/041/071 | 14.7 |  | 44.3 |  | 39.9 |  | 33 |  | 23 |  | 222 |  |
| PCB-042 | 6.88 |  | 21.5 |  | 18.9 |  | 15.9 |  | 10.3 |  | 113 |  |
| PCB-043 | 0.831 |  | 2.33 |  | 2.26 |  | 2.21 | U | 1.27 |  | 24.7 |  |
| PCB-044/047/065 | 27.6 |  | 82 |  | 78.4 |  | 61.9 |  | 44 |  | 437 |  |
| PCB-045/051 | 3.65 |  | 9.67 |  | 10.8 |  | 8.21 |  | 5.67 |  | 66.5 |  |
| PCB-046 | 1.03 |  | 3.21 |  | 3.3 |  | 2.62 |  | 1.76 |  | 22.2 |  |
| PCB-048 | 4.77 |  | 13.5 |  | 12.3 |  | 10.3 |  | 7.1 |  | 100 |  |
| PCB-049/069 | 20.3 |  | 64.5 |  | 65.3 |  | 49.7 |  | 34 |  | 293 |  |
| PCB-050/053 | 3.4 |  | 10.3 |  | 11.3 |  | 7.89 |  | 5.86 |  | 50.6 |  |
| PCB-052 | 38.7 |  | 103 |  | 114 |  | 80.9 |  | 55.6 |  | 583 |  |
| PCB-054 | 0.065 | U | 0.284 | J | 0.239 | J | 0.169 | J | 0.11 | J | 1.09 | U |
| PCB-055 | 0.788 |  | 2.8 |  | 2.23 |  | 1.85 |  | 1 |  | 20.5 |  |
| PCB-056 | 14.8 |  | 52 |  | 46.6 |  | 38.4 |  | 22.7 |  | 234 |  |
| PCB-057 | 0.212 | J | 0.827 | U | 0.8 |  | 0.491 |  | 0.376 | J | 2.73 | J |
| PCB-058 | 0.161 | U | 0.56 |  | 0.569 |  | 0.361 | U | 0.162 | U | 2.42 | U |
| PCB-059/062/075 | 2.76 |  | 8.28 |  | 8.01 |  | 6.06 |  | 4.48 |  | 43.8 |  |
| PCB-060 | 8.46 |  | 29.8 |  | 25.1 |  | 22.1 |  | 12.7 |  | 154 |  |
| PCB-061/070/074/076 | 64.5 |  | 218 |  | 193 |  | 163 |  | 95.9 |  | 977 |  |
| PCB-063 | 1.4 |  | 5.35 |  | 4.55 |  | 3.73 |  | 2.24 |  | 20.4 |  |
| PCB-064 | 11.8 |  | 34.3 |  | 31.4 |  | 25.8 |  | 17.7 |  | 217 |  |
| PCB-066 | 32.8 |  | 115 |  | 106 |  | 82.8 |  | 48.4 |  | 449 |  |
| PCB-067 | 1.06 |  | 4.23 |  | 3.88 |  | 3.08 |  | 1.98 |  | 16.7 |  |
| PCB-068 | 0.399 |  | 1.4 |  | 1.32 |  | 1.09 |  | 0.687 |  | 2.16 | J |
| PCB-072 | 0.583 |  | 2.11 |  | 2.06 |  | 1.33 |  | 0.985 |  | 5.51 | J |
| PCB-073 | 0.0519 | U | 0.411 |  | 0.43 |  | 0.538 | U | 0.0499 | U | 0.236 | U |


| Location ID | BB-07-S | Q | BB-08-S | Q | BB-09-S | Q | BB-10-S | Q | BB-11-S | Q | BB-12-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/10/2014 | 1356 | 9/9/2014 | 933 | 9/9/2014 | 1540 | 9/9/2014 | 1147 | 9/11/2014 | 1024 | 9/10/2014 | 1330 |
| PCB-077 | 4.27 |  | 15.7 |  | 12 |  | 10.6 |  | 6.98 |  | 36.7 |  |
| PCB-078 | 0.173 | U | 0.347 | U | 0.203 | U | 0.338 | U | 0.165 | U | 2.61 | U |
| PCB-079 | 0.9 |  | 2.47 |  | 1.96 |  | 1.86 |  | 1.01 |  | 16 |  |
| PCB-080 | 0.148 | U | 0.34 | U | 0.131 | U | 0.331 | U | 0.144 | U | 2.38 | U |
| PCB-081 | 0.184 | U | 0.71 | U | 0.456 |  | 0.507 | U | 0.351 | J | 2.55 | U |
| PCB-082 | 6.42 |  | 16.5 |  | 16.4 |  | 11.8 |  | 8.89 |  | 105 |  |
| PCB-083/099 | 32.1 |  | 96.6 |  | 86.8 |  | 73.7 |  | 54.8 |  | 456 |  |
| PCB-084 | 11.5 |  | 28.3 |  | 31 |  | 23.1 |  | 17 |  | 181 |  |
| PCB-085/116/117 | 9.43 |  | 27.7 |  | 26.5 |  | 21.5 |  | 15.2 |  | 144 |  |
| PCB-087/097/108/119/125 | 33.2 |  | 90.2 |  | 83.9 |  | 70.7 |  | 49.9 |  | 552 |  |
| PCB-088/091 | 6.47 |  | 17.6 |  | 19.3 |  | 15.3 |  | 10.7 |  | 97.6 |  |
| PCB-089 | 0.428 |  | 1.12 | J | 1.24 |  | 0.989 |  | 0.7 | U | 8.05 |  |
| PCB-090/101/113 | 47.3 |  | 134 |  | 131 |  | 108 |  | 75.1 |  | 825 |  |
| PCB-092 | 9.63 |  | 24.6 |  | 23.1 |  | 19.3 |  | 13.9 |  | 142 |  |
| PCB-093/095/098/100/102 | 33.2 |  | 84.9 |  | 93.5 |  | 72.5 |  | 50 |  | 535 |  |
| PCB-094 | 0.248 | U | 0.585 |  | 0.592 |  | 0.4 |  | 0.294 | J | 3.31 | U |
| PCB-096 | 0.255 | J | 0.714 |  | 0.828 |  | 0.529 |  | 0.46 | U | 3.62 | J |
| PCB-103 | 0.631 |  | 1.89 |  | 2.04 |  | 1.52 |  | 1.11 |  | 5.39 | U |
| PCB-104 | 0.0619 | U | 0.068 | J | 0.0642 | U | 0.0854 | U | 0.0643 | U | 0.164 | U |
| PCB-105 | 24.7 |  | 69 |  | 55.3 |  | 51.1 |  | 35.3 |  | 299 |  |
| PCB-106 | 0.167 | U | 0.57 | U | 0.155 | U | 0.183 | U | 0.14 | U | 2.56 | U |
| PCB-107/124 | 2.46 |  | 6.59 |  | 5.84 |  | 5.44 |  | 3.18 |  | 28.5 |  |
| PCB-109 | 6.01 |  | 15.1 |  | 11.8 |  | 11.5 |  | 7.01 |  | 57.3 |  |
| PCB-110/115 | 56.3 |  | 163 |  | 160 |  | 131 |  | 84.8 |  | 870 |  |
| PCB-111 | 0.096 | U | 0.212 | J | 0.233 | U | 0.148 | J | 0.145 | U | 0.517 | U |
| PCB-112 | 0.0802 | U | 0.429 | U | 0.255 | J | 0.593 |  | 0.0741 | U | 0.494 | U |
| PCB-114 | 1.39 |  | 3.69 |  | 2.87 |  | 2.79 |  | 1.84 |  | 16.7 |  |
| PCB-118 | 55.2 |  | 167 |  | 141 |  | 124 |  | 82 |  | 726 |  |
| PCB-120 | 0.328 | J | 0.921 |  | 0.772 | U | 0.857 |  | 0.564 | U | 1.85 | J |
| PCB-121 | 0.0812 | U | 0.118 | U | 0.091 | J | 0.112 | U | 0.0811 | U | 0.495 | U |
| PCB-122 | 0.957 |  | 1.79 |  | 1.96 |  | 1.75 |  | 1.11 | U | 9.67 |  |
| PCB-123 | 1.3 |  | 3.42 |  | 3.01 |  | 2.6 |  | 1.85 |  | 12.2 | U |
| PCB-126 | 0.421 |  | 1.34 |  | 1.1 |  | 0.933 |  | 0.505 | U | 3.41 | U |
| PCB-127 | 0.168 | U | 0.606 | U | 0.156 | U | 0.251 | J | 0.162 | U | 3.13 | U |
| PCB-128/166 | 0.228 | U | 37.3 |  | 32.6 |  | 33.5 |  | 18.9 |  | 155 |  |
| PCB-129/138/160/163 | 74.2 |  | 215 |  | 167 |  | 171 |  | 107 |  | 983 |  |
| PCB-130 | 3.67 | J | 11.7 | U | 9.02 |  | 8.78 |  | 7.56 | U | 64.3 |  |
| PCB-131 | 0.557 | J | 2.05 | U | 1.31 | J | 1.43 | U | 0.411 | U | 12.3 |  |
| PCB-132 | 20.2 |  | 54.9 |  | 47.7 |  | 43.7 |  | 30.6 |  | 302 |  |
| PCB-133 | 1.29 | U | 3.21 | U | 2.15 | U | 2.14 |  | 1.16 | U | 13.1 |  |
| PCB-134/143 | 3.5 | U | 8.55 |  | 7.13 |  | 6.65 |  | 4.31 | J | 44.2 |  |
| PCB-135/151/154 | 17.6 |  | 41.8 |  | 42.5 |  | 35.8 |  | 24.4 |  | 235 |  |
| PCB-136 | 6.13 |  | 12.2 |  | 14.5 |  | 10.9 |  | 7.47 |  | 83.7 |  |
| PCB-137 | 3.38 | U | 9.18 |  | 7.6 | U | 6.37 |  | 4.57 | U | 47.2 |  |
| PCB-139/140 | 1.34 | J | 2.93 | J | 2.43 | J | 2.51 |  | 1.7 | U | 15.4 |  |
| PCB-141 | 10.6 |  | 22.3 |  | 19.3 |  | 18.7 |  | 13 |  | 157 |  |
| PCB-142 | 0.246 | U | 0.71 | U | 0.827 | U | 0.37 | U | 0.436 | U | 2.02 | U |
| PCB-144 | 2.22 | U | 5.25 |  | 5.17 |  | 4.3 |  | 3.08 | J | 36.1 |  |
| PCB-145 | 0.0492 | U | 0.0478 | U | 0.156 | U | 0.0492 | U | 0.085 | U | 0.0478 | U |
| PCB-146 | 11 |  | 26.2 |  | 23.7 |  | 20.5 |  | 14.7 |  | 127 |  |
| PCB-147/149 | 45.3 |  | 113 |  | 105 |  | 102 |  | 65.7 |  | 626 |  |
| PCB-148 | 0.213 | U | 0.674 | U | 0.409 | U | 0.387 | J | 0.185 | J | 0.991 | J |
| PCB-150 | 0.0492 | U | 0.492 | J | 0.634 | U | 0.419 | J | 0.0491 | U | 1.81 | J |
| PCB-152 | 0.0492 | U | 0.0478 | U | 0.083 | U | 0.093 | U | 0.087 | U | 0.588 | J |
| PCB-153/168 | 57.1 |  | 151 |  | 129 |  | 117 |  | 82.9 |  | 741 |  |
| PCB-155 | 0.0557 | U | 0.0919 | U | 0.13 | U | 0.119 | U | 0.082 | U | 0.17 | U |
| PCB-156/157 | 8.59 |  | 21.9 |  | 18.7 |  | 17.7 |  | 11.4 | J | 96.8 |  |
| PCB-158 | 6.82 |  | 16.8 |  | 14 |  | 14.1 |  | 8.91 |  | 100 |  |
| PCB-159 | 0.863 | U | 1.87 | U | 1.74 | J | 2.07 |  | 1.02 | J | 5.85 |  |
| PCB-161 | 0.172 | U | 0.47 | U | 0.594 | U | 0.245 | U | 0.307 | U | 1.51 | U |
| PCB-162 | 0.201 | U | 1.19 | J | 0.791 | U | 0.563 | U | 0.347 | U | 2.29 | U |


| Location ID | BB-07-S | Q | BB-08-S | Q | BB-09-S | Q | BB-10-S | Q | BB-11-S | Q | BB-12-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/10/2014 | 1356 | 9/9/2014 | 933 | 9/9/2014 | 1540 | 9/9/2014 | 1147 | 9/11/2014 | 1024 | 9/10/2014 | 1330 |
| PCB-164 | 4.92 | J | 11.3 |  | 9.22 |  | 8.32 |  | 6.29 |  | 59.7 |  |
| PCB-165 | 0.198 | U | 0.545 | U | 0.7 | U | 0.284 | U | 0.342 | U | 1.72 | U |
| PCB-167 | 3.1 |  | 7.99 |  | 6.69 |  | 6.21 |  | 4.11 |  | 32 |  |
| PCB-169 | 0.186 | U | 0.793 | U | 0.696 | U | 0.466 | U | 0.404 | U | 1.72 | U |
| PCB-170 | 15.4 |  | 39.4 |  | 33.5 |  | 35.5 |  | 19.3 |  | 138 |  |
| PCB-171/173 | 4.64 |  | 13 |  | 10.8 |  | 11 |  | 5.81 |  | 51.9 |  |
| PCB-172 | 3.1 |  | 7.4 |  | 6.53 |  | 6.21 |  | 3.68 |  | 28.7 |  |
| PCB-174 | 13.1 |  | 41.6 |  | 33.5 |  | 35.9 |  | 19.2 |  | 153 |  |
| PCB-175 | 0.647 |  | 2.45 |  | 1.79 |  | 1.98 |  | 1.13 |  | 7.27 |  |
| PCB-176 | 1.8 |  | 6.53 |  | 5.27 |  | 5.94 |  | 2.84 |  | 22.1 |  |
| PCB-177 | 8.91 |  | 28.1 |  | 22.4 |  | 23.5 |  | 13.3 |  | 95.4 |  |
| PCB-178 | 3.87 |  | 13.8 |  | 10.2 |  | 10.8 |  | 5.88 |  | 34.7 |  |
| PCB-179 | 5.54 |  | 22.9 |  | 18.6 |  | 19.6 |  | 9.92 |  | 65.7 |  |
| PCB-180/193 | 31.8 |  | 92.1 |  | 69.1 |  | 80 |  | 38.9 |  | 278 |  |
| PCB-181 | 0.181 | J | 0.496 | U | 0.387 | J | 0.355 | J | 0.27 | J | 2.84 | U |
| PCB-182 | 0.0726 | U | 0.621 |  | 0.42 |  | 0.44 |  | 0.265 | U | 0.0478 | U |
| PCB-183/185 | 10.7 |  | 36.7 |  | 26.8 |  | 31 |  | 16.2 |  | 117 |  |
| PCB-184 | 0.0492 | U | 0.301 | J | 0.306 | U | 0.315 | U | 0.157 | J | 0.296 | U |
| PCB-186 | 0.0556 | U | 0.114 | U | 0.0597 | U | 0.154 | U | 0.0773 | U | 0.0478 | U |
| PCB-187 | 20.4 |  | 82.3 |  | 56.1 |  | 66.3 |  | 33.4 |  | 207 |  |
| PCB-188 | 0.081 | J | 0.407 | U | 0.272 | J | 0.323 | J | 0.164 | J | 0.437 | U |
| PCB-189 | 0.714 | U | 1.87 |  | 1.59 |  | 1.6 |  | 0.958 |  | 4.76 | J |
| PCB-190 | 2.96 | U | 9.23 |  | 6.06 |  | 7.46 |  | 3.58 | J | 24 |  |
| PCB-191 | 0.592 |  | 1.77 |  | 1.31 |  | 1.36 |  | 0.821 |  | 7.27 |  |
| PCB-192 | 0.0705 | U | 0.108 | U | 0.0507 | U | 0.145 | U | 0.0738 | U | 0.0478 | U |
| PCB-194 | 11 |  | 33.4 |  | 19.7 |  | 29.5 |  | 13 |  | 50.5 |  |
| PCB-195 | 3.86 |  | 10.9 |  | 8.37 |  | 10.1 |  | 4.46 |  | 18 |  |
| PCB-196 | 5.7 |  | 16.6 |  | 10.6 |  | 16.3 |  | 6.51 |  | 27.5 |  |
| PCB-197/199 | 1.41 |  | 6.72 |  | 4.78 |  | 5.83 |  | 2.56 |  | 10.1 |  |
| PCB-198/201 | 14.8 |  | 54.5 |  | 31.6 |  | 45.3 |  | 19.6 |  | 75.7 |  |
| PCB-200 | 1.58 | U | 5.89 |  | 4.11 |  | 6.16 |  | 2.18 |  | 10.6 |  |
| PCB-202 | 3.42 |  | 11.9 |  | 7.73 |  | 10.9 |  | 4.52 |  | 18.6 |  |
| PCB-203 | 8.97 |  | 27 |  | 17.8 |  | 29.2 |  | 10.5 |  | 39.5 |  |
| PCB-204 | 0.0492 | U | 0.192 | U | 0.0507 | U | 0.119 | J | 0.0491 | U | 0.0478 | U |
| PCB-205 | 0.439 |  | 1.46 |  | 0.971 |  | 1.22 |  | 0.544 |  | 2.74 | U |
| PCB-206 | 10.2 |  | 32.8 |  | 21.5 |  | 30.6 |  | 12.3 |  | 21.1 |  |
| PCB-207 | 1.28 | J | 4.47 |  | 3.33 | J | 5.05 |  | 1.72 | U | 3.15 | J |
| PCB-208 | 4.1 |  | 14.6 |  | 9.93 |  | 12.2 |  | 5.11 |  | 13.1 |  |
| PCB-209 | 10.9 |  | 37.6 |  | 34.1 |  | 30.5 |  | 12.9 |  | 30.8 | U |
| Total PCBs* | 1290 |  | 4300 |  | 3810 |  | 3430 |  | 2040 |  | 17700 |  |
| PCB TEQ (0 DL) | 0.0454 |  | 0.144 |  | 0.118 |  | 0.101 |  | 0.00493 |  | 0.0389 |  |
| PCB TEQ (1/2 DL) | 0.0482 |  | 0.156 |  | 0.129 |  | 0.108 |  | 0.0362 |  | 0.236 |  |
| PCB TEQ (1 DL) | 0.0510 |  | 0.168 |  | 0.139 |  | 0.115 |  | 0.0675 |  | 0.433 |  |
| PCB TEQ (KM) | 0.0458 |  | 0.145 |  | 0.119 |  | 0.101 |  | 0.0560 | L | 0.385 | L |

*total PCBs represents the sum of all detected congeners
U-the analytie was analyzed for, but was not detected above the reported sample quantitation limi
$J$-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sampl UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximat $\epsilon$ L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detect KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicate

Table F-3. Summary of Bellingham Bay Polychlorinated Biphenyl (PCB) Congener Data (continued).

| Location ID | BB-13-S | Q | BB-14-S | Q | BB-15-S | Q | BB-16-S | Q | BB-17-S | Q | BB-18-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1423 | 9/11/2014 | 853 | 9/9/2014 | 915 | 9/11/2014 | 1044 | 9/9/2014 | 1524 | 9/9/2014 | 1042 |
| PCB Congeners (ng/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |
| PCB-001 | 31.7 |  | 7.16 |  | 39.2 |  | 20.5 |  | 33.7 |  | 36.8 |  |
| PCB-002 | 21.6 |  | 8.91 |  | 26.2 |  | 10.8 |  | 23.3 |  | 16.7 |  |
| PCB-003 | 35.5 |  | 9.46 |  | 35.9 |  | 23.1 |  | 42.8 |  | 38.7 |  |
| PCB-004 | 9.59 |  | 4.9 |  | 26 |  | 9.84 |  | 13.7 |  | 16.1 |  |
| PCB-005 | 0.874 |  | 0.359 | J | 1.55 |  | 2.04 | U | 1.17 |  | 1.56 |  |
| PCB-006 | 9.27 |  | 5.8 |  | 19.8 |  | 12.1 |  | 16.6 |  | 18.7 |  |
| PCB-007 | 3.39 |  | 1.21 | U | 4.99 |  | 1.85 | U | 4.13 |  | 4.51 |  |
| PCB-008 | 53.5 |  | 21 |  | 103 |  | 43.1 |  | 63.5 |  | 71.3 |  |
| PCB-009 | 1.97 |  | 1 |  | 4.67 |  | 1.79 | U | 2.44 |  | 3.4 |  |
| PCB-010 | 0.412 |  | 0.237 | U | 0.926 |  | 1.74 | U | 0.521 |  | 0.694 |  |
| PCB-011 | 35.6 |  | 26 |  | 43.1 |  | 23 |  | 37 |  | 34.8 |  |
| PCB-012/013 | 10.5 |  | 6.04 |  | 15 |  | 10.5 |  | 16.2 |  | 15.7 |  |
| PCB-014 | 0.801 | U | 0.596 | U | 1.02 | U | 1.91 | U | 1.69 | U | 1.13 | U |
| PCB-015 | 64.9 |  | 23.9 |  | 87.7 |  | 41.8 |  | 73.7 |  | 70.7 |  |
| PCB-016 | 10.9 |  | 6.43 | U | 37.9 |  | 9.53 |  | 13.4 |  | 15 |  |
| PCB-017 | 16.5 |  | 10.8 |  | 43.2 |  | 18 |  | 23.2 |  | 24.2 |  |
| PCB-018/030 | 27.1 |  | 17.7 |  | 75.6 |  | 29.4 |  | 41.4 |  | 43.8 |  |
| PCB-019 | 2.54 |  | 2.2 | U | 8.5 |  | 3.06 |  | 4.05 |  | 4.29 |  |
| PCB-020/028 | 98 |  | 57.7 |  | 224 |  | 95.9 |  | 128 |  | 152 |  |
| PCB-021/033 | 36.6 |  | 21.9 |  | 108 |  | 31.8 |  | 46.1 |  | 64.3 |  |
| PCB-022 | 27.3 |  | 15.4 |  | 75.5 |  | 24.9 |  | 34 |  | 38.2 |  |
| PCB-023 | 0.235 | J | 0.461 | U | 0.393 | J | 0.141 | J | 0.276 | J | 0.348 | U |
| PCB-024 | 0.436 |  | 0.331 | U | 1.6 |  | 0.418 |  | 0.511 |  | 0.568 |  |
| PCB-025 | 9.18 |  | 5.86 | J | 18.6 |  | 10.9 |  | 16.9 |  | 16.9 |  |
| PCB-026/029 | 18 |  | 12.2 |  | 37.7 |  | 23.2 |  | 32.4 |  | 33.3 |  |
| PCB-027 | 3.19 |  | 2.31 | J | 7.36 |  | 3.8 |  | 5.46 |  | 5.43 |  |
| PCB-031 | 75.9 |  | 45 |  | 196 |  | 72.4 |  | 107 |  | 118 |  |
| PCB-032 | 11.4 |  | 7.28 |  | 31 |  | 12.3 |  | 16.7 |  | 21.3 |  |
| PCB-034 | 0.559 |  | 0.447 | U | 1.04 |  | 0.544 |  | 0.72 |  | 0.789 |  |
| PCB-035 | 4.14 |  | 1.62 | J | 6.27 |  | 3.29 |  | 5.23 |  | 4.98 |  |
| PCB-036 | 1.64 |  | 0.895 | J | 1.57 |  | 0.801 |  | 1.46 |  | 1.14 |  |
| PCB-037 | 26.4 |  | 16.4 |  | 62.6 |  | 24.3 |  | 32 |  | 38 |  |
| PCB-038 | 0.675 |  | 0.461 | U | 0.603 |  | 0.385 | J | 0.75 | U | 0.8 |  |
| PCB-039 | 0.718 |  | 0.44 | U | 1.22 |  | 0.583 |  | 0.83 |  | 1.04 |  |
| PCB-040/041/071 | 30.2 |  | 18.8 |  | 59.5 |  | 29.1 |  | 38.3 |  | 46.9 |  |
| PCB-042 | 13.8 |  | 8.78 |  | 26.5 |  | 13.5 |  | 18 |  | 23 |  |
| PCB-043 | 1.77 |  | 1.21 |  | 3.63 |  | 1.66 |  | 2.31 |  | 2.66 |  |
| PCB-044/047/065 | 54.4 |  | 36.9 |  | 91.2 |  | 54.2 |  | 70.9 |  | 103 |  |
| PCB-045/051 | 6.55 |  | 4.63 |  | 14.4 |  | 7.75 |  | 9.51 |  | 10.9 |  |
| PCB-046 | 2.21 |  | 1.37 |  | 4.87 |  | 2.24 |  | 2.92 |  | 3.73 |  |
| PCB-048 | 9.29 |  | 6.01 |  | 20.9 |  | 9.13 |  | 11.5 |  | 14.6 |  |
| PCB-049/069 | 39.4 |  | 27.4 |  | 61.1 |  | 42.2 |  | 55.6 |  | 78.6 |  |
| PCB-050/053 | 6.23 |  | 4.26 |  | 12 |  | 7.42 |  | 9.85 |  | 11.8 |  |
| PCB-052 | 68.2 |  | 46.5 |  | 91.4 |  | 68.8 |  | 96.3 |  | 159 |  |
| PCB-054 | 0.127 | J | 0.127 | J | 0.246 | J | 0.172 | J | 0.225 | J | 0.247 | J |
| PCB-055 | 1.86 |  | 0.785 | U | 4.19 |  | 1.4 |  | 2.37 |  | 2.13 |  |
| PCB-056 | 31.6 |  | 18.9 |  | 62.8 |  | 27.5 |  | 40.6 |  | 50.5 |  |
| PCB-057 | 0.414 |  | 0.234 | J | 0.879 |  | 0.454 |  | 0.654 |  | 0.712 |  |
| PCB-058 | 0.315 | J | 0.0808 | U | 0.31 | J | 0.24 | U | 0.468 |  | 0.508 | U |
| PCB-059/062/075 | 5.46 |  | 3.55 |  | 10.2 |  | 5.63 |  | 7.48 |  | 9.2 |  |
| PCB-060 | 18.8 |  | 0.0829 | UJ | 40.8 |  | 15.1 |  | 23.1 |  | 28 |  |
| PCB-061/070/074/076 | 128 |  | 79.3 |  | 221 |  | 116 |  | 171 |  | 248 |  |
| PCB-063 | 2.92 |  | 1.88 |  | 5.45 |  | 2.7 |  | 4.11 |  | 5.19 |  |
| PCB-064 | 22.5 |  | 14.5 |  | 40.7 |  | 21.6 |  | 29.4 |  | 39.4 |  |
| PCB-066 | 70.1 |  | 42.2 |  | 118 |  | 57.3 |  | 94.5 |  | 115 |  |
| PCB-067 | 2.48 |  | 1.54 |  | 5.17 |  | 2.32 |  | 3.56 |  | 4.08 |  |
| PCB-068 | 0.832 |  | 0.61 |  | 1.02 |  | 0.762 |  | 1.3 |  | 1.32 |  |
| PCB-072 | 1.12 |  | 0.729 |  | 1.49 |  | 1.08 |  | 1.86 |  | 2.2 |  |
| PCB-073 | 0.508 |  | 0.049 | U | 0.593 |  | 0.0623 | U | 0.397 |  | 0.483 |  |


| Location ID | BB-13-S | Q | BB-14-S | Q | BB-15-S | Q | BB-16-S | Q | BB-17-S | Q | BB-18-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1423 | 9/11/2014 | 853 | 9/9/2014 | 915 | 9/11/2014 | 1044 | 9/9/2014 | 1524 | 9/9/2014 | 1042 |
| PCB-077 | 9.97 |  | 5.73 |  | 14.1 |  | 7.27 |  | 10.8 |  | 14.9 |  |
| PCB-078 | 0.199 | U | 0.0798 | U | 0.263 | U | 0.245 | U | 0.213 | J | 0.35 | U |
| PCB-079 | 1.45 |  | 0.812 |  | 1.48 |  | 1.05 |  | 1.81 |  | 3.77 |  |
| PCB-080 | 0.175 | U | 0.365 | J | 0.258 | U | 0.213 | U | 0.188 | U | 0.343 | U |
| PCB-081 | 0.432 |  | 0.275 | J | 0.643 |  | 0.29 | J | 0.439 | U | 0.615 |  |
| PCB-082 | 13.7 |  | 7.93 |  | 14.6 |  | 12.8 |  | 16.6 |  | 27.5 |  |
| PCB-083/099 | 70.3 |  | 45.1 |  | 65.3 |  | 61.6 |  | 78.2 |  | 158 |  |
| PCB-084 | 20.8 |  | 15.4 |  | 20.9 |  | 22 |  | 25.2 |  | 64.9 |  |
| PCB-085/116/117 | 20.6 |  | 13.1 |  | 20.6 |  | 17 |  | 22.8 |  | 46.2 |  |
| PCB-087/097/108/119/125 | 67 |  | 44.4 |  | 63.9 |  | 57.7 |  | 74 |  | 174 |  |
| PCB-088/091 | 12.7 |  | 9.09 |  | 13.3 |  | 13 |  | 16.4 |  | 35.7 |  |
| PCB-089 | 0.921 |  | 0.612 |  | 1.2 |  | 0.87 |  | 1.08 |  | 2.62 | J |
| PCB-090/101/113 | 92.8 |  | 63.7 |  | 91.8 |  | 86.1 |  | 108 |  | 265 |  |
| PCB-092 | 18 |  | 11.6 |  | 16.5 |  | 16.1 |  | 20.1 |  | 49 |  |
| PCB-093/095/098/100/102 | 63.9 |  | 45.5 |  | 62.5 |  | 60.6 |  | 72.8 |  | 188 |  |
| PCB-094 | 0.453 | U | 0.308 | U | 0.404 | U | 0.402 |  | 0.424 |  | 1.1 |  |
| PCB-096 | 0.527 |  | 0.329 | U | 0.597 |  | 0.496 | U | 0.605 |  | 1.26 |  |
| PCB-103 | 1.14 |  | 0.891 |  | 1.27 |  | 1.18 |  | 1.56 |  | 3.4 |  |
| PCB-104 | 0.048 | U | 0.049 | U | 0.0816 | U | 0.0543 | U | 0.072 | U | 0.068 | J |
| PCB-105 | 44.8 |  | 27.2 |  | 44.5 |  | 37.7 |  | 51.3 |  | 103 |  |
| PCB-106 | 0.282 | U | 0.165 | U | 0.489 | U | 0.158 | U | 0.291 | U | 0.385 | U |
| PCB-107/124 | 4.53 |  | 2.74 |  | 3.9 |  | 3.8 |  | 5.08 |  | 11.3 |  |
| PCB-109 | 9.46 |  | 5.82 |  | 9.17 |  | 7.27 |  | 10.5 |  | 21.9 |  |
| PCB-110/115 | 114 |  | 68.6 |  | 104 |  | 98.5 |  | 136 |  | 311 |  |
| PCB-111 | 0.21 | J | 0.15 | J | 0.171 | U | 0.121 | U | 0.219 | J | 0.204 | U |
| PCB-112 | 0.252 | U | 0.0612 | U | 0.21 | U | 0.0654 | U | 0.261 | J | 0.387 |  |
| PCB-114 | 2.34 |  | 1.44 |  | 2.45 |  | 2.09 |  | 2.72 |  | 5.7 |  |
| PCB-118 | 104 |  | 64.2 |  | 101 |  | 89.1 |  | 126 |  | 256 |  |
| PCB-120 | 0.747 |  | 0.395 |  | 0.762 |  | 0.509 | U | 0.728 | U | 1.07 | U |
| PCB-121 | 0.102 | U | 0.067 | U | 0.156 | U | 0.0716 | U | 0.107 | U | 0.092 | U |
| PCB-122 | 1.57 |  | 0.844 |  | 1.4 |  | 1.34 |  | 1.62 |  | 3.46 |  |
| PCB-123 | 2.4 |  | 1.62 |  | 2.09 |  | 2.24 |  | 2.78 |  | 4.7 |  |
| PCB-126 | 0.735 |  | 0.487 |  | 0.848 |  | 0.446 |  | 0.732 | U | 1.37 |  |
| PCB-127 | 0.302 | U | 0.17 | U | 0.521 | U | 0.182 | U | 0.289 | U | 0.547 |  |
| PCB-128/166 | 23.1 |  | 0.236 | UJ | 23.2 |  | 0.474 | U | 28.9 |  | 75.7 |  |
| PCB-129/138/160/163 | 132 |  | 85.6 |  | 136 |  | 122 |  | 155 |  | 386 |  |
| PCB-130 | 7.44 |  | 5.42 | UJ | 9 |  | 8.37 | U | 9.09 |  | 20.2 |  |
| PCB-131 | 1.19 | J | 0.975 | J | 1.44 | U | 1.59 | U | 1.06 | U | 4.07 | U |
| PCB-132 | 36 |  | 25.3 |  | 44.1 |  | 35.9 |  | 38 |  | 99.8 |  |
| PCB-133 | 1.32 | U | 1.48 | U | 2.64 |  | 2.48 | U | 1.88 | U | 5.55 |  |
| PCB-134/143 | 4.78 |  | 2.77 | UJ | 6.15 |  | 5.32 | J | 5.41 | U | 13.9 |  |
| PCB-135/151/154 | 27.6 |  | 19.5 |  | 37.6 |  | 30.2 |  | 32.2 |  | 74.9 |  |
| PCB-136 | 9.39 |  | 6.08 |  | 12.2 |  | 9.4 |  | 10.5 |  | 26.7 |  |
| PCB-137 | 5.44 |  | 3.81 | J | 6.67 |  | 6.07 | U | 7.99 |  | 17.5 |  |
| PCB-139/140 | 1.7 | J | 1.6 | J | 2.81 |  | 2.3 | J | 1.84 | J | 6.29 |  |
| PCB-141 | 13.5 |  | 11.3 |  | 16.3 |  | 16.7 |  | 17.3 |  | 43.4 |  |
| PCB-142 | 0.644 | U | 0.262 | U | 0.476 | U | 0.526 | U | 0.688 | U | 0.755 | U |
| PCB-144 | 3.42 |  | 2.57 | J | 4.56 |  | 3.62 | J | 4.03 |  | 10.3 |  |
| PCB-145 | 0.056 | U | 0.185 | U | 0.158 | U | 0.072 | U | 0.127 | U | 0.173 | U |
| PCB-146 | 19.3 |  | 11.7 |  | 23.3 |  | 14.9 |  | 19.2 |  | 43.8 |  |
| PCB-147/149 | 74.2 |  | 47.4 |  | 98.4 |  | 73.8 |  | 95.5 |  | 203 |  |
| PCB-148 | 0.323 | U | 0.223 | U | 0.339 | J | 0.259 | U | 0.518 | J | 1.11 | U |
| PCB-150 | 0.278 | U | 0.049 | U | 0.329 | U | 0.0501 | U | 0.311 | U | 0.852 | J |
| PCB-152 | 0.049 | U | 0.049 | U | 0.142 | U | 0.0501 | U | 0.127 | U | 0.0469 | U |
| PCB-153/168 | 96.8 |  | 67.6 |  | 115 |  | 91.2 |  | 109 |  | 244 |  |
| PCB-155 | 0.138 | U | 0.065 | U | 0.0901 | U | 0.075 | U | 0.126 | U | 0.212 | J |
| PCB-156/157 | 14.3 |  | 8.69 |  | 12.7 |  | 12.3 |  | 17.5 |  | 39 |  |
| PCB-158 | 10.7 |  | 6.99 | U | 11.6 |  | 10.3 |  | 12.8 |  | 37.3 |  |
| PCB-159 | 1.13 | U | 0.642 | U | 1.27 | U | 1.29 | J | 1.65 | J | 3.17 | J |
| PCB-161 | 0.463 | U | 0.184 | U | 0.33 | U | 0.371 | U | 0.494 | U | 0.5 | U |
| PCB-162 | 0.616 | U | 0.267 | U | 0.641 |  | 0.48 | U | 0.659 | U | 1.44 | J |


| Location ID | BB-13-S | Q | BB-14-S | Q | BB-15-S | Q | BB-16-S | Q | BB-17-S | Q | BB-18-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1423 | 9/11/2014 | 853 | 9/9/2014 | 915 | 9/11/2014 | 1044 | 9/9/2014 | 1524 | 9/9/2014 | 1042 |
| PCB-164 | 6.62 | U | 5.87 | U | 7.52 |  | 7.71 |  | 7.96 |  | 20.4 |  |
| PCB-165 | 0.545 | U | 0.205 | U | 0.392 | U | 0.413 | U | 0.582 | U | 0.579 | U |
| PCB-167 | 5.01 |  | 3.23 |  | 4.47 |  | 4.27 |  | 6.13 |  | 13.6 |  |
| PCB-169 | 0.518 | U | 0.289 | U | 0.285 | U | 0.406 | U | 0.604 | U | 1.01 | U |
| PCB-170 | 24.2 |  | 14.9 |  | 24.1 |  | 20.5 |  | 28 |  | 60 |  |
| PCB-171/173 | 8.23 |  | 5.28 |  | 8.13 |  | 6.96 |  | 9.01 |  | 21.2 |  |
| PCB-172 | 4.51 |  | 2.9 |  | 4.17 | U | 3.84 |  | 5.52 |  | 11.2 |  |
| PCB-174 | 22.8 |  | 15.6 |  | 23.5 |  | 23.1 |  | 28.1 |  | 67.2 |  |
| PCB-175 | 1.38 |  | 0.836 | U | 1.37 |  | 1.21 |  | 1.66 |  | 3.56 |  |
| PCB-176 | 3.43 |  | 2.33 |  | 3.68 |  | 3.4 |  | 4.15 |  | 10.5 |  |
| PCB-177 | 16.7 |  | 11 |  | 17.4 |  | 14.7 |  | 20.1 |  | 40.1 |  |
| PCB-178 | 7.82 |  | 4.97 |  | 7.7 |  | 6.33 |  | 8.15 |  | 20.3 |  |
| PCB-179 | 12.5 |  | 7.85 |  | 13.5 |  | 11.6 |  | 15.3 |  | 36.9 |  |
| PCB-180/193 | 49.9 |  | 31 |  | 48 |  | 41.6 |  | 58.6 |  | 137 |  |
| PCB-181 | 0.301 | J | 0.259 | U | 0.274 | U | 0.31 | U | 0.372 | J | 0.853 |  |
| PCB-182 | 0.368 | J | 0.277 | U | 0.3 | U | 0.285 | U | 0.373 | J | 1.11 |  |
| PCB-183/185 | 19.2 |  | 12.9 |  | 19.5 |  | 18.3 |  | 22.2 |  | 59.3 |  |
| PCB-184 | 0.215 | J | 0.115 | U | 0.228 | J | 0.111 | U | 0.232 | U | 0.215 | J |
| PCB-186 | 0.0896 | U | 0.0601 | U | 0.162 | U | 0.102 | U | 0.0948 | U | 0.083 | U |
| PCB-187 | 41.7 |  | 28.2 |  | 46 |  | 35.6 |  | 47.8 |  | 125 |  |
| PCB-188 | 0.228 | U | 0.146 | J | 0.17 | U | 0.166 | U | 0.237 | J | 0.762 |  |
| PCB-189 | 1.06 |  | 0.703 |  | 1.09 |  | 0.994 |  | 1.31 |  | 2.61 |  |
| PCB-190 | 4.49 | U | 2.49 | J | 4.73 |  | 4.29 | J | 5.38 |  | 14.5 |  |
| PCB-191 | 0.988 |  | 0.641 |  | 0.888 |  | 0.858 |  | 1.22 |  | 2.78 |  |
| PCB-192 | 0.0863 | U | 0.0582 | U | 0.153 | U | 0.0975 | U | 0.09 | U | 0.0783 | U |
| PCB-194 | 15.9 |  | 10.2 |  | 16.4 |  | 12.3 |  | 16.5 |  | 50.5 |  |
| PCB-195 | 5.74 |  | 3.26 |  | 5.43 |  | 4.47 |  | 6.34 |  | 15.6 |  |
| PCB-196 | 8.28 |  | 5.07 |  | 8.81 |  | 6.69 |  | 9.55 |  | 27.4 |  |
| PCB-197/199 | 3.17 |  | 1.9 |  | 3.33 |  | 2.51 |  | 3.93 |  | 10.8 |  |
| PCB-198/201 | 24.5 |  | 15.2 |  | 30.5 |  | 18.4 |  | 28.9 |  | 84 |  |
| PCB-200 | 3.15 |  | 1.62 |  | 3.14 |  | 2.28 |  | 3.27 |  | 9.14 |  |
| PCB-202 | 5.97 |  | 3.6 |  | 6.31 |  | 4.47 |  | 6.53 |  | 18.9 |  |
| PCB-203 | 12.7 |  | 8.51 |  | 13.2 |  | 11.4 |  | 15.4 |  | 45.7 |  |
| PCB-204 | 0.056 | U | 0.09 | U | 0.171 | U | 0.126 | U | 0.0495 | U | 0.0469 | U |
| PCB-205 | 0.757 |  | 0.405 |  | 0.877 |  | 0.578 | U | 0.906 |  | 1.96 |  |
| PCB-206 | 19.9 |  | 9.7 |  | 19.1 |  | 12.4 |  | 16.8 |  | 46.2 |  |
| PCB-207 | 3.21 |  | 1.53 | U | 2.72 | U | 2.13 | J | 2.54 | U | 6.7 |  |
| PCB-208 | 8.56 |  | 4.14 |  | 8.23 |  | 5.53 |  | 7.81 |  | 22.3 |  |
| PCB-209 | 43.5 |  | 11.5 |  | 19.8 |  | 14.8 |  | 21.8 |  | 54 |  |
| Total PCBs* | 2700 |  | 1600 |  | 3830 |  | 2380 |  | 3290 |  | 6030 |  |
| PCB TEQ (0 DL) | 0.0798 |  | 0.0526 |  | 0.0915 |  | 0.0499 |  | 0.00731 |  | 0.151 |  |
| PCB TEQ (1/2 DL) | 0.0876 |  | 0.0569 |  | 0.0957 |  | 0.0560 |  | 0.0530 |  | 0.167 |  |
| PCB TEQ (1 DL) | 0.0954 |  | 0.0612 |  | 0.100 |  | 0.0621 |  | 0.0988 |  | 0.182 |  |
| PCB TEQ (KM) | 0.0805 |  | 0.0530 |  | 0.0922 |  | 0.0505 |  | 0.0814 | L | 0.153 |  |

*total PCBs represents the sum of all detected congeners
U-the analytie was analyzed for, but was not detected above the reported sample quantitation limi
J-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sampl UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximat $\epsilon$ L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detect KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicate

Table F-3. Summary of Bellingham Bay Polychlorinated Biphenyl (PCB) Congener Data (continued).

| Location ID | BB-19-S | Q | BB-20-S | Q | BB-20-D | Q | BB-21-S | Q | BB-22-S | Q | BB-23-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1326 | 9/10/2014 | 1530 | 9/10/2014 | 1530 | 9/9/2014 | 1453 | 9/9/2014 | 1026 | 9/10/2014 | 942 |
| PCB Congeners (ng/kg DW) |  |  |  |  |  |  |  |  |  |  |  |  |
| PCB-001 | 50.9 |  | 19.6 |  | 14.9 |  | 28.7 |  | 67.2 | J | 90.7 |  |
| PCB-002 | 30.6 |  | 13.3 |  | 9.68 |  | 17.7 |  | 43.1 | J | 30.8 |  |
| PCB-003 | 51.3 |  | 21.8 |  | 19.7 |  | 37.1 |  | 73.4 | J | 98.7 |  |
| PCB-004 | 26.6 |  | 11 |  | 9.65 |  | 8.09 |  | 32.1 |  | 99.3 |  |
| PCB-005 | 1.73 |  | 0.718 |  | 0.743 |  | 1.11 |  | 2.55 |  | 4.95 |  |
| PCB-006 | 36.1 |  | 15.7 |  | 13 |  | 10.7 |  | 31.3 |  | 154 |  |
| PCB-007 | 4.88 |  | 2.72 |  | 2.12 |  | 3.54 |  | 8.03 |  | 14.4 |  |
| PCB-008 | 109 |  | 44.4 |  | 37.7 |  | 53 |  | 139 |  | 357 |  |
| PCB-009 | 4.34 |  | 2.24 |  | 1.8 |  | 2.22 |  | 5.59 |  | 14.2 |  |
| PCB-010 | 0.848 |  | 0.44 |  | 0.366 | J | 0.359 | J | 1.1 |  | 3.29 |  |
| PCB-011 | 37.7 |  | 28.6 |  | 25 |  | 40.8 |  | 51.3 |  | 48.5 |  |
| PCB-012/013 | 26.7 |  | 12.9 |  | 10.8 |  | 13.5 |  | 30.5 |  | 90.7 |  |
| PCB-014 | 1.24 |  | 0.555 | U | 0.491 |  | 1.19 |  | 2.3 |  | 1.25 |  |
| PCB-015 | 92.1 |  | 51.7 |  | 43.8 |  | 70.5 |  | 154 |  | 259 |  |
| PCB-016 | 26.8 |  | 11.2 |  | 9.89 |  | 10.4 |  | 47.5 |  | 102 |  |
| PCB-017 | 49.1 |  | 20.3 |  | 17.8 |  | 16.6 |  | 76.6 |  | 200 |  |
| PCB-018/030 | 97.8 |  | 33.2 |  | 28.5 |  | 28.8 |  | 121 |  | 358 |  |
| PCB-019 | 10.1 |  | 3.98 |  | 3.33 |  | 2.53 |  | 13 |  | 39.2 |  |
| PCB-020/028 | 224 |  | 135 |  | 111 |  | 111 |  | 351 |  | 863 |  |
| PCB-021/033 | 79 |  | 45 |  | 40.6 |  | 45.8 |  | 122 | J | 307 |  |
| PCB-022 | 60 |  | 34.9 |  | 30.6 |  | 30.1 |  | 98.3 |  | 221 |  |
| PCB-023 | 0.381 | J | 0.422 | U | 0.53 | U | 0.251 | J | 0.603 | J | 0.623 |  |
| PCB-024 | 1.28 |  | 0.461 |  | 0.385 | J | 0.485 |  | 1.64 |  | 4.13 |  |
| PCB-025 | 36.5 |  | 16.7 |  | 13.6 |  | 11 |  | 32.4 |  | 155 |  |
| PCB-026/029 | 74.7 |  | 32.2 |  | 27.6 |  | 21.7 |  | 67.7 |  | 315 |  |
| PCB-027 | 13.3 |  | 4.51 |  | 3.85 |  | 3.17 |  | 13.9 |  | 52.9 |  |
| PCB-031 | 224 |  | 99.3 |  | 80.1 |  | 86.3 |  | 253 |  | 836 |  |
| PCB-032 | 34.5 |  | 15.9 |  | 13.8 |  | 12 |  | 45.1 | J | 140 |  |
| PCB-034 | 1.29 |  | 0.701 |  | 0.537 |  | 0.725 |  | 1.52 |  | 5.31 |  |
| PCB-035 | 5.82 |  | 4.31 |  | 3.51 |  | 5.14 |  | 10.6 |  | 14.7 |  |
| PCB-036 | 1.41 |  | 1.04 |  | 0.761 |  | 2.02 |  | 2.19 |  | 1.29 |  |
| PCB-037 | 45.9 |  | 32.7 |  | 26.7 |  | 29 |  | 78.9 |  | 166 |  |
| PCB-038 | 0.771 |  | 0.654 |  | 0.454 |  | 0.888 | U | 1.43 |  | 1.31 |  |
| PCB-039 | 1.32 |  | 0.872 |  | 0.658 |  | 0.985 |  | 2.55 |  | 4.54 |  |
| PCB-040/041/071 | 74 |  | 35.6 |  | 29.2 |  | 31.6 |  | 129 |  | 246 |  |
| PCB-042 | 33.8 |  | 17.1 |  | 14 |  | 15.2 |  | 62.9 |  | 119 |  |
| PCB-043 | 4.05 |  | 2.17 |  | 1.74 |  | 1.83 |  | 6.59 |  | 16.5 |  |
| PCB-044/047/065 | 146 |  | 66.4 |  | 55.6 |  | 57.7 |  | 245 |  | 493 |  |
| PCB-045/051 | 22.4 |  | 8.7 |  | 7.76 |  | 6.46 |  | 35 |  | 87.1 |  |
| PCB-046 | 6.74 |  | 2.79 |  | 2.26 |  | 2.31 |  | 11.6 |  | 23.8 |  |
| PCB-048 | 20.6 |  | 11.6 |  | 9.16 |  | 10.1 |  | 41.6 |  | 77.4 |  |
| PCB-049/069 | 129 |  | 54.5 |  | 44.4 |  | 43.8 |  | 173 |  | 458 |  |
| PCB-050/053 | 25 |  | 9.24 |  | 7.52 |  | 6.37 |  | 32.9 |  | 94.5 |  |
| PCB-052 | 234 |  | 91.6 |  | 78.9 |  | 73.4 |  | 309 |  | 738 |  |
| PCB-054 | 0.522 |  | 0.189 | J | 0.161 | U | 0.12 | J | 0.524 |  | 1.91 |  |
| PCB-055 | 3.71 |  | 1.92 |  | 1.46 |  | 1.81 |  | 4.98 |  | 13.4 |  |
| PCB-056 | 64.3 |  | 40.4 |  | 30.3 |  | 35.5 |  | 112 |  | 218 |  |
| PCB-057 | 1.23 |  | 0.601 |  | 0.479 |  | 0.454 |  | 1.29 |  | 4.03 |  |
| PCB-058 | 0.81 | U | 0.165 | U | 0.21 | U | 0.433 | U | 1.11 |  | 2.59 |  |
| PCB-059/062/075 | 14.5 |  | 6.9 |  | 5.73 |  | 5.86 |  | 23.4 |  | 50.8 |  |
| PCB-060 | 33.5 |  | 22.5 |  | 16.4 |  | 20.5 |  | 64.2 |  | 104 |  |
| PCB-061/070/074/076 | 297 |  | 168 |  | 131 |  | 148 |  | 490 |  | 941 |  |
| PCB-063 | 6.16 |  | 3.92 |  | 2.94 |  | 3.47 |  | 9.76 |  | 23.1 |  |
| PCB-064 | 59.7 |  | 27.6 |  | 22.6 |  | 24.8 |  | 101 |  | 196 |  |
| PCB-066 | 157 |  | 90 |  | 68.6 |  | 81.1 |  | 247 |  | 544 |  |
| PCB-067 | 5.2 |  | 3.52 |  | 2.67 |  | 2.84 |  | 8.66 |  | 21.3 |  |
| PCB-068 | 1.69 |  | 1.11 |  | 0.841 |  | 0.974 |  | 2.32 |  | 6.32 |  |
| PCB-072 | 2.86 |  | 1.62 |  | 1.2 |  | 1.28 |  | 3.59 |  | 11.3 |  |
| PCB-073 | 1.39 |  | 0.0545 | U | 0.049 | U | 0.344 | J | 0.583 |  | 5.05 |  |


| Location ID | BB-19-S | Q | BB-20-S | Q | BB-20-D | Q | BB-21-S | Q | BB-22-S | Q | BB-23-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1326 | 9/10/2014 | 1530 | 9/10/2014 | 1530 | 9/9/2014 | 1453 | 9/9/2014 | 1026 | 9/10/2014 | 942 |
| PCB-077 | 15.8 |  | 10.5 |  | 8.22 |  | 11.9 |  | 25.5 |  | 45.9 |  |
| PCB-078 | 0.288 | U | 0.178 | U | 0.226 | U | 0.305 | U | 0.617 | U | 0.925 | U |
| PCB-079 | 3.49 |  | 1.92 |  | 1.84 |  | 1.83 |  | 6.13 |  | 8.3 |  |
| PCB-080 | 0.253 | U | 1.11 | J | 0.193 | UJ | 0.268 | U | 0.55 | U | 0.883 | U |
| PCB-081 | 0.608 | U | 0.504 |  | 0.401 |  | 0.511 |  | 1.14 | U | 1.73 |  |
| PCB-082 | 31 |  | 15.7 |  | 14.3 |  | 13.1 |  | 43.7 |  | 62.8 |  |
| PCB-083/099 | 178 |  | 73.9 |  | 64.6 |  | 75.4 |  | 266 |  | 420 |  |
| PCB-084 | 68.8 |  | 24.7 |  | 23.3 |  | 20.7 |  | 104 |  | 157 |  |
| PCB-085/116/117 | 50.3 |  | 22.1 |  | 18.7 |  | 22.3 |  | 77.3 |  | 106 |  |
| PCB-087/097/108/119/125 | 184 |  | 72.5 |  | 64.5 |  | 69.8 |  | 275 |  | 365 |  |
| PCB-088/091 | 45 |  | 15.2 |  | 13.4 |  | 13.3 |  | 58 |  | 109 |  |
| PCB-089 | 2.43 |  | 0.881 |  | 0.802 |  | 0.935 |  | 3.64 |  | 5.89 |  |
| PCB-090/101/113 | 245 |  | 108 |  | 96 |  | 101 |  | 414 |  | 568 |  |
| PCB-092 | 47.1 |  | 19.8 |  | 17.8 |  | 19.1 |  | 74.1 |  | 109 |  |
| PCB-093/095/098/100/102 | 201 |  | 71.7 |  | 65.2 |  | 64.4 |  | 325 |  | 448 |  |
| PCB-094 | 1.13 |  | 0.514 |  | 0.392 | U | 0.453 |  | 1.61 |  | 2.56 |  |
| PCB-096 | 1.79 |  | 0.594 |  | 0.469 |  | 0.483 |  | 2.26 |  | 4.72 |  |
| PCB-103 | 3.42 |  | 1.52 |  | 1.31 |  | 1.25 |  | 5.08 |  | 10.3 |  |
| PCB-104 | 0.142 | J | 0.0543 | U | 0.049 | U | 0.0512 | U | 0.111 | U | 0.469 |  |
| PCB-105 | 97.4 |  | 53.4 |  | 45.8 |  | 48.8 |  | 164 | J | 223 |  |
| PCB-106 | 0.287 | U | 0.244 | U | 0.205 | U | 0.277 | U | 0.355 | U | 1.27 | U |
| PCB-107/124 | 10.1 |  | 5.97 |  | 4.95 |  | 4.98 |  | 16 |  | 21.7 |  |
| PCB-109 | 20.5 |  | 12.7 |  | 10.7 |  | 10.4 |  | 30.2 |  | 49.9 |  |
| PCB-110/115 | 329 |  | 128 |  | 113 |  | 120 |  | 480 |  | 744 |  |
| PCB-111 | 0.15 | U | 0.176 | J | 0.083 | U | 0.176 | J | 0.437 |  | 0.395 | J |
| PCB-112 | 0.553 |  | 0.0548 | U | 0.0684 | U | 0.361 | J | 0.121 | U | 1.96 |  |
| PCB-114 | 5.51 |  | 2.95 |  | 2.57 |  | 2.55 |  | 8.02 |  | 13 |  |
| PCB-118 | 244 |  | 130 |  | 111 |  | 114 |  | 384 |  | 607 |  |
| PCB-120 | 0.728 | U | 0.726 | U | 0.495 |  | 0.834 | U | 1.77 |  | 2.2 |  |
| PCB-121 | 0.14 | U | 0.0555 | U | 0.0693 | U | 0.087 | U | 0.195 | U | 0.191 | U |
| PCB-122 | 3.28 |  | 1.82 |  | 1.72 |  | 1.66 |  | 5 |  | 7.56 |  |
| PCB-123 | 5.16 |  | 2.65 |  | 2.28 |  | 2.4 |  | 7.76 | U | 12.1 |  |
| PCB-126 | 1.1 |  | 0.693 |  | 0.625 |  | 0.968 |  | 2 |  | 1.74 |  |
| PCB-127 | 0.445 |  | 0.246 | U | 0.207 | U | 0.31 | J | 0.882 |  | 1.26 | U |
| PCB-128/166 | 58 |  | 0.759 | UJ | 24.7 | J | 25.7 |  | 75 |  | 133 |  |
| PCB-129/138/160/163 | 346 |  | 162 |  | 142 |  | 139 |  | 465 |  | 877 |  |
| PCB-130 | 22.6 |  | 10.9 |  | 8.79 |  | 8.38 | U | 29.4 |  | 48.5 |  |
| PCB-131 | 5.01 |  | 1.72 | J | 1.51 | J | 1.15 | U | 4.95 |  | 9.06 |  |
| PCB-132 | 116 |  | 45.9 |  | 45.3 |  | 37.9 |  | 120 | J | 244 |  |
| PCB-133 | 3.53 |  | 1.82 | U | 1.84 | J | 1.83 | J | 7.09 |  | 7.77 |  |
| PCB-134/143 | 19.2 |  | 7.71 |  | 6.86 |  | 5.84 |  | 19.5 |  | 40.4 |  |
| PCB-135/151/154 | 83.6 |  | 39.3 |  | 35.6 |  | 31.4 |  | 127 |  | 193 |  |
| PCB-136 | 35.3 |  | 11.5 |  | 11.3 |  | 9.51 |  | 42.8 |  | 78 |  |
| PCB-137 | 17.7 |  | 8.09 |  | 7.69 |  | 5.28 |  | 22.4 |  | 34.3 |  |
| PCB-139/140 | 7.22 |  | 2.52 | J | 2.51 | J | 1.87 | J | 7.84 |  | 14.3 |  |
| PCB-141 | 43.3 |  | 22.1 |  | 19.1 |  | 14.9 |  | 63 |  | 96.8 |  |
| PCB-142 | 0.409 | U | 0.842 | U | 0.508 | U | 0.678 | U | 0.36 | U | 2.04 | U |
| PCB-144 | 11 |  | 4.68 | J | 4.46 | J | 3.87 | J | 16.5 |  | 23.6 |  |
| PCB-145 | 0.163 | J | 0.104 | U | 0.286 | U | 0.153 | U | 0.191 | J | 0.484 | U |
| PCB-146 | 35.2 |  | 21.2 |  | 20 |  | 20.7 |  | 62.9 |  | 87.8 |  |
| PCB-147/149 | 269 |  | 91.9 |  | 81.7 |  | 77.4 |  | 301 |  | 641 |  |
| PCB-148 | 0.546 |  | 0.379 | U | 0.291 | U | 0.349 | U | 1.19 | U | 1.16 | U |
| PCB-150 | 1.27 |  | 0.284 | U | 0.054 | U | 0.266 | U | 1.24 |  | 2.88 |  |
| PCB-152 | 0.346 | J | 0.0484 | U | 0.173 | U | 0.107 | U | 0.422 |  | 0.523 | U |
| PCB-153/168 | 205 |  | 130 |  | 111 |  | 110 |  | 373 |  | 511 |  |
| PCB-155 | 0.222 | U | 0.09 | J | 0.058 | U | 0.113 | U | 0.214 | U | 0.306 | J |
| PCB-156/157 | 34.2 |  | 18.2 |  | 16.2 |  | 15.1 |  | 52.5 |  | 76.8 |  |
| PCB-158 | 33.1 |  | 14.1 |  | 12.9 |  | 11.7 | U | 41.6 |  | 67.1 |  |
| PCB-159 | 2.55 |  | 1.43 | U | 1.35 | J | 1.35 | U | 3.62 |  | 6.92 |  |
| PCB-161 | 0.28 | U | 0.593 | U | 0.379 | U | 0.488 | U | 0.247 | U | 1.47 | U |
| PCB-162 | 1.22 |  | 0.67 | U | 0.658 | U | 0.65 | U | 1.84 |  | 1.95 | U |


| Location ID | BB-19-S | Q | BB-20-S | Q | BB-20-D | Q | BB-21-S | Q | BB-22-S | Q | BB-23-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1326 | 9/10/2014 | 1530 | 9/10/2014 | 1530 | 9/9/2014 | 1453 | 9/9/2014 | 1026 | 9/10/2014 | 942 |
| PCB-164 | 22 |  | 9.93 |  | 8.36 |  | 7.14 |  | 28 |  | 51.2 |  |
| PCB-165 | 0.331 | U | 0.661 | U | 0.432 | U | 0.574 | U | 0.343 | J | 1.73 | U |
| PCB-167 | 11.6 |  | 6.28 |  | 5.54 |  | 5.61 |  | 17.7 |  | 26.9 |  |
| PCB-169 | 0.31 | U | 0.653 | U | 0.443 | U | 0.52 | U | 0.562 | U | 1.43 | U |
| PCB-170 | 46.6 |  | 30.8 |  | 26.4 |  | 29.7 |  | 84.7 |  | 115 |  |
| PCB-171/173 | 22 |  | 9.95 |  | 8.21 |  | 10 |  | 26.5 |  | 46.9 |  |
| PCB-172 | 8.7 |  | 6.51 |  | 5 |  | 5.71 |  | 14.7 |  | 20.5 |  |
| PCB-174 | 57.4 |  | 26.7 |  | 22.6 |  | 30 |  | 82 |  | 142 |  |
| PCB-175 | 2.49 |  | 1.4 |  | 1.16 |  | 1.49 |  | 4.39 |  | 5.82 |  |
| PCB-176 | 8.86 |  | 3.69 |  | 3.24 |  | 4.74 |  | 0.0487 | U | 22.7 |  |
| PCB-177 | 41.5 |  | 18.8 |  | 15.7 |  | 20.9 |  | 51.8 |  | 96.3 |  |
| PCB-178 | 12.2 |  | 7.42 |  | 6.34 |  | 9.81 |  | 22.2 |  | 29.6 |  |
| PCB-179 | 32.6 |  | 11.9 |  | 10.1 |  | 16.1 |  | 39.3 |  | 80.6 |  |
| PCB-180/193 | 92.9 |  | 61.8 |  | 51.7 |  | 63.6 |  | 191 |  | 230 |  |
| PCB-181 | 1.04 | U | 0.355 | J | 0.351 | J | 0.411 |  | 1.06 |  | 2.36 |  |
| PCB-182 | 0.499 |  | 0.326 | U | 0.288 | J | 0.343 | J | 1.13 |  | 1.17 |  |
| PCB-183/185 | 36.7 |  | 22.2 |  | 18.4 |  | 24.7 |  | 70.8 |  | 88.7 |  |
| PCB-184 | 0.198 | J | 0.125 | U | 0.101 | U | 0.309 | J | 0.378 | J | 0.359 | J |
| PCB-186 | 0.101 | U | 0.0751 | U | 0.0625 | U | 0.0686 | U | 0.0487 | U | 0.19 | U |
| PCB-187 | 75.6 |  | 43.8 |  | 35.7 |  | 52.7 |  | 124 |  | 181 |  |
| PCB-188 | 0.498 |  | 0.186 | U | 0.141 | U | 0.222 | J | 0.781 |  | 1.23 |  |
| PCB-189 | 1.97 |  | 1.4 |  | 1.11 |  | 1.25 |  | 3.66 |  | 5.03 |  |
| PCB-190 | 12.6 |  | 5.52 | J | 5.19 | J | 5.74 |  | 17 |  | 30.4 |  |
| PCB-191 | 2.13 |  | 1.36 |  | 1.12 |  | 1.19 |  | 3.45 |  | 5.01 |  |
| PCB-192 | 0.097 | U | 0.0952 | U | 0.0792 | U | 0.0661 | U | 0.0487 | U | 0.18 | U |
| PCB-194 | 29 |  | 22.6 |  | 17.8 |  | 20.1 |  | 66.3 |  | 71.7 |  |
| PCB-195 | 12.4 |  | 7.73 |  | 6.59 |  | 7.77 |  | 21.2 |  | 31 |  |
| PCB-196 | 14.9 |  | 10.4 |  | 8.26 |  | 10.8 | J | 34.1 |  | 35.3 |  |
| PCB-197/199 | 5.74 |  | 3.36 |  | 2.83 |  | 4.32 |  | 10.5 | J | 14.9 |  |
| PCB-198/201 | 41.9 |  | 29.3 |  | 20.7 |  | 30.7 |  | 92.8 |  | 95.3 |  |
| PCB-200 | 4.31 |  | 3.26 |  | 2.57 |  | 3.89 |  | 11.4 |  | 9.88 |  |
| PCB-202 | 10.9 |  | 6.12 |  | 4.85 |  | 7.05 |  | 24.1 |  | 23.1 |  |
| PCB-203 | 23.3 |  | 17.6 |  | 11.8 |  | 16.4 |  | 57.6 |  | 53.4 |  |
| PCB-204 | 0.0863 | U | 0.121 | U | 0.229 | U | 0.131 | U | 0.158 | U | 0.147 | U |
| PCB-205 | 1.25 |  | 0.96 |  | 0.756 |  | 1.03 |  | 2.95 |  | 3.49 |  |
| PCB-206 | 24.1 |  | 19.2 |  | 13.4 |  | 16 |  | 82.7 |  | 101 | J |
| PCB-207 | 3.3 |  | 2.59 | J | 1.63 | J | 2.46 | J | 10.6 |  | 17.3 | J |
| PCB-208 | 13.6 |  | 8.37 |  | 6.16 |  | 7.4 |  | 30.2 |  | 35.4 |  |
| PCB-209 | 45.3 |  | 20.3 |  | 15 |  | 19.7 |  | 63.2 |  | 135 |  |
| Total PCBs* | 6670 |  | 3180 |  | 2720 |  | 2940 |  | 10000 |  | 18700 |  |
| PCB TEQ (0 DL) | 0.124 |  | 0.0769 |  | 0.0690 |  | 0.104 |  | 0.221 |  | 0.208 |  |
| PCB TEQ (1/2 DL) | 0.128 |  | 0.0867 |  | 0.0756 |  | 0.112 |  | 0.230 |  | 0.229 |  |
| PCB TEQ (1 DL) | 0.133 |  | 0.0965 |  | 0.0823 |  | 0.119 |  | 0.239 |  | 0.251 |  |
| PCB TEQ (KM) | 0.125 |  | 0.0737 | Avg | -- |  | 0.105 |  | 0.224 |  | 0.212 |  |

*total PCBs represents the sum of all detected congeners
U-the analytie was analyzed for, but was not detected above the reported sample quantitation limi
$J$-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sampl UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximat $\epsilon$ L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detect KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicate

Table F-3. Summary of Bellingham Bay Polychlorinated Biphenyl (PCB) Congener Data (continued).

| Location ID | BB-24-S | Q | BB-24-D | Q | BB-25-S | Q | BB-26-S | Q | BB-27-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1353 | 9/9/2014 | 1353 | 9/11/2014 | 950 | 9/10/2014 | 1012 | 9/11/2014 | 923 |
| PCB Congeners (ng/kg DW) |  |  |  |  |  |  |  |  |  |  |
| PCB-001 | 83.2 |  | 67.9 |  | 68.6 |  | 38 |  | 9 |  |
| PCB-002 | 35.2 |  | 29.8 |  | 25.3 |  | 14 |  | 6.32 |  |
| PCB-003 | 80.8 |  | 72.7 |  | 73.4 |  | 38 |  | 9.96 |  |
| PCB-004 | 15.7 |  | 13.9 |  | 42.9 |  | 31.6 |  | 4.52 |  |
| PCB-005 | 1.73 |  | 2.16 | U | 3.85 |  | 1.64 |  | 0.352 | J |
| PCB-006 | 15.8 |  | 12.4 |  | 49 |  | 36.7 |  | 5.09 |  |
| PCB-007 | 6.61 |  | 5.14 |  | 9 |  | 5.36 |  | 1.16 | U |
| PCB-008 | 90.9 |  | 81.9 |  | 188 |  | 107 |  | 18.5 |  |
| PCB-009 | 3.51 |  | 2.82 | J | 7.85 |  | 4.44 |  | 0.971 |  |
| PCB-010 | 0.525 |  | 1.89 | U | 1.69 |  | 1.24 |  | 0.17 | U |
| PCB-011 | 43.4 |  | 41.2 |  | 37.6 |  | 27.1 |  | 17.1 |  |
| PCB-012/013 | 17.5 |  | 15.4 |  | 31 |  | 22.9 |  | 4.74 |  |
| PCB-014 | 1.99 | U | 2.03 | U | 1.16 | U | 0.741 | U | 0.253 | U |
| PCB-015 | 104 |  | 107 |  | 139 |  | 79.5 |  | 20.5 |  |
| PCB-016 | 12.6 |  | 12.2 |  | 56.2 |  | 33.7 |  | 5.05 |  |
| PCB-017 | 20.1 |  | 20.2 |  | 96.3 |  | 56.1 |  | 8.85 |  |
| PCB-018/030 | 33.1 |  | 33.2 |  | 149 |  | 112 |  | 14.4 |  |
| PCB-019 | 3.28 |  | 3.09 |  | 14.7 |  | 10.8 |  | 1.64 |  |
| PCB-020/028 | 125 |  | 118 |  | 476 |  | 231 |  | 48.2 |  |
| PCB-021/033 | 45.4 |  | 43.1 |  | 195 |  | 93.9 |  | 16.6 |  |
| PCB-022 | 32.5 |  | 32 |  | 148 |  | 64.6 |  | 14 |  |
| PCB-023 | 0.535 |  | 0.462 |  | 0.497 |  | 0.272 | J | 0.086 | U |
| PCB-024 | 0.566 |  | 0.418 |  | 2.6 |  | 1.43 |  | 0.193 | J |
| PCB-025 | 13.1 |  | 12.6 |  | 45.1 |  | 35.2 |  | 5.19 |  |
| PCB-026/029 | 26.2 |  | 26.1 |  | 99.4 |  | 69.4 |  | 11.2 |  |
| PCB-027 | 4.3 |  | 4.19 |  | 16.6 |  | 13 |  | 1.85 |  |
| PCB-031 | 96.1 |  | 92.6 |  | 392 |  | 200 |  | 34.8 |  |
| PCB-032 | 14.4 |  | 14.3 |  | 75.9 |  | 42.4 |  | 6.29 |  |
| PCB-034 | 0.804 |  | 0.764 |  | 2.97 |  | 1.39 |  | 0.282 | U |
| PCB-035 | 5.83 |  | 5.6 |  | 10.2 |  | 5.08 |  | 1.6 |  |
| PCB-036 | 2.02 |  | 1.86 |  | 1.1 |  | 0.828 |  | 0.477 |  |
| PCB-037 | 30.2 |  | 29 |  | 97 |  | 44.6 |  | 12.6 |  |
| PCB-038 | 0.922 | U | 0.952 |  | 0.922 |  | 0.577 |  | 0.262 | U |
| PCB-039 | 1.1 |  | 1.08 |  | 2.68 |  | 1.38 |  | 0.316 | J |
| PCB-040/041/071 | 36.7 |  | 34.3 |  | 139 |  | 75.7 |  | 17 |  |
| PCB-042 | 17 |  | 16.6 |  | 68.7 |  | 36.2 |  | 7.57 |  |
| PCB-043 | 2.03 |  | 2.18 |  | 11.3 |  | 5.36 |  | 1.3 |  |
| PCB-044/047/065 | 67.2 |  | 64.8 |  | 288 |  | 145 |  | 32.5 |  |
| PCB-045/051 | 8.59 |  | 7.65 |  | 37.8 |  | 24.3 |  | 4.27 |  |
| PCB-046 | 2.65 |  | 2.62 |  | 11.4 |  | 6.92 |  | 1.31 |  |
| PCB-048 | 11.3 |  | 11 |  | 52.3 |  | 24.1 |  | 5.21 |  |
| PCB-049/069 | 52.3 |  | 48.9 |  | 228 |  | 121 |  | 24 |  |
| PCB-050/053 | 8.5 |  | 7.76 |  | 34.4 |  | 24.5 |  | 4.2 |  |
| PCB-052 | 91.6 |  | 82.4 |  | 401 |  | 210 |  | 42.1 |  |
| PCB-054 | 0.181 | J | 0.186 | J | 0.616 | U | 0.457 |  | 0.0732 | U |
| PCB-055 | 2.23 |  | 2.18 |  | 6.82 |  | 3.98 |  | 0.855 |  |
| PCB-056 | 38.7 |  | 36 |  | 136 |  | 67.2 |  | 15.2 |  |
| PCB-057 | 0.687 |  | 0.557 |  | 1.89 |  | 1.07 |  | 0.227 | J |
| PCB-058 | 0.473 |  | 0.356 | U | 0.413 | U | 0.682 |  | 0.177 | U |
| PCB-059/062/075 | 7.06 |  | 6.59 |  | 28.3 |  | 14.6 |  | 3.11 |  |
| PCB-060 | 22.2 |  | 20.4 |  | 70.3 |  | 35.5 |  | 8.89 |  |
| PCB-061/070/074/076 | 163 |  | 152 |  | 598 |  | 291 |  | 66 |  |
| PCB-063 | 3.73 |  | 3.42 |  | 13.3 |  | 6.97 |  | 1.37 |  |
| PCB-064 | 28.5 |  | 27.3 |  | 114 |  | 57.8 |  | 13 |  |
| PCB-066 | 86.5 |  | 81.2 |  | 294 |  | 153 |  | 33.2 |  |
| PCB-067 | 3.21 |  | 3.13 |  | 11.7 |  | 6.18 |  | 1.32 |  |
| PCB-068 | 1.16 |  | 1.12 |  | 3.23 |  | 1.92 |  | 0.386 | J |
| PCB-072 | 1.55 |  | 1.46 |  | 5.6 |  | 3.33 |  | 0.583 |  |
| PCB-073 | 0.439 | U | 0.353 | U | 0.0379 | U | 1.16 |  | 0.073 | U |


| Location ID | BB-24-S | Q | BB-24-D | Q | BB-25-S | Q | BB-26-S | Q | BB-27-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1353 | 9/9/2014 | 1353 | 9/11/2014 | 950 | 9/10/2014 | 1012 | 9/11/2014 | 923 |
| PCB-077 | 11.6 |  | 10.9 |  | 27.4 |  | 13 |  | 4.18 |  |
| PCB-078 | 0.342 | U | 0.35 | U | 0.407 | U | 0.486 | U | 0.175 | U |
| PCB-079 | 2.07 |  | 2.04 |  | 5.75 |  | 2.79 |  | 0.76 |  |
| PCB-080 | 0.155 | U | 0.307 | U | 0.378 | U | 0.464 | U | 0.162 | U |
| PCB-081 | 0.604 |  | 0.516 | U | 1.01 |  | 0.56 | U | 0.186 | J |
| PCB-082 | 15.3 |  | 13.9 |  | 51.6 |  | 25.5 |  | 6.95 |  |
| PCB-083/099 | 89.5 |  | 79.7 |  | 313 |  | 142 |  | 37.5 |  |
| PCB-084 | 26.2 |  | 22.9 |  | 123 |  | 53.4 |  | 15.8 |  |
| PCB-085/116/117 | 25.8 |  | 22.5 |  | 80.4 |  | 40.2 |  | 11 |  |
| PCB-087/097/108/119/125 | 82.7 |  | 72.4 |  | 319 |  | 134 |  | 39.1 |  |
| PCB-088/091 | 16.8 |  | 14.8 |  | 68.5 |  | 33 |  | 8.25 |  |
| PCB-089 | 1.3 |  | 1.05 |  | 4.65 |  | 1.98 |  | 0.708 |  |
| PCB-090/101/113 | 121 |  | 107 |  | 508 |  | 204 |  | 57.2 |  |
| PCB-092 | 22.8 |  | 20.1 |  | 91.4 |  | 38.1 |  | 10.9 |  |
| PCB-093/095/098/100/102 | 79.2 |  | 71.4 |  | 356 |  | 152 |  | 44.7 |  |
| PCB-094 | 0.533 |  | 0.596 |  | 2.14 |  | 0.864 |  | 0.255 | J |
| PCB-096 | 0.669 |  | 0.546 |  | 2.65 |  | 1.21 |  | 0.327 | J |
| PCB-103 | 1.75 |  | 1.53 |  | 7.05 |  | 3.1 |  | 0.747 |  |
| PCB-104 | 0.068 | U | 0.0486 | U | 0.085 | J | 0.12 | J | 0.0586 | U |
| PCB-105 | 54 |  | 47.8 |  | 176 |  | 79.9 |  | 22.1 |  |
| PCB-106 | 0.22 | J | 0.272 | U | 0.624 | U | 0.43 | U | 0.237 | U |
| PCB-107/124 | 5.51 |  | 5.07 |  | 19.1 |  | 8.33 |  | 2.5 |  |
| PCB-109 | 10.9 |  | 10.3 |  | 38.8 |  | 16 |  | 4.81 |  |
| PCB-110/115 | 145 |  | 124 |  | 514 |  | 259 |  | 61.2 |  |
| PCB-111 | 0.19 | J | 0.247 | J | 0.461 |  | 0.126 | J | 0.103 | U |
| PCB-112 | 0.318 | U | 0.188 | U | 0.12 | U | 0.742 |  | 0.0984 | U |
| PCB-114 | 3.09 |  | 2.56 |  | 9.56 |  | 4.46 |  | 1.22 |  |
| PCB-118 | 131 |  | 116 |  | 446 |  | 199 |  | 52.4 |  |
| PCB-120 | 0.868 |  | 0.833 |  | 2.28 |  | 0.869 |  | 0.323 | J |
| PCB-121 | 0.126 | U | 0.107 | U | 0.131 | U | 0.095 | U | 0.108 | U |
| PCB-122 | 1.97 |  | 1.62 |  | 5.51 |  | 2.81 |  | 0.685 |  |
| PCB-123 | 2.96 |  | 2.56 |  | 7.73 |  | 4.44 |  | 0.998 |  |
| PCB-126 | 1.06 |  | 0.782 |  | 1.42 |  | 0.655 |  | 0.252 | U |
| PCB-127 | 0.321 | J | 0.291 | U | 0.643 | U | 0.426 | U | 0.245 | U |
| PCB-128/166 | 35.8 |  | 25.6 |  | 95.2 |  | 44.3 |  | 0.789 | U |
| PCB-129/138/160/163 | 176 |  | 142 |  | 581 |  | 261 |  | 68.6 |  |
| PCB-130 | 9.99 |  | 8.87 |  | 38.1 |  | 14.8 |  | 4.7 | J |
| PCB-131 | 1.43 | J | 0.944 | J | 1.4 | U | 3.22 | U | 0.824 | U |
| PCB-132 | 46.7 |  | 37.9 |  | 178 |  | 72 |  | 20.8 |  |
| PCB-133 | 2.51 | U | 1.85 | U | 7.7 |  | 2.96 | U | 0.976 | U |
| PCB-134/143 | 6.32 |  | 5.04 | U | 27.5 |  | 11.2 |  | 3.22 | J |
| PCB-135/151/154 | 43 |  | 30.7 |  | 136 |  | 58 |  | 16.2 |  |
| PCB-136 | 12.2 |  | 10.3 |  | 46.8 |  | 22.4 |  | 6.19 |  |
| PCB-137 | 7.24 |  | 5.21 |  | 24.9 |  | 12 |  | 3.36 | J |
| PCB-139/140 | 2.78 |  | 1.92 | J | 8.77 |  | 4.21 |  | 0.917 | U |
| PCB-141 | 21.5 |  | 16.1 |  | 79.7 |  | 28.6 |  | 9.17 |  |
| PCB-142 | 0.406 | U | 0.802 | U | 1.33 | U | 0.592 | U | 0.875 | U |
| PCB-144 | 5.48 |  | 3.61 | J | 20 |  | 7.12 |  | 2.21 | J |
| PCB-145 | 0.101 | U | 0.083 | J | 0.0379 | U | 0.169 | J | 0.051 | U |
| PCB-146 | 26.4 |  | 20.3 |  | 77.7 |  | 30.3 |  | 9.44 |  |
| PCB-147/149 | 112 |  | 84.1 |  | 360 |  | 165 |  | 40 |  |
| PCB-148 | 0.586 | U | 0.269 | U | 0.926 | J | 0.499 | U | 0.261 | U |
| PCB-150 | 0.356 | U | 0.346 | U | 0.0379 | U | 0.609 | U | 0.051 | U |
| PCB-152 | 0.053 | J | 0.133 | U | 0.332 | U | 0.124 | U | 0.139 | U |
| PCB-153/168 | 149 |  | 111 |  | 449 |  | 176 |  | 55.2 |  |
| PCB-155 | 0.124 | U | 0.128 | J | 0.122 | J | 0.085 | U | 0.0669 | U |
| PCB-156/157 | 17.5 |  | 15.5 |  | 60.4 |  | 26.7 |  | 7.57 |  |
| PCB-158 | 15.3 |  | 12.1 |  | 55.1 |  | 21.1 |  | 6.04 | J |
| PCB-159 | 3.45 | UJ | 1.36 | J | 5.01 |  | 2.26 |  | 0.685 | U |
| PCB-161 | 0.292 | U | 0.577 | U | 0.997 | U | 0.426 | U | 0.617 | U |
| PCB-162 | 0.588 | J | 0.768 | U | 1.87 | J | 0.812 | J | 0.696 | U |


| Location ID | BB-24-S | Q | BB-24-D | Q | BB-25-S | Q | BB-26-S | Q | BB-27-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1353 | 9/9/2014 | 1353 | 9/11/2014 | 950 | 9/10/2014 | 1012 | 9/11/2014 | 923 |
| PCB-164 | 9.84 |  | 7.86 |  | 39.2 |  | 14.4 |  | 3.6 | J |
| PCB-165 | 0.343 | U | 0.679 | U | 1.14 | U | 0.501 | U | 0.687 | U |
| PCB-167 | 6.2 |  | 5.58 |  | 21 |  | 8.66 |  | 2.54 |  |
| PCB-169 | 1.59 | U | 0.666 | U | 1.1 | U | 0.469 | U | 0.7 | U |
| PCB-170 | 40.3 |  | 29.1 |  | 98.4 |  | 41.1 |  | 13.4 |  |
| PCB-171/173 | 14.3 |  | 9.88 |  | 37 |  | 13.2 |  | 4.65 |  |
| PCB-172 | 8.71 |  | 5.78 |  | 19.2 |  | 7.42 |  | 2.66 |  |
| PCB-174 | 56.4 | J | 29.5 | J | 106 |  | 40.1 |  | 13.2 |  |
| PCB-175 | 2.8 | J | 1.62 | J | 5.72 |  | 2.28 |  | 0.809 |  |
| PCB-176 | 8.73 | J | 4.52 | J | 15.6 |  | 6.16 |  | 2.06 |  |
| PCB-177 | 31 |  | 19.6 |  | 66.4 |  | 25.4 |  | 9.4 |  |
| PCB-178 | 16.5 | J | 9.44 | J | 26.2 |  | 10.4 |  | 3.94 |  |
| PCB-179 | 33.5 | J | 15.4 | J | 44.6 |  | 22.2 |  | 6.26 |  |
| PCB-180/193 | 141 | J | 63.1 | J | 208 |  | 77.9 |  | 28.2 |  |
| PCB-181 | 0.43 |  | 0.373 | J | 1.56 |  | 0.659 |  | 0.182 | J |
| PCB-182 | 0.533 |  | 0.452 |  | 1.08 |  | 0.471 |  | 0.209 | U |
| PCB-183/185 | 55.6 | J | 24.2 | J | 86.1 |  | 30.4 |  | 11.4 |  |
| PCB-184 | 0.331 | J | 0.272 | J | 0.252 | J | 0.197 | J | 0.113 | U |
| PCB-186 | 0.0898 | U | 0.0756 | U | 0.078 | U | 0.124 | U | 0.104 | U |
| PCB-187 | 122 | J | 53.2 | J | 155 |  | 61.6 |  | 22.9 |  |
| PCB-188 | 0.347 | J | 0.319 | J | 0.547 |  | 0.338 | U | 0.153 | U |
| PCB-189 | 1.56 |  | 1.39 | U | 4.5 |  | 1.77 |  | 0.601 |  |
| PCB-190 | 8.93 |  | 6.53 |  | 19.7 |  | 8.07 |  | 2.59 | U |
| PCB-191 | 1.91 |  | 1.22 |  | 4.74 |  | 1.73 |  | 0.544 |  |
| PCB-192 | 0.0865 | U | 0.0728 | U | 0.0639 | U | 0.118 | U | 0.101 | U |
| PCB-194 | 81.9 | J | 22.4 | J | 58.5 |  | 22.6 |  | 10.8 |  |
| PCB-195 | 22.2 | J | 8.11 | J | 20.5 |  | 8.52 |  | 3.28 |  |
| PCB-196 | 44.3 | J | 11.7 | J | 26.8 |  | 12.3 |  | 5.29 |  |
| PCB-197/199 | 17.5 | J | 4.66 | J | 11.3 |  | 4.85 |  | 1.75 |  |
| PCB-198/201 | 130 | J | 32.4 | J | 79.8 |  | 34.4 |  | 14.6 |  |
| PCB-200 | 16.9 | J | 4.29 | J | 9.74 |  | 4.11 | U | 1.69 |  |
| PCB-202 | 29.7 | J | 7.81 | J | 18.7 |  | 8.33 |  | 3.33 |  |
| PCB-203 | 95.9 | J | 19.8 | J | 49.1 |  | 20 |  | 9.05 |  |
| PCB-204 | 0.132 | U | 0.264 | U | 0.071 | U | 0.0517 | U | 0.056 | U |
| PCB-205 | 2.82 | J | 1.08 | J | 2.66 |  | 1.05 |  | 0.408 |  |
| PCB-206 | 92.3 | J | 21.5 | J | 36.3 |  | 21.9 |  | 12.3 |  |
| PCB-207 | 15 | J | 3.67 | J | 5.9 |  | 2.83 |  | 1.71 | J |
| PCB-208 | 28.8 | J | 8.54 | J | 22.9 |  | 11.5 |  | 4.9 |  |
| PCB-209 | 34.9 | J | 19.8 | J | 51 |  | 34.4 |  | 13.3 |  |
| Total PCBs* | 4340 |  | 3250 |  | 11900 |  | 5680 |  | 1410 |  |
| PCB TEQ (0 DL) | 0.114 |  | 0.0850 |  | 0.167 |  | 0.0765 |  | 0.0031 |  |
| PCB TEQ (1/2 DL) | 0.138 |  | 0.0951 |  | 0.183 |  | 0.0837 |  | 0.0262 |  |
| PCB TEQ (1 DL) | 0.162 |  | 0.105 |  | 0.200 |  | 0.0908 |  | 0.0493 |  |
| PCB TEQ (KM) | 0.100 | Avg | -- |  | 0.170 |  | 0.0779 |  | 0.0286 | L |

*total PCBs represents the sum of all detected congeners
U-the analytie was analyzed for, but was not detected above the reported sample quantitation limi
J-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sampl UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximat $\epsilon$ L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detect KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicate

Table F-3. Summary of Bellingham Bay Polychlorinated Biphenyl (PCB) Congener Data (continued).

| Location ID | BB-28-S | Q | BB-29-S | Q | BB-30-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1129 | 9/11/2014 | 1103 | 9/9/2014 | 1507 |
| PCB Congeners (ng/kg DW) |  |  |  |  |  |  |
| PCB-001 | 63.5 |  | 26.9 |  | 36.3 | J |
| PCB-002 | 37 |  | 16.6 |  | 18.5 | J |
| PCB-003 | 58.2 |  | 30.4 |  | 41.1 | J |
| PCB-004 | 37 |  | 15.9 |  | 10.1 |  |
| PCB-005 | 2.15 | J | 0.842 | U | 1.06 |  |
| PCB-006 | 42.8 |  | 22.8 |  | 12 |  |
| PCB-007 | 7.96 |  | 3.14 |  | 3.46 |  |
| PCB-008 | 174 |  | 68 |  | 53.7 |  |
| PCB-009 | 7.59 |  | 2.67 |  | 2.1 |  |
| PCB-010 | 2.07 | U | 0.541 | J | 0.359 | J |
| PCB-011 | 41.6 |  | 27 |  | 32.2 |  |
| PCB-012/013 | 28.2 |  | 16.9 |  | 11.3 |  |
| PCB-014 | 2.64 | J | 0.587 | U | 0.922 |  |
| PCB-015 | 141 |  | 65.1 |  | 66.5 |  |
| PCB-016 | 63.4 |  | 14.1 |  | 10.4 |  |
| PCB-017 | 87.3 |  | 29 |  | 17.8 |  |
| PCB-018/030 | 190 |  | 53.4 |  | 29 |  |
| PCB-019 | 20.2 |  | 5.45 |  | 2.68 |  |
| PCB-020/028 | 490 |  | 141 |  | 91.6 |  |
| PCB-021/033 | 210 |  | 48.2 |  | 33.1 |  |
| PCB-022 | 133 |  | 34.8 |  | 24.4 |  |
| PCB-023 | 0.75 | U | 0.251 | U | 0.31 | J |
| PCB-024 | 2.19 |  | 0.589 | U | 0.408 |  |
| PCB-025 | 46.2 |  | 19 |  | 10.5 |  |
| PCB-026/029 | 97.5 |  | 40.5 |  | 21 |  |
| PCB-027 | 21.2 |  | 6.49 |  | 3.54 |  |
| PCB-031 | 402 |  | 113 |  | 65.1 |  |
| PCB-032 | 84.9 |  | 21.3 |  | 10.2 |  |
| PCB-034 | 2.59 | J | 0.685 | J | 0.555 |  |
| PCB-035 | 15.5 |  | 3.34 |  | 3.99 |  |
| PCB-036 | 2.37 |  | 0.232 | U | 1.38 |  |
| PCB-037 | 110 |  | 33.3 |  | 25 |  |
| PCB-038 | 1.77 | U | 0.416 | U | 0.574 | U |
| PCB-039 | 4.03 |  | 0.786 | U | 0.765 |  |
| PCB-040/041/071 | 198 |  | 42.7 |  | 28.6 |  |
| PCB-042 | 92.5 |  | 20.3 |  | 14.3 |  |
| PCB-043 | 13 |  | 2.58 |  | 1.67 |  |
| PCB-044/047/065 | 365 |  | 88.5 |  | 56.6 |  |
| PCB-045/051 | 53.4 |  | 11.5 |  | 7.23 |  |
| PCB-046 | 16.9 |  | 3.18 |  | 2.17 |  |
| PCB-048 | 59.3 |  | 12.7 |  | 9.09 |  |
| PCB-049/069 | 292 |  | 72.1 |  | 44.7 |  |
| PCB-050/053 | 49.6 |  | 12.6 |  | 7.1 |  |
| PCB-052 | 527 |  | 123 |  | 72.1 |  |
| PCB-054 | 0.82 | U | 0.27 | U | 0.201 | J |
| PCB-055 | 9 |  | 0.806 | U | 1.41 |  |
| PCB-056 | 180 |  | 48.5 |  | 29.2 |  |
| PCB-057 | 2.23 | U | 0.719 | U | 0.421 |  |
| PCB-058 | 1.89 | J | 0.749 | U | 0.33 | J |
| PCB-059/062/075 | 41.3 |  | 8.35 |  | 5.33 |  |
| PCB-060 | 104 |  | 25.8 |  | 15.8 |  |
| PCB-061/070/074/076 | 734 |  | 211 |  | 129 |  |
| PCB-063 | 15.2 |  | 4.18 |  | 2.63 |  |
| PCB-064 | 151 |  | 34.2 |  | 23 |  |
| PCB-066 | 386 |  | 110 |  | 68 |  |
| PCB-067 | 12.6 |  | 3.39 |  | 2.32 |  |
| PCB-068 | 2.99 | U | 0.998 | J | 0.784 |  |
| PCB-072 | 6.03 |  | 1.6 | J | 1.06 |  |
| PCB-073 | 3.84 | U | 0.0486 | U | 0.189 | J |


| Location ID | BB-28-S | Q | BB-29-S | Q | BB-30-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1129 | 9/11/2014 | 1103 | 9/9/2014 | 1507 |
| PCB-077 | 37.5 |  | 9.65 |  | 10.1 |  |
| PCB-078 | 1.41 | U | 0.809 | U | 0.206 | U |
| PCB-079 | 10.1 |  | 2.07 |  | 1.54 |  |
| PCB-080 | 1.33 | U | 0.737 | U | 0.183 | U |
| PCB-081 | 1.63 | U | 0.744 | U | 0.43 | U |
| PCB-082 | 69.4 |  | 16.7 |  | 9.41 |  |
| PCB-083/099 | 387 |  | 110 |  | 68.3 |  |
| PCB-084 | 169 |  | 37 |  | 22.9 |  |
| PCB-085/116/117 | 123 |  | 28.7 |  | 20.3 |  |
| PCB-087/097/108/119/125 | 423 |  | 102 |  | 69.5 |  |
| PCB-088/091 | 92.9 |  | 23 |  | 14.2 |  |
| PCB-089 | 8.14 |  | 1.6 | J | 0.965 |  |
| PCB-090/101/113 | 670 |  | 165 |  | 104 |  |
| PCB-092 | 118 |  | 28 |  | 18.4 |  |
| PCB-093/095/098/100/102 | 519 |  | 108 |  | 76.3 |  |
| PCB-094 | 3.08 |  | 0.477 | U | 0.329 | J |
| PCB-096 | 3.86 |  | 0.727 | U | 0.52 |  |
| PCB-103 | 7.25 |  | 1.86 | U | 1.27 |  |
| PCB-104 | 0.112 | J | 0.093 | U | 0.123 | U |
| PCB-105 | 245 |  | 60.9 |  | 48.5 |  |
| PCB-106 | 0.714 | U | 0.462 | U | 0.241 | U |
| PCB-107/124 | 26 |  | 5.62 |  | 4.73 |  |
| PCB-109 | 54.9 |  | 12.2 |  | 8.92 |  |
| PCB-110/115 | 739 |  | 186 |  | 125 |  |
| PCB-111 | 0.541 | U | 0.141 | U | 0.167 | U |
| PCB-112 | 0.963 |  | 0.135 | U | 0.0851 | U |
| PCB-114 | 13.5 |  | 3.14 |  | 2.17 |  |
| PCB-118 | 601 |  | 164 |  | 114 |  |
| PCB-120 | 2.75 |  | 0.691 | J | 0.667 | U |
| PCB-121 | 0.198 | J | 0.135 | U | 0.0844 | U |
| PCB-122 | 8.37 |  | 1.85 | J | 1.41 |  |
| PCB-123 | 12.7 |  | 3.25 |  | 2.54 | U |
| PCB-126 | 2.44 |  | 0.598 | J | 2.43 |  |
| PCB-127 | 1.41 |  | 0.564 | U | 0.284 | J |
| PCB-128/166 | 130 |  | 30.2 |  | 30.6 |  |
| PCB-129/138/160/163 | 786 |  | 176 |  | 176 |  |
| PCB-130 | 39.9 |  | 11.2 |  | 11.2 |  |
| PCB-131 | 8.72 | U | 2.19 |  | 1.87 |  |
| PCB-132 | 212 |  | 54.5 |  | 49.4 |  |
| PCB-133 | 9.61 |  | 2.55 |  | 2.53 |  |
| PCB-134/143 | 36.2 |  | 7.98 |  | 7.45 |  |
| PCB-135/151/154 | 186 |  | 41.7 |  | 42 |  |
| PCB-136 | 66.5 |  | 14.9 |  | 15 |  |
| PCB-137 | 30.5 |  | 8.91 |  | 7.96 |  |
| PCB-139/140 | 12.2 |  | 3.54 |  | 2.7 |  |
| PCB-141 | 94.6 |  | 22.7 |  | 23.8 |  |
| PCB-142 | 1.22 | U | 0.463 | U | 0.178 | U |
| PCB-144 | 27.3 |  | 5.7 |  | 5.43 |  |
| PCB-145 | 0.488 | U | 0.148 | U | 0.071 | U |
| PCB-146 | 88.4 |  | 24.2 |  | 22.7 |  |
| PCB-147/149 | 496 |  | 108 |  | 107 |  |
| PCB-148 | 1.41 | J | 0.296 | U | 0.312 | U |
| PCB-150 | 1.83 | U | 0.603 | U | 0.394 |  |
| PCB-152 | 0.871 | J | 0.131 | U | 0.123 | J |
| PCB-153/168 | 532 |  | 134 |  | 130 |  |
| PCB-155 | 0.188 | U | 0.13 | U | 0.228 | U |
| PCB-156/157 | 70.1 |  | 19.2 |  | 19.7 |  |
| PCB-158 | 68.7 |  | 15.2 |  | 15.3 |  |
| PCB-159 | 6.28 | U | 1.14 | J | 1.22 |  |
| PCB-161 | 0.811 | U | 0.346 | U | 0.122 | U |
| PCB-162 | 2.44 | J | 0.491 | U | 0.742 |  |


| Location ID | BB-28-S | Q | BB-29-S | Q | BB-30-S | Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date/Time | 9/9/2014 | 1129 | 9/11/2014 | 1103 | 9/9/2014 | 1507 |
| PCB-164 | 44.1 |  | 10.6 |  | 11.7 |  |
| PCB-165 | 0.94 | U | 0.394 | U | 0.141 | U |
| PCB-167 | 25.4 |  | 6.66 |  | 8.1 |  |
| PCB-169 | 1.16 | U | 0.412 | U | 0.235 | U |
| PCB-170 | 105 |  | 28.3 |  | 35 |  |
| PCB-171/173 | 38.1 |  | 10.7 |  | 10.6 |  |
| PCB-172 | 19.6 |  | 5.99 |  | 5.52 |  |
| PCB-174 | 117 |  | 30.3 |  | 28.7 |  |
| PCB-175 | 6.81 |  | 1.51 | J | 1.62 | U |
| PCB-176 | 19.8 |  | 4.43 |  | 3.83 |  |
| PCB-177 | 67 |  | 20.3 |  | 19.4 |  |
| PCB-178 | 33.6 |  | 8.06 |  | 7.36 |  |
| PCB-179 | 68.3 |  | 13.8 |  | 12.3 |  |
| PCB-180/193 | 241 |  | 57.9 |  | 64.9 |  |
| PCB-181 | 1.37 |  | 0.0486 | U | 0.43 | U |
| PCB-182 | 1.44 |  | 0.0486 | U | 0.41 |  |
| PCB-183/185 | 103 |  | 24.1 |  | 21.5 |  |
| PCB-184 | 0.502 |  | 0.117 | U | 0.267 | J |
| PCB-186 | 0.135 | U | 0.0486 | U | 0.082 | U |
| PCB-187 | 204 |  | 49.1 |  | 38.2 |  |
| PCB-188 | 2.5 |  | 0.243 | U | 0.255 | U |
| PCB-189 | 4.77 |  | 1.52 | J | 1.66 |  |
| PCB-190 | 23.4 |  | 5.14 |  | 6.26 |  |
| PCB-191 | 4.72 |  | 1.38 | U | 1.33 | U |
| PCB-192 | 0.127 | U | 0.0486 | U | 0.0476 | U |
| PCB-194 | 66.9 |  | 16.5 |  | 16.3 |  |
| PCB-195 | 24.3 |  | 4.78 |  | 5.95 |  |
| PCB-196 | 42.5 |  | 7.94 |  | 8.2 |  |
| PCB-197/199 | 15.9 |  | 3.71 |  | 2.68 | J |
| PCB-198/201 | 126 |  | 24.2 |  | 23 |  |
| PCB-200 | 15.7 |  | 3.28 |  | 2.91 |  |
| PCB-202 | 30.6 |  | 6.36 |  | 5.86 |  |
| PCB-203 | 74.8 |  | 13.8 |  | 13 |  |
| PCB-204 | 0.404 | U | 0.051 | U | 0.0476 | U |
| PCB-205 | 3.05 |  | 0.917 | J | 0.8 | U |
| PCB-206 | 167 | J | 18.4 |  | 15.1 |  |
| PCB-207 | 32.7 | J | 2.53 |  | 2.32 |  |
| PCB-208 | 83.2 |  | 8.09 |  | 6.12 |  |
| PCB-209 | 207 |  | 21.8 |  | 18 |  |
| Total PCBs* | 15200 |  | 3860 |  | 2960 |  |
| PCB TEQ (0 DL) | 0.277 |  | 0.0685 |  | 0.250 |  |
| PCB TEQ (1/2 DL) | 0.295 |  | 0.0748 |  | 0.253 |  |
| PCB TEQ (1 DL) | 0.312 |  | 0.0811 |  | 0.257 |  |
| PCB TEQ (KM) | 0.281 |  | 0.0696 |  | 0.251 |  |

*total PCBs represents the sum of all detected congeners
U-the analytie was analyzed for, but was not detected above the reported sample quantitation limi
J-the analyte was positively identified; the associated numerical value is the approximate concentration of the analyte in the sampl UJ-the analyte was not detected above the reported sample quantitation limit. However, the reported quantitation limit is approximat $\epsilon$ L-the detection frequency of compounds within a sample was $<50 \%$, the numeric value indicates the number of non-detect KM-Kaplan-Meier DW-dry weight Q-qualifier TEQ-toxicity equivalent DL-detection limit Avg-average of field duplicate

## Appendix G. Statistical Evaluation of the Bellingham Bay Data Set

## Appendix G: Statistical Evaluation of the Bellingham Bay Data Set

Various statistical methods were used to evaluate the Bellingham Bay data set. This appendix represents a record of the analyses performed. The analyses in Section G. 1 were conducted to determine the most appropriate distribution for each analyte. All data were evaluated in Section G. 2 to identify if any statistical outliers existed for any of the individual analytes. The correlation analyses of Section G. 3 were conducted to further explore patterns in the data and identify any unusual results not obvious from the univariate analysis. Multivariate exploratory analyses can be used to differentiate natural and regional background distributions, but was not necessary with this study because Bellingham Bay sediments clearly differed from natural background for dioxin/furan congeners and cPAHs.

Any samples flagged by the exploratory analyses outside of Sections G. 1 and G. 2 are not considered outliers for the purpose of calculating the 90/90 upper tolerance limits (90/90 UTL).

## G. 1 Distributional Analysis

The distributions for each chemical analyte plus total organic carbon (TOC) were evaluated individually in R ( R Core Team 2014) to generate normal-probability (also referred to as quantile-quantile [QQ]) plots and boxplots, and in ProUCL 5.0 (USEPA 2013) for formal goodness of fit (GOF) tests against the standard distributions (i.e., normal and gamma). The normal probability plots for each of the analytes are shown in Figure G-1, and for TOC in Figure G-2.

These plots were used in part to evaluate the shape of the data distributions and help identify unusual values. Several stations were identified: two with elevated concentrations, and two with concentrations below natural background for dioxin/furan TEQ. The physical location of these samples, either near areas presumed to represent natural background or historical sources, led to the decision to exclude these samples from the calculation of regional background values for one or more analytes. Details for these samples are provided below.

- Station BB-28 was identified as an outlier for lead and cPAHs, and also had a somewhat elevated level of the PCB TEQ (Figure G-1). Review of the potential sites and sources near this sampling location identified a former refueling station that was a likely source for all three of these analytes. Therefore, data for these analytes from Station BB-28 were excluded from the calculation of regional background.
- Station BB-12 was identified as an outlier for TOC (Figure G-2) and a large amount of wood waste was observed at this station during the field investigation. These
observations are indicative of a direct anthropogenic influence. The wood waste in the sample caused elevated detection limits for some analytes, particularly the PCB congeners. One of the elevated detection limits resulted in the PCB TEQ being identified as a statistical outlier at this station. Other analytes were also affected, although less severely. Due to the interference noted in the chemical analyses of PCBs and the wood waste as evidence of anthropogenic influence, all results for this station were excluded from the calculation of regional background.
- Stations BB-14 and BB-27 had concentrations of dioxin/furan congeners that had recovered to levels below natural background (less than 4 ng TEQ $/ \mathrm{kg}$ ). These stations were located adjacent to the boundary of the area that represented dioxin/furan natural background. Consequently, the area occupied by these stations was considered an extension of the dioxin/furan natural background exclusion of the AOI. Dioxin/furan congeners from these two samples were excluded from the calculation of regional background.

After excluding the sample results specified above, probability plots were generated for the remaining data (Figure G-3). Results are shown in Table G-1.

There were no additional statistical outliers found after excluding the samples specified in Table $\mathrm{G}-1$. The data distributions are described below.

- Lead - the data were right skewed. The log-normal distribution had the highest QQ plot correlation coefficient ( $\mathrm{r}=0.967$ ), however, the right tail had a small secondary mode for concentrations greater than $14 \mathrm{mg} / \mathrm{kg}$. A parametric estimate of the upper tolerance limit (UTL) based on the log-normal distribution would underestimate this upper tail. Consequently, non-parametric bootstrap estimates were preferred.
- cPAH TEQ - these data were also right skewed, with a pattern similar to that of lead. Non-parametric bootstrap estimates were preferred.
- Dioxin/Furan TEQ - these data were best described by a normal distribution (QQ plot correlation $\mathrm{r}=0.976$; Shapiro-Wilk's test $p=0.19$ ).
- PCB TEQ - The gamma distribution was a good fit to these data ( QQ plot correlation $\mathrm{r}=$ 0.992 ; goodness of fit tests $p>0.05$ ).


## G. 2 Univariate Investigation of Extreme Values for each Analyte

The term "extreme value" is a statistical term used to define a result that deviates from the population mean but may still be representative of background (SCUM II, Ecology 2013a \& 2015). Extreme values are common in data sets that lack sufficient sample sizes to capture the full range of values. Essentially, it is unknown whether an extreme value is an outlier unless
more samples are available. The individual analytes were screened for extreme values using a variety of approaches:

- Single extreme values were identified using Rosner's formal outlier test (for $\mathrm{n} \geq 25$ ) or Dixon's outlier test (for $\mathrm{n}<25$ ) for data that fit a normal distribution (see Table G-1). Both tests were conducted in ProUCL with $\alpha=0.05$.
- Extreme values were also identified when the single highest concentration sample was the difference between a normal and a non-normal distribution (i.e. if the high concentration sample was excluded, the distribution would be normal).
- High values were flagged as extreme based on the boxplots in Figure G-4, using Tukey's rule of thumb ${ }^{1}$.

Screening for extreme values was done to determine whether specific samples may be skewing the distribution for an individual analyte. A summary of the presence of extreme values for each analyte in the Bellingham Bay data set follows:

- Lead presented three extreme values based on Tukey's rule of thumb (Figure G-4). However, these three samples were part of the small secondary mode of five samples with lead concentrations greater than $14 \mathrm{mg} / \mathrm{kg}$, so these are not considered outliers. These stations are adjacent to the shoreline, and are presumed to represent the upper tail of a skewed Bellingham Bay data set distribution. Note that the entire Bellingham Bay data set is contained within the range of natural background.
- cPAH TEQ presented three extreme values based on Tukey's rule of thumb (Figure G-4). These samples represent the small secondary mode of three samples with cPAH values greater than $60 \mu \mathrm{TEQ} / \mathrm{kg}$. These stations are adjacent to the shoreline, and are presumed to represent the upper tail of a skewed Bellingham Bay data set distribution.
- Dioxin/furan TEQ presented no extreme values (Figures G-3 and G-4). This was the only analyte for which Dixon's test could be used, and no formal outliers were identified.
- PCB TEQ presented one extreme value based on Tukey's rule of thumb (Figure G-4). However, this rule of thumb is most appropriate for symmetric data, and these data are best described by the gamma distribution. This value appears to be part of the upper tail of a skewed distribution. Note that the entire Bellingham Bay data set is contained within the range of natural background.

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## G. 3 Correlation between Sediment Chemistry Concentrations and TOC

The following bivariate investigation evaluated the distribution of each analyte in relation to percent TOC. Anthropogenic contamination tends to be associated with TOC. As a result there is typically a positive correlation between TOC and sediment chemistry. Deviations from the general trend can be indicative of a different source, or a different geochemical/depositional environment. This exploratory investigation provided information for understanding some of the possible sediment dynamics in Bellingham Bay and may be useful for future investigations. However, these results did not contribute to any decision towards excluding samples as outliers.

Correlations between percent fines and each analyte were also considered. However these correlations were not conducted because there was so little variability in the level of percent fines (i.e., values ranged from 90 to 100 percent, after excluding the samples specified in Section G.1).

The relationships between total organic carbon (as percent TOC) and sediment chemical concentrations are presented graphically as scatterplots in Figure G-4. The TOC for the samples in this data set were fairly evenly distributed throughout the observed range of 1 to 3.7 percent. The scatterplots show that except for dioxin/furan congeners ( $p=0.218$ ), all correlations are fairly good between each analyte and percent TOC (all statistically significant with $p<0.01$, Table 4, Section 4.3.1). There were several samples with concentrations of one or more analytes that were higher than expected given the amount of TOC in the sediments. These samples are circled in Figure G-4, and include:

- Sample BB-22 for lead, cPAH and PCB TEQs.
- Sample BB-23 for lead, dioxin/furan and PCB TEQs.
- Sample BB-25 for cPAH and dioxin/furan TEQ.
- Samples BB-08 and BB-24 for dioxin/furan TEQ.
- Sample BB-30 for PCB TEQ.

Table G-1. Best Fit distribution for each analyte in the Bellingham Bay data set.

|  | Correlation <br> Best Fit <br> Analyte | QQ-Plot for the <br> best fit <br> distribution(s) | Detection <br> Frequency | Samples <br> Excluded |
| :--- | :---: | :---: | :---: | :---: |
| Lead | None | 0.967 (lognormal) | $\mathbf{2 8 / 2 8}$ | BB-12, BB-28 |
| cPAH TEQ | None | 0.969 (lognormal) | $\mathbf{2 8 / 2 8}$ | BB-12, BB-28 |
| Dioxin/Furan <br> TEQ | Normal | 0.976 | $\mathbf{2 0 / 2 0}$ | BB-12, BB-14, BB-27 |
| PCB TEQ | Gamma | 0.992 | $\mathbf{2 7 / 2 8}$ | BB-12, BB-28 |



Figure H-1. Normal probability plots for analytes in Bellingham Bay sediments. Note: Censored (non-detect) data points are shown on the graphs as open circles, their quantiles are estimated from the distribution of the detected data using regression on order statistics. The quantiles observed from the sample data ( $y$-axis) are plotted against the corresponding quantiles from the standard normal distribution (bottom $x$-axis). The top axis and vertical grid lines show the percent chance of exceedance for the normal distribution (e.g., a 5 percent chance of exceedance is the $95^{\text {th }}$ percentile, which intersects the estimated distribution for dioxin/furan TEQ, represented by the line through the data, at a concentration of approximately $14 \mathrm{ng} \mathrm{TEO} / \mathrm{kg}$ ). When the data points fall along a straight line, they are approximately normally distributed. Extreme values and systematic deviations from normality can be observed on these plots based on where the data points fall relative to the line.

## TOC (percent)



Figure H-2. Normal probability plot for total organic carbon in Bellingham Bay sediments.


Figure H-3. Probability plots for analytes in Bellingham Bay sediments, excluding samples flagged in Figure H-1. Note: the quantiles for lead and cPAHs are log-scaled, indicating that the log-normal distribution was the best-fit distribution for these data; quantiles for the PCB TEQ are plotted relative to the theoretical gamma quantiles, the best-fit distribution for these data.


Figure H-4. Boxplots showing the distribution of sediment concentrations in Bellingham Bay sediments.
Note: The 1st, 2nd, and 3rd quartiles are estimated using Kaplan-Meier for censored data; horizontal lines indicate the level of the highest detection limit among the two data sets. The sample sizes for the Bold plus natural background analytes were $n=96$ (lead), $n=76$ (cPAH TEQ), $n=91$ (dioxin/furan TEQ), and $n=70$ (PCBs TEQ).


- Non-Detected • Detected

Figure H-5. Relationship between percent total organic carbon and each analyte in Bellingham Bay sediments. Note: Non-detected values (turquoise data points) are shown at the reported detection limit. The ordinary least squares regression line is shown for the data in each panel. Extreme samples identified in Section 1.0 have been omitted from these plots. Circled samples have higher concentrations than expected for the amount of TOC in the sample (see text for more details).


[^0]:    ${ }^{1}$ Tukey's rule of thumb flags values as extreme when they exceed the median $+/-1.5^{*} \mathrm{IQR}(\mathrm{IQR}=$ interquartile range $=3^{\text {rd }}$ quartile minus the $1^{\text {st }}$ quartile). These points fall outside the whiskers on the boxplots.

