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Port Gardner Bay Regional Background Sediment Characterization

Everett, WA

Supplemental Sampling and Analysis Plan

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Port Gardner Bay Regional Background Sediment Characterization

Supplemental Sampling and Analysis Plan

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**With support from
TerraStat Consulting Group and Avocet Consulting**

For

Toxics Cleanup Program
Washington State Department of Ecology
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List of Acronyms

COPC	chemicals of potential concern
cPAH	carcinogenic polycyclic aromatic hydrocarbon
CSL	Cleanup Screening Level
DGPS	differential global positioning system
Ecology	Washington State Department of Ecology
EIM	Environmental Information Management
FM	Field Manager
GPC	gel permeation chromatography
GPM	Government Project Manager
HSP	Health and Safety Plan
LCS/LCSD	laboratory control sample/laboratory control sample duplicates
MS/MSD	matrix spike/matrix spike duplicate
MTCA	Model Toxics Control Act
NAD83	1983 North American Datum
cPAH	carcinogenic polycyclic aromatic hydrocarbon
PCB	polychlorinated biphenyl
PPE	personal protective clothing
PQL	practical quantitation limit
PSEP	Puget Sound Estuary Program
QA/QC	Quality Assurance/Quality Control
QAPP	Quality Assurance Project Plan
RPD	relative percent difference
SAP	Sampling and Analysis Plan
SIM	select ion monitoring
SMARM	Sediment Management Annual Review Meetings
SMS	Sediment Management Standards
SOP	standard operating procedure
SCO	Sediment Cleanup Objective
TCDD	2,3,7,8-Tetrachlorodibenzodioxin
TEF	toxic equivalent factor
TEQ	toxic equivalent quotient
TOC	total organic carbon
UCL	upper confidence limit
USEPA	U.S. Environmental Protection Agency
UTL	upper tolerance limit
WAAS	Wide Area Augmentation System
WAC	Washington Administrative Code
WHO	World Health Organization

Introduction

The scope of work described in this Supplemental Sampling and Analysis Plan (SAP) for Port Gardner Bay is intended to supplement the surface sediment data collected in 2013 for establishment of sediment regional background. The original Port Gardner Bay SAP was finalized on March 19, 2013 (Ecology 2013a), sampling was conducted in Port Gardner Bay on March 26-29, 2013, and a data package containing the data, graphs and figures, and limited interpretation was provided to the stakeholders on August 5, 2013. To review the 2013 SAP and data package, see Ecology (2013a) and visit the website:

http://www.ecy.wa.gov/programs/tcp/sites_brochure/psi/everett/pg-sed.html

Port Gardner Bay is one of several embayments identified for focused sediment investigation, cleanup, and source control under the Washington State Department of Ecology (Ecology) Toxics Cleanup Program's Puget Sound Initiative. Lessons learned during the evaluation of these supplemental results will help to inform sample location placement and study design for future regional background characterization work in other areas of the state.

Regional Background Definition

The Sediment Management Standards Chapter 173-204 WAC (SMS) includes a provision for regional background as defined in WAC 173-204-505(16) and parameters for establishing regional background (WAC 173-204-560(5)):

“Regional Background” means the concentration of a contaminant within a department defined geographic area that is primarily attributable to diffuse sources, such as atmospheric deposition or storm water, not attributable to a specific source or release.

The SMS revisions are intended to provide flexibility to establish regional background on a case by case basis and do not prescribe specifically how regional background should be established. The approach and methods contained in the original Port Gardner Regional Background SAP (Ecology 2013a) were developed by Ecology to estimate regional background concentrations for selected hazardous substances (arsenic, cadmium, mercury, carcinogenic polycyclic aromatic hydrocarbons [cPAHs], dioxins/furans, and polychlorinated biphenyls [PCBs]) in Port Gardner Bay.

Stakeholder Discussions

In 2013, Ecology received a number of comments from stakeholders dealing with the initial Port Gardner Regional Background SAP (Ecology 2013a) and the North Olympic Peninsula Regional Background SAP (Ecology 2013c), some of which were incorporated into the final SAPs. Many stakeholders requested that for future regional background characterizations, they would like to

work with Ecology before SAPs were drafted and submitted for public comment. In response, Ecology engaged stakeholders earlier in the process for the Elliott Bay and the Lower Duwamish River regional background work. This included conducting a series of interviews with key regional stakeholders in June 2013 to prepare for a September 2013 technical workshop to 1) discuss whether to establish regional background in Elliott Bay and/or the Lower Duwamish River, 2) share information and data, including stakeholder and Ecology presentations, and 3) collaboratively work on the sampling design, which included discussion of alternative sampling approaches, for both areas. In addition, Ecology received a number of follow up comment letters after this technical workshop.

Based on all of the collective comments and discussions, Ecology determined that some modifications to the original sampling design used to establish regional background were appropriate. This Supplemental SAP describes the additional sampling that will be carried out to implement this revised approach for Port Gardner Bay. In addition, this revised approach will be updated in the final Sediment Cleanup Users Manual II guidance (SCUM II, Ecology 2013b).

The following modifications have been incorporated into this Supplemental SAP:

- **Rationale and “Conceptual Bay Model.”** Future SAPs will contain a more complete discussion of the selected analytes, rationale and existing information informing development of the sampling area, and the rationale for the selected sampling method(s) and/or modeling approach (if used). These choices will be based on a “conceptual bay or site model” and key features of the bay or area that influenced these decisions. This will include, as appropriate, known sites and sources, existing chemistry data, existing modeling information, hydrodynamic information, bathymetry, etc. A conceptual bay model for Port Gardner Bay has been added to this Supplemental SAP to guide the sampling design.
- **Sampling Area.** The area in which sediment samples will be collected has been modified to be more consistent with the SMS definition of regional background (WAC 173-204-505(16)). This will entail sampling closer to the shoreline, sources, and sites, while remaining outside areas of direct influence. It will, however, no longer use a default distance from these areas. Instead, bay-specific information has been used, where available, to determine areas associated with depositional zones of outfalls or other point sources and areas directly affected by sites. Accordingly, this Supplemental SAP includes collection of samples in nearshore areas that were not sampled for the 2013 regional background characterization (Ecology 2013a). This will be done in a way that data can be statistically combined with the samples collected in 2013.

Project Scope

The purpose of this Supplemental SAP is to describe the manner and methods by which data collection and analyses will be performed. The results of this supplemental sampling effort will be used in conjunction with the original data collected in 2013 (Ecology 2013a) to determine regional background sediment concentrations in Port Gardner Bay. For clarity, the following nomenclature is used through the rest of the document:

- Phase I refers to the original 2013 regional background sampling (Ecology 2013a).
- Phase II refers to the 2014 supplemental regional background sampling, for which this Supplemental SAP has been written.

Both Phase I and Phase II include baseline and secondary samples. Baseline samples are those collected and analyzed for the full suite of contaminants of concern. Secondary samples are collected but archived in case additional samples are needed to meet project precision goals.

The bioaccumulative contaminants of concern for this investigation include:

- metals (arsenic, cadmium, and mercury),
- carcinogenic polycyclic aromatic hydrocarbons (cPAHs),
- dioxins/furans congeners, and
- polychlorinated biphenyls congeners (PCBs).

The Supplemental SAP for this study, which includes the sampling and analytical procedures and methods, was prepared in accordance with the SMS requirements and draft SCUM II/SAPA guidance (Ecology 2013b/2008) and WAC 173-340-830.

Project Team and Responsibilities

NewFields and associated subcontractors will implement the Supplemental SAP under the direction of Ecology. The following sections describe the key roles and responsibilities of the project team.

Project Planning and Coordination

Chance Asher of Ecology will serve as the Government Project Manager (GPM) who will oversee the overall project coordination, supply government-furnished data and services, review reports, and coordinate with contractors. Tim Hammermeister will serve as the NewFields project manager and be responsible for executing the approved Supplemental SAP, overseeing the collection and analysis of field samples, and reporting analytical results.

Ecology

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Sample Collection

Dr. Will Hafner of NewFields will serve as field manager (FM) responsible for collecting and processing samples in accordance with the Supplemental SAP, and transporting samples to the analytical laboratory for analysis and testing. The FM will ensure accurate station positioning and reporting.

Laboratory Sample Preparation and Analysis

Dr. Will Hafner of NewFields will serve as laboratory coordinator responsible for subcontracting state-certified laboratories, delivering samples to the analytical and biological laboratories, and ensuring observation of established protocols for decontamination, sample preservation, holding times, chain-of-custody documentation, and laboratory reporting.

QA/QC Management

Dr. Will Hafner will serve as the NewFields quality assurance/quality control (QA/QC) manager providing quality assurance oversight for the laboratory programs ensuring that the laboratory analytical and QA/QC data are considered valid, and that procedures meet the required analytical quality control limits.

Health and Safety Manager

Preston Martin will serve as the designated NewFields Health and Safety Manager. The Health and Safety Manager is responsible for ensuring that all personnel are properly trained, fully aware of potential site hazards, conduct all work in a safe manner, wear appropriate personal protective clothing (PPE), and abide by the conditions set forth in the site-specific Health and Safety Plan (HSP).

Subcontractor Support

The NewFields project team will consist of the following subcontractors and external support to assist in the data collection activities and provide analytical laboratory services:

- Sampling Vessel
Bio-Marine Enterprise
R/V Kittiwake
Charles Eaton

2717 3rd Ave. N
Seattle, WA 98109
Phone: (206) 714-1055
cmeaton@msn.com

- Analytical Chemistry (metals, cPAHs, sediment conventionals)

Analytical Resources, Incorporated

Cheronne Oreiro
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Tukwila, WA 98166
Phone: (206) 695-6214
cheronneo@arilabs.com

- Dioxin/Furan and PCB Congener Analysis

Axys Analytical Services, Ltd.

Cynthia Tomey
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Sidney, BC V8L 3S8
CANADA
Phone: (250) 655-5800
Fax: (250) 655-5811
ctomey@axys.com

- Statistical Guidance

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- Data Validation

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Schedule

The proposed schedule for field activities is two days between April 22 and April 25, 2014. The sampling schedule was set to not coincide with any dredged material disposal at the Port Gardner open-water disposal site.

Port Gardner Bay Conceptual Model

Port Gardner Bay is an embayment of Puget Sound's Whidbey Basin, bounded to the east by the City of Everett (Figure 1). Port Gardner Bay has had a wide variety of commercial and industrial uses, multiple potential point sources of contamination, and an overall history of contamination. Since the early 1900s, the lower Snohomish River has been used for commercial and industrial purposes, often related to timber and maritime industries (saw mills, paper production, boat building, and waste disposal). The area of Port Gardner Bay located between Hat (Gedney) Island and the City of Everett was the original Phase I area of interest (AOI) for evaluating regional background, but has been modified for Phase II based on the Conceptual Bay Model, as described below.

Hydrology and Bathymetry

The Snohomish River system, the second largest river discharge into Puget Sound, empties into Port Gardner Bay at the City of Everett waterfront and provides approximately 30 percent of the freshwater discharge to the Whidbey Basin. Originating in the Cascade Mountains, tributaries of the Snohomish River drain a variety of forested, agricultural, and industrial properties. The mouth of the Snohomish River's main channel is bounded to the west by Jetty Island, a manmade island composed of sediment from continual maintenance dredging of the river channel (Figure 1). Currently, the lower Snohomish River Estuary is home to numerous environmental restoration projects focused on tideland recovery and habitat restoration.

Many of the samples from Phase I were collected from the Snohomish River Delta. While the Snohomish River provides much of the total deposition to Port Gardner Bay, it may not be completely representative of regional background, because it is relatively coarse-grained and low in chemical concentrations. The influence of the Snohomish River can be clearly seen in a map of grain size for the region (Figure 2). The coarse particulate deposition from the Snohomish River corresponded approximately to the 30 percent fines contour. Therefore, Ecology decided to exclude areas north of this contour from the regional background Phase II AOI. The delta north of the 30 percent fines contour was primarily comprised of sediments with less than 1 percent measured total organic carbon (TOC; Figure 3).

In addition, deeper areas of the northwest corner of the Phase II AOI appear, from a bathymetric/hydrological standpoint, to be more consistent with outer areas of Possession Sound than the Everett nearshore area that may be representative of regional background. Therefore, the northernmost boundary of the Phase II AOI was drawn at the contour corresponding to less than 30 percent fines, excluding areas north and northwest of that line. Samples collected during Phase I within the excluded area will be excluded from the Phase II regional background calculations.

Sources, Sites, and Areas of Influence

Port Gardner Bay has been the focus of several sediment cleanup projects. Ten sites within the region of Port Gardner Bay have recently been identified as PSI sites for focused sediment cleanup and source control (Figure 1). Over the last 25 years, several sediment investigations have detected chlorinated aromatics, PAHs, metals, miscellaneous extractables, pesticides, phenols, and phthalates at levels exceeding current SMS criteria in various locations at these sites. These sites include (with primary COPCs listed):

- **East Waterway**
 - Metals – arsenic, copper, lead, mercury and zinc.
 - PAHs – acenaphthene, anthracene, benzo(a)pyrene, benzo(g,h,i)perylene, chrysene, dibenz(a,h)anthracene, fluoranthene, fluorene, naphthalene, phenanthrene, pyrene, high molecular weight PAHs, and low molecular weight PAHs.
 - Semivolatile Organic Compounds – 1,4-dichlorobenzene, 2,4-dimethylphenol, 2-methylnaphthalene, 2-methylphenol, 4-methylphenol, benzoic acid, benzyl alcohol, bis(2-ethylhexyl)phthalate, butylbenzylphthalate, dibenzofuran, di-n-octyl phthalate, hexachlorobenzene, N-nitrosodiphenylamine, pentachlorophenol, and phenol.
 - Total PCBs.
 - Dioxins/Furans – elevated with maximum concentrations greater than 300 ng/kg toxicity equivalents (TEQ).
- **Weyerhaeuser Mill A**
 - Metals – arsenic, cadmium, copper, lead, mercury, nickel, and zinc.
 - PAHs – acenaphthene, anthracene, benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene, chrysene, fluoranthene, fluorene, naphthalene, phenanthrene, and low molecular weight PAHs.
 - Semivolatile Organic Compounds (SVOCs) – 2,4-dimethylphenol, 2-methylnaphthalene, 2-methylphenol, 4-methylphenol, benzoic acid, benzyl alcohol, bis(2-ethylhexyl)phthalate, dibenzofuran, and phenol.
 - Total PCBs.
 - Dioxins/Furans.
- **Jeld-Wen**
 - Dioxins/furans; PCBs; PAHs; total petroleum hydrocarbons (TPH); VOCs, including benzene and toluene; SVOCs, including 2-methylnaphthalene, naphthalene, 2,4-dimethylphenol, carbazole, dibenzofuran, and m,p-cresol.

- **BayWood Products**
 - Dioxins/furans, PAHs, PCBs, TPH.
- **TC Systems Inc.**
 - Metals, including arsenic, copper, and lead; PAHs; SVOCs, including pentachlorophenol, dibenzofuran, naphthalene, and 1-methylnaphthalene; TPH, and PCBs.
- **Everett Shipyard**
 - PCBs; PAHs; metals, including antimony, arsenic, lead, mercury, silver, copper, zinc, and nickel; dioxins/furans; organotins, including tributyltin; bis(2-ethylhexyl) phthalate, and 4-methylphenol.
- **Kimberly Clark Worldwide**
 - Dioxins/furans; PAHs; metals, including lead, arsenic, nickel, copper, zinc, and mercury; TPH; PCBs; and VOCs, including xylene and benzene.
- **Exxon Mobil ADC**
 - PAHs, TPH, and lead.
- **West End**
 - Arsenic, copper, TPH, and PAHs.
- **Ameron/Hulbert**
 - Arsenic, copper, lead, antimony, PAHs, and TPH.

In addition to these known sites, there were three other locations that required further evaluation to determine the boundaries of the Phase II AOI: 1) a large deepwater diffuser outfall southwest of the Weyerhaeuser Mill site, which once discharged wastewater from pulp mills along the shoreline and now discharges municipal stormwater, 2) a historical disposal area dated 1954-1966 on the bathymetry map, and 3) the DMMP disposal site in Port Gardner Bay.

The chemistry results from previous studies (including the Phase I results) were reviewed to identify areas primarily influenced by these sites and sources, as any such areas should not be included in the definition of regional background. All available chemistry results for Port Gardner Bay were downloaded from EIM. The data were plotted and interpolated using a Geographic Information System (GIS; Environmental Systems Research Institute's ArcGIS) to better visualize the spatial distribution of the COPCs (Figures 4–11; areas without data shown in purple). A description of the data set and a data summary are presented in Appendix A.

Several of the cleanup sites discussed above are east of Jetty Island or near the navigational channel, outside the AOI for Port Gardner Bay. The waterway there is also, in general, too narrow for identification of a sampling area not primarily influenced by the neighboring sites.

The boundaries of the sediment cleanup sites in and adjacent to East Waterway have not been fully delineated. However, based on the chemical contouring, concentrations of metals and dioxins/furans are elevated throughout the entire East Waterway (Figures 4–11). PAHs are also high, but these elevated concentrations extend further out into Port Gardner Bay. Regional

background concentrations for PCBs will be calculated based on congener data. However, existing congener data within Port Gardner Bay is spatially limited (Figures 9 and 10). An interpolated plot of PCB Aroclor concentrations is shown in Figure 11 to demonstrate the extent of known PCB contamination in the vicinity of the East Waterway.

The elevated concentration gradients for several COPCs extended beyond the mouth of the East Waterway to the vicinity of the Weyerhaeuser Mill site just south of the waterway. The area from the southwest corner of Mill A north to the naval wharf, along with East Waterway, is considered to be primarily influenced by known sites (sources or releases). Therefore, they are not included in the Phase II AOI to remain consistent with the definition and intent of regional background (WAC 173-204-505(16) and 173-204-560(5)). This preliminary determination is solely for the purposes of establishing an AOI for the Phase II sampling to determine regional background, and should not be construed as a site boundary or any other regulatory determination.

Interpolations of existing and Phase I data for the target COPCs (Figures 11) did not demonstrate any chemical gradient in the vicinity of or moving away from the deepwater diffuser outfall. In addition, there was no record of the historical disposal site having been used. As with Phase I regional background sampling, no sample locations are targeted within the DMMP open water disposal site. As a result, there did not appear to be any area of primary influence from a site or point source along the southern shoreline, and the southeastern Phase II AOI boundary was moved closer to shore to the -6 ft MLLW bathymetric contour.

Summary of the Area of Interest Representative of Regional Background

The Phase II AOI was defined as a hydraulically connected marine environment including areas potentially affected by large-scale urban sources, but excluding areas directly or primarily influenced by sites and other sources. Based on the above conceptual bay model, the Phase II AOI was defined as follows:

- The southwestern boundary was established as in Phase I.
- The open-water DMMP dredged material disposal site was excluded;. Most recent disposal events (2013) were clean sand/gravel from the Snohomish River O&M dredging and would not be reflective of regional background.
- The northern boundary was roughly defined as the southern extent of coarse particulate deposition from the Snohomish River, corresponding to approximately the 30 percent fines contour line (Figure 2). All areas south of this line were retained.
- Deeper areas northwest of the AOI were not considered representative of potential regional COPC sources from the urbanized shoreline, and coarse-grained areas north of this line were considered representative of deposition from the Snohomish River.

- The southeast boundary was moved to -6 MLLW along the shoreline south of the Weyerhaeuser Mill Site. The -6 MLLW depth was initially set in Phase I to avoid sampling in shallow or shoaling area of the Snohomish River Delta. It was retained in Phase II to maintain a reasonable sampling distance from shore.
- East Waterway and areas outside of the mouth of East Waterway north of a line from the south end of the Weyerhaeuser Mill site extending to the western tip of the Everett Naval Station Pier were excluded.
- The AOI was not extended east of Jetty Island as much of this area is either intertidal or part of the navigational channel.

The Phase II AOI boundary is presented in Figure 1. The Phase I AOI boundary is included for comparison. Phase I samples that are no longer located within the Phase II AOI boundary will not be included for calculation of regional background.

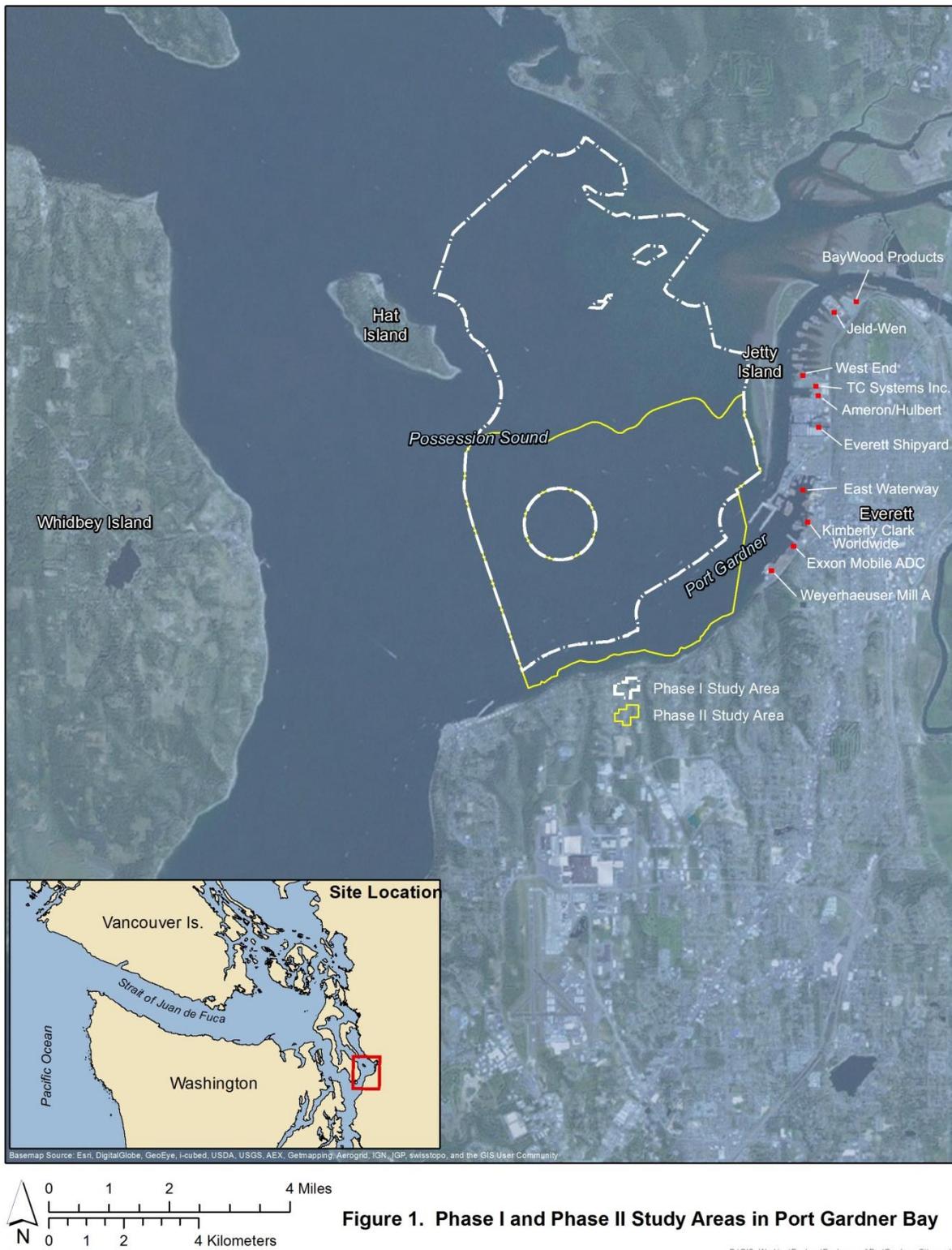


Figure 1. Phase I and Phase II Study Areas in Port Gardner Bay

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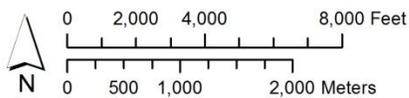
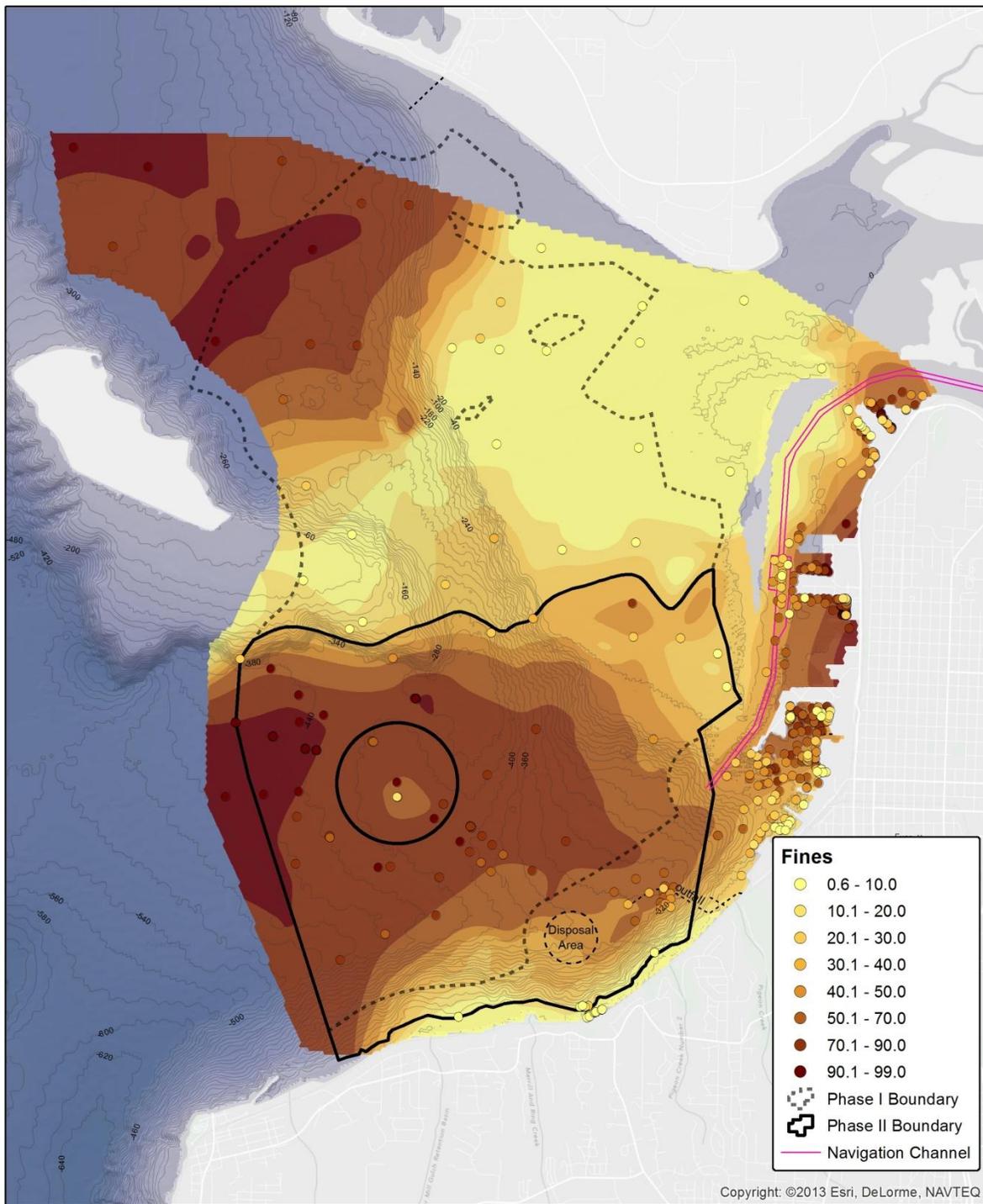
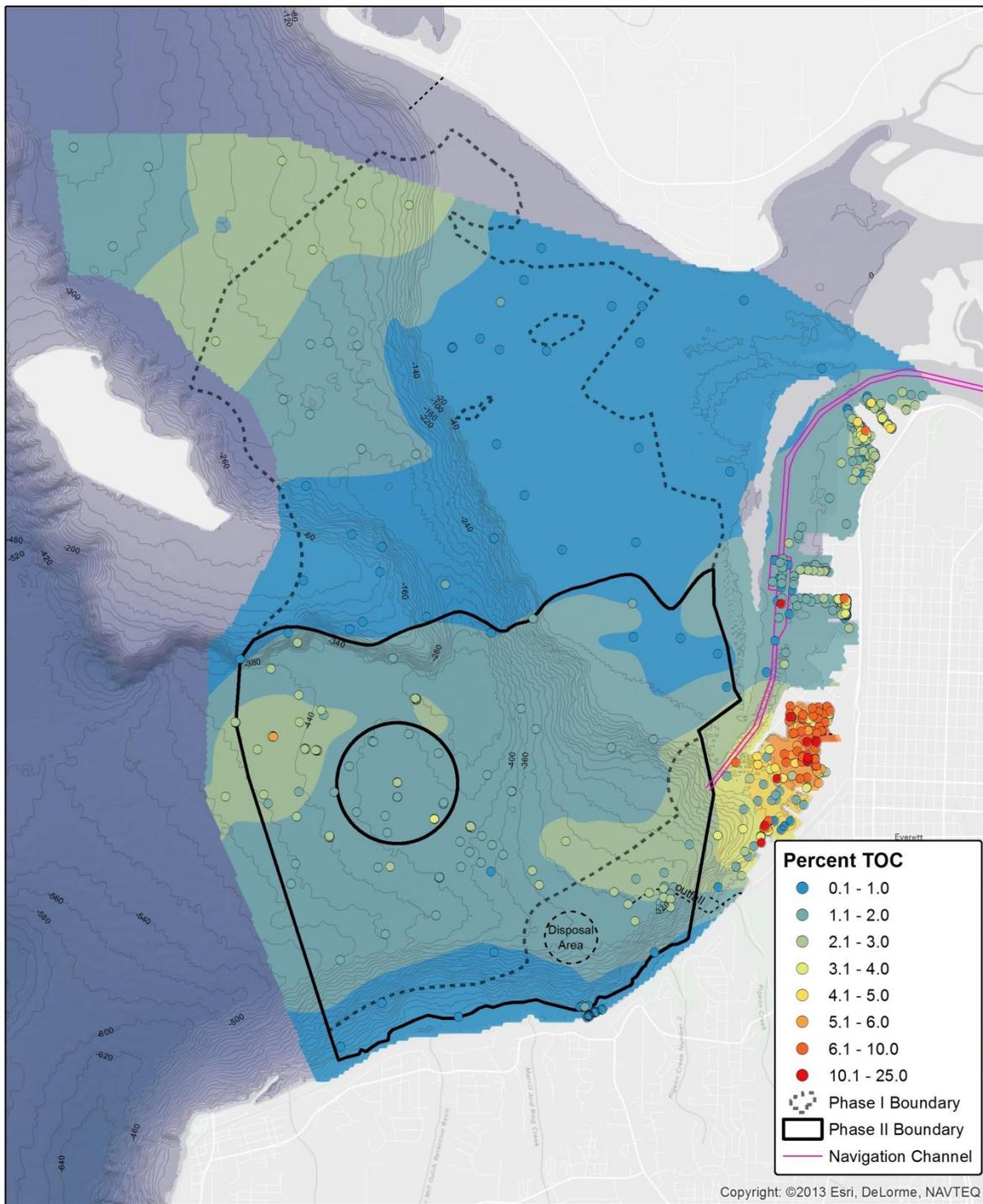


Figure 2. Interpolated Percent Fines for Port Gardner Bay.

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Copyright: ©2013 Esri, DeLorme, NAVTEQ

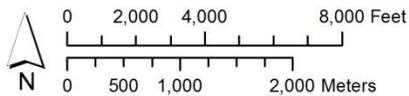


Figure 3. Interpolated Percent Total Organic Carbon for Port Gardner Bay

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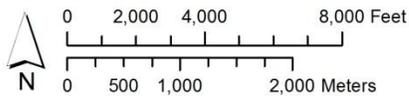
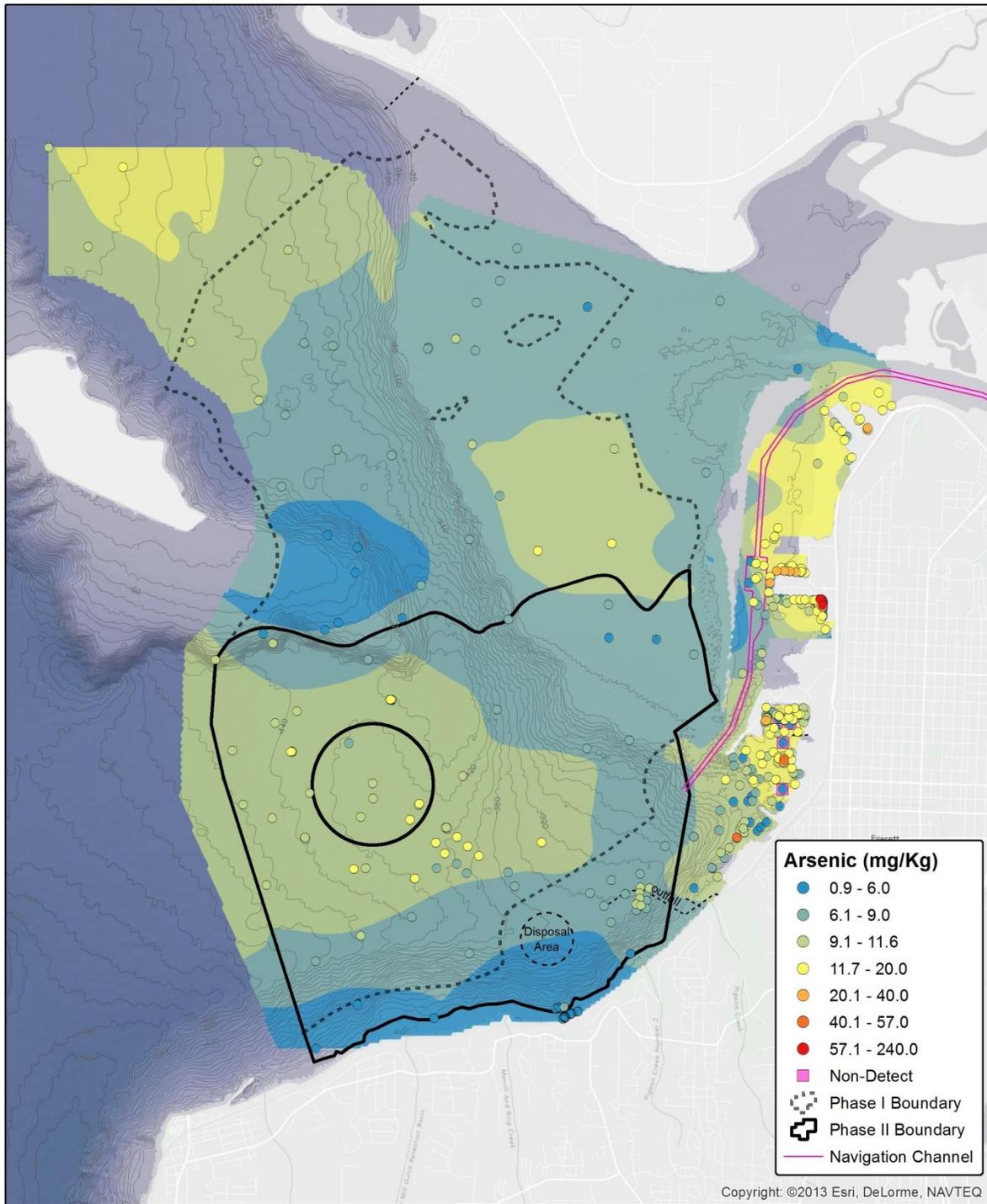


Figure 4. Interpolated Arsenic Concentrations for Port Gardner Bay

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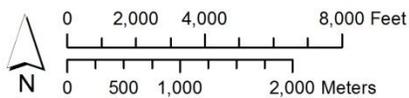
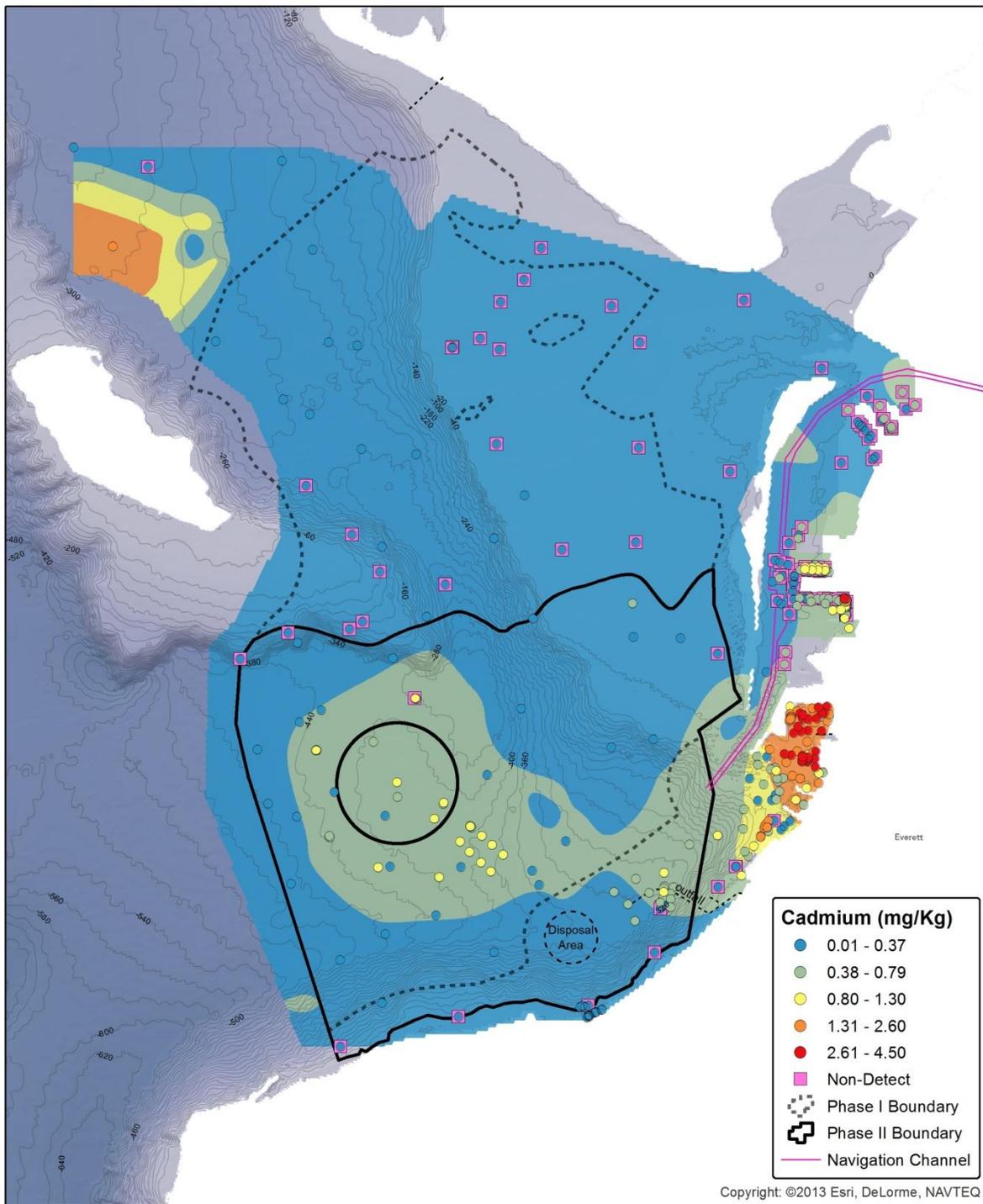


Figure 5. Interpolated Cadmium Concentrations for Port Gardner Bay

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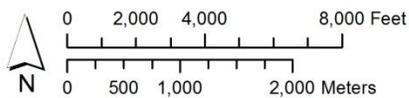
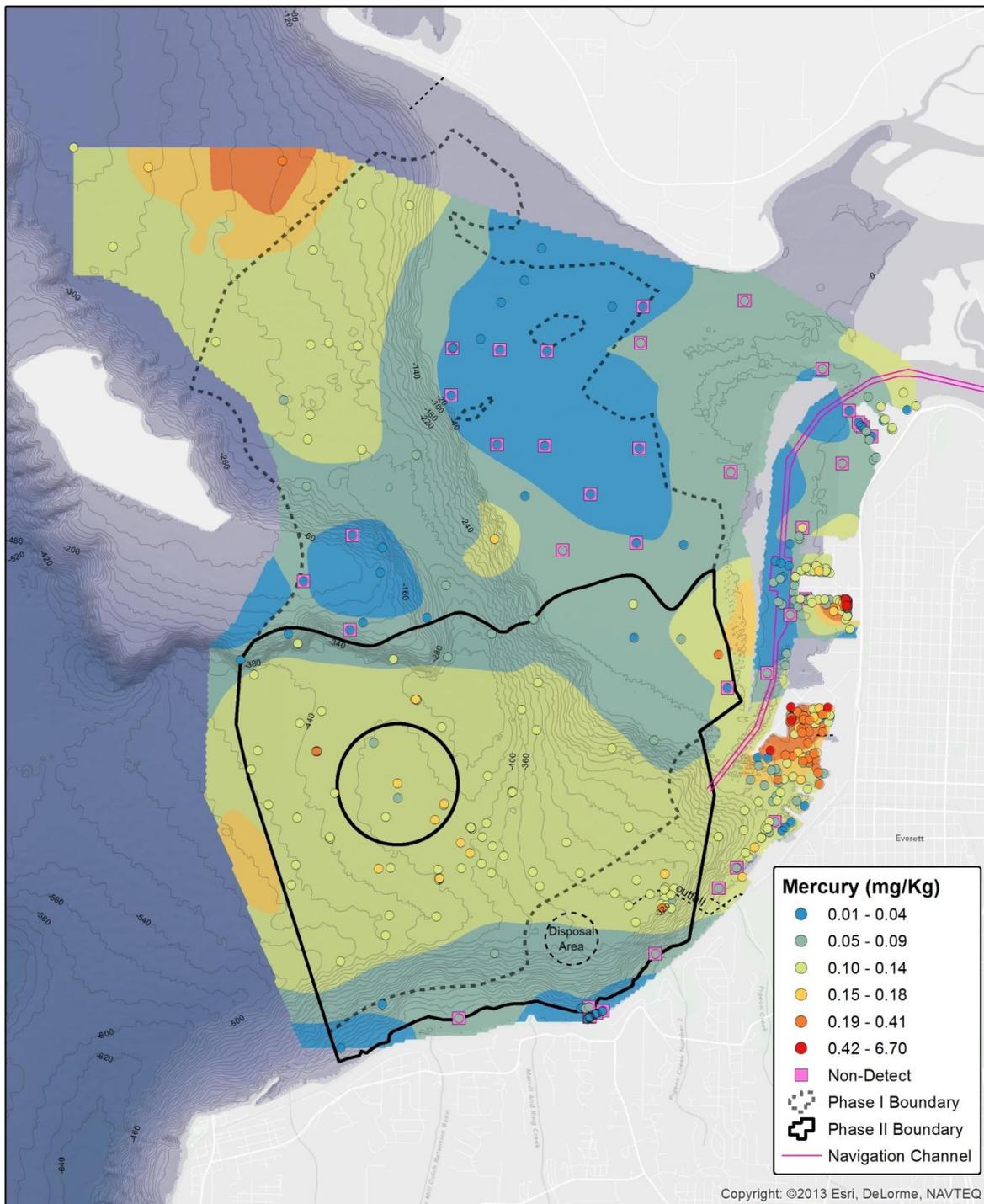


Figure 6. Interpolated Mercury Concentrations for Port Gardner Bay

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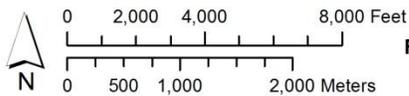
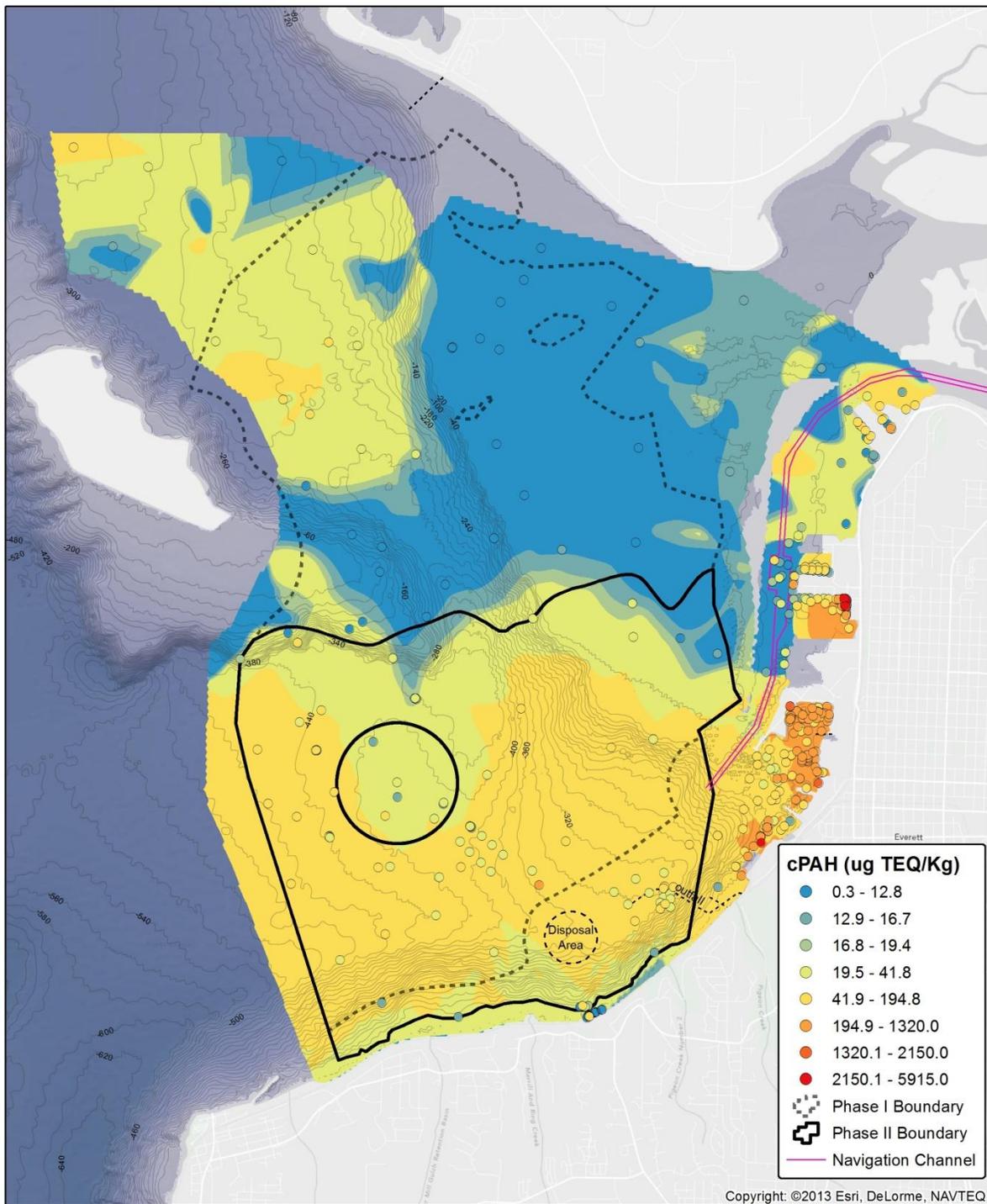


Figure 7. Interpolated Carcinogenic Polycyclic Aromatic Hydrocarbon (cPAH) Concentrations for Port Gardner Bay.

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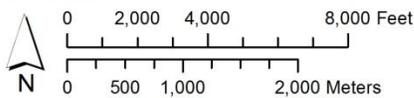
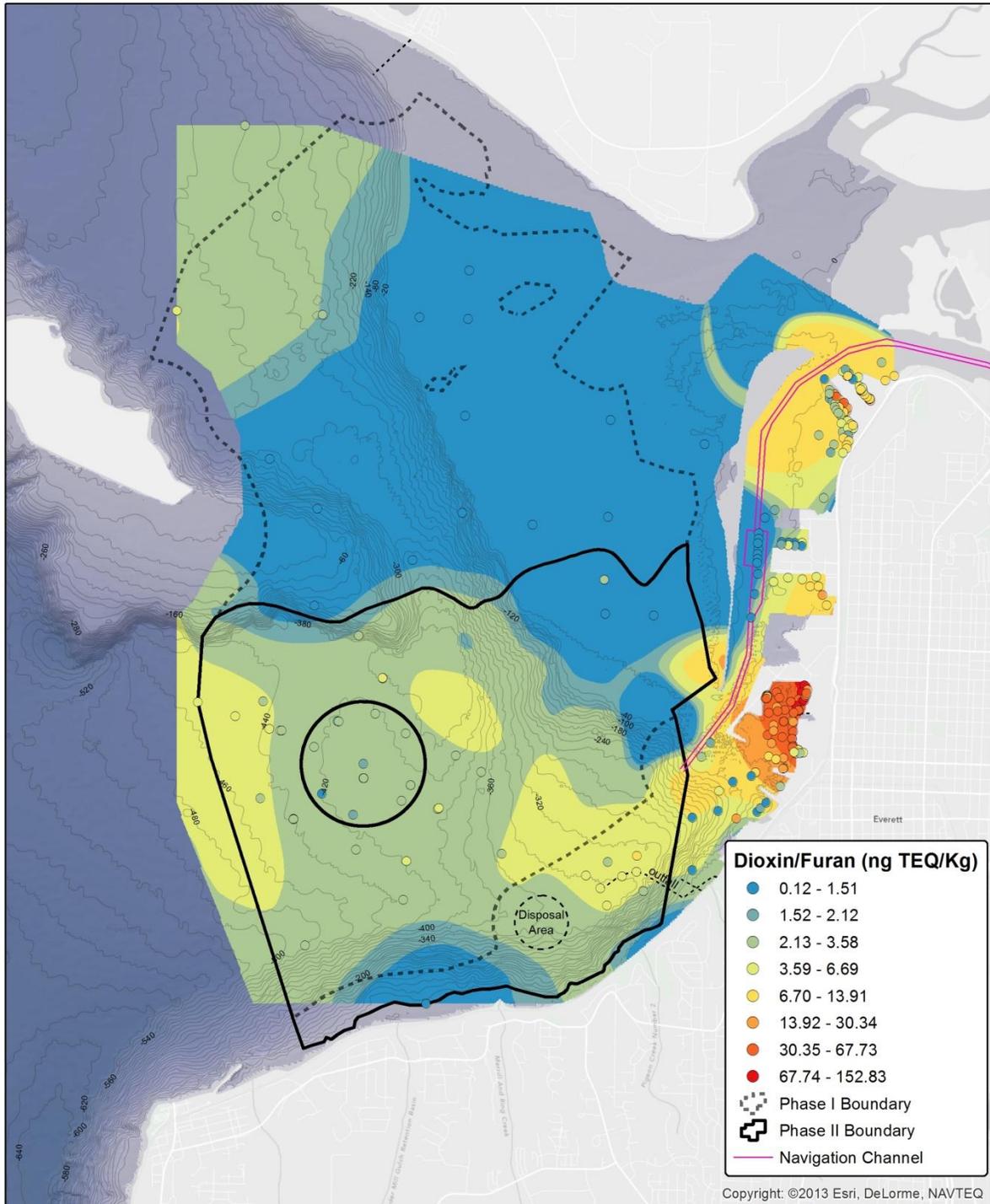


Figure 8. Interpolated Dioxin/Furan Congener TEQ Concentrations for Port Gardner Bay

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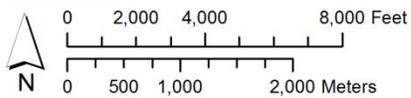
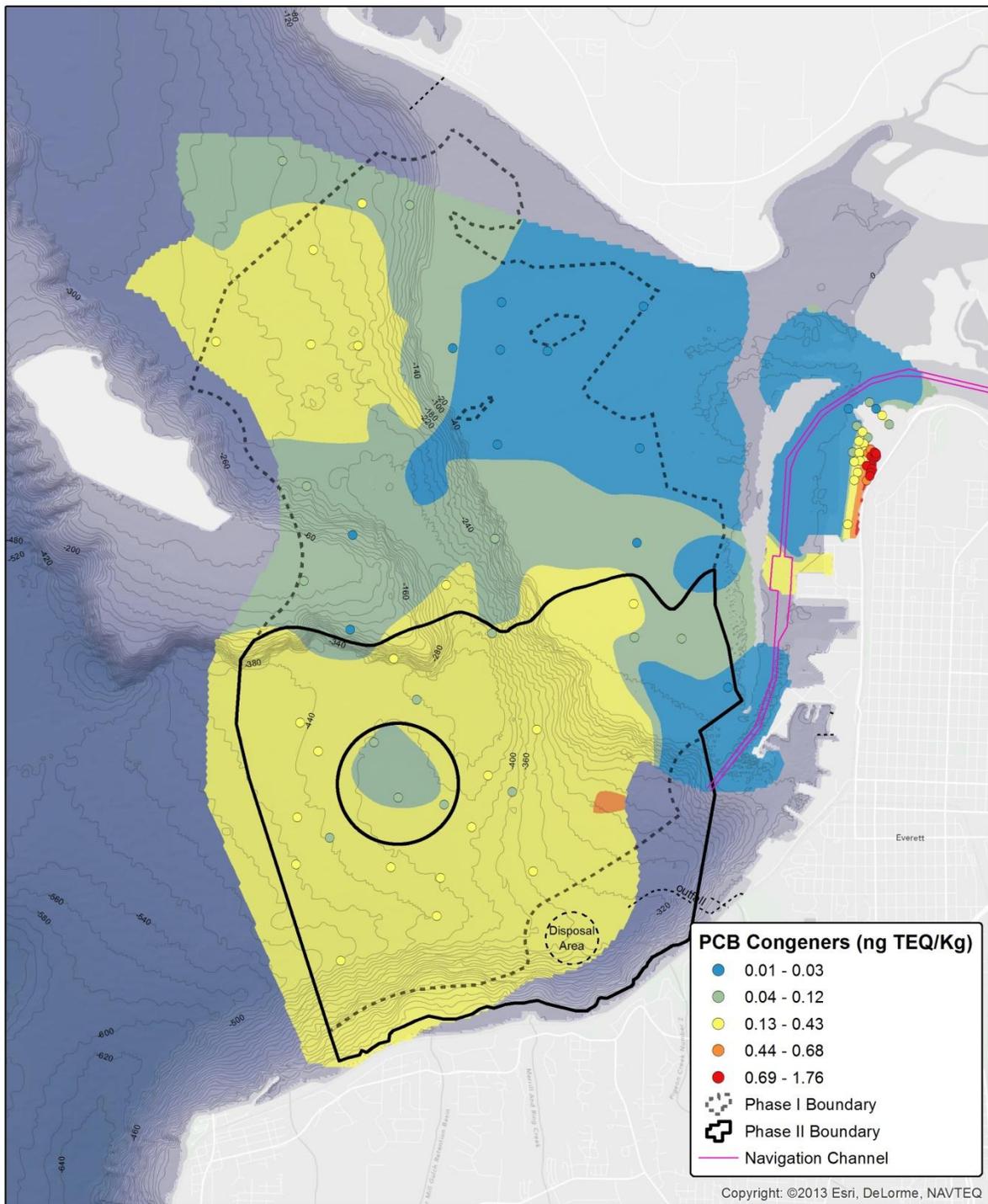


Figure 9. Interpolated Polychlorinated Biphenyl (PCB) TEQ Concentrations for Port Gardner Bay

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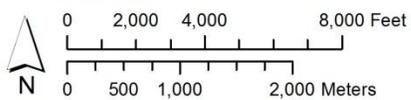
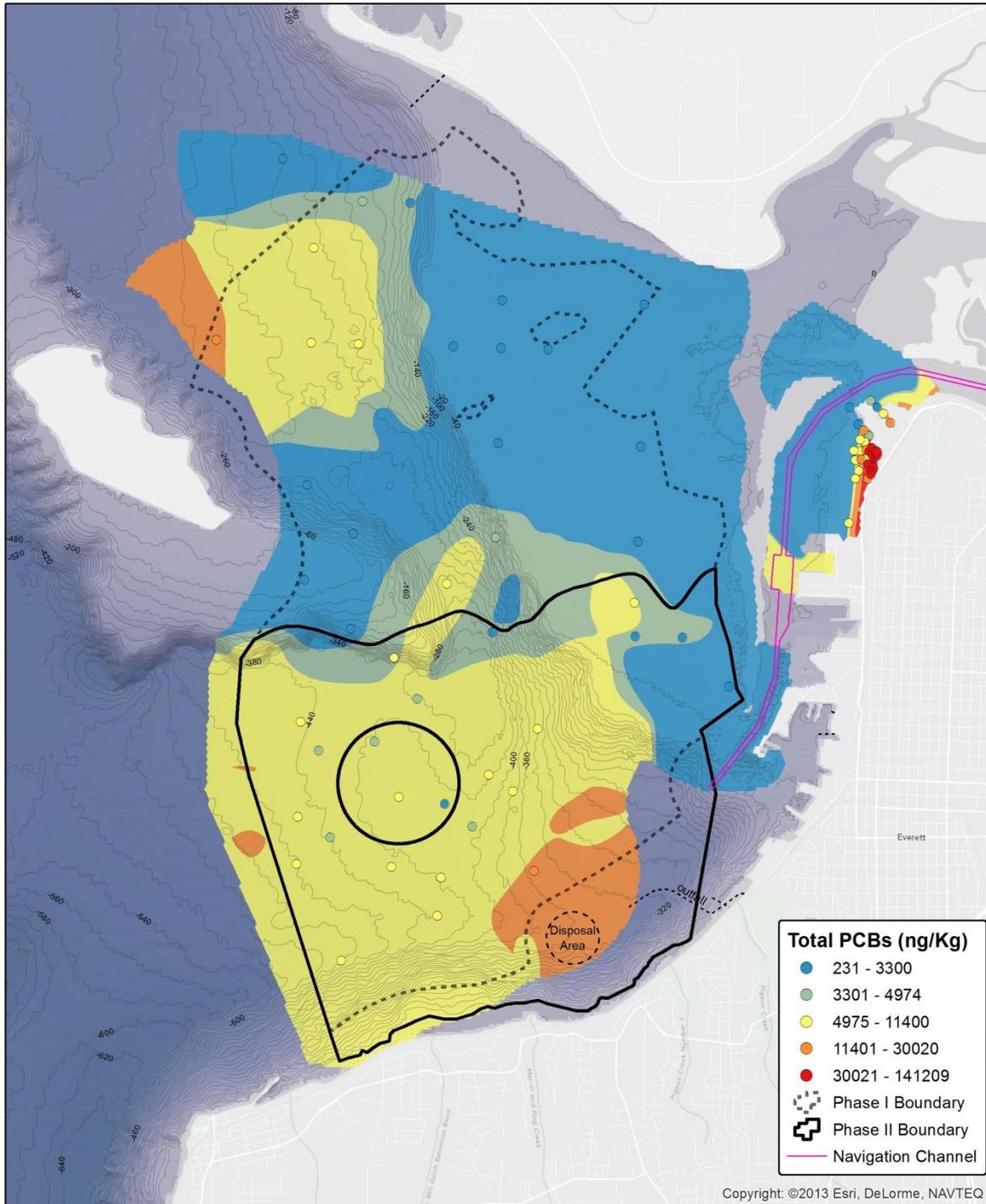


Figure 10. Interpolated Total Polychlorinated Biphenyl (PCB) Concentrations for Port Gardner Bay

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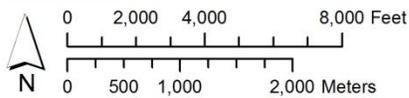
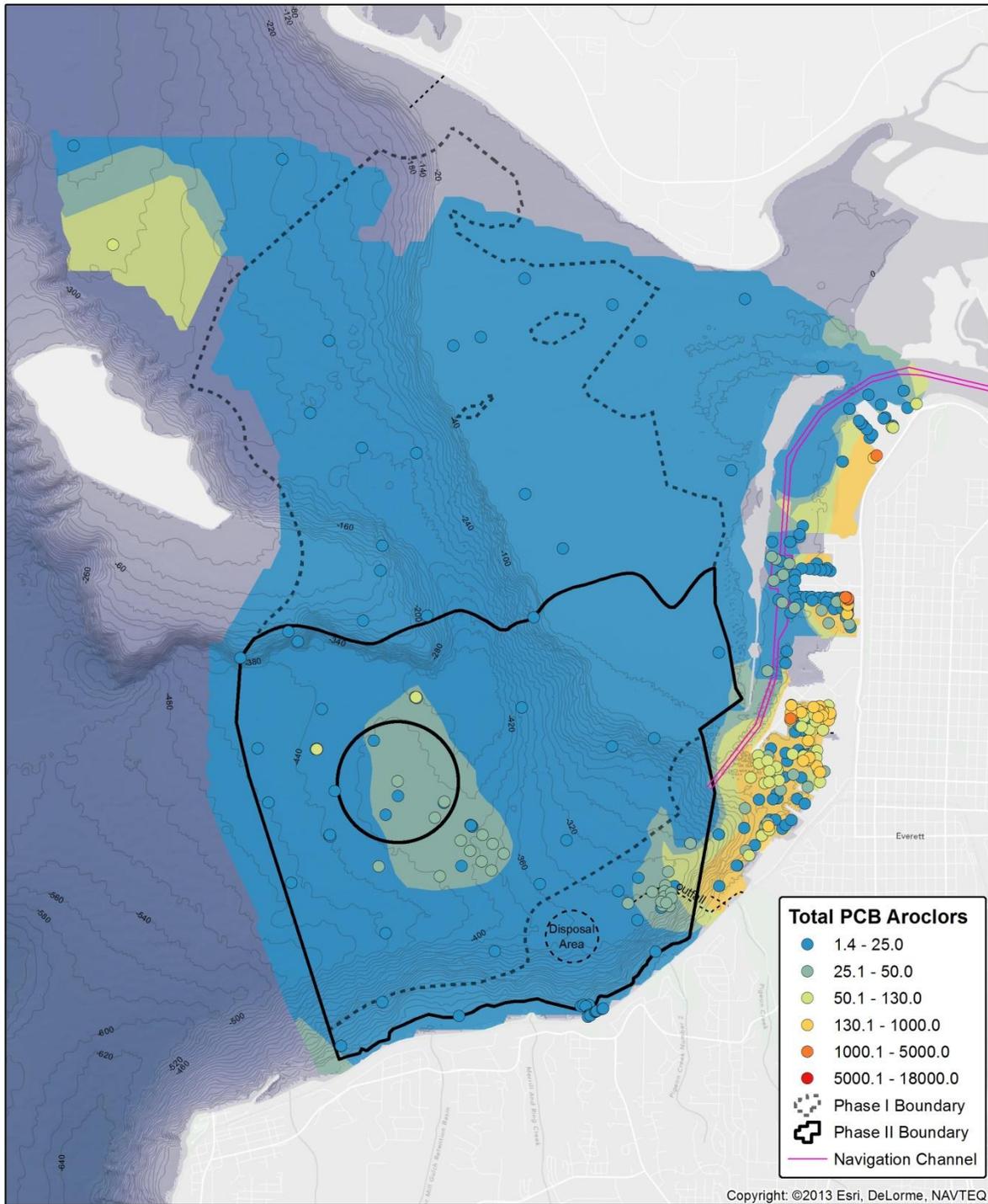


Figure 11. Interpolated Total Polychlorinated Biphenyl (PCB) Aroclor Concentrations for Port Gardner Bay

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Study Design

This section describes the study design for the Supplemental SAP including the new Phase II data collection in Port Gardner Bay. Several key study objectives were taken into consideration in the development of the study design, given the Phase II AOI developed in the previous section:

- Determine the minimum number of Phase II baseline samples needed to supplement existing Phase I data (Ecology 2013a) for calculation of regional background concentrations.
- Define the minimum distance between sampling locations to ensure results are independent among locations.
- Randomly select sediment sampling locations in such a manner that they can be statistically combined with the original Phase I data.

The following sections discuss the development and the details of the study design.

Use of Existing and Phase I Data

The Phase I SAP included a limited evaluation of existing chemistry data in Port Gardner Bay to determine whether results from historical sampling events (data less than 10 years old) could be used to supplement the newly collected data for determination of regional background (Ecology 2013a). None of the existing data was incorporated into the Phase I regional background calculations, and will not be incorporated into the Phase II regional background calculations, for two primary reasons:

- For most COPCs, the number of Phase I baseline samples provided a sufficient sample size.
- There was a desire to use a synoptic data set to reduce variability. The sampling purposes and analytical methodologies for the existing data differed from those used as part of Phase I regional background sampling.

For these reasons, no historical data will be incorporated into the calculation of regional background. Instead, the new samples collected as part of Phase II will be used to supplement the Phase I results. Phase II samples will be collected in the same manner, analyzed by the same labs, and can be combined with the Phase I data set without loss of statistical integrity.

Sample Number and Density

The combined Phase I and II regional background data set will be used to characterize the concentration distributions of target analytes within the revised AOI (Figure 1). Based on an analysis of existing Phase I data, an initial sample size of at least 25 samples was expected to provide an indication of the shape of the concentration distributions, and preliminary estimates for the mean and variance of each analyte. The combined dataset from the two phases of the Port Gardner sediment characterization will consist of 27 samples, including:

- Eleven baseline samples collected as part of Phase I and analyzed for the full suite of COPCs.
- Four secondary samples collected as part of Phase I and analyzed in April 2013 for conventionals and mercury, and in May 2013 for PCB congeners. Remaining archived sediment from these four samples was submitted to the analytical laboratories on March 19, 2014 for analysis of dioxin/furan congeners, arsenic, cadmium, and cPAH.
- Twelve new baseline samples collected as part of Phase II and analyzed for the full suite of COPCs.

In addition to the 12 Phase II new baseline samples, three secondary sampling locations will be occupied during this investigation to archive sufficient sediment for further analysis as needed, with the exception of mercury. Due to the short holding time, all three of the secondary samples will be submitted for mercury analysis. Secondary samples will also be analyzed for grain size and total organic carbon (TOC).

From this data set, the regional background sediment concentrations will be calculated from the distribution (excluding any outlier(s)), utilizing the most appropriate parametric or non-parametric methods. Statistical precision of the data set will be measured as the width of the 95th percent upper confidence limit (UCL) on the mean as a percent of the mean, i.e., $(95 \text{ UCL} - \text{mean}) / \text{mean} \times 100$. A target value of 25 percent for the precision of the 95th percent UCL of the mean will be set. If the target value is exceeded for a given analyte, additional samples may be randomly selected from the set of archived secondary samples for analysis. Additional analyses will be conducted when deemed most appropriate for reducing uncertainty in the dataset. See the Data Analysis and Reporting section for more details on the calculation of regional background and the process for determining whether secondary sample analysis is needed.

Sample placement followed the same criteria outlined in Phase I (Ecology 2013a; Appendix D). Autocorrelation analyses were used to assess the similarity of sample results within increasing distance intervals for the existing Port Gardner Bay data set, dating from 2004 to 2010 (Appendix D). The smallest testable autocorrelation distance was 200 m. Data observations within 200 m of each other were determined to be significantly correlated (Pearson's $r = 0.74$, $p = 0.048$, $n = 6$); observations spaced between 200 m and 400 m apart were not correlated (Pearson's $r = -0.26$, $n = 15$). A minimum sampling interval of 400 m was expected to be

adequate to achieve an independent dataset. Ultimately, a slightly larger interval of 500 m was used as the distance between sampling locations to ensure an independent dataset, but allow for a sufficient number of samples to be located within the study AOI.

A GIS algorithm was used to ensure a minimum distance of 500 m between sampling locations and spatial evenness throughout the Phase II AOI. Fifteen previously analyzed Phase I locations were located within the Phase II AOI, and twelve new locations were targeted for this investigation. The expanded portion of the Phase II AOI (the southeastern portion in Figure 1) comprised about 20 percent of the total AOI. A sample size allocation that was approximately proportional to the relative size of each area was achieved with six of the twelve new locations within the expanded portion, and six within the original portion of the revised AOI for a total of 22 percent (6/27) and 78 percent (21/27), respectively. Using the same algorithm, the three secondary locations were randomly placed throughout the Phase II AOI (one fell within the expanded portion and two within the original AOI).

Sediment Sampling Locations

Twelve new baseline locations and three new secondary locations will be sampled as part of the Phase II investigation. The Phase II baseline samples are listed in Table 1. For reference, the Phase I samples that were analyzed and will be used in the Port Gardner Regional Background dataset are included in Table 2.

Phase II Baseline Sediment Sample Locations

The purpose of baseline sediment sample locations is to provide a minimum number of randomly placed sampling locations, at least 500 m apart, and to provide balanced spatial coverage for an area pre-determined to be representative of regional background conditions. The following data collection activities have been identified and are summarized in Table 1:

- Collect twelve surface sediment (0–10 cm) grab samples using a spatially balanced random design placed within the perimeter of the Phase II AOI designated in Figure 1.
- Submit Phase II sediment samples for analysis of the following bioaccumulative contaminants:
 - Metals (arsenic, cadmium, and mercury)
 - Dioxin/furan congeners
 - PCB congeners
 - cPAHs
- Submit sediment samples for analysis of sediment conventionals (grain size distribution, total volatile solids, and total organic carbon).
- Archive sediment from each location for additional analysis or re-analysis as needed.

The baseline sediment sample locations are presented in Figure 12. Target coordinates are provided in Table 4.

Table 1. Phase II Baseline Sediment Sample Locations and Analyses.

Sampling Location	Sediment Conventionals ¹	Metals ²	cPAH	Dioxin/Furan Congeners	PCB Congeners	Archive ³
PG-51	X	X	X	X	X	A
PG-52	X	X	X	X	X	A
PG-53	X	X	X	X	X	A
PG-54	X	X	X	X	X	A
PG-55	X	X	X	X	X	A
PG-56	X	X	X	X	X	A
PG-57	X	X	X	X	X	A
PG-58	X	X	X	X	X	A
PG-59	X	X	X	X	X	A
PG-60	X	X	X	X	X	A
PG-61	X	X	X	X	X	A
PG-62	X	X	X	X	X	A

Notes:

X – analyze, **A** – archive, **cPAH** – carcinogenic polycyclic aromatic hydrocarbons, **PCBs** – polychlorinated biphenyls.

¹ Sediment conventionals include total organic carbon (TOC), total volatile solids (TVS), and grain size distribution.

² Metals include arsenic, cadmium, and mercury.

³ Sediment archived for potential analysis or reanalysis.

Table 2. Phase I Sediment Sample Locations and Analyses Already Collected.

Sampling Location	Sediment Conventionals ¹	Mercury	Arsenic/ Cadmium	cPAH	Dioxin/Furan Congeners	PCB Congeners
Available Phase I Sample Results						
PG-01	X	X	X	X	X	X
PG-04	X	X	X	X	X	X
PG-05	X	X	X	X	X	X
PG-08	X	X	X	X	X	X
PG-09	X	X	X	X	X	X
PG-10	X	X	X	X	X	X
PG-12	X	X	X	X	X	X
PG-15	X	X	X	X	X	X
PG-17	X	X	X	X	X	X
PG-21	X	X	X	X	X	X
PG-23	X	X	X	X	X	X
PG-27	X	X	S	S	S	X
PG-28	X	X	S	S	S	X
PG-31	X	X	S	S	S	X
PG-34	X	X	S	S	S	X

Notes

X – analyzed as part of Phase I.,

S – submitted for analysis on March 19, 2014. These were secondary samples pulled from archive during Phase I to supplement the Phase I PCB congener data set to meet the precision target. Additional archived sediment was submitted for analysis of the remaining analytes March 2014 to complete the analysis suite.

cPAH – carcinogenic polycyclic aromatic hydrocarbons.

PCBs – polychlorinated biphenyls.

Secondary samples pulled from archive to supplement the Phase I PCB congener data set to meet the precision target.

¹ Sediment conventionals include total solids, total sulfides, total organic carbon (TOC), total volatile solids (TVS), and grain size distribution.

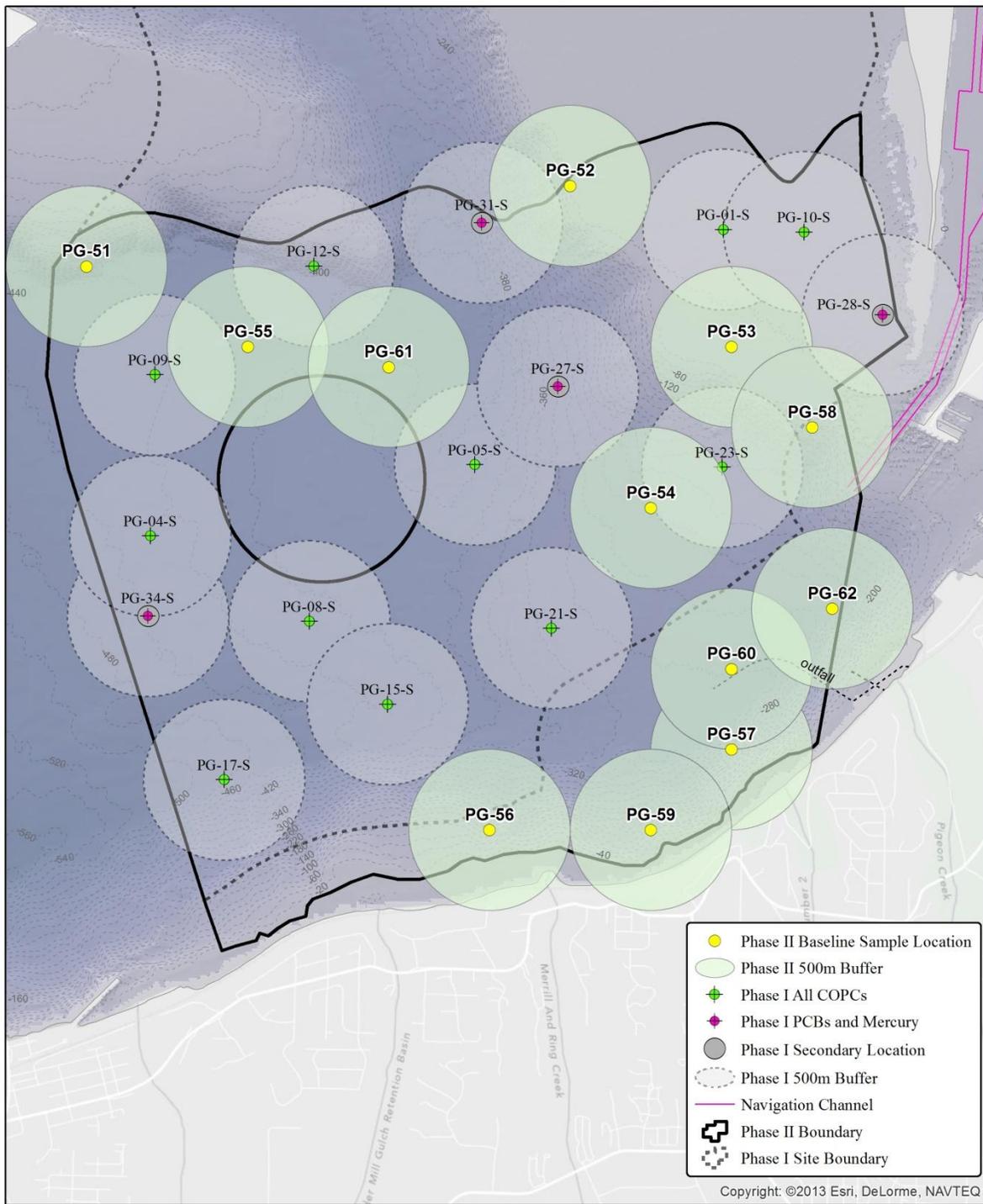


Figure 12. Baseline Sediment Sampling Locations in the Phase II Study Area

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Phase II Secondary Sediment Sample Locations

The purpose of the secondary sediment sample locations is to archive additional samples from randomly placed sampling locations that meet the minimum criteria of 500 m from the nearest baseline location. These archived samples would then be available, as needed, for chemical analysis to ensure a sufficient number of usable data are available for calculation of regional background concentrations. The following data collection activities have been identified and are summarized in Table 3:

The secondary sediment data collection activities include:

- Collect three surface sediment (0-10 cm) grab samples using a spatially balanced random design placed within the perimeter of the Phase II AOI designated in Figure 1.
- Analyze all three samples for mercury due to short holding time.
- Analyze all three samples for grain size, TVS, and TOC to better characterize the collected sediment.
- Archive sediment for all other analyses.

The secondary sediment sample locations are presented in Figure 13. Target coordinates are provided in Table 5.

Table 3. Phase II Secondary Sediment Sample Locations and Analyses.

Sampling Location	Sediment Conventionals ¹	Hg	As, Cd	cPAH	Dioxin/Furan Congeners	PCB Congeners	Archive
PG-63	X	X	A	A	A	A	A
PG-64	X	X	A	A	A	A	A
PG-65	X	X	A	A	A	A	A

Notes:

cPAHs – carcinogenic polycyclic aromatic hydrocarbons.

PCBs – polychlorinated biphenyls.

¹ Sediment conventionals include total organic carbon (TOC), total volatile solids (TVS), and grain size distribution.

X: sample to be collected and submitted for analysis/testing;

A: sample to be archived

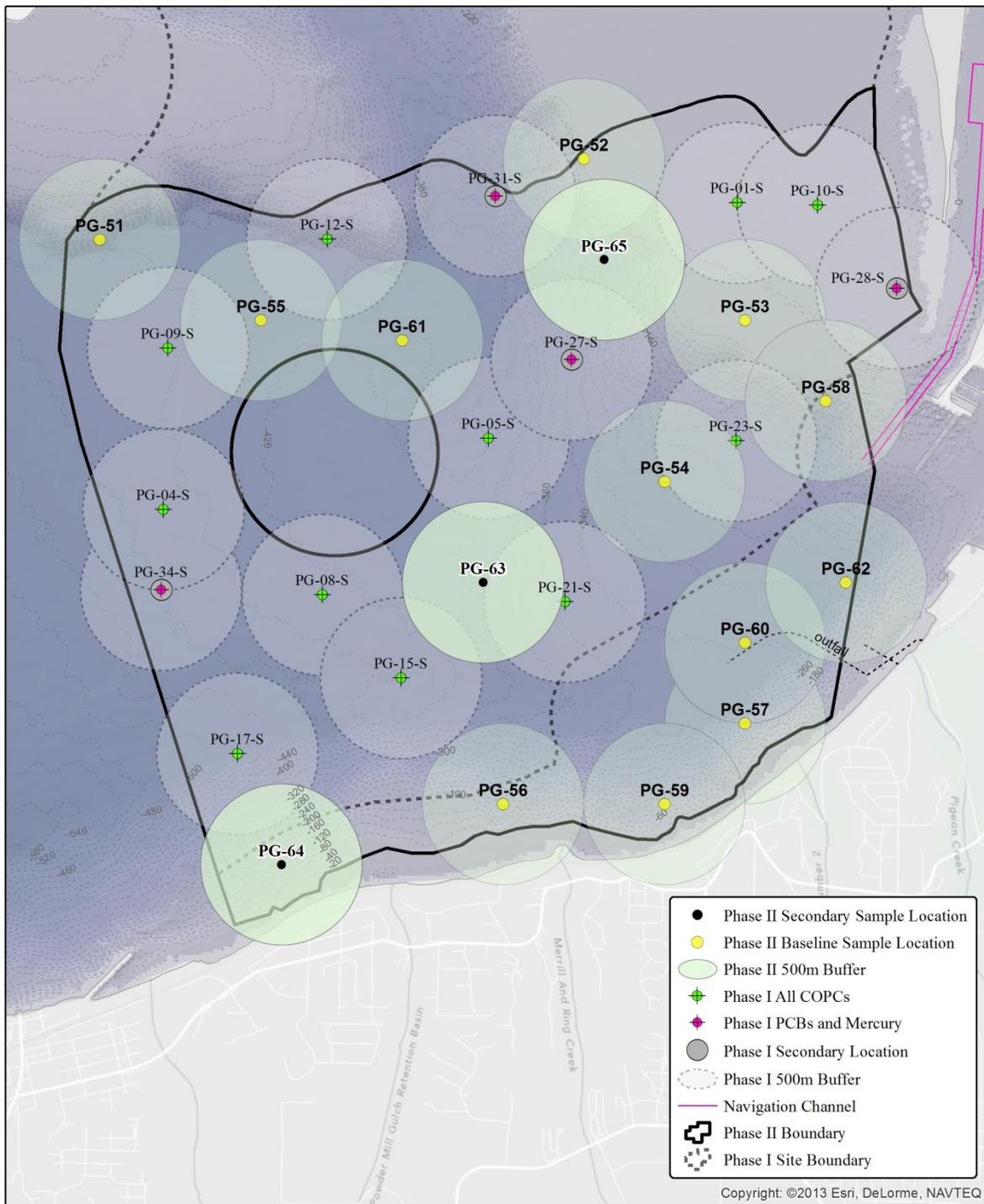


Figure 13. Secondary Sediment Sampling Locations in the Phase II Study Area

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Sample Collection and Handling Methods

This section describes the methodology for positioning, sample collection, processing, identification, documentation, equipment decontamination, and waste handling for the field investigation. Section 4.0 presents the laboratory methods for chemical analysis.

Sampling Platforms

The R/V *Kittiwake*, owned and operated by Mr. Charles Eaton of Bio-Marine Enterprises will be used for the surface sediment grabs in Port Gardner Bay.

Station Positioning and Navigation

NewFields will ensure that vessel navigation provides accurate station positioning, and that sample locations and water depths are recorded. A differential global positioning system (DGPS) will be used aboard the R/V *Kittiwake* for station positioning. A U.S. Coast Guard differential correction signal will be utilized to obtain a minimum accuracy of ± 3 meters. The DGPS receiver will be placed above the block on the sampling device deployment boom to accurately record the position of the sampling device.

Sampling location target coordinates will be provided in advance and programmed into the R/V *Kittiwake*'s navigation system. Upon sampling device deployment, the actual position will be recorded once the device reaches the seafloor and the deployment cable is in a vertical position. Latitude and longitude station coordinates will be recorded in degrees decimal minutes using the 1983 North American Datum (NAD83). Water depths will be measured using the winch meter wheel and verified by the ship's fathometer. In the event a successful grab cannot be obtained at the target location, additional attempts will be made within a 50-m radius of the target location. If a target location cannot be sampled due to unforeseen conditions (i.e., shoaling, rocky substrate, etc.), a new location that meets the same criteria as the baseline locations will be occupied. Tables 3 and 4 provide the target coordinates for the baseline and secondary sample locations, respectively.

Table 4. Target Coordinates for Phase II Baseline Sampling Locations.

StationID	Easting	Northing	Latitude	Longitude
	(SPN NAD83)	(SPN NAD83)	(NAD83)	(NAD83)
PG-51	1281598.61	365433.97	47.992455	-122.29885
PG-52	1291441.13	367074.39	47.997458	-122.258783
PG-53	1294721.97	363793.55	47.988631	-122.24514
PG-54	1293081.55	360512.71	47.979556	-122.251591
PG-55	1284879.45	363793.55	47.988129	-122.285327
PG-56	1289800.71	353951.03	47.961404	-122.264486
PG-57	1294721.97	355591.45	47.96615	-122.244525
PG-58	1296362.39	362153.13	47.984217	-122.238319
PG-59	1293081.55	353951.03	47.961571	-122.251097
PG-60	1294721.97	357231.87	47.970646	-122.244648
PG-61	1287750.19	363383.44	47.987153	-122.273574
PG-62	1296772.50	358462.18	47.974121	-122.23637

Notes

SPN NAD83: Washington State Plane North, North American Datum 1983

NAD83: North American Datum 1983

Table 5. Target Coordinates for Phase II Secondary Sampling Locations.

StationID	Easting	Northing	Latitude	Longitude
	(SPN NAD83)	(SPN NAD83)	(NAD83)	(NAD83)
PG-63	1289390.61	358462.18	47.973748	-122.266502
PG-64	1285289.56	352720.71	47.957801	-122.282801
PG-65	1291851.24	365023.86	47.991859	-122.256954

Notes

SPN NAD83: Washington State Plane North, North American Datum 1983

NAD83: North American Datum 1983

Sediment Sample Collection

Phase II surface sediment samples will be collected at 15 locations in Port Gardner Bay. Chemical analyses for the twelve baseline locations are summarized in Table 1. Chemical analyses for the three secondary locations are summarized in Table 4. In addition, four archived samples from Phase I were submitted for analysis (Table 2). Table 6 lists the sediment samples to be collected and analyzed, the number of QA/QC samples, sample container, volume, and preservation requirements.

Table 6. Sediment Sample Analysis, Containers, and Holding Times.

Analyses	Grain Size	TOC, TVS	SIM cPAH	Dioxin/Furan &/or PCB Congeners	Metals	Mercury	Archive
Container(s)	16 oz HDPE	8 oz glass	16 oz glass	8 oz amber glass	4 oz glass		16 oz glass ⁴
Preservative	4°C	4°C/-18°C ²	4°C/-18°C	4°C/-18°C	4°C/-18°C	-18°C	-18°C
Holding Time	6 months	14 days ³ / 6 months	14 days/ 1year	14 days/ 1year	14 days/ 1year	28 days	1year
Baseline	12X	12X	12X	12X	12X	12X	12A ⁴
Secondary	3X	3X	3A	3A	3A	3X	3A ⁴
QA/QC Samples							
Duplicates ¹	1X	1X	1X	1X	1X	1X	-
Triplicates ^{1,2}	1X	1X	-	-	-	-	-
Equipment Rinsate ^{1,3}	-	-	1X	-	1X	1X	-
Rinsate Blank ³	-	-	1X	-	1X	1X	-
CRM	-	-	-	1X	-	-	-
Rinsate Totals	-	-	2	-	2	2	-
Sample Totals	17	17	13	13	13	16	15

Notes

X: sample to be collected and submitted for analysis/testing;

A: sample to be archived

-: not applicable

HDPE: high density polyethylene

1. Frequency of analysis is one per 20 samples (5%).

2. Triplicate analysis for sediment conventional parameters only.

3. Equipment rinsate and rinsate blanks conducted for organics and metals only.

4. One 16-oz glass jar to be collected for archive at all sampling locations.

Surface Sediment Grabs

Surface sediment grabs will be collected for chemical analysis from the R/V *Kittiwake* using a stainless steel 0.2 m² dual van Veen (0.1m² per bucket).

Established deployment and recovery procedures for the grab sampling gear, described in PSEP, will be followed to ensure recovery of the best possible samples and minimize risks to personnel and equipment (PSEP 1997a). Once a grab sample is retrieved, the overlying water will be carefully siphoned off one side of the sampler. If the sample is judged to be acceptable according to PSEP specifications, the penetration depth will be measured with a decontaminated stainless steel ruler, and sample quality, color, odor, and texture will be described in the sample log (Appendix B).

If needed, multiple grab samples will be collected and composited for each sampling location to provide sufficient volume for chemical analysis. The general procedure for collecting sediment using a grab sampler is as follows:

1. Make logbook and field form entries as necessary throughout the sampling process to ensure accurate and thorough record-keeping.
2. Position the sampling vessel at the targeted sampling location.
3. Set the sampler jaws in the open position, place the sampler over the edge of the boat, and lower the sampler to the bottom.
4. Record the location using the DGPS; measure and record the water depth.
5. Retrieve the sampler and place it securely in the sampling vessel.
6. Examine the sample for the following sample acceptance criteria:
 - a. The sampler is not overfilled with sample so that the sediment surface is pressing against the top of the sampler.
 - b. The sample does not contain large foreign objects (i.e., trash or debris). A sample that is rock/gravel fill will be rejected in favor of depositional material (i.e., sand/silt/clay).
 - c. Overlying water is present indicating minimal leakage.
 - d. Overlying water is not excessively turbid indicating minimal sample disturbance.
 - e. Sediment surface is relatively flat and/or intact without any indications of disturbance or winnowing.
 - f. A penetration depth has been achieved that allows the collection of the upper 10 cm of sediment. In instances where 10 cm penetration is not possible due to sandy sediments, the maximum penetration depth will be recorded.
 - g. If sample acceptance criteria are not achieved, the sample will be rejected and another sample collection attempt will be made.
 - h. If multiple attempts within 50 m of a given target location do not produce an acceptable sample, the sampling location will be relocated to the nearest pre-

selected random sampling location.

7. Siphon off any overlying surface water.
8. Measure and collect the top 10 cm with a stainless steel spoon, avoiding any sediment that is in contact with the inside surface of the grab sampler, then place the sediment into a stainless steel bowl and cover with aluminum foil.
9. Record the following observations of sediment sample characteristics on the field form (Appendix B); repeat steps 4 through 11 if more sample volume is required.
 - a. Texture
 - b. Color
 - c. Biological organisms or structures (i.e., shells)
 - d. Presence of debris (i.e., natural or anthropogenic objects)
 - e. Presence of oily sheen or obvious contamination
 - f. Odor (e.g., hydrogen sulfide, petroleum)
10. Wash excess sediment back into the water away from any areas remaining to be sampled.
11. Percent fines will be determined by rinsing a known volume of sediment through a 63.5 micron sieve until the water is clear. Percent fines are equal to 1 minus the volume of remaining sediment divided by the initial volume.
12. Once sufficient sediment volume has been collected, samples should be placed in the appropriate, pre-cleaned, labeled sample containers, placed in a cooler maintained at 4°C, and prepared for shipment to the analytical laboratory.
13. Confirm all relevant documentation has been completed, entries are accurate, and paperwork has been signed.
14. Decontaminate all sampling equipment before proceeding to the next sampling location.

A single replicate sample will be collected from each target sampling location, with the exception of field duplicates and QA/QC samples to be collected randomly at the field supervisor’s discretion. Secondary samples collected for archival will be processed in the same manner as the baseline samples. Secondary samples will be submitted to their respective laboratories for archival under the conditions listed in Table 6.

Sample Identification, Containers, and Labels

Samples will be identified based on the project, sampling area, location, and sample type. All samples collected during will be labeled clearly and legibly. Secondary samples will be differentiated from the baseline samples by the Location Number. Each sample will be labeled with a unique alphanumeric sample identification number that identifies characteristics of the sample as follows:

Project	Study Area	Location Number	Sample Type
RB14-	PG-	51-	S

Project consists of four characters describing the project (RB14 = Regional Background 2014).

Study Area consists of two characters describing the sampling area (PG=Port Gardner Bay)

Location Number consists of two characters identifying the station location number

Sample Type consists of one to two characters indicating the sample type. S denotes a sediment sample. Sediment QA/QC samples are further identified with D = duplicate, T = triplicate, ER = equipment rinsate, RB = rinsate blank. For labeling rinsate blanks and equipment rinsates, substitute a six digit date (031214) for the location number.

Sample labels will be self-adhering, waterproof material. An indelible pen will be used to fill out each label. Each sample label will contain the project name (Phase II Port Gardner Bay Regional Background Sediment Characterization), sample identification, date and time of collection, analyses, preservative (as applicable), and the initials of the person preparing the sample. In addition, a unique, sequentially numbered jar tag will be placed on each sample container for tracking purposes. Jar tag numbers will be recorded in a Sample Container Logbook (Appendix B). Sample labels and jar tags will be protected by packaging tape wrapped around the entire jar to prevent loss or damage of the labels during handling and storage.

Sample Storage and Delivery

All samples will be stored in insulated coolers and preserved by cooling to a temperature of 4°C or as required by analytical methods. Maximum sample holding and extraction times will be strictly adhered to by field personnel and the analytical and testing laboratories.

Preparation of jars for shipment will be performed in the following manner:

1. Wipe or decontaminate the outside of filled, capped sample bottles to ensure there is no sample residual on the outside of the container. Secure sample lid jars with electrical tape or placed jar in Ziploc bag to prevent leakage.
2. Label jars with prepared labels. Each set of samples will have a unique sample ID and jar tag number.
3. Secure labels with clear packaging tape.
4. Record the samples in Sample Container Logbook (see Appendix B) and the Chain of Custody forms.
5. Place sample containers in plastic bubble-pack bags, or wrap in bubble pack and secure with packaging tape.
6. Prepare an empty insulated cooler by placing three to four ice packs in a garbage bag at the bottom of the cooler. Place sample containers in a garbage bag and fill with the sample bottles. Add additional bags of ice as needed to surround the bag containing the samples.
7. Seal the cooler with strapping tape and a custody seal. Samples for chemical analyses will be shipped via overnight courier to the analytical laboratory once per day or

whenever a cooler is filled, and accompanied by the chain-of-custody record, which identifies the shipment contents. The chain-of-custody will be signed by the individual relinquishing samples to the onsite laboratory representative. The field personnel will be responsible for:

- a. Packaging the samples;
- b. Signing the chain-of-custody before placing inside the cooler to be sealed;
- c. Applying a shipping label, an air bill, a custody seal, and strapping tape to the cooler; and
- d. Shipping the samples in accordance with the maximum holding time allowed for the analyses to be performed.

A separate chain-of-custody form will be filled out for each analytical laboratory. The chain-of-custody will be signed by the individuals relinquishing the samples and will be placed inside the cooler before it is sealed.

All sediment samples will be retained for a minimum of 6 months from the time they were received using standard laboratory handling procedures. They may be removed from the laboratory prior to the end of the 6-month period only at the direction of the contractor project manager in consultation with Ecology.

Field Documentation

A complete record of field activities will be maintained. Documentation necessary to meet QA objectives for this project include field notes and field forms (Appendix B), sample container labels, and chain-of-custody forms. The field documentation will provide descriptions of all sampling activities, sampling personnel, and weather conditions, and will record all modifications, decisions, and/or corrective actions to the study design and procedures identified in this work plan.

Field Notebooks

All handwritten documentation must be legible and completed in permanent waterproof ink. Corrections must be marked with a single line, dated, and initialed. All documentation, including voided entries, must be maintained within project files.

The Field Manager will keep the field logbook(s) on site during field operations. Daily activities will be recorded in a bound field logbook of water-resistant paper. Separate logbooks consisting of bound, paginated field forms will be kept for surface sediment grab descriptions, and an inventory of sample containers (separate from COPC documentation). Examples of the various field forms to be used are presented in Appendix B. All entries will be made legibly, in indelible ink, and will be signed and dated. Information recorded will include the following:

- Date, time, place, and location of sampling;

- Onsite personnel and visitors;
- Daily safety discussion and any safety issues;
- Quality control samples (e.g., duplicate samples, field blanks, etc.);
- Calibration of field equipment (including make and model of equipment);
- Field measurements and their units;
- Observations about site, location, and samples (i.e., weather, current, odors, appearance); and
- Equipment decontamination verification.

Field logbooks are intended to provide sufficient data and observations to enable participants to reconstruct events that occur during project field activities. Entries should be factual, detailed, and objective. Unless restricted by weather conditions, all original data recorded in field logbooks and on sample identification tags, chain-of-custody records, and field forms will be written in waterproof ink. If an error is made, the individual responsible may make corrections simply by crossing out the error and entering the correct information. The erroneous information should not be obliterated. All corrections must be initialed and dated.

Chain-of-Custody Procedures

The field crew will retain samples at all times until contractor personnel deliver samples to the appropriate laboratory. All samples will be held and transported in coolers with ice or frozen gel-packs at approximately 4°C.

Chain-of-custody forms will be initiated at the time of sample collection to ensure that all collected samples are properly documented and traceable through storage, transport, and analysis. When all line items on the form are completed or when the samples are relinquished, the sample collection custodian will sign and date the form, list the time, and confirm the completeness of all descriptive information contained on the form. Each individual who subsequently assumes responsibility for the sample will sign the chain-of-custody form and provide the reason for assuming custody. The field chain-of-custody terminates when the laboratory receives the samples. The FM should retain a copy of the completed, signed, chain-of-custody form(s) for project files.

Equipment Decontamination

Sample processing equipment (i.e., spoons, bowls, and reusable containers from which samples are transferred to sample jars) will be washed with a laboratory-grade detergent (e.g., Liquinox, Alconox) and water solution, rinsed with site or tap water, and will undergo a final distilled water rinse prior to field operations. Decontaminated equipment will be wrapped or covered with aluminum foil. Sub-sampling and processing equipment will be decontaminated before use at each station in order to prevent cross-contamination of samples. Any deviations from these procedures will be documented in the field notebook.

Personal non-disposable field equipment (i.e., boots, waterproof gloves, and garments) will be rinsed with water and brushed clean prior to leaving the immediate vicinity of the sample collection area. Special attention will be given to removing mud that may adhere to boot treads.

Waste Disposal

During the field investigation, field personnel will be responsible for securing appropriate waste containers, and placing wastes in labeled storage containers, performing appropriate testing, preparing wastes for disposal, and proper disposition of wastes.

Excess sediment sample not submitted to the laboratories, and disposable protective clothing, sampling equipment, and packaging are the two types of waste the activities described in this work plan will generate.

Sediment Samples

Small quantities of excess sediment and rinsate water generated during sample processing will be returned to the site. Care will be taken to not dispose of sediment and/or rinsate at locations targeted for subsequent sampling.

Disposable Protective Clothing and Sampling Equipment

Used PPE, such as protective Tyvek suits or gloves, and sampling equipment, such as aluminum foil and paper towels, and any packaging material that cannot be recycled, will be placed in plastic storage bags and disposed of as municipal waste.

Laboratory Analytical Methods

All of the analytical methods used in this program will be performed in accordance with the PSEP guidelines. The laboratory analysis will be consistent with PSEP guidelines (PSEP 1997a, b, c), any recent modifications proposed during the SMARM, and/or the most current laboratory recommendations. Each laboratory participating in this program will institute internal QA/QC plans. Analyses will be required to conform to accepted standard methods and internal QA/QC checks prior to final approval.

Chemical Analyses

Analytical Resources, Inc., and AXYS Analytical Services will conduct the chemical analysis. Table 6 presents the sample preparation methods, analytical methods, and practical quantitation limits (PQL) for the target conventionals, metals, cPAH compounds, and dioxin/furan congeners. Each sample will also be analyzed for the full list of PCB congeners. The congeners and congener pairs obtainable by EPA method 1668A are presented in Table 7.

The analytical results from this supplemental investigation will be used for the determination of regional background concentrations. Accordingly, the data quality objectives are to obtain as few non-detects and as many unqualified results as possible. The PQLs required for analysis in this study are lower than most standard methods provide. Efforts were made for many of the analytes to find methods that provide lower PQLs.

Few requirements exist for the selection of PQLs. MTCA guidance does stipulate that where the PQL is used as a cleanup level, it must meet the more stringent of the following conditions (WAC 173-340-707(2)(a) and (b)):

- The PQL is no greater than ten times the method detection limit (MDL).
- The PQL is no greater than that established by the U.S. EPA and used to establish requirements in 40 CFR 136, 40 CFS 141-143, or 40 CFR 260-270.

For all target analytes, the PQL is within a factor of ten of the MDL.

cPAHs will be analyzed in select ion monitoring (SIM) mode. Metals will be analyzed by inductively coupled plasma-mass spectrometry (ICP-MS) for sediment samples to achieve lower PQLs for arsenic and cadmium. An additional low end point will be added to the calibration standard for dioxin/furan and PCB congener analysis to provide for lower reporting limits. The PQL values listed for dioxin/furan and PCB congeners in Tables 6 and 7 are lower method calibration limits (LMCL), which is defined by the lower limit of the calibration curve. The LMCL is equivalent to the PQL in that it meets the definition provided in WAC 173-204-505: "PQL means the lowest concentration that can be reliably measured within specified limits of

precision, accuracy, representativeness, completeness, and comparability during routine laboratory operating conditions, using department approved methods”.

The PQLs listed may be subject to modification due to elevated sample concentrations, heterogeneous samples (sediment), and potential matrix interferences that may preclude obtaining the desired quantification limit. Specifically:

- Two possible ions are used for the quantification of arsenic. The specific ion used is dependent upon the matrix and interferences. As a result, two reporting limits are listed for arsenic in Table 6.
- The standard reporting limit for cPAH compounds is 5.0 µg/kg using EPA method 8270 SIM. If necessary, ARI can achieve a PQL as low as 0.5 µg/kg using a low level version of 8270 SIM, but only in samples where the concentration is below the standard reporting limit. ARI will prescreen a solvent shake-out of the sample for potential high concentrations and analyze using the appropriate method.

In the event either laboratory is unable to meet the PQLs additional clean-up measures may be used. If the PQLs still cannot be met, the reasons for the deviation will also be reported.

Table 7. Target Analytes, Methods, and Practical Quantitation Limits.

Analyte	Preparation Method	Analytical Method	PQL
Conventional Parameters			
Grain size	---	PSEP	---
Total organic carbon (%)	---	PSEP	0.10
Total Volatile Solids (%)	---	PSEP	0.1
Metals (mg/kg DW)			
Arsenic	EPA 3050B/3051	EPA 200.8	0.2/0.5†
Cadmium	EPA 3050B/3051	EPA 200.8	0.1
Mercury	EPA 7471A	EPA 7471A	0.025
cPAHs (µg/kg DW)			
Benzo(a)pyrene	EPA 3546	8270-SIM PAH*	5.0
Benz(a)anthracene	EPA 3546	8270-SIM PAH*	5.0
Benzo(b)fluoranthene	EPA 3546	8270-SIM PAH*	5.0
Benzo(k)fluoranthene	EPA 3546	8270-SIM PAH*	5.0
Chrysene	EPA 3546	8270-SIM PAH*	5.0
Dibenz(a,h)anthracene	EPA 3546	8270-SIM PAH*	5.0
Indeno(1,2,3-cd)pyrene	EPA 3546	8270-SIM PAH*	5.0
Dioxin/Furan Congeners (ng/kg DW)‡			
2,3,7,8-TCDD	EPA 1613B/3540C	EPA 1613B (CS-0.2)	0.2
1,2,3,7,8-PeCDD	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
1,2,3,4,7,8-HxCDD	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
1,2,3,6,7,8-HxCDD	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
1,2,3,7,8,9-HxCDD	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
1,2,3,4,6,7,8-HpCDD	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
OCDD	EPA 1613B/3540C	EPA 1613B (CS-0.2)	2.0
2,3,7,8-TCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	0.2
1,2,3,7,8-PeCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
2,3,4,7,8-PeCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
1,2,3,4,7,8-HxCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
1,2,3,6,7,8-HxCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
1,2,3,7,8,9-HxCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
2,3,4,6,7,8-HxCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
1,2,3,4,6,7,8-HpCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
1,2,3,4,7,8,9-HpCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	1.0
OCDF	EPA 1613B/3540C	EPA 1613B (CS-0.2)	2.0

Notes

SIM-select ion monitoring **PQL**-practical quantitation limit **DW**-dry weight

CS-0.2-additional low level calibration point **cPAH**-carcinogenic polycyclic aromatic hydrocarbons

† PQL for arsenic dependent on the quantified ion, which is in turn dependent on the matrix and interferences

* Samples will undergo a prescreening process at ARI and low concentration samples will be analyzed by a low level variant of 8270-SIM with a reporting limit of 0.5 µg/kg.

‡ Values listed for dioxin/furan congeners are the lower method calibration limits (LMCL) defined by the lowest concentration on the calibration curve. The LMCL is functionally equivalent to the PQL.

Table 8. PCB Congener Methods and Practical Quantitation Limits (PQL).

Analyte	Preparation Method	Analytical Method	PQL ng/kg
*PCB-156/157	EPA 1668A	EPA 1668A (CS-0.2)	0.8
see below	EPA 1668A	EPA 1668A (CS-0.2)	0.4
PCB Congeners and Congener Pairs (ng/kg DW) for Prep Method EPA 1668A and PQL of 0.4			
PCB-1	PCB-48	PCB-110/115	PCB-164
PCB-2	PCB-50/53	PCB-111	PCB-165
PCB-3	PCB-52	PCB-112	PCB-167
PCB-4	PCB-54	PCB-113/90/101	PCB-169
PCB-5	PCB-55	PCB-114	PCB-170
PCB-6	PCB-56	PCB-117/116/85	PCB-171/173
PCB-7	PCB-57	PCB-118	PCB-172
PCB-8	PCB-58	PCB-120	PCB-174
PCB-9	PCB-59/62/75	PCB-121	PCB-175
PCB-10	PCB-60	PCB-122	PCB-176
PCB-11	PCB-61/70/74/76	PCB-123	PCB-177
PCB-12/13	PCB-63	PCB-126	PCB-178
PCB-14	PCB-64	PCB-127	PCB-179
PCB-15	PCB-66	PCB-128/166	PCB-180/193
PCB-16	PCB-67	PCB-130	PCB-181
PCB-17	PCB-68	PCB-131	PCB-182
PCB-19	PCB-69/49	PCB-132	PCB-183/185
PCB-21/33	PCB-72	PCB-133	PCB-184
PCB-22	PCB-73	PCB-134/143	PCB-186
PCB-23	PCB-77	PCB-136	PCB-187
PCB-24	PCB-78	PCB-137	PCB-188
PCB-25	PCB-79	PCB-138/163/129/160	PCB-189
PCB-26/29	PCB-80	PCB-139/140	PCB-190
PCB-27	PCB-81	PCB-141	PCB-191
PCB-28/20	PCB-82	PCB-142	PCB-192
PCB-30/18	PCB-83/99	PCB-144	PCB-194
PCB-31	PCB-84	PCB-145	PCB-195
PCB-32	PCB-88/91	PCB-146	PCB-196
PCB-34	PCB-89	PCB-147/149	PCB-197/200
PCB-35	PCB-92	PCB-148	PCB-198/199
PCB-36	PCB-94	PCB-150	PCB-201
PCB-37	PCB-95/100/93/102/98	PCB-151/135/154	PCB-202
PCB-38	PCB-96	PCB-152	PCB-203
PCB-39	PCB-103	PCB-153/168	PCB-204
PCB-41/40/71	PCB-104	PCB-155	PCB-205
PCB-42	PCB-105	PCB-156/157*	PCB-206
PCB-43	PCB-106	PCB-158	PCB-207
PCB-44/47/65	PCB-107/124	PCB-159	PCB-208
PCB-45/51	PCB-108/119/86/97/125/87	PCB-161	PCB-209
PCB-46	PCB-109	PCB-162	--

Notes

PQL-practical quantitation limit **DW**-dry weight **CS-0.2**-additional low level calibration point

Analytical Laboratory Reporting

Analytical laboratory reports will be accompanied by sufficient backup data and QC results to enable independent reviewers to evaluate the quality of the data results. Analytical data will be reported in the units specified by the PQLs listed in Tables 6 and 7.

All PQLs will be met. If matrix interferences exist that prevent meeting the listed PQL, the reason will be listed in the laboratory narrative. All non-detect sample results for cPAH will be reported to the method detection limit and detected results less than the target PQL will be qualified. All non-detect results for metals will be reported at the PQL. Metals data are not qualified below the PQL.

Non-detect results for dioxin/furan and PCB congeners will be reported at the sample specific detection limit (SDL). All detected congener results less than the LMCL/PQL will be qualified.

The analytical laboratory deliverables will include the following:

- Case narrative (including any problems encountered, protocol modifications, and/or corrective actions taken);
- Sample analytical and QA/QC results with units;
- All protocols used during analyses;
- Any protocol deviations from the approved sampling plan;
- Surrogate recovery results;
- MS/MSD results;
- Laboratory duplicate/triplicate results;
- Blank results;
- Sample custody records (including original chain-of-custody forms); and
- Electronic analytical results in Ecology's Environmental Information Management (EIM) format.

Quality Assurance Project Plan

The purpose of the project QA/QC is to provide confidence in the project data results through a system of quality control performance checks with respect to data collection methods, laboratory analysis, data reporting, and appropriate corrective actions to achieve compliance with established performance and data quality criteria. This section presents the QA/QC procedures to ensure that the investigation data results are defensible and usable for their intended purpose.

Measurements of Data Quality

The tolerable limits for the data reported by the laboratory will be measured with regard to precision, accuracy, representativeness, completeness, and comparability.

Precision is a measure of mutual agreement among individual measurements of the same property under prescribed conditions. Precision will be assessed by the analysis of matrix spike/matrix spike duplicates (MS/MSDs), field duplicate and triplicates, and laboratory control sample/laboratory control sample duplicates (LCS/LCSD). The calculated relative percent differences (RPDs) for field duplicates and triplicates and MS/MSD pairs will provide information on the precision of sampling and analytical procedures, and the RPDs for LCS/LCSD pairs will provide information on precision of the analytical procedures.

Accuracy is the degree to which an observed measurement agrees with an accepted reference or true value. Accuracy is a measure of the bias in the system and is expressed as the percent recoveries of spiked analytes in MS/MSD and LCS/LCSD samples. Accuracy will also be evaluated through the surrogate spikes in each sample. The laboratory control limits for surrogates will be used for the project.

Representativeness expresses the degree to which data accurately and precisely represent an actual condition or characteristic at a particular sampling point. Representativeness is achieved by collecting samples representative of the matrix at the time of collection. Representativeness can be evaluated using replicate samples, additional sampling locations, and blanks.

Completeness refers to the amount of measurement data collected relative to that needed to assess the project's technical objectives. It is calculated as the number of valid data points achieved divided by the total number of data points requested by virtue of the study design. For this project, completeness objectives have been established at 95 percent.

Comparability is based on the use of established USEPA-approved methods for the analysis of the selected parameters. The quantification of the analytical parameters is based on published methods, supplemented with well-documented procedures used in the laboratory to ensure reproducibility of the data.

Quality Assurance and Quality Control for Chemistry Sediment Samples

Field and laboratory QA/QC samples will be used to evaluate the data precision, accuracy, representativeness, and comparability of the analytical results.

Field QA/QC for Sediment Chemistry

Field QC samples will be collected during sampling to quantitatively measure and ensure the quality of the sampling effort and the analytical data. Field QC samples include field duplicates, equipment rinsate, and rinsate blanks. Field QA/QC sample counts are presented in Table 5. QC samples are to be handled in the same manner as the environmental samples collected. Brief descriptions of the field QC samples are provided below.

Field Duplicates and Triplicates

Field duplicates and triplicates are collected at the same time as the original sample using identical sampling techniques. Field duplicate sample results (triplicates for sediment conventional parameters) are used to assess the precision of the sample collection process and to help determine the representativeness of the sample. Field duplicates/triplicates will be collected at a five percent frequency. The duplicates/triplicates will be designated for the same analysis as the original samples. The field duplicates/triplicates will be collected from the same homogenate as the original sample.

Equipment Rinsate and Rinsate Blanks

The equipment rinsate blank and decontamination water (rinsate) blank provide a quality control check on the potential for cross contamination by measuring the effectiveness of the sampling and processing decontamination procedures. The equipment rinsate sample consists of de-ionized water rinsed across sample collection and processing equipment after they have been used to collect a sample and have been decontaminated for use at the next sampling location. The decontamination water blank is an unadulterated sample of the de-ionized water used to create the rinsate blank, analyzed to ensure no contaminants were present in the rinse water. Equipment blank samples will not be required when using disposable sample equipment.

Laboratory QA/QC for Sediment Chemistry

One laboratory matrix spike and matrix spike duplicate will be analyzed for every 20 samples submitted or for each analytical batch of samples (if less than 20 submitted) for the analysis of cPAHs, metals, and TOC. The combination of these spiked samples will provide information on the accuracy and precision of the chemical analysis, and to verify that the extraction and measured concentrations are acceptable. The MS/MSDs will be analyzed in accordance with USEPA methods for each respective analyte.

One laboratory replicate will be analyzed for all constituents for every 20 samples submitted or for each analytical batch of samples (if less than 20 submitted). Laboratory triplicates will be analyzed for grain size, TOC, and total solids. These QA/QC samples will be analyzed in accordance with the respective USEPA method and will be used to evaluate the precision of the analytical method.

One laboratory method blank and LCS will be analyzed for all constituents (except grain size and total solids) for each analytical batch of 20 samples to assess potential laboratory contamination and accuracy. An LCSD will be analyzed if required by the method, or if the laboratory does not have enough sample volume to prepare an MS/MSD.

Laboratory control samples, ongoing precision and recovery samples, and surrogate spikes will be used as defined by the analytical methods and equipment calibration requirements.

The Puget Sound Sediment Reference Material Sample will be analyzed for dioxin/furan and PCB congeners. Reference material results will be evaluated in the context of the guidance document available at:

<http://www.nws.usace.army.mil/Missions/CivilWorks/Dredging/SRM.aspx>.

Data Validation

The data generated as part of this investigation will undergo an independent quality assurance review and data validation. A QA2 (USEPA Stage 4) chemistry data review will be conducted that examines the complete analytical process from calculation of instrument and method detection limits, practical quantification limits, final dilution volumes, sample size, and wet-to-dry ratios to quantification of calibration compounds and all analytes detected in blanks and environmental samples (PTI 1989a; PTI 1989b; USEPA 2009).

The independent data validation will be conducted by EcoChem, Inc. of Seattle, WA.

Data Analysis and Reporting

This section describes the data analysis and reporting requirements for the data collection activities described in this supplemental SAP.

Treatment of Sediment Chemistry Data

The sediment chemistry data will be summarized and presented in tables indicating sediment locations, detected contaminants, detection limits that exceed target PQLs, and data qualifiers assigned by the laboratory or during the data validation efforts. Concentrations of relevant COPCs will be mapped to better demonstrate the spatial distributions.

For reporting, dioxin/furan and PCB congeners will be normalized to the toxicity of 2,3,7,8-TCDD (tetrachlorodibenzo-p-dioxin) using toxic equivalent factors (TEFs) updated by the WHO in 2005 (Van den Berg et al. 2006). TEQs for dioxin/furan congeners and PCB congeners will be reported separately. The toxic equivalent quotient (TEQ) is equivalent to the sum of the concentrations of individual congeners multiplied by their TEF (potency relative to 2,3,7,8-TCDD). Non-detected values will be treated following the process described in the draft SCUM II guidance, Chapter 6 (Section 6.1.2, Tables 6-1 through 6-4) and Appendices G and M. This treatment is essentially a Kaplan-Meier estimation of concentrations for samples with non-detected congeners (Ecology 2013b, Helsel 2010).

PCB congeners will also be reported as total PCBs. Total PCBs is the sum of all detected congeners. In a rare case where all PCB congeners are not detected, the highest detection limit will be used to represent the total PCB concentration.

The concentrations for cPAH will be determined by normalizing individual cPAH to the toxicity of benzo(a)pyrene using TEFs present in Ecology's guidance document *Evaluating the Toxicity and Assessing the Carcinogenic Risk of Environmental Mixtures Using Toxicity Equivalency Factors* (Ecology 2007). Non-detected values will be treated following the process described in the draft SCUM II guidance, Chapter 6 (Section 6.1.2, Tables 6-1 through 6-4) and Appendices G and M, i.e., Kaplan-Meier estimation of concentrations for samples with non-detected congeners (Ecology 2013b, Helsel 2010).

Potential Analysis of Secondary Samples

All sediment from the secondary locations will be archived after sample collection. Analysis of these samples will be conducted if a larger sample size is needed to supplement the baseline results. The flow chart in Figure 14 outlines the process followed for determining whether or not to analyze the secondary samples.

The first step is to evaluate the precision of the 95 percent UCL on the mean. A precision value of 25 percent was selected as a guideline. If this target is met, no additional analysis will be needed.

If the calculated precision exceeds the 25 percent target, the baseline data will be compared with natural background as proposed by Ecology (see Ecology 2013b [Chapter 11 and Appendix L which includes the R/B Bold and additional Puget Sound reference area data-sets]). If the regional background 90/90 UTL is less than that of the natural background 90/90 UTL, the additional analysis is unlikely to change regional background. In this case, the regional background values would default to natural background.

If the 90/90 UTL for regional background is less than the study PQLs, analysis of a few additional samples is unlikely to change the distribution enough to increase the regional background value above the PQLs. If criteria 2 and 3 from Figure 14 are not met, the final factor for deciding whether to analyze secondary samples is to determine whether a larger sample size would improve confidence in the upper tail of the distribution (i.e., the 90/90 UTL). If this is the case, then the number of secondary samples to be analyzed will be determined by either budgetary constraints or an estimate of the number of samples needed to bring the precision of the mean below 25 percent.

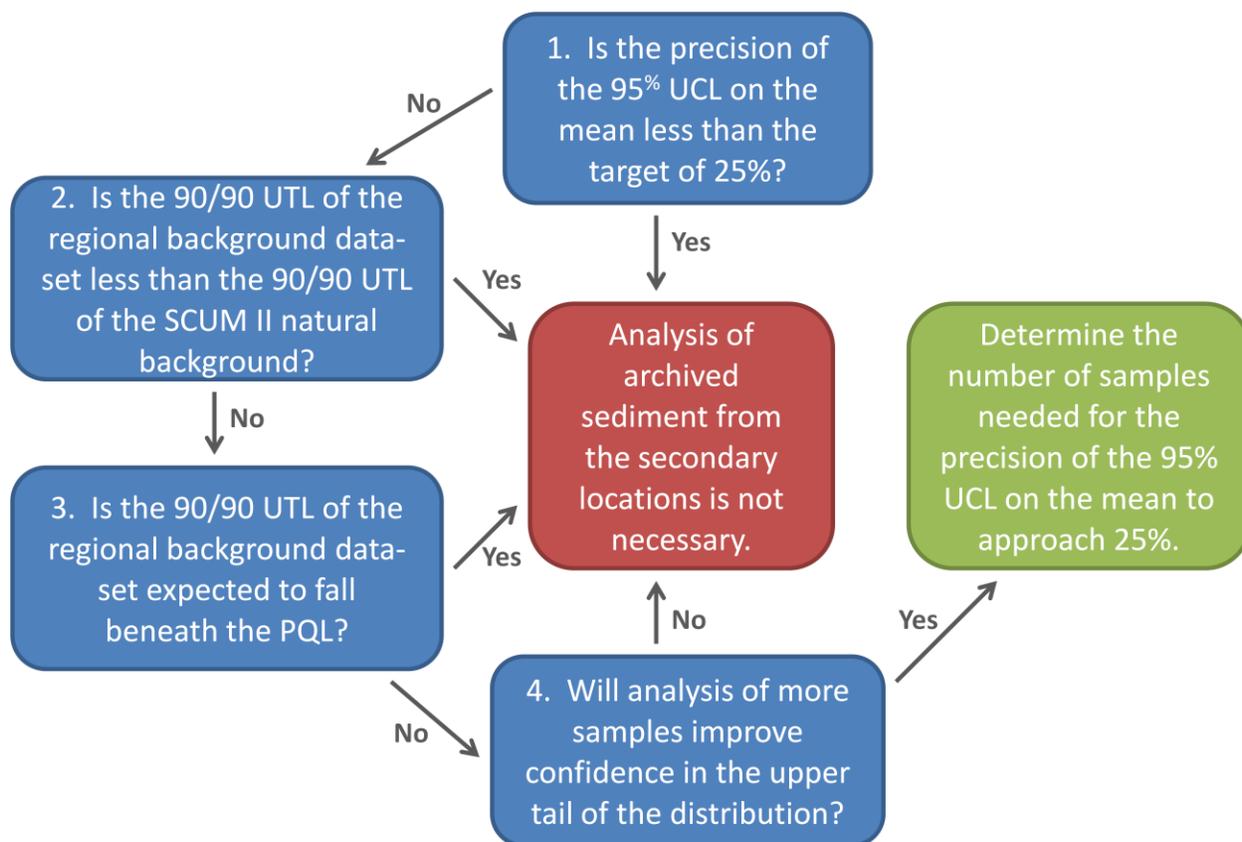


Figure 14. Process for Evaluating Secondary Samples for Analysis.

Statistical Analysis and Calculation of the Regional Background Concentration

The statistical metric Ecology will use to calculate regional background for each COPC is the 90 percent upper confidence bound on the 90th percentile (90/90 UTL). This value should be calculated on a single data distribution, excluding outliers.

The distribution of the data will be evaluated for the presence of potential outliers, or for a mixture of distinct populations using graphical tools (e.g., boxplots and Q-Q plots) in conjunction with formal outlier tests (e.g., Dixon's or Rosner's test) for identifying samples with extreme concentrations for a single analyte or TEQ. Some samples may not have extreme concentrations for individual analytes, but may still exhibit very different multivariate patterns within the suite of PCBs, cPAHs, or dioxin/furan congeners. The presence of samples with very different congener patterns may signify unique contaminant sources.

Multivariate outlier investigations will use Mahalanobis distances. Mahalanobis distance is a metric very similar to Euclidean distance (i.e., the familiar metric used to calculate the distance between two points on a line). To identify multivariate outliers, the Mahalanobis distance is calculated as the distance between each observation and the center of mass for the remaining observations, scaled to the covariance among congeners in the direction of that observation. A large distance in a direction of high covariance is more likely than a moderate distance in a direction of very low covariance. An observation that doesn't fall within the "cloud" of other data points is identified as a potential multivariate outlier.

Outliers will contaminate a distribution, producing higher variance estimates and subsequently greater uncertainty in the tolerance limit. Any outliers identified will be discussed with the project team, and the regional background will be calculated both with and without the outlier(s) to determine their impact. Any concentrations deemed likely to be from outside of regional background will be excluded from the calculations.

After the removal of any outliers, goodness-of-fit tests and graphical displays (e.g., Q-Q plots) will be used to identify the best-fit distributional form for the data. For the calculation of the 90/90 UTLs, if the assumption of a particular parametric distribution is not deemed appropriate for the data, non-parametric methods such as the bootstrap or order statistics will be used as appropriate. In all cases where concentrations are present below detection, methods appropriate for left-censored data will be used. All of the statistical tools required are available in ProUCL 5.0.00 (US EPA 2013), with the exception of the multivariate pattern analysis. The multivariate analysis tools are available in Scout version 1.00.01 (US EPA 2008, runs only up to Windows 98), and R (R Development Core Team 2013); as well as in other commercially available statistical software.

Reporting Procedures

A written data report documenting all activities associated with collection, transportation, and chemical analyses of sediment samples will be prepared. The chemical and QA/QC reports will be included as appendices. At minimum, the Final Report will include:

- A summary of the purpose of the investigation;
- Description of sampling and analysis activities;
- Protocols used during sampling and testing, and an explanation of any deviations from the sampling plan protocols or the approved work plan;
- Methods used for station positioning, sample collection locations reported in latitude and longitude to the nearest tenth of a second (NAD83);
- Maps showing actual locations of sampling stations;
- Maps and data tables of sediment chemistry;
- Chain-of-custody records;
- Analytical laboratory reports;
- Copies of field and sampling logs as appendices;
- QA/QC summary;
- Data validation reports;
- Data analysis and interpretation for determination of background concentrations; and
- Summary statistics, outliers, and uncertainties associated with calculation of regional background.

A summary of the data, background calculations, and statistical results will be presented to the stakeholders prior to finalization of the report.

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Appendices

Appendix A. Existing Chemistry Data Summary Figures

Historical Data Summary

Ecology provided NewFields downloads of all historical data in Port Gardner Bay from the EIM database. Results from the Phase I regional background characterization has not been uploaded into EIM, but were combined with the historical data set. While all sample depths were included, surface sediment samples from 0 to 10 cm comprised nearly half of all results.

Nearly 600 historical samples were present. However, not all COPCs were analyzed in each sample. Table A-1 summarizes the sample counts by year and by COPC. The level of QA/QC on the studies was variable, though most were thought to be above a level QA1 data validation. There was no consistent reporting for non-detect results. Some results were reported at the MDL, while some were at the PQL.

TEQ values were calculated for cPAH compounds and dioxin/furan congeners using Ecology guidance (Ecology 2007). A concentration of one half the reported MDL/PQL was used for non-detected results. A separate TEQ was calculated using the same process, but only incorporated the non-detected results.

All data interpolations were conducted in ArcGIS.

Table A-1. Summary of Available Historical Data Present in Port Gardner Bay.

Year‡	TOC	Fines	Arsenic*	Cadmium	Mercury	Dx/F TEQ†	PAH TEQ†	PCB TEQ†	Total PCBs
1988	35	34	26	26	26		24		
1989		7							
1990									
1991									
1992	15	29	13	13	13	4	9		
1993	5	5	1	5	5		5		
1994	6	6	6	6	6		6		
1995	38	37	1	38	38		37		
1996	4			4	4		4		
1997									
1998	1	1	1	1	1		1		
2000	3	3	3	3	3		2		
2001									
2002	1	1	1	1	1		1		
2003	7		7	7	7		7		
2004	17	1	10	17	17	3	17		
2005	13	13	13	13	13		8		
2006	16	10	3	4	4	12	3		
2007	27	15	20	26	24		27		
2008	36	29	20	32	32	8	15		3
2009	105	105	55	69	77	19	56		
2010	45	29	17	18	18	22	15	6	9
2011	24	24	5	6	6	20	4		
2012	93	40	43	49	49	61	49	18	21
2013	57	37	6	15	15	50	15	14	15
Historical Total	548	426	251	353	359	199	305	38	48
Phase I RB Samples	35	35	25	25	50	25	25	35	35

Notes

‡if multiple samples were collected at one location, only the most recent results were retained

*excludes all non-detect values greater than 5 mg/kg

†excluded all results where undetected congeners comprised >25% of the total TEQ

Appendix B. Field Forms

Appendix C. Health and Safety Plan

1.0 Introduction

This Site-Specific Health & Safety Plan (HASP) has been developed as part of the Port Gardner Bay Regional Background Sediment Characterization. This plan is intended to incorporate sampling activities in support of the bay-wide sediment collection effort, and should be revised if the scope of work is changed.

The procedures and protocols in this plan have been established to ensure that a mechanism is in place to address project personnel in the event that hazards from field work or site contamination are encountered during the project. This plan addresses typical on-site activities such as collection of contaminated sediment samples and marine vessel use. This HASP is not designed to replace existing procedures or to address all health and safety procedures that could be required during typical emergency response activities.

Compliance with this HASP is required from all authorized NewFields project personnel, project support personnel, and visitors who enter the work areas of this project. No field work will be conducted without meeting the requirements of this HASP.

The content of this HASP may change or undergo revision based upon unexpected field conditions, modifications to the technical scope of work or additional information made available to health and safety (H&S) personnel. Any proposed changes must also be reviewed and approved by designated NewFields personnel.

1.1 Project Location

Project field work will be conducted in Port Gardner Bay, WA. All sediment sampling activities will be conducted aboard the R/V *Kittiwake*, embarking and disembarking from the Everett marina area.

1.2 Personnel and Emergency Contact Information

Table 1 lists relevant project personnel and local emergency contact information. Additional detailed emergency information is found in Section 6.0 along with written hospital directions and accompanying maps.

All project personnel, project support personnel, and visitors present during field work must sign in the space provided in Table 1 prior to initiating project work. A signature below indicates commitment to implement this plan and to ensure that project fieldwork is conducted safely. A signature below also indicates review and approval of the plan and agreement that the anticipated hazards are correct and that planned hazard controls are sufficient.

2.0 Health and Safety Personnel

The following briefly describes the health and safety designations and general responsibilities for this project.

2.1 Project Manager – NewFields

The Project Manager or designee has overall executive responsibility for all activities and personnel on the site during all project activities described in this HASP.

2.2 On-Site Health and Safety Officer

The HSO is responsible for the development of safety protocols and procedures, pursuant to the all hazardous aspects of this project, implementation and enforcement of this HASP. The HSO has the authority to modify this HASP based on actual site working conditions and procedures. The HSO will also be responsible for the resolution of any outstanding health and safety issues which arise during the conduct of site work.

Health and safety-related duties and responsibilities will be assigned only to qualified individuals by the HSO. The HSO has stop-work authorization, which will be executed upon determination of an imminent safety hazard, emergency situation, or other potentially dangerous situation, such as extreme weather conditions. An Authorization to Proceed with work will be issued by the HSO after such action. The HSO or designee will initiate and execute contact with support facilities and personnel when this action is appropriate. The HSO may periodically conduct QA/QC surveys of the health and safety procedures implemented onsite.

3.0 Site and Project Description

Port Gardner Bay is an approximately 18 square mile coastal embayment located within Possession Sound, an extension of Puget Sound near Everett, WA. The bay is protected from oceanic swells entering from the Strait of Juan de Fuca, while limited fetch and terrain restricts the growth of local wind waves to less than 3 feet for all but gale conditions. Local weather patterns are generally characterized by daily sea breeze formation throughout the summer months, which transition to southerly wind and the passage of frontal systems from late-fall through spring. Water depths within the bay range from intertidal near the mouth of the Snohomish River, to ~600 feet along the SW extent of the embayment. Circulation within the bay is generally weak and variable, with mixed semi-diurnal tides having a mean diurnal range of 11 feet. The Snohomish River delta is a dominant feature within the bay, acting as a large source of sand and suspended sediment north of the Port of Everett. Distributary channels and shoaling areas of the delta may represent navigational hazards, but are not being targeted for this study.

The current investigation will involve sediment chemistry collection throughout the Bay to determine regional background concentrations of the analytes of concern.

3.1 Scope of Work

Under direction of the Washington Department of Ecology (Ecology), NewFields will conduct a sediment chemistry evaluation at several locations throughout the bay. The objective of the study is to evaluate the Regional Background sediment concentrations of selected bioaccumulative contaminants for Port Gardner Bay. Sediment sampling is proposed at 45 locations in Port Gardner Bay. A more detailed description of the scope of work and maps of the sample locations can be found in the Sampling and Analysis Plan (SAP).

Surface sediment samples will be collected for chemical analyses using a dual van Veen grab sampler deployed from the R/V *Kittiwake*, operated by Charles Eaton of Bio-marine enterprises. Hazards associated with grab sampling are primarily physical in nature. Slipping/tripping hazards are present on the sampling vessel when the deck is wet. There are numerous pinch points on the sampling equipment as well as the vessel itself. All personnel will be trained in the operation and deployment of the field gear, and will receive a vessel-specific safety briefing from Charles Eaton, owner and operator of the R/V *Kittiwake*. All members of the sampling crew will wear slip-resistant boots, safety glasses, nitrile gloves, personal flotation devices (PFDs).

Sediment samples will be processed on deck once the vessel is repositioned to minimize pitch and roll. The sediment will be sub sampled for chemistry analysis using a stainless steel spoon and pre-cleaned laboratory jars. All sampling equipment will be cleaned according to decontamination procedures outlined in the SAP. Sampling and decontamination includes potential contact with bottom sediments and decontamination chemicals.

4.0 Hazard Assessment

This section summarizes hazards that may exist during project related tasks.

4.1 Task Specific Hazard Assessment

For the field sampling tasks described in Section 3, the overall hazard level is low. Hazards encountered during this sampling program are due to physical safety hazards associated with the field operations. Types of potential hazards associated with the field sampling effort are summarized in Table 2. Potential hazards while working at the site include, but are not limited to, the following:

- Physical hazards from use of sampling equipment and operations on a vessel and beach areas
- Physical hazards from working conditions (e.g., slips/trips/falls, drowning, hypothermia).
- Physical hazards from operating a motor vehicle to transit to and from the work site.

As described below, protective equipment and safe working procedures will help prevent accidents caused by these hazards. Exposure to harmful microbial organisms or other organisms in the sediments is not expected during this program.

Table 2. Sediment Sampling – Types of Potential Hazards

Physical Hazards			
Name of Physical Hazard	Source	Exposure Level/ Potential	Exposure Limit
Boating Operations	boat deck	Likely	N/A
Heat (ambient)	sun	Unlikely	N/A
Cold Weather Operations	boat deck area	Likely	N/A
Heavy Manual Lifting/Moving	van Veen grab	Likely	N/A
Slips/Trips/Falls	boat deck area	Likely	N/A
Inclement Weather – Snow, rain	boat deck area	Likely	N/A
Material Handling	sediment	Likely	N/A
Vehicular Travel	van shuttle	Likely	N/A
Working Over Water	boat deck area	Likely	N/A
Biological Hazards			
Name of Biological Hazard	Source	Exposure Level/ Potential	Exposure Limit
Insect bites and stings	Boat launch area	Unlikely	N/A
Control Measures Used			
Engineering Controls:			
Level of PPE: D			
Location: on boat deck, stream/intertidal	PPE Equipment: Chemical-resistant steel toe boots or waders, PVC Bib-style overalls (and jacket with hood as necessary), splash-proof safety goggles, nitrile gloves, PFD Type III. Cold-weather clothing and insulated gloves are recommended.		
Work Practices:	Frequent changes of disposable nitrile gloves Wash hands and face with soap and water after each sampling event Take shower at end of workday		

NA = Not applicable.

4.2 Physical Hazards

The following is a general discussion of the hazards that may be encountered on site. Information on any contaminants encountered during this project may be found in standard health and safety references, such as the NIOSH "Pocket Guide to Chemical Hazards." Internet site:

<http://www.cdc.gov/niosh/npg/npg.html>

4.2.1 Sampling Vessel Operations

The physical hazards associated with the deployment and retrieval of sampling equipment result from their weight and the method of deployment. Only appropriate personnel whose presence is required will be deploying and retrieving sampling gear. Under circumstances of potentially dangerous waves or winds, the sampling leader will employ best professional judgment to ensure safe field operations.

To avoid injuries from slipping on wet surfaces, rubber boots or waders with appropriate tread will be worn when working on the work deck or loading/unloading heavy equipment from the vessel. No overhead gear will be deployed. However, hard hats will be worn if overhead hazards exist. Sample handling equipment, containers, and deck lines, not in immediate use will be kept clear of walkways and work areas until needed. Each time operations at a given location have been completed, excess sediment on the deck will be washed overboard to prevent slipping, minimize personnel exposure to potentially contaminated sediment, and limit cross-contamination between sample locations.

Life vests will be provided for and worn by all personnel working on the deck, or as directed by the Site Safety Officer or vessel operator.

In a man overboard situation, sampling personnel will immediately notify the captain, while a designated person maintains constant visual contact with the victim. A life ring will be available to throw to the victim, while other buoyant materials may be thrown into the water to assist with the return and recovery of the victim. The captain will inform the sampling crew of vessel-specific hazards and safety procedures on the first day of sampling operations.

4.2.2 Motor Vehicle Operation

Motor vehicles will be used to transport field personnel, equipment, and supplies to the sampling sites or laboratories. Only sampling team personnel with valid driver's licenses and liability insurance (per local state laws) will operate motor vehicles required for work activities. All field staff will use best professional judgment at all times to ensure safe operation of motor vehicles, including:

- Operators are to practice defensive driving and drive in a courteous manner
- Be aware of pedestrians and give them the right-of-way
- All vehicles are to be operated in a safe manner and in compliance with statutory traffic regulations and ordinances

- Verifying safety seat belts are in proper operating order
- Seat belts are to be worn by the driver and all passengers whenever the vehicle is in motion
- No persons are allowed to ride in the back of any vehicles, unless equipped with seatbelts
- Vehicles are to be driven in conformance with local speed limits
- Avoid excessively long driving periods
- Personnel who are impaired by fatigue, illness, alcohol, illegal or prescription drugs, or who are otherwise physically unfit, are not allowed to drive
- Personnel are to avoid using cellular phones or engaging in other distractions while driving
- Motor vehicle accidents are to be reported to the responsible law enforcement agency, the NewFields manager, and the NewFields HSO.

4.2.3 Weather

If severe weather occurs that may affect the safety of site workers, the NewFields PM or their designee shall stop affected field operations. The PM or their designee will resume operations when weather conditions improve to acceptable levels.

4.2.4 Heat and Cold Stress

Depending on the time of year and weather conditions, cold or heat stress may be a potential problem. The PM will ensure that the heat and cold stress programs are implemented and that adequate rest breaks and liquid (i.e., water, Gatorade) consumption occur.

Proposed work/rest schedules will be dependent upon the weather conditions encountered and the level of personal protective equipment being utilized by on-site personnel. The PM or designee will establish work/rest schedules prior to the commencement of the project tasks and will adjust as needed.

4.2.5 Illumination

If work activities occur before sunrise and/or after sunset, lighting will be provided at each work area to meet the requirements of 29 CFR 1910.120(m). The Standard states that while any work is in progress, the general site areas shall be lighted to not less than 5 foot-candles; excavation, waste areas, access ways, active storage areas, loading platforms, and field maintenance areas shall be lighted to not less than 3 foot-candles; and first aid stations not less than 30 foot-candle.

4.2.6 Slip, Trip and Fall Hazards

As in any work area, it is expected that the ground may be uneven, the surface may be unreliable due to surface evenness, debris may be present, work is being performed on poly sheeting, and wet or muddy areas may exist. Therefore, the potential for slipping, tripping, and falling is present, especially considering that encapsulating suits and respiratory protection will which can impede vision. Severe trip hazards will be identified prior to commencement of project activities and demarcated by flags or caution tape.

4.2.7 Manual Lifting

Manual lifting of heavy objects such as coolers with samples may be required. Failure to follow proper lifting technique can result in back injuries and strains. Site workers will be instructed to use power equipment to lift heavy loads whenever possible and to evaluate loads before trying to lift them (i.e. they should be able to easily tip the load and then return it to its original position). Carrying heavy loads with a buddy and proper lifting techniques, 1) make sure footing is solid, 2) make back straight with no curving or slouching, 3) center body over feet, 4) grasp the object firmly and as close to your body as possible, 5) lift with legs, and 6) turn with your feet, don't twist, will be stressed. Back injuries are a serious concern as they are the most common workplace injury, often resulting in lost or restricted work time, and long treatment and recovery periods. In addition, hand digging for pipes may present lifting/ergonomic hazards.

4.2.8 Other Physical Hazards

Incorporating the following basic safety procedures can prevent many of the most common causes of injury or accident during field sampling:

- Implement good housekeeping practices, including immediate cleanup of spills and safe storage of all materials. All equipment or materials not in immediate use will be removed from the immediate work area.
- Use proper lifting and moving techniques to prevent back or muscle strain or injury. Any heavy equipment, boxes, coolers etc. should be tested before lifting and if it is too heavy, the equipment should be broken into smaller components or assistance requested. Lifting should be done with the legs, not the back.
- Use extra caution when handling sharp tools or sampling devices and when possible, wear protective gloves.

4.2.9 Biological Hazards

The project location and timing of proposed fieldwork is such that risks from biological hazards are low.

5.0 Work Clothing and Levels of Personnel Protection

The PM or designee will recommend appropriate levels of protective clothing to be worn in the event that hazardous materials are encountered. The sediment and water field sampling activities described in this site-specific HASP will be performed in Level D or modified Level D PPE, as specified in Table 2. If site conditions include hazards that exceed the protection of Level D or modified Level D PPE, work will be halted and personnel will immediately exit the area while site conditions and PPE levels are re-evaluated by the Site Supervisor and HSO.

5.1 Definition of Levels of Protection:

- Level D: Work coveralls
 Gloves
 Appropriate work boots
 Hardhat (if overhead gear is present)
 Safety glasses with side shields or splash goggles as needed
 A respirator is not required.
- Level C: Chemical-resistant disposable coveralls
 Chemical-resistant outer gloves
 Chemical-resistant inner gloves
 Appropriate leather work boots with chemically resistant outer boots or
 chemically resistant rubber boots
 Hardhat
 Full or Half face, Air Purifying Respirator (APR) with combination HEPA -
 P,O,N 100 (dusts, fumes, aerosols) and chemical cartridge as appropriate for
 hazard.
- Level B: Chemical-resistant disposable coveralls
 Chemical-resistant outer gloves
 Chemical-resistant inner gloves
 Appropriate leather work boots with chemically resistant outer boots or
 chemically resistant rubber boots
 Hardhat
 Supplied air - air line or self-contained breathing apparatus (SCBA).
- Level A: Fully encapsulating chemical-resistant/gas tight suit
 Attached chemical-resistant outer gloves
 Chemical-resistant inner gloves
 Attached chemical-resistant boots.
 Self-contained breathing apparatus.

5.2 Donning and Doffing

Manufacturers procedures for donning and removing PPE ensembles will be followed in order to prevent damage to PPE, reduce and eliminate migration from the work area and a transfer of contaminants to the wearer's body or others.

5.2 Storage and Inspection

Protective equipment will be stored and maintained in the company vehicles on site or in the work trailer. Items such as gloves, protective suits, and hearing protection will be kept within a suitable storage area. Table 3 lists PPE storage and cleaning procedures.

Employees are responsible for inspecting personal protective equipment prior to donning, during use and at the end of the shift. Defective equipment shall be removed from service and reported to the PM. All reusable equipment will be maintained in a sanitary condition, in accordance with the manufacturer's recommendations.

Table 3. Level D Storage and Cleaning Procedures.

<p><u>Level D Storage Procedures:</u></p> <p>In the Field laboratory, decontamination solutions such as nitric acid, methanol and acetone will be stored in dedicated cabinets and the outside doors labeled with flammable and acid labels respectively. Alconox soap powder does not require special storage and will be placed on a shelf. Any plastic containers containing Alconox will be labeled as such.</p>
<p><u>Level D Cleaning Procedures:</u></p> <p>Cleaning procedures for PPE require that hard hats, nitrile gloves, rain gear, boots, and personal floatation devices be brushed thoroughly with a solution of Alconox and rinsed with tap water after each sampling event.</p>

6.0 Emergency Plan

Emergency situations can be characterized as an accident or injury to the field personnel. Emergency phone numbers are listed in Section 1 of this Health and Safety Plan. In case of emergency, it is important that the following Incident Reporting Procedure be observed:

It is important to assure the rapid and accurate transfer of information appropriate personnel in the event of an emergency situation. To simplify the procedure, emergency situations can be reported by dialing 911. This includes incidents requiring police assistance, fire department, or medical emergencies.

Be sure to provide the following information to the dispatcher:

1. Caller full name
2. The nature of the incident (i.e. "Fire")
3. The location of the incident (i.e., "Street location and nearest intersection"), the more specific the better.
4. What you need (i.e. "Fire Department and First Aid")
5. If you are able, where you will meet emergency responders (i.e. At end of West Street, near train tracks)
6. If applicable, a call back number or your cell phone number (e.g., "I'll be at the scene; my cell phone number is 123-4567").
7. Status of the situation. (e.g., is the situation stabilized or "I have the fire under control")
8. If anyone is injured or in need of emergency assistance (e.g., "A mechanic working on a pump was burned.")

6.1 Site Emergency Coordinator

Site Emergency Coordinator: Preston Martin (HSO)

6.2 Personnel Injury

In the event of an emergency situation, the local emergency response group will be called. In case of a life-threatening situation, emergency first aid may be applied on-site as deemed necessary. The individual should be cleaned up and/or decontaminated and then transported to the nearest medical facility if needed.

The local rescue squad shall be contacted for transport as necessary in an emergency. Since some situations may require transport of an injured party by other means, transportation by automobile may be required.

6.3 Personnel Exposure Treatment

SKIN CONTACT: Use copious amounts of soap and water. Wash and/or rinse affected area thoroughly, then provide appropriate medical attention. Eyes should be thoroughly rinsed with water for at least 15 minutes.

INHALATION: Move to fresh air and, if necessary, decon/transport to hospital.

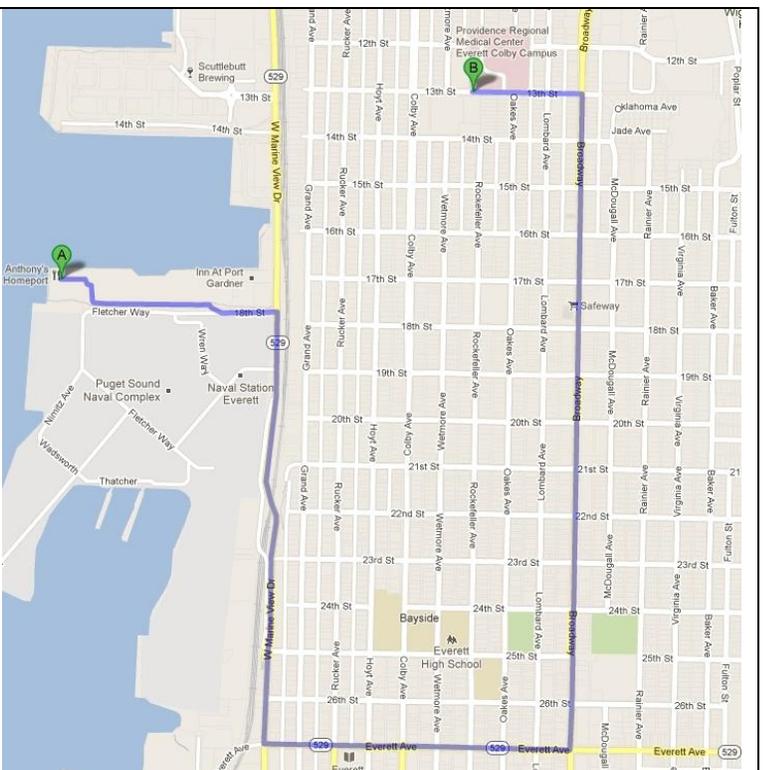
INGESTION: Decontaminate and transport to emergency medical facility.

PUNCTURE WOUND OR LACERATION: Decontaminate, if possible, and transport to emergency medical facility.

6.4 Hospitals

The following map and table show the location and driving directions to the nearest hospital.

Table 4. Nearest Hospital to the Port Gardner Bay Moorage Site.

<p>Providence Regional Medical Center 1726 W Marine View Dr. Everett, WA 98201</p>	
<p>General: (425) 261-2000</p>	
<p>Emergency: (425) 261-3024</p> <ol style="list-style-type: none">1) From Anthony's Homeport parking lot, head east on 18th St.2) Turn right onto W Marine View Dr.3) Take the 3rd left onto Everett Ave.4) Turn left onto Broadway.5) Turn left at WA-303S/NW Waaga Way.6) Turn left onto 13th St.7) The Cymbaluk Medical Tower will be on the left.	

Note: For non-emergency treatment, an urgent care walk-in clinic is located nearby:

The Everett Clinic (Gunderson Building)
3927 Rucker Ave.
Everett, WA 98201

Appendix D. Background Memorandum

Characterizing Natural or Regional Background Populations for Washington Department of Ecology – Final Draft – February 18, 2013

1.0 Introduction

Washington Department of Ecology (Ecology) contracted with TerraStat Consulting Group to help identify statistical tools that could be used to distinguish between populations of Natural Background and Regional Background, as well as how to best summarize available data for background for the different objectives required under Sediment Cleanup programs. Statistical tools that can be used to separate a dataset into distinct populations with overlapping concentrations generally rely on either *a priori* description of the underlying characteristics for one of the populations so that separation of a mixture distribution can be accomplished; or data sets with sufficient spatial coverage such that the concentration surfaces can be modeled and locations where changes in the concentrations occur can be identified. Neither of these scenarios is met by the data that are generally available for Puget Sound background sites, so this work product ultimately became the description of a process for describing and generating appropriate summary statistics from existing Regional Background and/or Local Natural Background data sets.

TerraStat reviewed existing approaches to the description of Background that have been used by various agencies; these definitions and approaches are briefly touched on in Section 2. Knowledge from Ecology staff, from other agency approaches, and about the statistical tools appropriate for the type of data available for Puget Sound Background areas was integrated to describe a process for identifying and characterizing Background. A description of the process and important considerations to this process are included in Section 3; applications to three Puget Sound case studies are included in Section 4. Final recommendations regarding the description of Background are provided in Section 5.

2.0 Approaches to Background

TerraStat reviewed approaches to Background as described in USEPA guidance (US EPA 1995, US EPA 2002), and as used by WDOE for several local sites (summarized by NewFields 2011), by OR DEQ for the lower Willamette River, and by the ACOE for DMMP sites (DMMP 2011). These approaches all started with the *a priori* selection of the geographic boundaries that constitute background and then proceeded to summarize the background data set in different ways for different purposes.

Some working definitions of background in use by EPA in the CERCLA (Superfund) program (US EPA 2002) are “substances or locations that are not influenced by the releases from a site, and are usually described as naturally occurring or anthropogenic:

- 1) Naturally occurring - substances present in the environment in forms that have not been influenced by human activity [*matches Ecology’s definition for ‘Natural Background’*]; and,
- 2) Anthropogenic - natural and human-made substances present in the environment as a result of human activities (but not specifically related to the CERCLA site in question)” [*matches Ecology’s definition of ‘Regional Background’*].

In the EPA documents reviewed, the background data set is not extracted and statistically separated from a larger dataset, but rather it is acquired via appropriate sampling within the boundaries of the area expected to meet the background definition. EPA emphasizes that background sampling is a strategic, biased sampling event with sampling to occur “in areas expected to be outside the area influenced by the site.” So information about the fate and transport of contamination from the site, as well as best professional judgment (BPJ) must be used to identify the background areas. It is also recommended that possible concentration outliers should be identified and decisions regarding these outliers be made by the project team (US EPA 2010).

The reviewed approaches used by Washington DOE, Oregon DEQ, and Puget Sound DMMP all define the geographic boundaries for the background population using BPJ and knowledge about the site, which is consistent with EPA’s recommendations. Once the background population has been identified, the summary statistics suited to the intended application may be calculated (e.g., central tendency value such as 95UCL on the mean for exposure related questions, or an upper threshold of the distribution such as 90/90 UTL for a not-to-exceed value for station-by-station comparisons).

3.0 Background Considerations & Recommendations

As with any sampling or analysis plan, it is important to clearly state the objective up front. Sampling Objective: To characterize a “Background” population in order to set clean-up goals for a site, and to delineate site clean-up boundaries. Caveats:

- Current DOE guidance utilizes several definitions of “Background”, including Puget Sound-wide Natural Background, Local Natural Background, and Regional Background. The method(s) proposed herein will not separate the available data into these separate populations; but they will allow the proper estimation of summary statistics from whatever boundaries are considered to constitute the relevant background population.
- We provide no statistical recommendations for how to separate Local Natural from Regional Background –pattern matching requires knowledge of the source pattern; and

existing sampling locations may be insufficient to adequately describe the spatial patterns needed to statistically identify boundaries of the different types of background.

- Identifying the site-influenced areas needs to be done first with site knowledge combined with spatial contouring. Areas near known point sources, or areas adjacent to those point sources with elevated concentrations, are excluded *a priori* from the possible background population. If there are obvious trends away from the site, then a boundary can be determined with the help of contour mapping.
- Methods to identify univariate and multivariate outliers should be used, and samples that are clearly different from the others should probably be excluded from the background population.

3.1 Designing a Sampling Plan to Characterize Background

The ideal situation where a new survey is designed specifically to characterize background involves first determining an appropriate boundary for the Local Natural or Regional Background area, and then taking a spatially-balanced random sample that uniformly covers the area using a systematic random sample (e.g., random samples within a grid). Other methods for achieving a spatially-balanced random sample can also be used (e.g., generalized random tessellation stratified [GRTS] sampling used by US EPA [2011], and implemented via the RRQR algorithm in GIS), but are not generally necessary for this situation. If there are areas that are found to have elevated concentrations and are suspected to be due to a point source, the boundary of the regional background area can be adjusted after sampling (i.e., the areas with high concentrations and their corresponding data can be excluded). If the gridded samples are found to be spatially autocorrelated, the gridded data points can be sub sampled or methods that account for autocorrelation can be used to generate summary statistics. Otherwise, the set of samples taken on the grid can be assumed to be an independent random sample, and the appropriate summary statistics can be easily generated in ProUCL (for example).

3.2 Working With Existing Data to Characterize Background

The following recommendations address situations when compilations of existing data sets are being used to characterize background. In these cases, the “ideal” design may not have been used to generate the background data set, and the background area may have been sampled non-uniformly and non-randomly.

For the case studies examined in this report, sampling points were located unevenly through space and time, and collected for different purposes. In these case studies, the boundary of the characterized area is delineated by the locations of the existing samples and these existing sampling locations may be insufficient to fully characterize the background population.

Within the compilation of data available to use for background, there may be non-randomness, non-independence, and more than one population represented. The simplest statistics (e.g., 95th UCL on the mean) assume an independently and identically distributed (i.i.d.) random sample. If the dataset violates these basic assumptions, then the simple statistics may be biased, and the variance poorly estimated. These basic assumptions are relevant to a number of available methods for summarizing data (i.e., bootstrapping, generating a trend surface via interpolation, kriging, etc.).

Given a data set consisting of multiple studies, there are three main steps to be followed to define a background concentration distribution and produce unbiased estimates of summary statistics. These are:

- Step 1. Delineate Background by excluding areas near known point sources, and areas suspected to be of a different population based on proximity to local influences (e.g., developed shorelines). Initial evaluation of the compiled data should include identification of possible outliers.
- Step 2. Determine the extent of autocorrelation and/or trend in samples from the background area.
- Step 3. Generate upper bound estimates for the regional background concentration distribution using an independent subset of the data based on the results from Steps 1 and 2.

3.2.1 Step 1. Drawing Background Boundaries

Unless new data are collected, the background population will be defined by the area that has been sampled. However, the sample locations should be examined to ensure that areas near known point sources are excluded. In this stage it is also important to identify possible outliers indicative of an unsuspected source signal, potential mixture distributions due to the presence and blending of two or more strong and different local signals (e.g., Bellingham Bay), and to determine spatial gaps in the background data set. For example are there internal areas which have not been sampled, but which should be included in the background population? Is the existing boundary too limited, and more sampling is required? Or, is the existing boundary too broad, and should some areas with unusual contamination patterns be excluded until more information is available?

If the samples were not all collected using a random or systematic random design within the total area (e.g., some samples were targeted to address questions regarding local sources), then adjustments to the data set are needed to reduce the risk of bias in the summary statistics from over-sampling sub-regions of the population. We recommend two possibilities for this adjustment: 1) adjust the boundary of the background population to be a union of circles surrounding each sampled point. If the circles do not overlap, the

samples then comprise a systematic sample of the population thus defined. Or, 2) use a spatial interpolation method (e.g., kriging, or area weighted averaging) to estimate a concentration surface for a larger background boundary. There may not be enough data to accomplish the latter alternative, and the boundary for the former alternative may not be acceptable. In that case, more data must be collected. These two alternatives are discussed in more detail in Section 3.2.3.

3.2.2 Step 2. Examine Trend Characteristics and Autocorrelation of Samples within the Background Data Set

The presence of trends, differences in mean concentrations, and spatial autocorrelation within a data set require special attention. Ignoring trends or autocorrelation can result in biased estimates of population parameters and summary statistics that are not representative of the entire background population.

Spatial autocorrelation is important to identify so that only the independent samples may be used. Clusters of samples that targeted a particular sub-region of the background area should not be allowed to overly emphasize conditions of that sub-region in the description of the entire background area.

If there are concentration trends, or areas with clearly different concentrations (i.e., separate strata), these should be removed prior to estimating autocorrelation (i.e., autocorrelation is estimated from data that exhibit no trend and have a zero mean). In this report, a relatively simplistic approach to evaluating trends is used; an in-depth evaluation and description of a trend surface is beyond the scope of this report. For each case study in Section 4.0, several surface concentration models were used to evaluate potential trends in concentrations. Least squares polynomial surface models of orders 0 to 4 (i.e., from no trend up to a 4th order polynomial) were considered. (The total number of samples for the case studies is 26 to 27, which is probably too few for adequate fitting of the 4th order polynomial model. However, it is considered for illustration purposes.) The five polynomial regression models were compared using Aikake Information Criterion corrected for sample size (AIC_c , Burnham and Anderson, 2002). Note that the AIC_c is based on the maximum likelihood, which is a function of the residuals, and this metric may be somewhat compromised if autocorrelation is present in the residuals. For final trend models, the process is iterative – trend should be re-evaluated after autocorrelation is removed. Autocorrelation is not expected to have a large effect on the trend evaluation in this context of these case studies, mainly due to limitations of the spatial distribution of the small data sets.

There are many methods for evaluating spatial autocorrelation (e.g., using GIS or other spatial statistics packages). The simple method used here could be done manually in MS

Excel, or in R (R Development Core Team, 2011). The case studies do not have a regularly spaced grid of samples, so the boundary of autocorrelation is estimated by evaluating correlation among pairs of points within a certain distance of each other. An autocorrelation boundary can be estimated if there are a reasonable number of points that are close enough together to be autocorrelated. Pairs of sample points are grouped into bins of similar distances. For example, if there are at least six pairs of points within 200m of each other, the distance bins could be 0-200m, then 200-400m, etc. Theoretically, any existing positive autocorrelation would be highest in the first bin. The autocorrelation is estimated by Pearson's linear correlation coefficient between concentrations for all possible station pairs within a distance bin.

The presence of autocorrelation should be tested on the residuals from the best-fit trend model (i.e., the de-trended data). The closest distance that could be tested for autocorrelation was the smallest distance yielding at least six pairs. In spatial statistics literature, six is a small number of pairs on which to test the autocorrelation (e.g., Journel and Huijbregts, 1978), and is considered to be a bare minimum for a correlation test. For this small sample size, a significance test of the autocorrelation within each distance bin used $\alpha = 0.20$ in order to limit Type II errors (i.e., failing to reject the null hypothesis when autocorrelation is present). This binned hypothesis testing approach is useful given the data limitations (i.e., insufficient pairs of samples at sequentially increasing distances) and the objective of estimating the minimum distance between independent samples. If the data were to be used to estimate a kriged trend surface, then a smoothed autocorrelation function is required, where autocorrelation is described as a continuous function of distance.

3.2.3 Step 3. Calculate Summary Statistics

In the preceding two steps, the valid background samples and the background population boundary have been identified, along with potential trends and autocorrelations within the data set. At this point, we consider two methods for generating appropriate summary statistics in the presence of autocorrelation and/or trends:

Method 1. Adjust the background boundary to be simply the union of the set of independent circles (radius > autocorrelation range) surrounding the existing sampling locations, and treat samples from these independent circles as an independent data set;
or

Method 2. Generate a concentration surface for the defined background boundary and use the surface for generating upper bound estimates (e.g., 95 UCL on the mean and 90/90 UTL).

For Method 1, the boundary of the background population is redefined to be exactly the area that has been independently and systematically sampled. This is simply a union of circles around each sampled point, with the radius of the circles greater than or equal to the autocorrelation range. Only one observation within each circle can be used, so that the data set is a uniform independent sample from the defined population. The radius can be larger than the autocorrelation range in order to make the background boundary larger, but this may cause overlapping circles and require subsampling to maintain a uniform and independent systematic sample of the defined population. It also increases the uncertainty, because now a single observation is representative of a larger (unsampled) area, where the concentrations are unknown. If there are unsampled gaps among the sampled locations, these areas are not actually part of the background population as defined. Without information about whether the concentrations in the unsampled gap areas are homogeneous or trending, it is perhaps an overstatement of the available information to assert that the concentrations remain constant within these unsampled areas. Consequently, the background population as defined may not be contiguous.

Method 1 is a fairly simple and reasonable choice if the area that has been sampled reflects an adequate boundary for background. This method requires minimal assumptions, and does not extrapolate beyond, or interpolate between points.

Method 2 may be desirable if the area described by the union of circles around each sampled point is not acceptable, and more samples cannot be collected. For this method, a concentration surface is generated by kriging or another surface contouring method to estimate concentrations for the areas between the sampled points. Method 2 can be complex and requires more assumptions about the behavior of the data, but it is the only way to estimate concentrations across a broader boundary area. For example, if a large unsampled area is located between two areas of high concentrations, then the surface model would predict concentrations in that area to be similar to neighboring concentrations, rather than simply excluding that area from the described population. Uncertainty in the estimate of the concentrations in interpolated areas reflects prediction error from the model, so the upper bound on the mean for the total background area will quickly increase as you spatially interpolate or extrapolate beyond your data. This approach is not advisable when the data are sparse. Note that the simplest trend surface model would use Thiessen polygons to divide the area into polygons represented by one sampled point. This is a model that assumes constant concentrations within each polygon, and estimates of uncertainty are not readily available (although bootstrapping could be used).

3.2.3.1 Process for Estimating Summary Statistics from Existing Data Using Method 1

The autocorrelation distance d is defined to be the smallest distance between data points for which independence can be assumed. This distance may be derived from site-specific data, results from similar data sets in other areas, or best professional judgment (BPJ). To estimate background summary *statistics using Method 1*:

- a. Establish a sample boundary radius $r \geq d$. The larger the radius, the larger the boundary of the background area, but the smaller the overall sample size may be (only one sample per circle is permitted). There is a tradeoff between a large background area with high uncertainty and a smaller, perhaps non-contiguous background area with lower uncertainty.
- b. Identify all samples greater than r away from all other samples. These samples are assumed to be independent, random observations given our estimated autocorrelation range. If this captures all of the samples (i.e., only one sample per circle), then the existing data are i.i.d. random samples from the background population and summary statistics can be generated on this dataset, with no need to subset the data. Otherwise:
- c. Identify clusters that have two or more samples within r of each other. Permute all possible ways of selecting one or more independent samples ($>r$ apart) from each cluster.
- d. Combine the samples from b) and c) to form all the possible combinations of samples from the data set that are $> r$ apart from every other sample. For each of these permutations, generate summary statistics (mean, variance, 95 UCL on the mean, 90th percentile, and 90/90 UTL).
- e. Use the distribution of each statistic (e.g., 95 UCL on the mean) to find the best summary statistic for the population (e.g., the maximum 95UCL on the mean will be most likely to capture the true background population mean). Each of the individual UCLs represents a slightly different background boundary, but each is a possible representation of the regional background. The permutation distributions of these summary statistics reflect part of the uncertainty in the background boundary. A highly variable distribution of background 95 UCLs indicates that different background boundaries can produce widely different results, an indicator of small-scale changes in concentration and the need, perhaps, for a re-evaluation of the first steps of the background boundary definitions (i.e., Steps a and b).

3.3 Sampling to Augment Existing Background Data

For existing background data sets, there are likely to be four features that could be improved by additional sampling: 1) temporal consistency; 2) spatial extent; 3) sampling density; and 4) sample size adequacy:

Temporal consistency: Placing new samples in the areas with outdated sample results may be a priority.

Spatial extent: For the spatial extent (boundaries) of the background population, BPJ is required to decide whether the outer boundaries are sufficient to fully capture the target background conditions. Addressing the boundary question uses BPJ and possibly geophysical modeling information about the mechanisms distributing the sediments throughout the area from anthropogenic contributions (e.g., what are the boundaries of influence from non-point source runoff) or natural sources (e.g., what are the boundaries of influence from river sediments).

Sampling density: If the sampling locations are sparsely distributed, then the union of overlapping circles for the point locations sampled may not result in a contiguous background area. Large unsampled or under-sampled areas within the background area should be sampled to reduce uncertainty. Existing data may indicate that trends are present, so sampling at fairly regular intervals along that trend is recommended. Bathymetry and hydrologic flow patterns provide information about sedimentation or disturbance patterns. This information can be used to identify areas within the background population that have potentially different contamination levels, and where additional sampling is recommended.

Overall Sample Size: If more data are needed to satisfy one or more of the three features described above, grid sampling is recommended. The minimum grid size should be d (the autocorrelation range). The actual grid size will be determined by budget in most cases, but sample sizes needed for a desired estimate of precision can be estimated using existing data. These calculations, however, assume that additional samples will have the same mean and variance as the existing data, which may not be a valid assumption when sampling from a patchy distribution or a trending surface. At best it provides an informed guess regarding the change in precision of the mean estimate (i.e., width of the 95UCL on the mean) with additional samples.

4.0 Case Study Examples

Ecology provided dioxin/furan TEQ concentrations (ng/kg, dry weight) for three case study sites: Fidalgo Bay, Port Gardner, and Bellingham Bay. For each of these case studies, we illustrate the process of defining the regional background boundaries, evaluating concentration trends in the data, estimating the autocorrelation distance, and using Method 1 (Section 3.2.3.1) to estimate upper bound summary statistics.

4.1 Fidalgo Bay Case Study

Ecology staff used BPJ to identify samples that were too near to point sources, or were from a different area-of-influence and therefore considered to be inappropriate for comparison to

Fidalgo Bay Site concentrations. The sampling locations are shown in Figure D-1; the grid overlaid on the map is a 0.5 km square grid used simply to illustrate the scale of the distance between samples. The latitude and longitude for the sampling locations and their associated TEQ concentrations (ng/kg, dry weight) are provided in this Appendix, Table A-1.

4.1.1 Trends and Autocorrelation in Fidalgo Bay

The first step in estimating the autocorrelation range is to remove any existing trends that may be present in the samples. For the Fidalgo Bay data set the first-order polynomial (i.e., a linear trend) fits the trend surface best (indicated by the lowest AICc for all trend surface models considered). There appears to be a linear increasing trend to the northwest (Figure D-2).

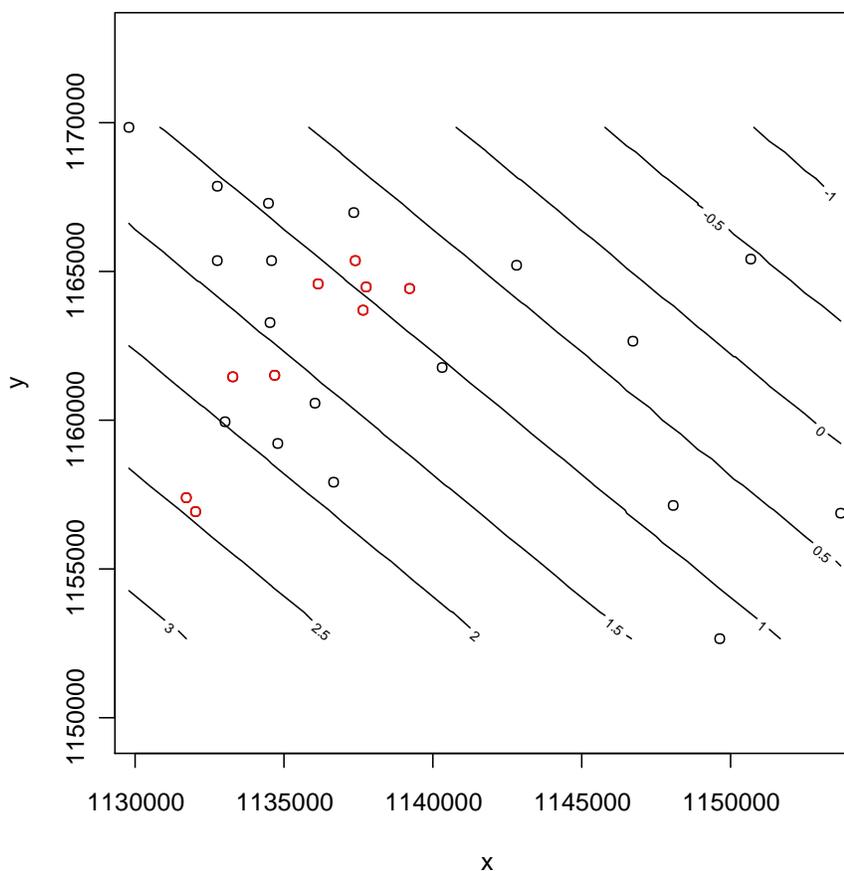


Figure D-2. Fidalgo Bay sample locations with contours based on first-order polynomial fit. Samples that are <460m from other samples are highlighted in red. The maximum distance between the six closest pairs of points in Fidalgo Bay is approximately 460m apart, so bin sizes are set to multiples of 460m. The data points within the first distance bin are highlighted with a different color in Figures D-1 and D-2. A distance of 460 m is the smallest autocorrelation range that we can test with these data. There may be autocorrelation present at smaller distances, but we cannot test whether the

correlation is significant because of insufficient numbers of data pairs available for that distance. The correlation results after removing the linear trend from the data are shown in Table D-1.

Table D-1. Autocorrelation Results for Fidalgo Bay.

Bin Endpoints (m)	N	Pearson's Correlation Coefficient	one-tailed p-value for parametric test
0-460	6	0.338	0.512
460 – 920	33	0.246	0.168
920-1380	42	0.0894	0.573
1380 – 1840	59	-0.00570	n/a

The autocorrelation coefficient decreases with distance (Table D-1). The autocorrelation in the first bin is not significantly different from zero ($\alpha=0.20$). Sample sizes after the first bin increase dramatically, so it is appropriate to use a lower α -level to assess significance in the second bin (e.g., $\alpha=0.10$ or 0.05), from which we would conclude that the autocorrelation is not significantly different from zero in this distance range (460 – 920 m). We varied the size of the first bin, and the autocorrelation was evaluated for pairwise distances from zero to increasingly larger maximum distances, up to 914m, but no strong correlations were seen at any of these distances. There are insufficient samples close to each other to estimate autocorrelation less than 460m. There may be autocorrelation at smaller ranges, but it is not a testable hypothesis on this dataset, so we could assume that the data set is roughly uncorrelated at the distances that were sampled. However, we don't want to underestimate autocorrelation, so we also compare the estimation results by subsampling the data (*Method 1*, Section 3.2.3.1) based on a minimum separation distance of approximately 460m (for convenience we round up to 500m).

4.1.2 Method 1 Applied to Fidalgo Bay

Subsets of independent samples were generated from the full data set, as described above. The autocorrelation range used in this approximation was 500m.

There were 27 samples from acceptable non-site affected stations in Fidalgo Bay. Of these 27 samples, 15 were more than 500m away from any other samples. The remaining 12 samples were grouped into three clusters of two or more samples each. All permutations were constructed of independent samples within each of these clusters and combined with the other 15 samples (24 possible permutations). For each of the permutations, summary statistics for TEQ values were generated (i.e., mean, variance, bootstrapped 95 UCL on the

mean, and 90/90 UTL for the best fit gamma distribution). The distributions of these TEQ summary statistics are shown in Figure D-3. The red lines indicate the values calculated by assuming that the data are uncorrelated at the distances that were sampled (i.e., we have an independent data set). We can see that the full data set (n=27) produced a 95 UCL on the mean (1.55 ng/kg TEQ) that was lower than some of the permutations. For the permutations, 95UCL values range from 1.4 to 1.8 ng/kg TEQ. For the 90/90 UTL, permutation values ranged from 2.5 to 3.6 ng/kg, and the observed data had a TEQ value of 3.1 ng/kg.

4.1.3 Fidalgo Bay Conclusions and Recommendations

The trend surface regression models indicate a linear trend in Fidalgo Bay background concentrations. For this case study, the observed data can be used to estimate background population characteristics as is, or, to ensure independence, after the data have been sub sampled to generate a set of independent observations using the autocorrelation range (at 500 m). From the permutation distribution of UCLs, we could choose the maximum value (1.9 ng/kg) as this is the value most likely to capture the true mean, even though the coverage will likely exceed 95%.

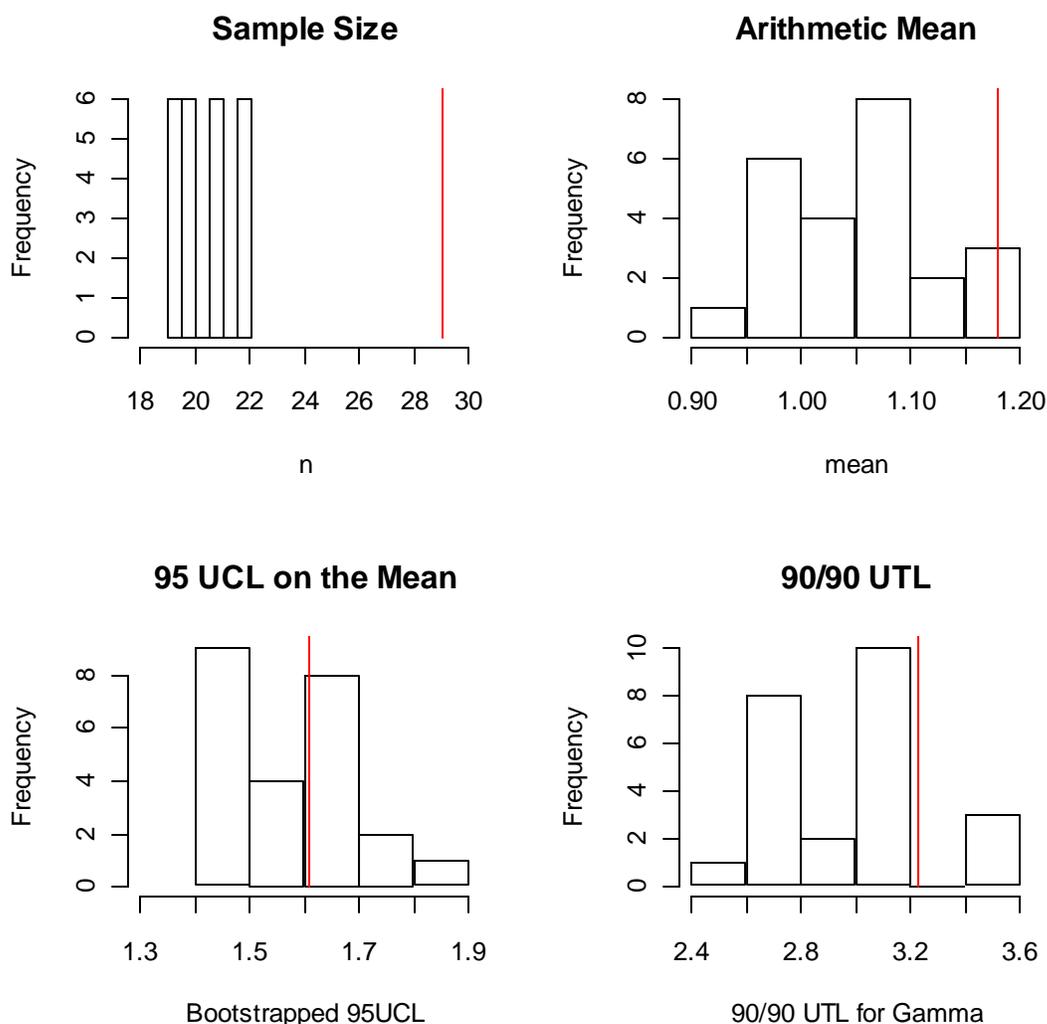


Figure D-3. Distribution of summary statistics on TEQ values (ng/kg, dry wt) for the permutations of independent samples (>500m) at Fidalgo Bay. The red lines indicate the values for the observed data set.

For sample adequacy, we consider the following:

Temporal consistency: The data within background areas are collected from 2007 and 2010, so time period is probably not an issue and all these data are useable.

Number of samples: The existing data are not significantly different from a gamma distribution (ProUCL, $\alpha=0.05$). We plot the number of samples vs. the width of the gamma confidence interval on the mean (Figure D-4). The figure shows that our sample size of 27 provides a UCL width that is 37% of the mean; we're not on the steepest part of the curve but are on a part of the curve where it's starting to flatten out for our sampled

population. Doubling the sample size is expected to decrease the UCL half-width to 24%, assuming that the mean and the variance stay the same. This assumption may not be realistic given that there is a trend in these data – samples collected from a different area will affect both the mean and the variance, so this graph provides simply a ballpark estimate of expected sample size adequacy.

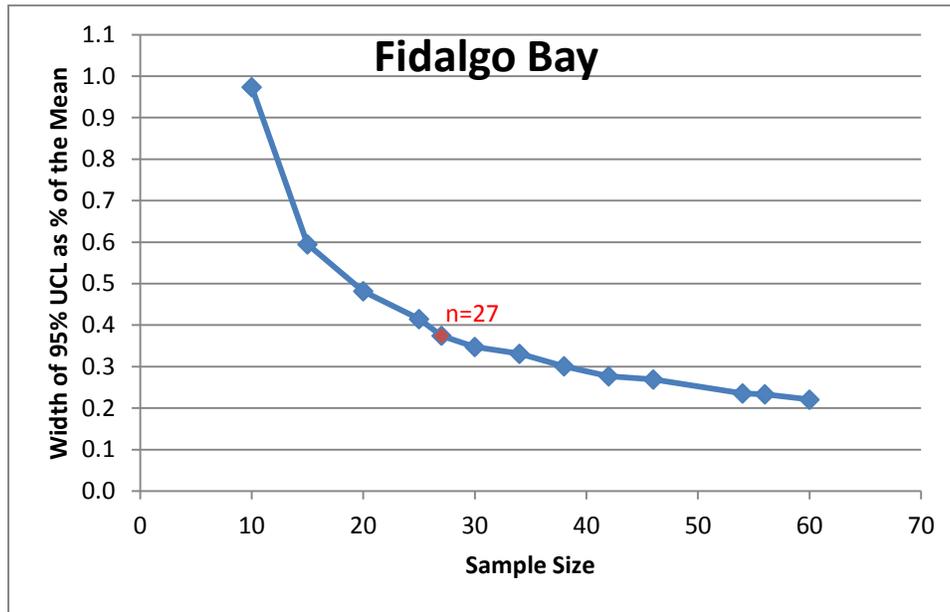


Figure D-4. Sample size vs. precision of the mean using Fidalgo Bay data, fit using a gamma distribution.

Spatial Extent: The best trend surface was a 1st order polynomial, decreasing to the northwest away from the Anacortes shoreline. Currently, most of the samples are closer to Anacortes in Fidalgo Bay, so these might describe Regional Background, whereas samples in Padilla Bay describe Local Natural Background, for example. There are rather few samples in Padilla Bay, so if this area is included as part of Regional or Local Natural Background, it would be wise to place more samples on that side of the Bay.

Sampling Density: Based on the autocorrelation range test, we recommend samples no closer than 500m apart (rounding up from 460m). This minimum spacing is expected to achieve independent samples. Any new samples would be placed as evenly as possible within the desired boundaries for the background population, and at least 500m away from any other new or existing samples.

- Option #1: Spatial extent for regional background is a line drawn from Anacortes west to Hat Island and south to March Point. Take 5-10 more samples within any of the available grid squares, trying to achieve uniform distribution of samples throughout the area and minimum separation between samples of 500m. Pros: this provides a good spatial coverage within the delineated regional background of

Fidalgo Bay. Cons: none, assuming that the boundary for background is sufficient.

- Option #2: Include Padilla bay in the Background characterization. Pursue Option #1, plus additional 5-10 samples from Padilla Bay. Use a grid approach to try to achieve a uniform distribution throughout the area with minimum separation of 500m. Pros: larger background area; provides data to test if Padilla Bay is a separate population. Cons: The higher cost over Option #1 may not be necessary; the shallower depths and the large Intertidal areas in Padilla Bay may make this an inappropriate background data set for the subtidal sites in Fidalgo Bay.
- Option #3: Include Padilla Bay in the Background characterization, and fill as many squares as possible in both Fidalgo and Padilla Bay, increasing the grid size to 1km (we exchange small scale accuracy for broad scale information). Pros: More information about both areas and a dataset that may be sufficient for drawing a surface contour map. Cons: Cost and potentially sampling overkill.

4.2 Port Gardner Case Study

Ecology staff used BPJ to identify samples that were too near to point sources, or were from a different area-of-influence and therefore considered to be inappropriate for comparison to Port Gardner site concentrations. The sampling locations are shown in Figure D-5; the grid overlaid on the map is a 0.5 km square grid used simply to illustrate the scale of the distance between samples. The latitude and longitude for the sampling locations and their associated TEQ concentrations (ng/kg, dry weight) are provided in this Appendix, Table A-2.

4.2.1 Trend and Autocorrelation in Port Gardner

For the Port Gardner data set, the trend surface is best fit by the third order polynomial (the model with the lowest AICc; Figure D- 6).

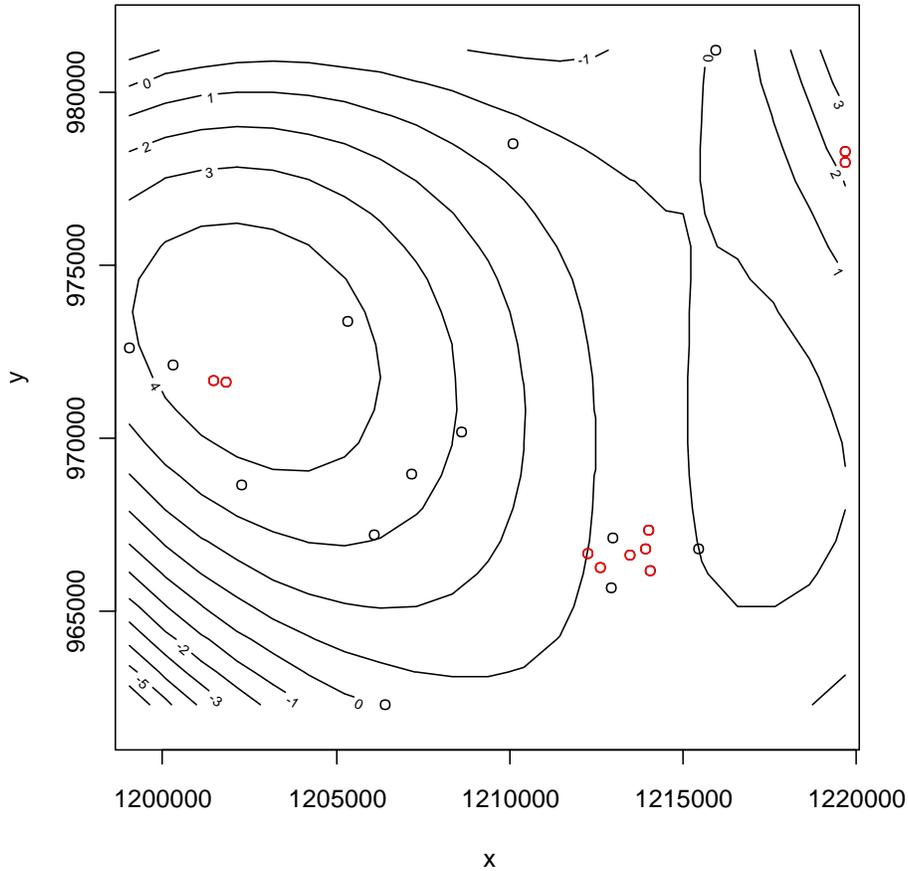


Figure D-6. Port Gardner sample locations with contours based on third-order polynomial fit. Samples that are <200 m from other samples are highlighted in red.

The maximum distance between the six closest pairs of points is 200m, so the bin sizes were set to multiples of 200 m. The autocorrelation results are shown in Table D-2.

Table D-2. Autocorrelation Results for Port Gardner data.

Bin Endpoints (m)	N	Pearson's Correlation Coefficient	one-tailed p-value for parametric test
0-200	6	0.735	0.048
200-400	15	-0.261	n/a
400-600	15	-0.0598	n/a
600-800	5	-0.300	n/a

With the polynomial trend removed, there is evidence that samples within 200m of each other are still correlated. We are limited by the number of samples and the distances among

them in our estimation process of the autocorrelation range. The data indicate that samples within 200m of one another should not be treated as independent samples. The minimum distance for independence is at least 200m. Samples between 200m and 400m apart were not correlated, though this appears to be strongly influenced by a single elevated sample in the cluster of stations near Weyerhaeuser (the cluster of stations close to shore in Figure D-5).

4.2.2 Method 1 Applied to Port Gardner

Subsets of independent samples were generated from the full data set, as described above. The autocorrelation range used in this approximation was 200m. There were 26 samples from acceptable non-site affected stations in Port Gardner. Of these 26 samples, 17 were more than 200m away from any other samples. The remaining nine samples were grouped into four clusters of two or more samples each. All permutations were constructed of independent samples within each of these clusters and combined with the other 17 samples (24 possible permutations).

There were a few samples that were just beyond 200m apart, so we also calculated results for autocorrelation range of 305m. For this range, there were 14 samples more than 305m away from any other samples. The remaining 12 samples were grouped into four clusters of two or more samples each, and all permutations were constructed of independent samples within each of these clusters and combined with the other 14 samples (28 possible permutations).

Sample sizes ranged from 18 to 20 for the permutations at 305m distance; and were always 21 for the permutations at 200m distance. Arithmetic mean values were higher for the 305m distance: ranging from 1.86 to 2.01ng/kg TEQ compared to a range of 1.77 to 1.84 ng/kg TEQ for the 200m distance. The distributions of the 95UCL and the 90/90 UTL values for the TEQ of these two sets of permutations are shown in Figure D-7. The larger values for the 305m distance partially reflects the smaller sample size (even for an identical distribution, a smaller n generates a larger UCL and UTL because of the greater uncertainty). But the larger values in the 305m distance permutations also indicate the presence of some small scale spatial variability, as observed in the tight cluster of samples near Weyerhaeuser (the cluster of stations close to shore in Figure D-5).

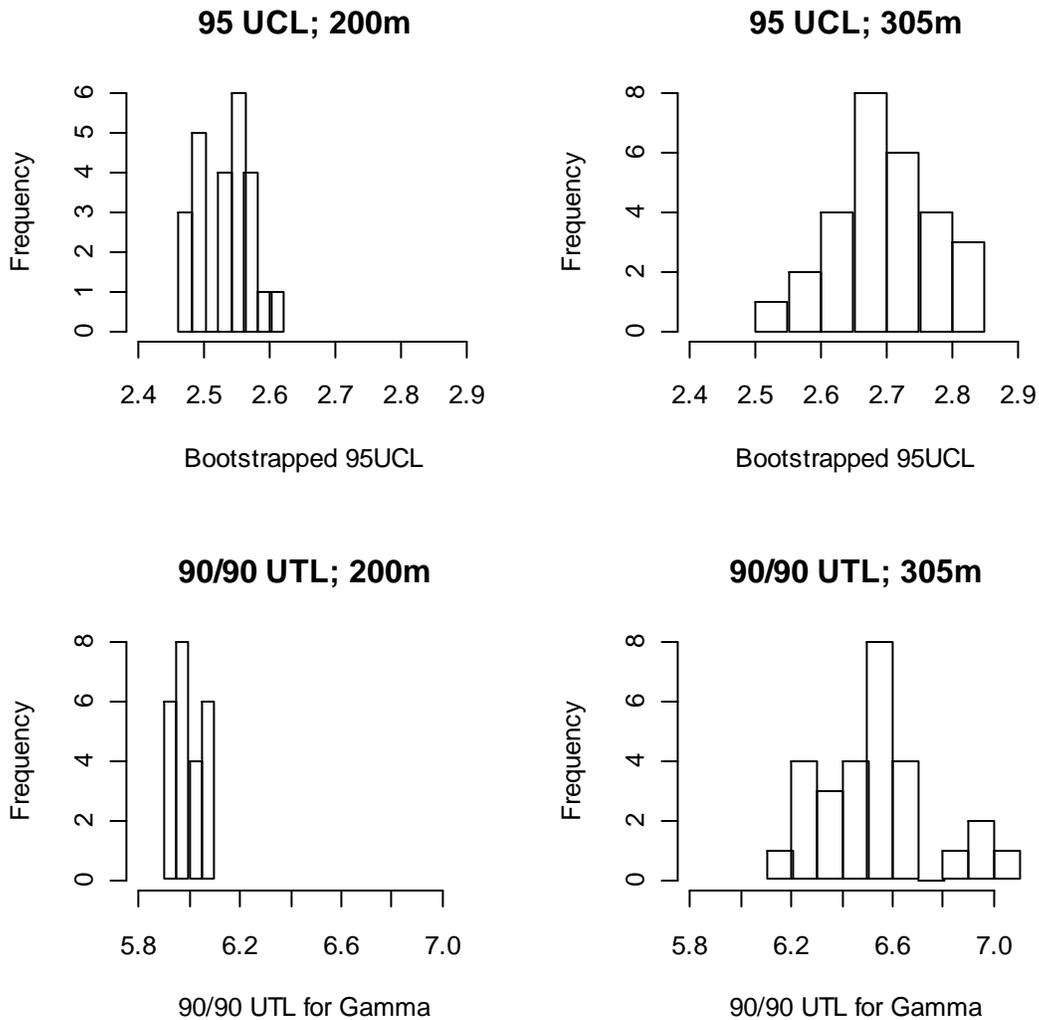


Figure D-7. Distribution of 95UCL and 90/90 UTL for TEQ values (ng/kg, dry wt) for the permutations of independent samples in Port Gardner using 200m autocorrelation range (left side), or 305m autocorrelation range (right side).

4.2.3 Port Gardner Conclusions and Recommendations

The autocorrelation investigation and the trend surface regression models indicate that there is trend and/or patchiness in Port Gardner background concentrations of TEQ values (ng/kg, dry weight). Using an independent subset of the existing data will describe the area sampled (the union of the circles around our points sampled), but there are large areas un-sampled with uncertainty about what concentrations might be found there. The differences in the results for the autocorrelation range of 200m vs. 305m indicate the patchy nature of TEQ concentrations in at least some of the areas (although this could only

be tested for the dense sample cluster near Weyerhaeuser). This means that interpolation could lead to erroneous conclusions about area averages.

For sample adequacy, we consider the following:

Temporal consistency: The data within background areas are collected from 2004 to 2010. The data from 2004 are all from the Weyerhaeuser sampling, which were also found to be highly influential in the subsetting exercise. Additional sampling in this area may be desirable to describe more current conditions in the area.

Number of samples: The existing data are bimodal, and are not well fit by any parametric distribution (ProUCL). For this data set we can't generate any assumptions about sample size adequacy.

Spatial Extent: The best trend surface was a 3rd order polynomial, indicating patchiness with low concentrations near Weyerhaeuser; higher concentrations near the DMMP disposal site. This site would benefit from additional samples placed out beyond the DMMP disposal site (to see how far out those concentrations extend), and the area between Jetty Island and the disposal site. It may also be desirable to sample the area north and northwest of the current northern boundary of the existing data in order to capture what influence the Snohomish River may have on the bay concentrations.

Sampling Density: Based on the autocorrelation tests, we recommend samples at least 400m – 500m apart. This sampling interval will miss some of the small scale spatial variability that is present, but would allow efficient description of a larger area. In addition, a grid spacing of 500m is expected to achieve independent samples. Any new samples would be placed as evenly as possible within the desired boundaries for the defined background population, and at least 500m away from any other new or existing samples.

- **Option #1:** Modify the spatial boundary for regional background to exclude everything north to northwest of the line drawn from the southern point of Jetty Island out into Possession Sound. Take at least 10 more samples within any of the available grid squares (500m grid), trying to achieve uniform distribution of samples throughout the area and minimum separation between samples of 500m. Pros: A smaller boundary allows a greater sampling density within the area considered representative of regional background. Excluding the areas on the Snohomish River delta may be justified if the project locations are not heavily influenced by the river. Cons: the spatial boundary may be too limited.

- Option #2: Modify the spatial boundary to exclude the deeper subtidal areas, and include only the areas strongly influenced by the Snohomish River. Pros: A smaller boundary allows a greater sampling density within the area considered representative of local background. Excluding the areas outside of the Snohomish River influence may be justified if the project locations are primarily influenced by the river. Cons: the spatial boundary may be too limited.
- Option #3: Combine Options 1 and 2 to describe a larger background area. Sample in as many grid squares as is affordable, increasing the grid size to 1km (exchange small scale accuracy for broad scale information). Try to achieve a uniform distribution throughout the area and minimum separation between all new and existing data of 0.5-1km. Pros: A broader area is defined that allows the description of an overall background average; if separate populations are present near the mouth of the Snohomish River vs. subtidal Possession Sound, it may be apparent by these data. Cons: Cost.

4.3 Bellingham Bay Case Study

Ecology staff used BPJ to identify samples that were too near to point sources, or were from a different area-of-influence and therefore considered to be inappropriate for comparison to Bellingham Bay site concentrations. The sampling locations are shown in Figure D-8; the grid overlaid on the map is a 0.5 km square grid used simply to illustrate the scale of the distance between samples. The latitude and longitude for the sampling locations and their associated TEQ concentrations (ng/kg, dry weight) are provided in this Appendix, Table A-3.

4.3.1 Trend and Autocorrelation in Bellingham Bay

For the Bellingham Bay data set, the second-order polynomial provides the best fit (the model with the lowest AICc) for the trend surface model (Figure D-9).

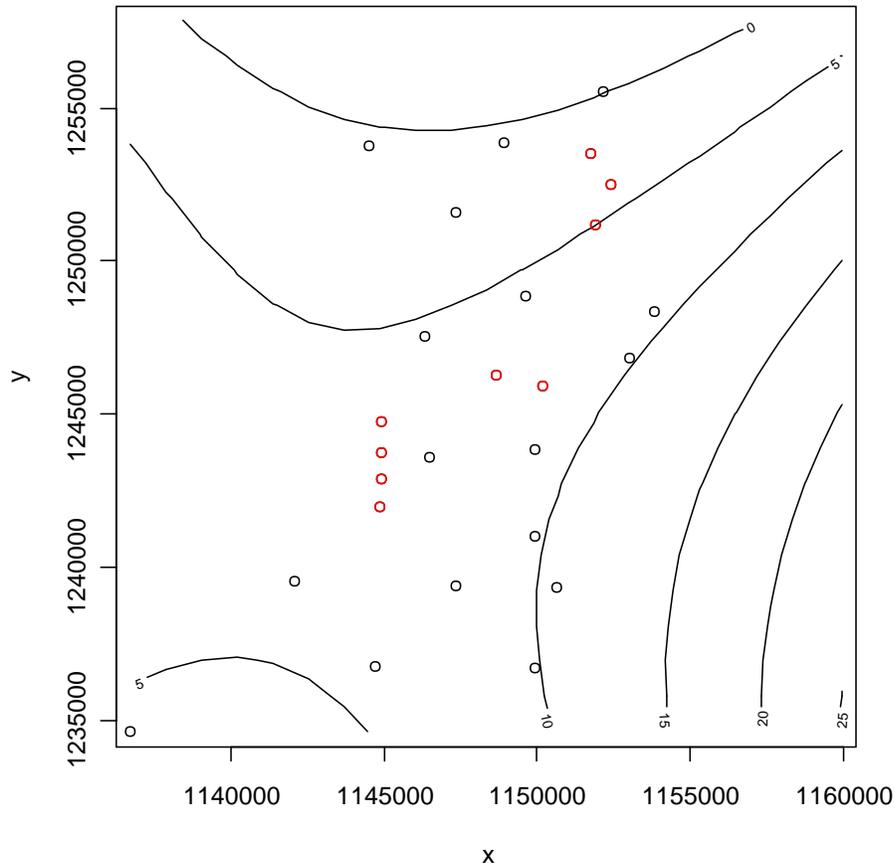


Figure D-9. Bellingham Bay sample locations with contours based on second-order polynomial fit. Samples that are <472 m from other samples are highlighted in red. The maximum distance between the six closest pairs of points is 472m, so the bin sizes are set to multiples of 472m. This is the smallest autocorrelation range that we can test with these data. There may be autocorrelation present at smaller distances, but we cannot test whether the correlation is significant because of insufficient numbers of data pairs available for that distance. The correlation results after removing the trend from the data are shown in Table D-3.

Table D-3. Autocorrelation results for Bellingham Bay data.

Bin Endpoints (m)	N	Pearson's Correlation Coefficient	one-tailed p-value for parametric test
0 – 472	6	0.315	0.543
472 – 945	23	-0.118	n/a
945 - 1416	41	-0.0704	n/a
1416 - 1890	42	-0.218	n/a

There is no evidence of autocorrelation in the 0-472m range. There may be autocorrelation at smaller ranges, but it is not a testable hypothesis on this dataset, so we could assume that the data set is roughly uncorrelated at the distances that were sampled. However, we don't want to underestimate autocorrelation, so we also test the estimation results by subsampling the data based on a minimum separation distance of 472m (round up to 500m).

4.3.2 Method 1 applied to Bellingham Bay

Subsets of independent samples were generated from the full data set, as described above, using an autocorrelation range of 500m, slightly larger than the smallest autocorrelation range that could be tested.

There were 26 samples from acceptable non-site affected stations in Bellingham Bay. Of these 26 samples, 16 were more than 500m away from any other samples. The remaining ten samples were grouped into three clusters of two or more samples each. All permutations were constructed of independent samples within each of these clusters and combined with the other 16 samples (12 possible permutations). The distributions of summary statistics for these permutations are shown in Figure D-10. The red lines indicate the values calculated by assuming that the data are roughly uncorrelated at the distances that were sampled. We can see that the full data set (n=26) produced a lower 95 UCL on the mean (7.3 ng/kg TEQ, dry weight) than some of the random permutations. For the permutations, 95UCL values range from 7.0 to 7.7 ng/kg TEQ. For the 90/90 UTL, permutation values ranged from 14.5 to 15.9, and the observed data had a value of 14.6 ng/kg TEQ.

One of the clusters of samples had substantial variability in the reported TEQ values: 1.5, 1.6, and 6.3 ng/kg. The first two concentrations were reported for samples from the 0-12cm horizon; the last for a sample from the 0-55cm horizon. This was the only sample included in this background data set that was collected beyond the 0-12cm depth horizon. Permutation results excluding this deeper horizon sample had 95 UCL values ranging from

7.0 to 7.4 ng/kg, and 90/90 UTL values ranging from 14.5 to 15.3 ng/kg. So, this sample definitely had an effect on the upper range of the estimates, but it's unknown whether it reflects greater contamination at depth or overall small scale spatial variability in surface concentrations.

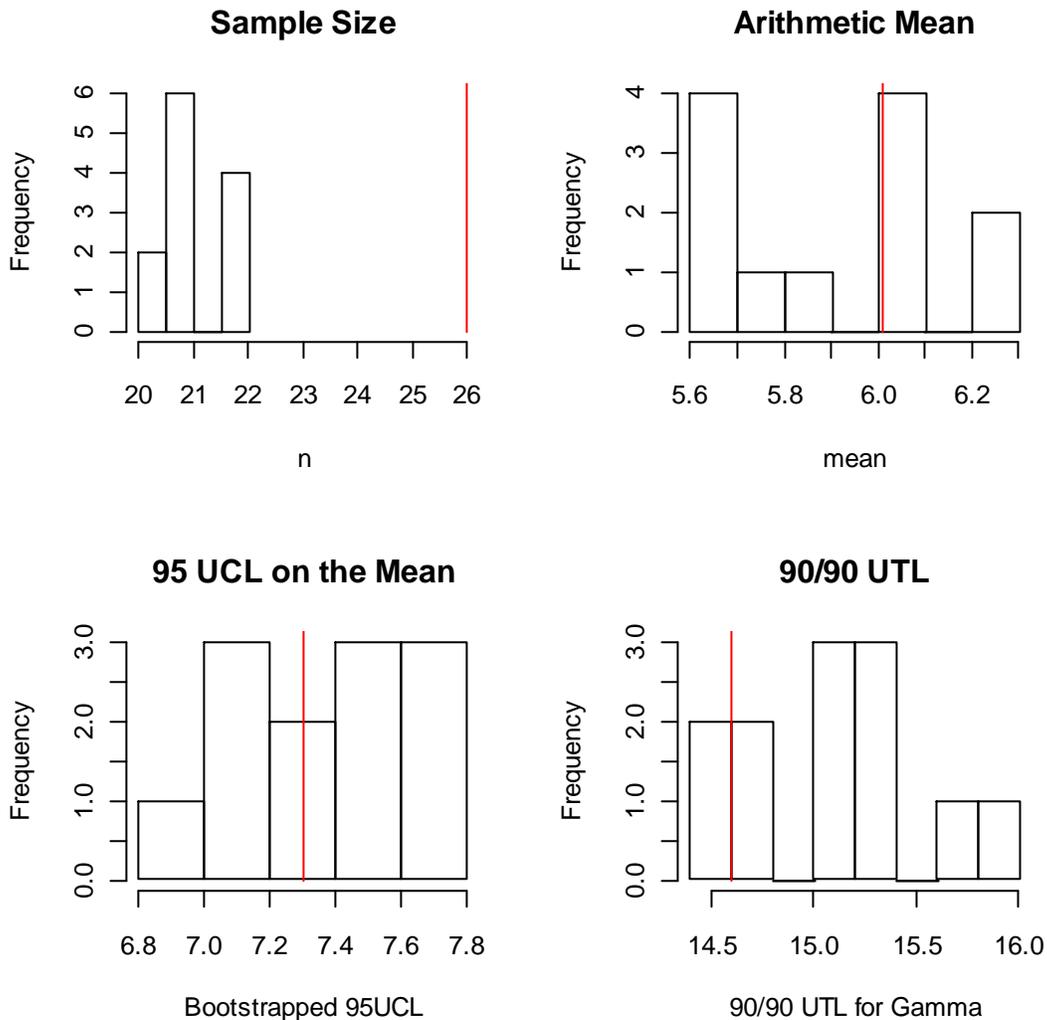


Figure D-10. Distribution of summary statistics TEQ values (ng/kg, dry weight) for the permutations of independent samples (>500m) at Bellingham Bay. Red lines indicate the values for the observed data set.

4.3.3 Bellingham Bay Conclusions and Recommendations

The trend surface regression models indicate that there is significant trend in Bellingham Bay background concentrations. The autocorrelation investigation indicated that the data could be considered independent as sampled, but there are large areas un-sampled with uncertainty about what concentrations might be found there.

Given the strong appearance of trends, and potentially two competing trends (one from the southern shoreline of Bellingham Bay, and another from the Nooksack River in the north), this site would benefit from additional samples. Where the boundary is drawn depends on BPJ regarding the relevance of the Nooksack River influence on the Regional Background concentrations. An independent sampling interval would be 500m, but the large area that needs to be sampled justifies using a larger sampling interval (e.g., 1000m) if needed. For sample adequacy, we consider the following:

Temporal consistency: The data within the background area are collected from 2007 to 2010, so time period is probably not an issue and all these data are useable.

Number of Samples: The existing data are not significantly different from either the normal or the gamma distributions (ProUCL, $\alpha=0.05$). For this sample size calculation, we use the gamma distribution because it allows for more potential skewness in the distribution and a more conservative sample size calculation. We plot the number of samples vs. the width of the gamma confidence interval on the mean (Figure D-11). The figure shows that our sample size of 26 provides a UCL width that is 32% of the mean for the area sampled; we're close to the part of the curve where it's starting to flatten out for our sampled population indicating incrementally smaller advantage from each additional sample. Doubling the sample size is expected to decrease the UCL half-width to about 20%, assuming that the mean and the variance stay the same. This assumption may not be realistic given that there is a trend in these data – samples collected from a different area will affect both the mean and the variance, so this graph provides simply a ballpark estimate of expected sample size adequacy.

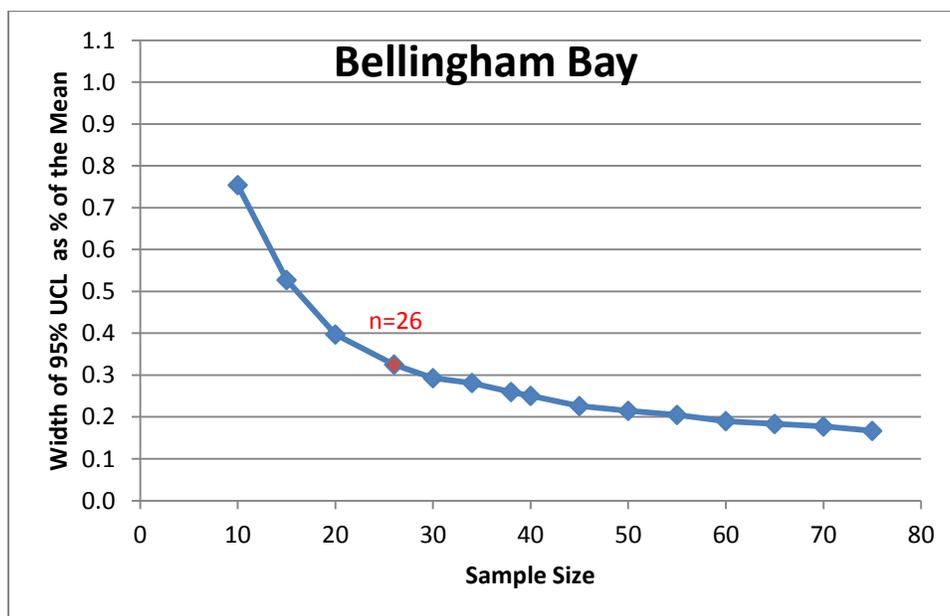


Figure D-11. Sample size vs. precision of the mean using Bellingham Bay data, fit with a gamma distribution.

Spatial Extent: The best trend surface was a 2nd order polynomial, showing a strong trend decreasing away from the Bellingham shoreline along the SE portion of the Bay, and a weaker trend that decreases approaching the northern portion of the Bay and the Nooksack River delta. Where the boundary for background is drawn depends on BPJ regarding the relevance of the Nooksack River influence on the Regional Background concentrations.

Sampling Density: Based on the autocorrelation tests, we recommend samples no closer than 0.5km apart to get a data set of independent samples.

- **Option #1:** Draw the spatial boundary to be just outside of the existing sampling locations, and take an additional 10 samples within any of the available 500m grid squares within that boundary. As much as possible try to achieve spatial evenness, and a minimum separation of 500m. Pros: Maximizes the use of the existing data and fills some data gaps for this background boundary. Cons: If the northern area influenced by the Nooksack River is more of a local natural background, this data set will be a combination of two blending populations.
- **Option #2:** Draw the spatial boundary to exclude some of the existing locations in the northern portion of the bay where the Nooksack River may be influencing concentrations. Take an additional 10 samples within any of the available 500m grid squares within this area, trying to achieve spatial evenness and a minimum separation of 500m. Pros: Same cost as Option 1, but a smaller boundary allows a greater sampling density within the area considered representative of regional background. Excluding the areas of the bay with strong Nooksack River influence may be justified if the project locations are more strongly regionally influenced similar to what's found in the southern portion of the Bay. Cons: the spatial boundary may be too limited (encompasses an area generally within 3km of the shoreline) and therefore may be focused too much on the upper concentration end of the trend.
- **Option #3:** Draw the spatial boundary further out into the Bay to try to identify where the two trends meet. The grid size could be increased to 1km (exchange small scale accuracy for broad scale information). Try to achieve a uniform distribution throughout the area and minimum separation of 0.5-1km. Pros: A broader area is defined that allows better understanding of the two local influences (i.e., the river and the urban area), and therefore a better description of an overall background average. Cons: Cost, and sampling overkill if the regional background is what's needed for project comparison.

5.0 General Recommendations

The approach used by agencies (US EPA, OR DEQ, WA DOE, and ACOE) for describing background involves the initial definition of the population. Given a narrative description of Regional Background, or Local Natural Background, the spatial boundaries for the appropriate background are a site-specific question and must be drawn using existing data from the area, information about fate and transport of contamination from the site, regional influences, as well as best professional judgment.

Once the data within the presumed background area have been compiled, it is fairly simple to fit a selection of modeled trend surfaces, and look for autocorrelation in the residuals. The trend surface provides some information about spatial variability and local patchiness in the concentrations, which can assist in helping determine the best locations for additional sampling. The presence of observable trends also indicates that a random sample (or even a random subset of the existing data) may inadequately describe background areas that have not been uniformly sampled. Treating a non-uniform sample from a trending population as if it were i.i.d. can result in biased estimates of the mean and the 95UCL on the mean. Existing data may be insufficient to detect the presence of a trend (i.e., just because we don't detect a trend doesn't mean it's not there). The optimal sampling design in the presence of trends is one that achieves spatial evenness. Since there is no harm in assuming that trends are present, the optimal design should always attempt to achieve uniform spatial coverage using systematic random samples, or more complex designs such as GRTS sampling (US EPA 2011).

The autocorrelation investigation helps uncover the magnitude of spatial autocorrelation in existing background data sets. Using the autocorrelation range estimate, we can assume that samples outside of this distance range can be considered effectively independent. The sampling density of the existing data set may be insufficient to measure the true autocorrelation range, but it should be sufficient to provide an approximate minimum separation distance to define the grid size for future sampling. If the existing data set does contain dense clusters of samples, then the autocorrelation investigation allows those data to be treated appropriately for the calculation of background summary statistics by selecting independent subsets of the data (i.e., Method 1).

Finally, we emphasize that Method 1 described herein only allows description of the areas that are directly sampled (the union of circles around the sampled data points). Any extrapolation beyond, or interpolation between the sampled locations is avoided. Spatial modeling (Method 2) can provide estimates for interpolated concentrations across a broader

area but has the disadvantages that it can be complex, requires more assumptions about the behavior of the data, and cannot be done adequately when the data are sparse.

6.0 References

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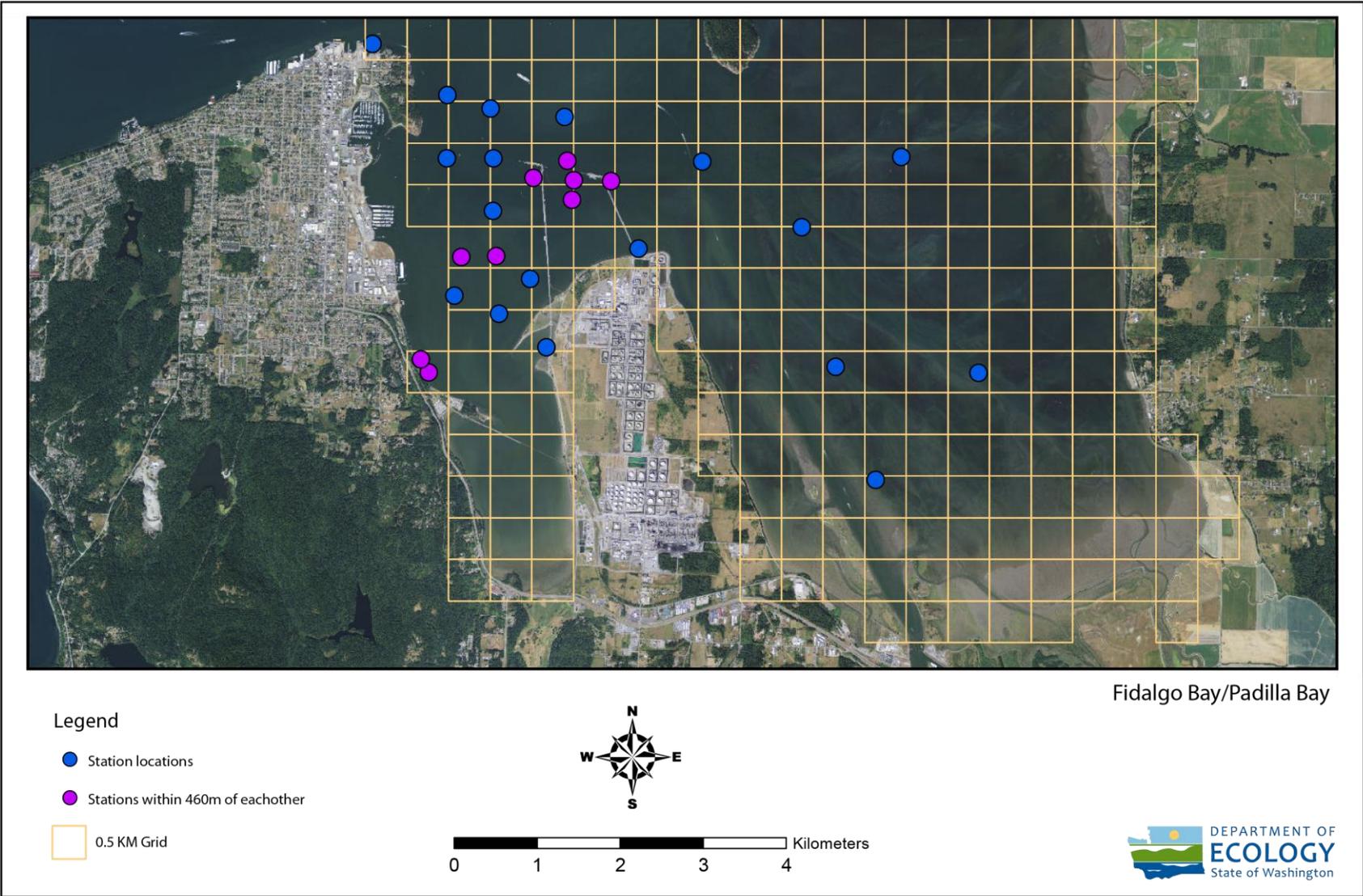


Figure 1. Map of Fidalgo Bay case study site, showing locations of existing data.

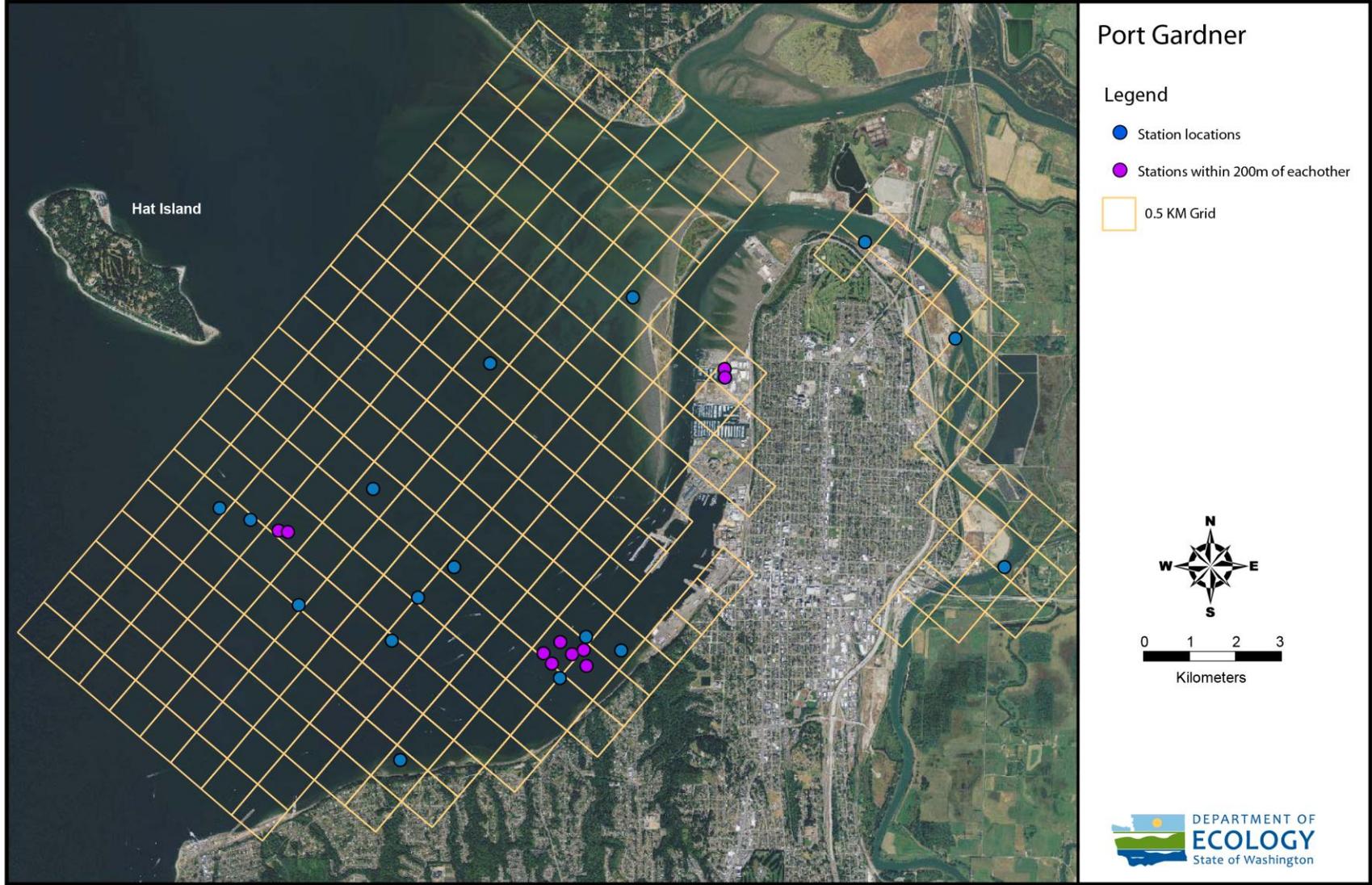


Figure 5. Map of Port Gardner case study site, showing locations of existing data.

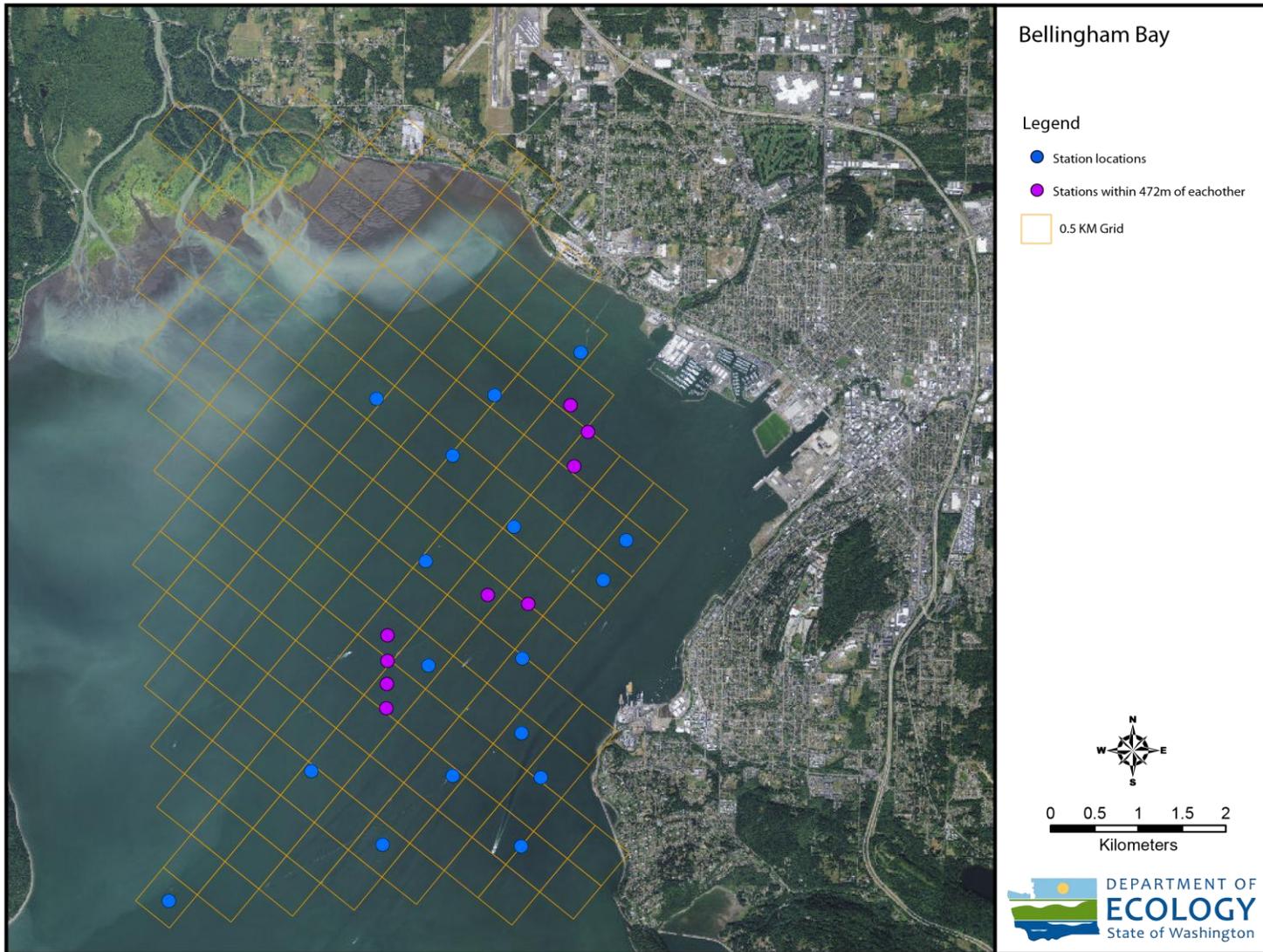


Figure 8. Map of Bellingham Bay case study site, showing locations of existing data.

Table A-1. Fidalgo Bay Case Study TEQ Data for Marine Sediments used as Regional Background

StudyID	LocationID	Study Location Name	FieldActivity		Latitude	Longitude	Study_Type	Location_Setting	TEQ Conc (pptr)
			StartDate	SampleID					
FBCPDX48	FSID6858-PB-10	PADILLABAY-10	6/8/2010	SDS-PB-10	48.476283	-122.5225	SiteInvestigation	Intertidal	0.56
FBCPDX48	FSID6858-CT-05	CLAMCOLLECTIONSITE-05	6/14/2010	SDS-CT-05	48.488004	-122.5969	SiteInvestigation	Intertidal	1.3
FIDALG08	FB-A3-42	FB-A3-42	9/4/2007	FB-A3-42	48.486639	-122.5956	InitialInvestigation	Estuary	3.4
FBCPDX48	FSID6858-PB-08	PADILLABAY-08	6/8/2010	SDS-PB-08	48.488383	-122.5295	SiteInvestigation	Intertidal	0.24
FBCPDX48	FSID6858-PB-09	PADILLABAY-09	6/8/2010	SDS-PB-09	48.48815	-122.5063	SiteInvestigation	Intertidal	0.13
FBCPDX48	FSID6858-FB-10	FIDALGOBAY-10	6/8/2010	SDS-FB-10	48.493153	-122.5844	SiteInvestigation	Intertidal	2
FIDALG08	FB-A3-41	FB-A3-41	9/4/2007	FB-A3-41	48.4897	-122.5766	InitialInvestigation	Intertidal	3.8
FBCPDX48	FSID6858-CPD-12	CUSTOMPLYWOODMILL-12	6/10/2010	SDS-CPD-12	48.495	-122.5918	SiteInvestigation	Intertidal	1.7
FIDALG08	FB-A2-38	FB-A2-38	8/30/2007	FB-A2-38	48.497	-122.5795	InitialInvestigation	Subtidal	1.9
FIDALG08	FB-A3-25	FB-A3-25	9/5/2007	FB-A3-25	48.4992	-122.5908	InitialInvestigation	Subtidal	1.9
FBCPDX48	FSID6858-FB-09	FIDALGOBAY-09	6/8/2010	SDS-FB-09	48.4994	-122.5851	SiteInvestigation	Intertidal	1.3
FIDALG08	FB-A2-35	FB-A2-35	8/30/2007	FB-A2-35	48.5006	-122.562	InitialInvestigation	Intertidal	0.72
FBCPDX48	FSID6858-PB-07	PADILLABAY-07	6/7/2010	SDS-PB-07	48.503367	-122.5356	SiteInvestigation	Intertidal	0.22
FBCPDX48	FSID6858-FB-07	FIDALGOBAY-07	6/8/2010	SDS-FB-07	48.504233	-122.5859	SiteInvestigation	Intertidal	0.79
FBCPDX48	FSID6858-FB-08	FIDALGOBAY-08	6/8/2010	SDS-FB-08	48.50565	-122.573	SiteInvestigation	Intertidal	0.67
FBCPDX48	FSID6858-FB-04	FIDALGOBAY-04	6/8/2010	SDS-FB-04	48.509783	-122.5935	SiteInvestigation	Intertidal	1.1
FBCPDX48	FSID6858-FB-05	FIDALGOBAY-05	6/8/2010	SDS-FB-05	48.509917	-122.586	SiteInvestigation	Intertidal	0.41
FBCPDX48	FSID6858-FB-06	FIDALGOBAY-06	6/8/2010	SDS-FB-06	48.510183	-122.5742	SiteInvestigation	Intertidal	0.33
FIDALG08	FB-A2-03	FB-A2-03	8/30/2007	FB-A2-03	48.5079	-122.5794	InitialInvestigation	Subtidal	1.8
FIDALG08	FB-A2-06	FB-A2-06	8/30/2007	FB-A2-06	48.5078	-122.5668	InitialInvestigation	Subtidal	1.4
FIDALG08	FB-A2-05	FB-A2-05	8/31/2007	FB-A2-05	48.5078	-122.5728	InitialInvestigation	Subtidal	2.7
FBCPDX48	FSID6858-PB-05	PADILLABAY-05	6/7/2010	SDS-PB-05	48.510167	-122.552	SiteInvestigation	Intertidal	0.57
FBCPDX48	FSID6858-PB-06	PADILLABAY-06	6/7/2010	SDS-PB-06	48.5112	-122.5197	SiteInvestigation	Intertidal	0.13
FBCPDX48	FSID6858-FB-02	FIDALGOBAY-02	6/8/2010	SDS-FB-02	48.515283	-122.5866	SiteInvestigation	Intertidal	0.51
FBCPDX48	FSID6858-FB-03	FIDALGOBAY-03	6/8/2010	SDS-FB-03	48.514567	-122.5746	SiteInvestigation	Intertidal	0.46
FBCPDX48	FSID6858-FB-01	FIDALGOBAY-01	6/8/2010	SDS-FB-01	48.516633	-122.5937	SiteInvestigation	Intertidal	0.31
FIDALG08	FB-A4-20	FB-A4-20	9/5/2007	FB-A4-20	48.5219	-122.6061	InitialInvestigation	Subtidal	1.4

Dioxin data downloaded from EIM. TEQs were calculated using TEFs from WAC Tables, found at: <http://apps.leg.wa.gov/wac/default.aspx?cite=173-340-900>. TEQs were calculated using substitution of non-detects at one-half the detection limit.

Table A-2. Port Gardner Case Study TEQ Data for Marine Sediments used as Regional Background

StudyID	LocationID	Study Location		FieldActivity		Latitude	Longitude	Location_Setting	TEQ Conc (pptr)
		Name	StartDate	SampleID					
DMMP_Dioxin_2005-07	DMMP-PGT15	PGT15	6/29/2006	PGT15-A		47.98630	-122.3020	SUBTIDAL	4.30
DMMP_Dioxin_2005-07	DMMP-PGT13	PGT13	6/29/2006	PGT13-A		47.98505	-122.2968	SUBTIDAL	4.20
DMMP_Dioxin_2005-07	DMMP-PGT11	PGT11	6/29/2006	PGT11-A		47.98392	-122.2921	SUBTIDAL	4.40
DMMP_Dioxin_2005-07	DMMP-PGP08_1	PGP08_1	6/29/2006	PGP08_10cm		47.98380	-122.2905	SUBTIDAL	3.90
DMMP_Dioxin_2005-07	DMMP-PGP07_1	PGP07_1	6/30/2006	PGP07_10cm		47.97562	-122.2885	SUBTIDAL	3.80
DMMP_Dioxin_2005-07	DMMP-PGP01_1	PGP01_1	6/30/2006	PGP01_10cm		47.98880	-122.2765	SUBTIDAL	5.00
DMMP_Dioxin_2005-07	DMMP-PGB01	PGB01	6/29/2006	PGB01_10cm		47.97192	-122.2728	SUBTIDAL	3.40
PortGardner_08	A1-46B	A1-46B	9/4/2008	A1-46B-S		47.95856	-122.2710	ESTUARY	0.18
DMMP_Dioxin_2005-07	DMMP-PGP09_1	PGP09_1	6/29/2006	PGP09_10cm		47.97679	-122.2686	SUBTIDAL	3.20
DMMP_Dioxin_2005-07	DMMP-PGB09_1	PGB09_1	6/30/2006	PGB09_10cm		47.98029	-122.2627	SUBTIDAL	3.00
PortGardner_08	A2-02	A2-02	9/4/2008	A2-02-S		48.00314	-122.2575	ESTUARY	0.18
KIMCLK04	KIMCLK04AKC-7	AKC-7	2/26/2004	AKC-7SD		47.97088	-122.2476	Subtidal	0.66
KIMCLK04	KIMCLK04AKC-3	AKC-3	2/26/2004	AKC-3SD		47.96972	-122.2461	Subtidal	0.28
KIMCLK04	KIMCLK04AKC-5	AKC-5	2/26/2004	AKC-5SD		47.97218	-122.2447	Subtidal	0.51
KIMCLK04	KIMCLK04AKC-6	AKC-6	2/26/2004	AKC-6SD		47.96814	-122.2447	Subtidal	1.20
KIMCLK04	KIMCLK04AKC-2	AKC-2	2/26/2004	AKC-2SD		47.97086	-122.2427	Subtidal	0.61
KIMCLK04	KIMCLK04AKC-1	AKC-1	2/26/2004	AKC-1SD		47.97131	-122.2408	Subtidal	0.42
KIMCLK04	KIMCLK04AKC-8	AKC-8	2/26/2004	AKC-8SD		47.97282	-122.2405	Subtidal	0.72
KIMCLK04	KIMCLK04AKC-4	AKC-4	2/26/2004	AKC-4SD		47.96955	-122.2403	Subtidal	0.21
PortGardner_08	A1-31B	A1-31B	9/4/2008	A1-31B-S		47.97136	-122.2346	ESTUARY	0.18
PortGardner_08	A2-08	A2-08	9/4/2008	A2-08-S		48.01088	-122.2340	ESTUARY	0.26
AODE6677	AO6677-462.1	A/H-SED-1	12/10/2010	SED-1		48.00310	-122.2184	Subtidal	2.55
AODE6677	AO6677-465	A/H-SED-4	12/10/2010	SED-4		48.00217	-122.2183	SUBTIDAL	2.10
PortGardner_08	A2-30	A2-30	9/12/2008	A2-30-S		48.01762	-122.1954	ESTUARY	0.42
PortGardner_08	A2-32	A2-32	9/4/2008	A2-32-S		48.00702	-122.1800	ESTUARY	0.16
PortGardner_08	A2-37B	A2-37B	9/4/2008	A2-37B-S		47.98164	-122.1710	ESTUARY	0.18

Dioxin data downloaded from EIM. TEQs were calculated using TEFs from WAC Tables, found at: <http://apps.leg.wa.gov/wac/default.aspx?cite=173-340-900>. TEQs were calculated using substitution of non-detects at one-half the detection limit.

Table A-3. Bellingham Bay Case Study TEQ Data for Marine Sediments used as Regional Background

StudyID	LocationID	Study Location Name	FieldActivity		Upper	Lower	Latitude	Longitude	Setting	TEQ Conc (pptr)
			StartDate	SampleID	Depth (cm)	Depth (cm)				
DMMP_Dioxin_2005-07	DMMP-BBB04	BBB04	7/19/2007	BBB04	0	10	48.6998	-122.5846	Subtidal	4.3 J
DMMP_Dioxin_2005-07	DMMP-BBP04	BBP04	7/19/2007	BBP04	0	10	48.7137	-122.5631	Subtidal	5.2 J
DMMP_Dioxin_2005-07	DMMP-BBP01	BBP01	7/19/2007	BBP01	0	10	48.7062	-122.5517	Subtidal	5.5 J
DMMP_Dioxin_2005-07	DMMP-BBP02	BBP02	7/19/2007	BBP02	0	10	48.7136	-122.5411	Subtidal	8.5 J
BELSEDDF	BBDIOX-10	BBDIOX-10	6/9/2010	BBDIOX-10	0	12	48.7064	-122.5303	Subtidal	11 J
DMMP_Dioxin_2005-07	DMMP-BBT05	BBT05	7/20/2007	BBT05	0	10	48.7254	-122.5517	Subtidal	7.2 J
DMMP_Dioxin_2005-07	DMMP-BBP03	BBP03	7/19/2007	BBP03	0	10	48.7204	-122.5517	Subtidal	7 J
DMMP_Dioxin_2005-07	DMMP-BBT04	BBT04	7/19/2007	BBT04	0	10	48.7230	-122.5517	Subtidal	7 J
DMMP_Dioxin_2005-07	DMMP-BBT06	BBT06	7/20/2007	BBT06	0	10	48.7281	-122.5518	Subtidal	6.8 J
BELSEDDF	UWI 32	UWI 32	6/10/2010	UWI 32	0	12	48.7250	-122.5453	Subtidal	2.6 J
BELSEDDF	BBDIOX-11	BBDIOX-11	6/9/2010	BBDIOX-11	0	12	48.7182	-122.5307	Subtidal	6.7 J
DMMP_Dioxin_2005-07	DMMP-BBB02	BBB02	7/20/2007	BBB02	0	10	48.7136	-122.5275	Subtidal	10 J
BELSEDDF	BBDIOX-9	BBDIOX-9	6/10/2010	BBDIOX-9	0	12	48.7260	-122.5309	Subtidal	10 J
BELSEDDF	UWI 277	UWI 277	6/9/2010	UWI 277	0	12	48.7359	-122.5462	Subtidal	5.7 J
Bellingham Bay										
Bellinghambay08	HART17_BBDXSS05	Dioxin BBDx-SS-05	9/18/2008	BBDX-SS-05	0	12	48.7326	-122.5365	Subtidal	12
BELSEDDF	BBDIOX-1A	BBDIOX-1A	6/15/2010	BBDIOX-1A	0	12	48.7317	-122.5302	Subtidal	11 J
BELSEDDF	BBDIOX-3A	BBDIOX-3A	6/15/2010	BBDIOX-3A	0	12	48.7527	-122.5545	Subtidal	0.57 J
BELSEDDF	BBDIOX-4	BBDIOX-4	6/11/2010	BBDIOX-4	0	12	48.7470	-122.5425	Subtidal	1.7 J
BELSEDDF	BBDIOX-6	BBDIOX-6	6/11/2010	BBDIOX-6	0	12	48.7397	-122.5327	Subtidal	3 J
Bellingham Bay										
Bellinghambay08	HART17_BBDXSS04	Dioxin BBDx-SS-04	9/19/2008	BBDX-SS-04	0	12	48.7344	-122.5186	Subtidal	13 J
BELSEDDF	UWI 29	UWI 29	6/9/2010	UWI 29	0	11	48.7386	-122.5153	Subtidal	6 J
BELSEDDF	UWI 35	UWI 35	6/11/2010	UWI 35	0	10	48.7534	-122.5363	Subtidal	1.4 J
BELSEDDF	BBDIOX-5	BBDIOX-5	6/15/2010	BBDIOX-5	0	12	48.7462	-122.5236	Subtidal	1.6 J
DMMP O&M Squalicum	Sq-15	Sq-15	9/7/2010	Sq-15	0	55	48.7498	-122.5216		6.29 J
Bellingham Bay										
Bellinghambay08	HART17_BBDXSS01	Dioxin BBDx-SS-01	9/19/2008	BBDX-SS-01	0	12	48.7526	-122.5244	Subtidal	1.5
BELSEDDF	BBDIOX-2	BBDIOX-2	6/10/2010	BBDIOX-2	0	9	48.7581	-122.5231	Intertidal	0.7 J

Dioxin data downloaded from EIM. TEQs were calculated using TEFs from WAC Tables, found at: <http://apps.leg.wa.gov/wac/default.aspx?cite=173-340-900>.
TEQs were calculated using substitution of non-detects at one-half the detection limit.