

Snohomish River Estuary Dry Season TMDL Study - Phase II

Water Quality Model Confirmation and
Pollutant Loading Capacity Recommendations

June 1997

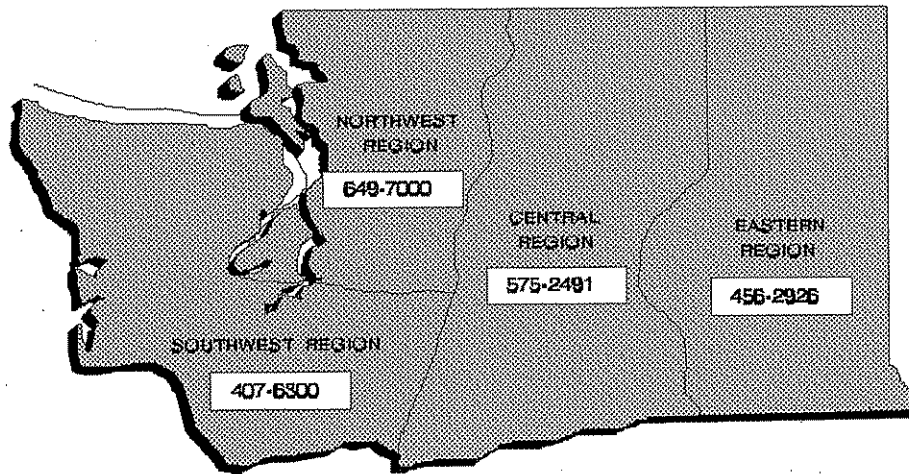
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Snohomish River Estuary Dry Season TMDL Study - Phase II

Water Quality Model Confirmation and
Pollutant Loading Capacity Recommendations

by
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Water Body Numbers
WA-07-0010, WA-07-1010, WA-07-1011, WA-07-1012,
WA-07-1015, WA-07-1020, WA-07-1030, WA-07-1050,
WA-07-1052, WA-07-1160, WA-07-1163

June 1997
Publication No. 97-325


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Abstract

A water quality modeling study of the Snohomish River Estuary system from Possession Sound to river mile 20 was conducted using an Environmental Protection Agency supported water quality simulation program (WASP5) model to simulate hydrodynamics and water quality. The WASP5 model was developed to assess the capacity of the estuary system to assimilate oxygen consuming pollutants from point and nonpoint sources. The model was calibrated using data collected in August and September 1993 and confirmed (verified) using data collected in August 1996. An interim report summarizing water quality in the study area and the development of the WASP5 model was published in 1995. The present document is the final report for this Total Maximum Daily Load (TMDL) study.

The water quality model predicted that natural conditions in a large portion of the estuary would be below dissolved oxygen standards under critical conditions. The model also predicted that point sources of oxygen-consuming pollutants would cause an exceedence of the 0.2 mg/L deficit allowed by the marine criteria when natural conditions are below the standard (the allowable deficit was also applied to the freshwater portion of the modeled area). The model was used to recommend waste load allocations (WLAs) for the following existing and proposed point sources of carbonaceous biochemical oxygen demand (CBOD) and ammonia BOD: the City of Snohomish Wastewater Treatment Plant (WWTP), Lake Steven's Sewer District WWTP, the City of Marysville WWTP, the City of Everett WWTP, and the proposed Smith Island WWTP. However, the recommended WLAs for ammonia are subject to changes based on mixing zone toxicity limitations that may be established in the future.

Acknowledgments

Many Ecology staff have provided valuable help on this project since it began in the spring of 1993. I would like to specifically thank the following: Ed Abbasi, Skip Albertson, Steve Butkus, Randy Coots, Mike Dawda, Lisa Eisner, Karol Erickson, Denis Erickson, Laura Fricke, Joe Jacobson, Will Kendra, Jan Newton, Deborah North, David Palazzi, Paul Pickett, Mary Rumpf, Dan Saul, Keith Seiders, Gerald Shervey, Dave Wright, and all the staff at Manchester Laboratory. In addition, I would like to thank Bill Fox, Cosmopolitan Engineering Group and Robert Waddle and other staff from the city of Everett, Public Works for their comments and help in collecting data. As always I want to acknowledge Barbara Tovrea for her work on word processing and report formatting. Finally, I would like to especially thank Greg Pelletier for assisting in field sampling and providing me with guidance and advice in developing the WASP5 model.

Introduction

Background Information

In June of 1993, the Watershed Assessments Section (WAS) distributed a proposal and quality assurance project plan (QAPP) for conducting a Snohomish River Estuary dry season total maximum daily load (TMDL) study (Cusimano, 1993). The project was requested by staff from Ecology's Northwest Regional Office (NWRO), because of their concern that population growth and development in the estuary watershed may cause adverse effects to water quality. Their major concern was with increased demands on the wastewater treatment plants (WWTPs) permitted to discharge to the river and sloughs.

Although the major focus of the TMDL study is on point sources of pollution, nonpoint sources are considered, but only as loads from tributaries entering the Snohomish River or sloughs. WAS is currently conducting a more detailed water quality assessment of the tributaries and their drainages to the Snohomish River. A final report of the tributary study is scheduled to be completed by September 15, 1997.

The Snohomish River Estuary TMDL study was initiated in the summer of 1993. Figure 1 shows the general location of the study area. Figure 2 shows the study area and the sampling locations, and Figure 3 is a more detailed map of the lower river, sloughs and near shore estuary within the study area. The first sampling survey was conducted on August 16 and 17, 1993. A second sampling survey, originally scheduled to be collected in the summer of 1994, was rescheduled to the summer of 1996 to adjust the project schedule to coincide with Ecology's newly adopted Basin Approach time frame. An interim report (Phase I) was published by Ecology in 1995 (Cusimano, 1995). The Phase I report summarized the 1993 sampling results, and the development and calibration of a water quality analysis simulation program or WASP5 (EPA, 1993) model of the area from Possession Sound to just downstream of the confluence with the Skykomish and Snoqualmie Rivers at River Mile 20.5. An analysis of the 1993 data presented in the Phase I report showed that impacts to the Skykomish River from the WWTPs at Monroe and Sultan would be minimal. In addition, a TMDL for biochemical oxygen demand in the Snoqualmie River was completed in 1994 (Joy, 1994). Therefore, the Skykomish and Snoqualmie Rivers were not included in the modeled area reported in the Phase I document, nor were they sampled during the 1996 survey.

This document is the final report for the TMDL study. The results and recommendations presented here will be used to help set permit limits for discharges to the river and estuary system beginning in 1999. The following topics are summarized:

1. Water quality data collected in 1996;
2. Data collected to improve the accuracy of the calibrated model reported in the Phase I report;

3. Re-calibration of the water quality model developed from data collected in 1993 and reported in the Phase I report;
4. Model verification (confirmation) using 1996 water quality data;
5. Modeling critical conditions; and
6. Waste load and load allocations based on critical condition water quality modeling predictions.

Further background information on problem definition; basin description; historical water quality; wastewater treatment plant descriptive information and effluent quality; 1993 sample collection and field measurement methods, and quality assurance/quality control; and hydrodynamic and water quality modeling development are reported in the QAPP (Cusimano, 1993) and Phase I report titled "Snohomish River Estuary Dry Season TMDL Study - Phase I Water Quality Model Calibration" (Cusimano 1995).

Project Goal and Objectives

The major goal of the Snohomish TMDL study was to develop waste load allocations (WLAs) and load allocations (LAs) for point and nonpoint sources of pollutants based on summer low river flow conditions. The specific objectives based on the project goal were as follows:

- Assess the potential for dissolved oxygen depletion from carbonaceous and nitrogenous biochemical oxygen demand (BOD) from known point sources;
- Identify other water quality variables which may need to be controlled (e.g., metals such as cadmium, copper, and mercury);
- Determine pollutant loading from point and nonpoint sources and background levels; and
- Determine loading capacity and recommend WLAs and LAs based on water quality modeling.

Sample Collection and Field Measurements Methods

An intensive survey was conducted on August 27 and 28, 1996, by four sampling teams to collect water quality samples at all major tributaries, selected river and slough stations, and WWTP effluents. Figure 2 shows the study area for both the 1993 calibration and 1996 confirmation surveys. The 1996 sampling network consisted of a subset of stations sampled during the 1993 survey: 16 stations in the river, sloughs, and estuary; 7 tributary stations; and 4 municipal WWTP effluents. In addition, Ecology's Ambient Monitoring Section (AMS) sampled at 7 stations along a transect from the lower river channel into Possession Sound (Figure 4). The data collection requirements for each of the following study components are discussed below: River, Sloughs, and Tributaries; Lower River and Possession Sound; and WWTP effluents.

The 1996 survey dates were selected to coincide with annual low river flow conditions and a tide sequence such that the low (minus tide) and high (approximately MHHW) occurred during the daylight hours. Appendix A contains graphs of the predicted Tulalip tides for the sampling dates. The Tulalip tides correspond to the WASP5 model seaward boundary. The 1996 survey conditions were similar to those sampled during 1993.

River, Sloughs, and Tributary Sampling:

Table 1 lists the parameters that were measured. Appendix B.1 contains the two day sampling schedules for the three sampling teams that sampled the river, sloughs, and tributaries. All stations were sampled at low tide, and at high tide from station 16 seaward to stations 34 and 35. All samples were collected from the center of the channel. At low tide, a single sample was collected from the surface (approximately 0.2 m below the surface at each station); while at high tide samples were collected from the surface and bottom (approximately one meter from the bottom).

Sampling teams 1 and 2 also collected surface and bottom conductivity measurements at three stations each between the low and high tide sampling periods: stations 21, 20, and 18 were sampled by team one; and stations 32, 31, and 23 by team two. In addition, teams one and two determined the extent of saltwater movement upstream at high tide in Ebey Slough and the mainstem of the river by taking surface and bottom conductivity measurements along a transect which saltwater was expected to reach.

Lower River and Possession Sound Sampling:

In conjunction with WAS, AMS collected vertical profile data with a CTD sensor at a series of stations extending from the lower river into Possession Sound (Figure 4). Parameters measured with the CTD were temperature, salinity, density, dissolved oxygen, fluorescence, and irradiance. The profiles were taken on August 27 and 28. Samples were collected and profiles taken during the later portion of the incoming tide and on the outgoing tide on each of the two days. At each

of the seven stations, a CTD profile was done and grab samples taken for dissolved oxygen, salinity, and chlorophyll α . At stations E1, E3, and E5 grab samples were collected for dissolved nitrate-nitrite, ammonia, and orthophosphate; at station E1 samples for total nitrogen and phosphorus were also collected. The AMS sampling team and team one both sampled at E7 (station 34).

WWTP Effluent Sampling:

Four WWTPs discharge treated wastewater within the 1996 study area (Figure 2): The City of Snohomish WWTP discharges to the mainstem at river mile (RM) 12, Lake Steven's Sewer District WWTP discharges to Ebey Slough at RM 6.5, the City of Marysville WWTP discharges to Steamboat Slough at RM 3.5, and the City of Everett WWTP discharges at two locations to the mainstem at RM 2.5 and RM 3.5. (Note: all RMs are approximate). Sampling of the WWTPs consisted of 24-hour composites as well as morning and afternoon grab samples of post-chlorination effluent from the five discharges. The type of samples and analyses conducted are listed in Appendix B.2.

Quality Assurance/Quality Control

Analytical procedures and associated precision or reporting limits are listed in Table 1. All samples for laboratory analysis were preserved as specified in Manchester Environmental Laboratory Lab Users Manual (MEL, 1994). Laboratory analyses were performed in accordance with MEL (1994). The data were reported by Ecology's Manchester Laboratory as usable with data qualifiers noted. (See the Phase I report for review of quality assurance/quality control for the data collected during the summer of 1993.)

Salinity/conductivity samples collected by WAS and AMS sampling teams in the lower river and sloughs and Possession Sound, were analyzed using an Autosal[®] salinometer by the Routine Chemistry Lab at the University of Washington.

Field sampling and measurement procedures used by WAS followed those specified in WAS (1993) for pH (Orion Model 250A meter and Triode™ pH electrode), conductivity (Beckman Model RB-5 and YSI33), dissolved oxygen (Winkler titration), and streamflow (Marsh-McBirney 2000). Temperature was measured using alcohol thermometers and YSI 33 probes. All meters were calibrated and post-calibrated according to manufacturer's instructions.

Field sampling procedures used by AMS followed those specified in Eisner *et al.*, 1994. The protocols include in-house and external CTD sensor calibrations, field-collected sensor verification samples (dissolved oxygen measurements were verified with the Winkler method), replicate samples, and routine cleaning and maintenance schedules. The current calibration coefficients were entered into the configuration file for use in processing the data with Sea-Bird Electronics CTD software. Data quality objectives and additional details on field and laboratory procedures can be found in Eisner *et al.* (1994) and Newton (1995).

Total variation or precision for field measurements and laboratory analysis was assessed by collecting replicate samples for approximately 10% of the total number of laboratory samples per parameter. Ten percent of field measurements for conductivity and dissolved oxygen were also replicated. Replicate precision for field measurements and laboratory analyses, calculated as the root mean square error of the coefficients of variation, are listed in Table 2. Replicate precision was acceptable for all variables given the analytical ranges measured.

Data Assessment Procedures

Data reduction, review, and reporting followed the procedures as outlined in MEL's Laboratory Users Manual (MEL, 1994). The principal investigator ensured that all data were validated before preparing a final project database by reviewing 100 percent of the data for possible errors.

Results and Discussion

Survey Data

The August 1996 and 1993 survey data results are listed in Appendix C. The major purpose of the 1996 survey was to collect data that could be used to confirm the WASP5 water quality model. Overall, the 1996 survey data were consistent with the 1993 survey data. Sampling results for the WWTP effluents are listed in Appendix C.3. (See the Phase I report for a summary of the data results collected during the summer 1993).

Data Collected to Improve Model Accuracy

In the Phase I report, five tasks were identified to be completed before confirming the water quality model results. The five tasks and the actions taken to complete each task are discussed below. In addition, other data and information collected after the Phase I report was published are discussed under number six:

1. Mixing characteristics of the estuary

The WASP5 model calibration was confirmed by comparing observed and simulated results for the 1993 survey period (Cusimano 1995). However, it was noted in the Phase I report that a number of assumptions related to the mixing characteristics of the estuary with respect to dissolved oxygen, during low river flow, needed to be assessed. Specifically, the high values of sediment oxygen demand (SOD) needed to calibrate the water quality model possibly reflected inaccurate assumptions about the mixing of marine water from Possession Sound with fresh river water in the lower river and sloughs. For example, since the maximum depth of the river and sloughs at high tide are less than 10 meters, the seaward (downstream) boundary concentrations for salinity and dissolved oxygen were assumed to be represented by the average of 0-10 meter profile data from Possession Sound. Because dissolved oxygen concentrations found in the river and sloughs were lower than the profile average concentration used for the downstream boundary, high model segment SOD levels were set during model calibration to adjust the model predicted dissolved oxygen concentrations to those measured.

In order to better define the mixing characteristics of the estuary during low river flow, CTD profile surveys of temperature, salinity, density, dissolved oxygen, fluorescence, and underwater irradiance were conducted during September 1995 and August 1996. Profiles were taken along a series of stations during low and high tide from the lower river into Possession Sound (Figure 4). Appendix D contains figures depicting contour plots of the 1996 profile data (1995 profiles are similar to the August 28, 1996 profiles). River mile zero

on the figures is set at Prieston Point. The results of the profile surveys show that during the flood tide, the lower river is strongly influenced by some portion of the marine water column layer in Possession Sound between 0-16.5 meters. This suggests that water with lower dissolved oxygen concentrations than originally assumed (*i.e.*, deeper water), can affect the dissolved oxygen concentration in the river and sloughs. (See WASP5 Model Re-calibration and Confirmation Sections for a further discussion of downstream boundary conditions.)

2. Model Geometry Characteristics

As reviewed in the Phase I report, an existing hydrodynamic model developed by Stein *et al.* (1991) was updated and used to simulate segment hydrodynamics. All of the model channel widths and depths (junction bottom elevations) were reviewed and modified during the original updating of the hydrodynamic model. River and slough channels were modified based on orthophoto maps and 63 surveyed cross-sections collected by Snohomish County Public Works (Snohomish County, 1989). In addition, the junction bottom elevations in the model segments representing Possession Sound were set at -5 meters MLLW based on the assumption that estuary freshwater and saltwater mixing would be confined to the upper layer of water. Since channel lengths were not reviewed and errors in model geometry can significantly affect model predictions due to differences in hydrodynamic characteristics, it was recommended in the Phase I report to re-evaluate all model geometry.

The hydrodynamic model geometry was reviewed in 1996. Assessing the model geometry was done by overlaying the model physical characteristics on USGS 1:24,000 scale maps, then comparing lengths on corresponding DNR 1:12,000 scale orthophoto maps, and NOAA 1:10,000 scale (for the lower river and Port Gardner) or 1:40,000 scale (for sloughs) navigational charts. Most of the channel lengths only varied 1-3% from the original value, but a few of the channels in Ebey Slough were found to be as much as 35% too long. The major changes made to the channel lengths were the combined reductions in the overall lengths of channels 97, 98, and 99 of 1,550 meters; and an increase in the length of channel 85 of 220 meters. Model channel widths and most junction depths were found to be accurate within 1-2% percent of those used in the model calibration input file listed in Appendix E of the 1995 Phase I report. However, a number of junction depths representing the area seaward of the mouth of Steamboat Slough (e.g., junctions 10, 11, 16, 17, 18, 22, 24, and 31) were reduced from -5 meters to more closely represent their actual depths. Even though some of the areas represented by the model in Possession Sound are exposed at low tide, the hydrodynamic model requires water to be in a channel at all times. Thus, tidal flat segment depths must be deeper than the lowest low tide. Also, in order to maintain model stability at the lowest low tide of approximately -0.4 meter, the minimum depth must be about 0.5 meter.

3. Critical Conditions for Point Sources of Pollution

A set of critical conditions based on defining a critical period of July through October (summer), were listed in the Phase I report. It was noted that critical conditions for point source loads should be reviewed by the permit manager(s) to make sure that critical design flows, concentrations, and loads are consistent with how future permit limits would be

derived. The NWRO permit manager, David Wright, has reviewed the critical conditions listed in the Phase I report and provided recommended changes. The most significant changes were increases in the design flows for the Marysville, Lake Steven's, and Snohomish WWTPs. (See Modeling Critical Conditions Table 7 for critical values).

4. Mixing Zone Studies

Another task identified in the Phase I report was to incorporate the results of mixing zone studies for Everett, Marysville, and Lake Steven's WWTPs, with respect to any water quality-based ammonia limits, into final dissolved oxygen model predictions. However, these studies are on-going and final results are not available. Consequently, all critical condition modeling uses effluent ammonia concentrations that could exceed final mixing zone permit limits.

5. 1996 Ambient and Effluent Data

The 1996 water quality survey was conducted on August 27 and 28, 1996. The data collected during the survey, and the 1996 profile data discussed above are used in this report to verify model predictions. However, one of the data tasks not accomplished was to establish discharge specific effluent ultimate CBOD/BOD₅ ratios. Effluent BOD samples were collected during the 1996 August survey. Laboratory ultimate CBOD test were also conducted in the summer of 1993 and 1995, but they were not successful. The results of the 1996 ultimate CBOD/BOD₅ analyses gave a range of ratios from 1.94 for Marysville's effluent to 5.14 for Lake Steven's effluent. Although the data were considered usable, they were not used because of possible problems with the nitrogen corrections used in the analytical method. Instead, ratios of 1.5 for Everett's mechanical (trickling filter treatment) plant and 2.0 for the other WWTP lagoon systems (secondary treatment) were used (Lund, 1993).

6. Other Data and Information Collected.

The Tulalip Landfill Superfund Site located on Ebey Island just west of Interstate 5 (see Figure 2) was identified as an additional loading source of ammonia, and added to the water quality model at segments 50 and 60 based on an estimated discharge rate of 2,971 lbs/year (EPA, 1996).

During the summer of 1995, the YSI 33 meters used to collect conductivity and salinity data in the river and sloughs were found to have a negative bias between 5-15%, depending on the meter and concentration range measured. During the 1996 confirmation survey, in addition to *in situ* measurements of conductivity/salinity, grab samples were collected and sent to the University of Washington, Marine Chemistry Laboratory, School of Oceanography for salinity analysis. The results were used to develop regression relationships with the YSI 33 meters that were used in both the 1993 and 1996 surveys. The regression corrections were applied to the 1993 conductivity/salinity estimates and used in the model re-calibration. Appendix C.2 contains the 1993 survey data, including the measured and corrected salinity estimates.

WASP5 Model Re-calibration

As reviewed in the Phase I report, the WASP5 sub-programs DYNHYD5 and EUTRO5 were used, respectively, to simulate the hydrodynamics and eutrophication kinetics in the Snohomish River Estuary system. The DYNHYD5 and EUTRO5 model networks extend from an upstream boundary just below the confluence of the Skykomish and Snoqualmie Rivers, to six seaward boundaries in Possession Sound (Figures 5 and 6). A description of the contents of the specific model "Data Groups" for each sub-program are contained in the Phase I report, which includes a review of all parameters, constants, and coefficients used in the model.

The DYNHYD5 program was used to simulate segment hydrodynamics (e.g., velocities, volumes, depths, etc.) under changing tidal conditions and steady state river flow.

The EUTRO5 water quality model was used to simulate simple eutrophication kinetics in the river and estuary system that include interactions of the following variables: ammonia nitrogen, nitrate nitrogen, inorganic phosphorus, phytoplankton carbon, carbonaceous BOD, dissolved oxygen, organic nitrogen, and organic phosphorus.

The EUTRO5 spatial network corresponds to the DYNHYD5 network such that each EUTRO5 segment corresponds exactly to a hydrodynamic volume element, or junction, and each segment interface corresponds exactly to a hydrodynamic link, or channel. The hydrodynamic model has additional junctions outside the EUTRO5 model network. These DYNHYD5 junctions correspond to EUTRO5 boundaries and are denoted in Figure 6 with segment number "0."

The hydrodynamic and water quality models were modified and re-calibrated to the 1993 water quality data based on the information and data collected in 1995 and 1996 (reviewed in the section--Data Collected to Improve Model Accuracy). The first step to recalibrate the model was to reassess the dispersion coefficients that were determined by matching model-predicted salinity values to the August 1993 survey salinity data. The corrected salinity values were used to adjust the dispersion coefficients. (The dispersion coefficients are contained in the EUTRO5 input file listed in Appendix E.)

The second step in re-calibrating the water quality model was to adjust the downstream dissolved oxygen boundary concentrations and model segment SOD values. In the model presented in the Phase I report, dissolved oxygen concentrations at the downstream boundary represented the average of 0-10 meter profile data collected in Possession Sound with a CTD on August 24th, 1993. The transect profiles collected in the summer of 1995 and 1996 showed that the marine water most likely influencing water quality in the main river channel is represented by some portion of the water column between 0.0-16.5 meters in Possession Sound. However, since transect profiles were not collected in 1993 (only two profiles were collected in Possession Sound) it was not possible to identify the specific portion of the water column that moved into the river and sloughs during the 1993 sampling surveys. Instead, the dissolved oxygen (and salinity) data collected at stations 34 and 35 were used to estimate the boundary concentration. The final boundary concentration was set to the best model fit with the water quality model segments corresponding to these sampling stations (*i.e.*, segments 29 and 31).

Overall, most of the contents of the model data groups are the same as those reported in the Phase I report. The major changes in the model were lower dissolved oxygen concentrations at the seaward boundary (from 8.5 to 7.3 mg/L), and the subsequent reduction of SOD in all model segments from a range of 0-9.5 g/m²-day to 0-4.5 g/m²-day, with values exceeding 2.8 only in the lower portion of Ebey Slough. (Reference the DYNHYD5 and EUTRO5 Critical Conditions model input files in Appendix E for specific model conditions.)

The model re-calibration was assessed by comparing observed and simulated results for the 1993 survey period. Table 3 lists the root mean square error (RMSE) between the predictions and observations for salinity, dissolved oxygen, and ammonia for both the re-calibrated and the calibrated model reported in the Phase I report. The re-calibrated model RMSEs compare favorably to the values for the original calibration.

WASP5 Model Confirmation

Hydrodynamic Model

The DYNHYD5 model used to confirm the water quality model was identical to the re-calibrated model except for the specific 1993 and 1996 inflows and tides that represent each sampling period.

Water Quality Model

Data from the 1996 survey were used to confirm the re-calibrated EUTRO5 model following the procedures described in the Phase I report. The boundary concentrations and summary of flows and concentrations used to calculate loading are listed in Table 4. Establishing these and other model characteristics used in the verification are discussed below.

River Flow and Tides -- The Snohomish River upstream boundary and Pilchuck River inflows were set at the average flow for the two day survey reported from USGS stations 12150800 and 12155300, respectively. Station 12150800 is located on the Snohomish River near the city of Monroe at about RM 20, and 12155300 is located on the Pilchuck River near the city of Snohomish. Inflows for other tributaries were set at the two day average flow measured by the survey team. WWTP discharges measured during the Class II inspections were also included as inflows. Ground water inflows were set at the same values reported in the Phase I report.

Boundary and Tributary Concentrations -- Upstream (freshwater) boundary concentrations were set at the average concentration measured during the two day August 1996 survey. The downstream (marine) boundary concentrations, except for dissolved oxygen and salinity, were set at the average of the measurements made at station E1. As with the re-calibration, downstream boundary concentrations for dissolved oxygen and salinity were estimated from sampling stations inside the model network, stations E3 and 35, which correspond approximately to the water quality model segments 24 and 31. (See Modeling Critical Conditions for further discussion of downstream boundary concentrations and profile data.)

Effluent Discharge Characteristics -- WWTP effluent water quality variable concentrations were derived by Norm Glenn, WAS, from the effluent survey results summarized in Appendix C.3.

Model Segment Characteristics -- The confirmation EUTRO5 eutrophication kinetic constants listed in Table 5 were the same as those used in the re-calibrated model and listed in the Phase I report (see Phase I report for review on developing rate kinetics). Dispersion coefficients and SOD were also the same as the re-calibrated model. Segment temperatures and initial conditions for all water quality parameters were set based on the average of the 1996 two day survey data.

Model Confirmation Results

As with the re-calibration, model confirmation was assessed by comparing observed and simulated results for the 1996 survey period. Figures 7-13 are representations of salinity and dissolved oxygen at selected segments which compare confirmation survey data against model predictions over the two-day August 1996 sampling period. Changing water depths also are shown on the graphs.

Table 6 lists the overall RMSE between the model estimates and measured values for salinity, dissolved oxygen, chlorophyll α , ammonia, and total phosphorus. For comparison, Table 3 lists the RMSE for the calibrated model reported in the Phase I report and the re-calibrated model. The estimate of model variation represented by the RMSE for dissolved oxygen and ammonia demonstrates a good model fit, relative to the ambient data range measured in the modeled system. A "good model fit" was defined in the Phase I report with respect to three qualitative criteria: (1) did the variable concentration increase or decrease as expected? (e.g., dissolved oxygen increases during low tide and decreases during high tide); (2) how did the RMSE compare to other reported model and observed deviations? (e.g., Pelletier (1993, 1994, 1997) reported dissolved oxygen RMSEs for models representing the Puyallup, Spokane, and Colville Rivers of 0.20, 0.38, and 0.4 to 0.8 mg/L, respectively; and ammonia RMSEs for the Spokane and Colville Rivers of 0.01 and 0.01 to 0.05 mg/L, respectively); and (3) how does the RMSE compare to temporal variability and measurement variability? (e.g., *in situ* data suggest that diurnal dissolved oxygen changes upstream of tidal effects were at least 0.2-0.4 mg/L, and the RMSEs for ammonia were close to the reporting limit).

The confirmation model estimates for salinity are less accurate than the re-calibrated model results for values less than 3 ppt. This is likely due to limitations of the DYNHYD5 model. For example, DYNHYD5 treats each segment as a rectangle with a corresponding rectangular cross-section, however, the actual properties of the cross-sections should be a function of depth and resemble a trapezoid. In the Snohomish Model, widths of each channel are set to approximate the channel width at mid-tide. Consequently, the channels in the lower river and sloughs are 20-30% wider (estimated at Langus Park boat launch near Everett's WWTP) than they should be at low tide and low river flow. The model channels allow saltwater to remain in the lower reaches of the river and sloughs at low tide because there is not enough freshwater to displace the saltwater, given that the channel bottoms are below MLLW and the channel widths are set to mid-tide

values. The re-calibrated model was more accurate at low salinities because there was significantly more freshwater river flow during the 1993 calibration survey (2,990 cfs) than in the 1996 confirmation survey (1,840 cfs).

In addition to confirming the model with the 1996 WAS survey data, the model was also confirmed by simulating *in situ* continuous monitoring data of salinity and dissolved oxygen at three locations in the lower river, and by simulating the results of a dye study of the city of Everett's WWTP discharge. The continuous monitoring and dye study data were collected by the city of Everett, Public Works.

The continuous monitoring data were collected for 72 hours (beginning at 1600 hours on August 26 and ending on August 29) using three YSI PC6000 probes at the locations shown on Figure 3. Figures 14-16 are plots of the continuous salinity and dissolved oxygen data for August 27 and 28. The grab sample data collected by both the city of Everett and WAS are also plotted. The model results fit the grab sample data well and compare favorably to the timing and changes in the *in situ* data for both salinity and dissolved oxygen. The differences in the *in situ* data, grab sample data, and model estimates are likely due to the following: 1) the model is a vertically averaged value, while the *in situ* measurements are taken near the surface, which may account for the steepness of the changes between high and low tide (i.e., *in situ* data only reflect part of the water column during periods that the water column is partially stratified so changes occur more quickly, while the model provides vertically averaged estimates so changes are more gradual); 2) *in situ* dissolved oxygen measurements appear to have some downward drift over time. Plus, the *in situ* monitors appear to measure salinity high at higher values when compared to the grab sample data.

The dye study was conducted on August 14-18, 1995, to determine the degree of mixing and the steady-state maximum concentration or "reflux" of the WWTP's effluent in the Snohomish River estuary (Cosmopolitan Engineering Group, 1996). Dye was injected into Everett's lagoon and mechanical plant outfalls for 24 hours (one tidal cycle) beginning at 18:00 August 14 and ending at 18:00 on August 15, 1995. Dye was measured in the river for five days, including the initial dye injection period.

Figure 17 shows both the measured dye concentration and the model simulation of the dye concentration versus time at model segment 37. The measured dye concentration represents a ten-minute moving average concentration taken in the river midway between the two outfalls. (Model segment 37 most closely corresponds to the dye study sampling location.) The dye tracking is part of the USGS superposition method for determining the daily maximum accumulation of effluent (Hubbard and Stamper, 1972). The method ascribes, for this situation, that the maximum accumulation of effluent is equal to the initial peak concentration during the first tidal cycle, plus the peak concentrations for every other cycle (Cosmopolitan Engineering Group, 1996).

The graph shows that the model predicted peak dye concentrations are similar to the measured dye concentrations except for the largest measured peak. According to the investigator who conducted the dye study, the differences between the model predicted peak values and the measured values during day one could be due to incomplete mixing of the dye during the injection

period (personal communication, William Fox, Cosmopolitan Engineering Group). In addition, the total dye retained within the model segment, as represented by the area under the time-concentration curve, is nearly identical to the dye measured (*i.e.*, 0.00617 fraction-days for the model, and 0.00612 fraction-days for the measured dye).

Overall, the model confirmation studies indicate that the model is a reliable predictor of dissolved oxygen and ammonia concentrations in the river and sloughs, and can be used to assess the impact of BOD loading to these areas. However, because the downstream boundary for dissolved oxygen had to be estimated from data inside the modeled system in both the re-calibration and confirmation, the effective use of the model to predict water quality under critical conditions was limited to the area between one segment seaward of the water quality model segments used to set the downstream boundary dissolved oxygen concentrations, and the upstream boundary. This limits the model use for predicting water quality to the area between segments 28 (possibly segment 24) and 76 in the mainstem of the Snohomish River, and between segments 18 outside the mouth of the sloughs upstream to segment 76.

Waste Load Allocations for BOD and Ammonia

Seasonal Permit Periods

Two seasonal permit periods: summer, low river flow; and winter, high river flow were chosen to regulate dischargers use of the assimilative capacity of the estuary. However, WLAs were only estimated for the summer low river flow season, because the model developed for the estuary is not appropriate to estimate water quality under high river flow conditions. Until an appropriate model is developed, permit limits for the high river flow season should be based on technology and mixing zone limits. The following two seasons were selected:

- July-October (low river flow)
- November-June (high river flow)

Waterbody Classification, Dissolved Oxygen Standard, and Targets for WASP5 Modeling

The Snohomish River from the mouth upstream to the confluence with the Skykomish and Snoqualmie Rivers is currently classified as Freshwater Class A, which requires a dissolved oxygen water quality criterion of 8 mg/L. From the river mouth upstream to the southern tip of Ebey Island a special condition has been established for fecal coliform: fecal coliform organism levels shall neither exceed a geometric mean value of 200 cfu/100 mL nor have more than 10 percent of the samples obtained for calculating the mean value exceeding 400 cfu/100 mL. Inner Everett Harbor northeast of a line bearing 121 degrees true from approximately 47°59'5"N and 122°13'44W (approximately southwest corner of pier--Figure 3) is Marine Class B. Possession Sound (North of Mukilteo) is Marine Class A.

Estuarine systems are transitional environments between saltwater and freshwater habitats. In recognizing this, WAC 173-201A-060(2) states:

In brackish waters of estuaries, where the fresh and marine water quality criteria differ within the same classification, the criteria shall be interpolated on the basis of salinity; except that the marine water quality criteria shall apply for dissolved oxygen when salinity is one part per thousand (‰) or greater and fecal coliform organisms when salinity is ten parts per thousand or greater.

In addition to these criteria and salinity conditional modifications, the marine classification in WAC 173-201A allows dissolved oxygen levels to be degraded up to 0.2 mg/L by human-caused activities when natural conditions are near or below 6.0 mg/L.

In the Phase I report, WAS proposed using the vertically-averaged salinity of 1 ‰ at Mean High Water (MHHW = 11.1 feet) during low river flow to define the estuary's upstream boundary as a fixed point, then applying the marine or freshwater criteria upstream or downstream of the defined point. However, for this report the dissolved oxygen criteria based on the WAC language listed above was used, such that criteria were applied based only on the 1 ‰ salinity designation. The vertically-averaged salinity at any time during the tidal cycle was used to define the 1 ‰ condition. Consequently, the marine or freshwater dissolved oxygen criteria apply at different points in the estuary depending on the tidal conditions. The WAC language was chosen over the Phase I report proposed fixed point 1 ‰ salinity boundary because the quasi-dynamic modeling results are easier to interpret with respect to changing fresh and marine criteria, and the WAC language better represents the actual physical and chemical processes occurring in the estuary (*i.e.*, dissolved oxygen concentrations changing due to tidal changes that brings higher salinity, lower dissolved oxygen water into the river and sloughs). In this report, the 0.2 mg/L allowance for marine waters below the criteria was also applied to the freshwater portion of the modeled system.

To estimate allowable BOD loading, the dissolved oxygen standard was assumed to be met if the WASP5 model predicted dissolved oxygen to be greater than the fresh or marine standard based on the salinity requirement discussed above, or if the human-caused depletion was less than 0.2 mg/L when dissolved oxygen was predicted to be below the standard.

Modeling Critical Conditions

Critical conditions were evaluated in the model to estimate the potential effects of current and future waste loading to the estuary. Critical conditions are those possible physical, chemical, and biological characteristics of the receiving water and pollutant loading sources that can increase the adverse effects of a pollutant of concern (e.g., low river flow and increased temperature would increase the effect of a given BOD load). For the Snohomish River Estuary, the critical period was set to July through October (summer) based on a review of ambient data for the drainage by Ehinger (1993) presented in the QAPP. The ambient data showed that the probable combined occurrence of high water temperature, low dissolved oxygen, and low river flow would be restricted to these months.

Table 7 lists estimated critical conditions for the modeled system. The following is a summary of how the critical condition values for river flow (and tides), boundary and tributary concentrations, effluent discharge characteristics, and model segment characteristics were established. The DYNHYD5 and EUTRO5 critical condition input files are contained in Appendix E. The Data Groups and individual model characteristics (e.g., channel depth, dispersion coefficients, etc.) are noted.

River Flow and Tides -- Critical river flows for the period were estimated as the 7-day average low flow with a recurrence interval of once every 20 years (7Q20) (WAS, 1996). Flow distribution estimates and flow statistic calculations for the Snohomish River were made using WQHYDRO (Aroner, 1992). Table 8 displays estimated critical flows listed in the Phase I report for the mainstem of the Snohomish River and its tributaries, including the annual 7Q10 and winter 7Q20 (November through June) statistics for comparison to the summer 7Q20. Derivation of these flows is discussed in the Phase I report.

Although no discharge was observed from the "pumped" tributaries (*i.e.*, French Creek, the Marshland, Deadwater Slough, and Swan Trail Slough) during the August 1993 and 1996 surveys, an estimated discharge was included for the critical condition scenario. The average daily (July-October) pumping rate for French Creek was estimated by Rod Denherder of the U.S. Department of Agriculture, Renton, Washington, from pumping records for 1993 and 1994. Since no records were available for the other pumped waters, an average daily value was estimated from the relative size of these drainages compared to French Creek.

Boundary and Tributary Concentrations

For most water quality variables, critical conditions for the upstream and downstream boundary, July-October period, were based on the 90th percentile (or 10th for chlorophyll) of data from Ecology's ambient stations near the City of Snohomish and near Gedney Island (stations 07A090 and PSS019, respectively). Upstream and downstream boundary temperatures were set at 19.8° and 15.9°C. Tributary critical conditions were set to the highest values (lowest for dissolved oxygen and chlorophyll) of those measured during the August 1993 and 1996 surveys, except for the Pilchuck River, which were based on percentiles of data from Ecology's ambient station on the Pilchuck near the City of Snohomish (07B055). In addition, dissolved oxygen at the upstream boundary was set at saturation, and temperatures for all other tributaries were set at the upstream boundary value of 19.8°C.

The seaward boundary conditions for dissolved oxygen and salinity were based on the most extreme monthly profile data collected from 1990-1996 for the July-October period at PSS019, which corresponded to the profile data collected on September 20, 1993. The dissolved oxygen recorded on this date ranged from 8.3 mg/L at the surface to 5.6 mg/L at 16.5 meters. (The profile depth range of 0-16.5 meters was selected as the maximum depth range that could influence water quality in the river and sloughs, based on the transect profile data collected in 1995 and 1996.) The median of the 10-meter running average profile data from 0 to 16.5 meters was used to set the seaward boundary at 6.0 mg/L for dissolved oxygen and 29.5‰ for salinity. A maximum running average depth of 10 meters was selected based on the assumption that no more than 10 meters of the water column would influence water quality in the channeled river and sloughs that have maximum depths <10 meters (see profile contour plots in Appendix D).

Effluent Discharge Characteristics

Effluent critical condition characteristics for flow, ammonia, and CBOD₅ were provided by NWRO permit writers responsible for managing NPDES permits. The critical condition flows for the Everett WWTP discharges were set at the summer low flow season permit limits. The flows used for the other WWTPs were set at the maximum monthly values listed in their respective annual permits. All other effluent characteristics were set to the highest value measured during the August 1993 Class II inspections and the August 1996 effluent sampling survey.

Model Segment Characteristics

All model segment characteristics, except temperature, were those established during model calibration and re-calibration. Segment temperatures were based on interpolating the 90th percentile temperatures measured at the following Ecology historical ambient stations located within the model network: PSS005, PSS008, PSS009, PSS015, PSS016, PSS018, PSS020 and 07A090.

Dissolved Oxygen Model Predictions Under Critical Conditions

The cumulative water quality impacts of NPDES permit loading alternatives are reviewed in this section. Alternatives for NPDES discharge limits are based on two possible scenarios: 1) existing discharges to the estuary, and 2) existing and proposed discharges to the estuary.

Alternative Total BOD Discharge Scenarios

Scenario I: Existing Discharges

Figures 18-20 show the results of the dissolved oxygen model predictions for the existing discharges to the system under critical conditions. The graphs represent the predicted minimum and maximum dissolved oxygen concentrations, with and without loading sources (*i.e.*, no point sources of BOD and nonpoint sources set at estimated background conditions--ammonia = 0.005 mg/L and CBOD = 1.2 mg/L) for the mainstem of the Snohomish River, Steamboat Slough, and Ebey Slough. The insert graph in Figure 18 shows how predicted values from the 24-day model run for each model segment were used to construct the dissolved oxygen profiles for the series of segments corresponding to river miles shown on the X-axis. The profile presented in Figure 18 corresponds to the model segments 2 (seaward boundary), 8, 14, 24, plus all the segments representing the mainstem of the river (25, 27-30, 34-48, and 71-75) to the headwaters at segment 76 (Figure 6). The profile in Figure 19 starts at segment 6 (seaward boundary) and includes segments 12, 18, 31, 32, 49, 50, 54-58. Figure 20 also starts at segment 6 and includes segments 12, 18, 31, 33, 59-64, 56, and 65-70. The minimum dissolved oxygen profiles

correspond to the time around high slack tide, while the maximum concentrations correspond to ebbing (or low) tide conditions.

The critical time, when dissolved oxygen concentrations are the lowest, is around high slack tide. The dissolved oxygen profiles for the mainstem and sloughs show that the predicted minimum dissolved oxygen values without loading are below the marine criteria for part of the modeled system. At high tide the marine dissolved oxygen criterion of 6.0 mg/L would apply to all of the segments in the lower river and sloughs (i.e., the 1 ‰ marine condition is predicted to reach an upstream point between segments 43 and 46 for all of the high tides during the 24 day model period). In order to compare the magnitude of the difference between both high and low dissolved oxygen concentrations with and without loading, the difference between the maximum dissolved oxygen concentrations with and without loading sources, and the difference between the minimum dissolved oxygen concentrations with and without loading sources (as presented in Figures 18-20) were calculated as the predicted dissolved oxygen deficits for each model segment. Table 9 presents the numeric results for the predicted dissolved oxygen deficits at low ebbing and high slack tide, and the percentage of the total deficit attributable to the specific BOD loading sources for model segments with the largest deficits. For example, Everett's WWTP BOD loading accounts for 77 percent of the projected critical condition oxygen deficit at segment 36 around high slack tide (salinity >1 ‰).

In order to establish WLAs, the model was run with reduced total BOD loads until all segments below the marine criterion met the 6.0 mg/L criterion at high tide, or caused no more than a 0.2 mg/L deficit. Table 10 shows the results of three different point source load reduction alternatives on the model predicted deficits. The first (I) BOD reduction was made by setting all ammonia point source discharges to 2 mg/L. The results show that this level of treatment would not be adequate to meet the criteria at all segments. The second (II) BOD reduction was made by setting ammonia to 2 mg/L and reducing BOD₅ to 33 mg/L. These results show that this level of treatment would be adequate to meet the criteria at all segments. The third (III) BOD reduction was made by setting ammonia at 5 mg/L and reducing BOD₅ 27 mg/L. This level of treatment also allows all the model segments to meet the criteria. Reductions of ammonia to 2 and 5 mg/L correspond to a typical effluent quality after activated sludge/nitrification-denitrification in separate stage and single stage treatment, respectively (Metcalf and Eddy, 1991). The BOD₅ levels needed to attain compliance with the dissolved oxygen criteria are also attainable with these levels of treatment (Metcalf and Eddy, 1991).

It should be noted that segment 57 (located near the midpoint of Steamboat Slough, at the junction with upper Ebey Slough) was found to be the "critical" model segment in the system, or the segment which requires the most load reduction to meet the criteria. Table 9 shows that this segment receives significant loading from all the major sources in the modeled area. It is also important to note that the level of BOD reduction alternatives listed in Table 10 were made equally to all point sources without regard to the percent contributions at any one segment, or the total loading of any one source. Generally, each loading source has its greatest effect on segments near its discharge, however, contributions to many segments are predicted.

It should also be noted that the minimum dissolved oxygen concentration in segment 57 with and without BOD loading was predicted to be 5.61 and 6.05 mg/L, respectively. The allowable 0.2 mg/L deficit was applied to the predicted deficit for this segment such that the final dissolved oxygen with treatment was 5.85 mg/L, which is 0.15 mg/L below the marine criterion. A more strict interpretation of the marine dissolved oxygen criteria would be to allow only a 0.05 mg/L deficit at segment 57. However, the 0.2 mg/L deficit was allowed because the criteria states that natural oxygen levels near or below 6.0 mg/L may be degraded by up to 0.2 mg/L.

Scenario II: Existing and Proposed Discharges

Two new pulp and paper facilities, a de-inking and recycling facility, are being planned for construction on the site of the former Weyerhaeuser bleached Kraft pulp mill in Everett (Figure 3). A cogeneration facility is also planned to be built on the site to supply steam for the de-inking facility. Process wastewater will be generated from the facilities, and the current plan is to send the wastewater to the Smith Island Treatment Plant (SITP) for treatment, then discharge the treated effluent into Steamboat Slough just downstream from the Union Slough junction, into the area represented by water quality model segment 49 (CH2M Hill, 1996).

The treated wastewater is planned to be intermittently discharged through a new outfall diffuser. Based on an analysis of the tidal conditions in Steamboat Slough, the effluent is planned to be discharged twice daily during the ebbing tides starting 30 minutes after high slack and ending 1 hour before low slack for a minimum period of 8 hours per day. The average flow rate for the intermittent discharge is projected to be 32.7 mgd (12.6 as a continuous daily flow) with a BOD₅ concentration of 63 mg/L and total ammonia concentration of 2 mg/L (CH2M Hill, 1996; Per. Comm. Robert York, Environmental Engineer, CH2M Hill). An estimate of the ultimate BOD/BOD₅ ratio of 1.27 was made from data collected to characterize the effluent of Sonoco Products Company waste paperboard mill in Sumner, Washington (Pelletier, 1993).

Figure 21 is an example of the modeled SITP intermittent discharge periods for the first 1.5 days of the tidal sequence used for the critical conditions simulations. Table 11 lists the results of model predicted dissolved oxygen concentrations for load reduction strategy II listed in Table 10, plus the same level of treatment with the SITP BOD loading (IV). The results show that the SITP discharge would cause additional deficits in the segments representing Steamboat Slough (*i.e.*, segments 49, 50, 55, 56, and 57) and just seaward of the slough (*i.e.*, segments 31 and 32), ranging from 0.04-0.07 mg/L. An additional dissolved oxygen deficit of 0.06 mg/L is predicted at the critical model segment 57. In order to meet the allowable 0.2 mg/L deficit at all locations with the SITP BOD loading, all point source dischargers would have to be reduced to 2 mg/L of ammonia and 26 mg/L of BOD₅ (Table 11, V). These ammonia and BOD₅ levels should be attainable for the municipal WWTPs with separate stage activated sludge/nitrification denitrification, however, it is not known whether these levels can be achieved by the proposed new discharge from the SITP.

WLAs and LAs

Table 12 presents the allowable effluent limits or WLAs for BOD₅ and ammonia that meet the dissolved oxygen criteria as the maximum daily average concentrations (in mg/L) and loads (in pounds/day). At this time, **no** allocation has been made for future growth (e.g., new discharges, increased WWTP capacity). Also, if monthly limits are needed, then the statistical procedure described in the EPA Technical Support Document should be used to establish the maximum average monthly limit (EPA, 1991). The maximum average monthly limit calculations need to account for effluent variation and the number of samples per month used to monitor effluent quality. The nonpoint LA is set at existing conditions (*i.e.*, no reductions are proposed for nonpoint sources). Nonpoint source pollutant controls will be addressed in the tributary report scheduled to be completed in September 1997.

Uncertainty and Sensitivity Analysis

The WASP5 (DYNHYD5 and EUTRO5) model used to simulate the Snohomish River Estuary system is a one-dimensional quasi-dynamic model. The DYNHYD5 sub-program provides real time variations in tidal heights and velocities (and channel flows) to the EUTRO5 eutrophication program. Although EUTRO5 averages concentrations of water quality variables vertically and laterally in each segment, it does simulate dynamic changes in a number of state variables. The resultant complexity of the model is such that a complete uncertainty or sensitivity analysis would be a monumental undertaking. However, because the economic and environmental impact of this project are considerable, a partial but reasonable assessment of model certainty and sensitivity has been conducted.

The calibration and confirmation of the model results reported in this document and the Phase I report using the RMSE of the model predicted and observed values as an estimator of model uncertainty is an acceptable method of assessing model uncertainty (Reckhow *et al.*, 1986). For example, the confirmation of the model suggests that the vertically averaged dissolved oxygen and ammonia concentrations within each model segment can be predicted with an average error of 0.39 and .016 mg/L, respectively. The correlation coefficients for the model and predicted values of salinity, dissolved oxygen and ammonia for the 1993 data were 0.93, 0.94, and 0.78, respectively, and for the 1996 data they were 0.98, 0.93, and 0.72. The RMSEs and correlation coefficients suggest that the model is a good predictor of these parameters in the modeled area.

A complete sensitivity analysis for the model at only a small subset of the segments would require a factorial design which is not possible without analysis software designed to link with the model. Instead, a one variable-at-a-time approach for some of the major input parameters was used to show what the effect would be on dissolved oxygen concentrations from changes to some of the input parameters. Table 13 lists the results for a subset of representative water quality model segments. The values listed in Table 13 are equal to the percent change in the model predicted dissolved oxygen concentration after a 10 percent change in each of the model input parameters listed in the table. For example, a 10 percent change in river flow (upstream boundary input to the model) would cause a 1.9 percent change in dissolved oxygen at low tide, and a 0.9 percent change at high tide at segment 34. The sensitivity analysis indicates that for all segments the model results are most sensitive to temperature, and at high tide the segments in the lower river and sloughs are most sensitive to the downstream (or seaward) boundary dissolved oxygen concentration.

Conclusions and Recommendations

- A one-dimensional quasi-dynamic water quality and quantity model of the Snohomish River Estuary was developed using WASP5. The model was calibrated and confirmed using data collected during August of 1993 and 1996, respectively. An uncertainty analysis of the differences between predicted and observed values for ammonia and dissolved oxygen indicate the model is an acceptable predictor for these variables. The model was developed to evaluate the capacity of the river to assimilate waste loads from point and nonpoint sources of BOD. The model was also developed to recommend WLAs for carbonaceous and nitrogenous BOD that would not violate the dissolved oxygen criteria.
- The WASP5 model of dissolved oxygen under critical conditions indicates that natural conditions in a large portion of the lower river and estuary would be below the Marine Class A dissolved oxygen criteria when salinity in the estuary exceeds 1 ‰. To meet the allowable anthropogenically-caused dissolved oxygen deficit of 0.2 mg/L, WLAs are recommended for the following point sources of ammonia and carbonaceous BOD: the City of Everett WWTP (two discharges), the City of Marysville WWTP, the Lake Steven's Sewer District WWTP, and the City of Snohomish WWTP. In addition, a WLA was determined for a proposed new discharge by Snohomish River Pulp Company. However, accommodating the new discharge would require more restrictive WLAs for all the existing dischargers.
- No reserve for future growth or water quality protection was determined. However, an allocation for a reserve could be considered during the public review period for this TMDL.
- Data analysis presented in the Phase I report showed that under critical conditions, the City of Monroe and Sultan WWTPs discharging to the Skykomish River would only have a small effect on dissolved oxygen concentrations. Permit limits for these discharges were recommended to be based on technology and mixing zone evaluations only. However, these WWTP discharges should be reassessed if increases in population are projected to significantly change the assumed critical effluent flows and concentrations used in the reported analyses.
- Mixing zone analysis should be completed for all permitted discharges to the river estuary system, and the results incorporated into the TMDL study and recommended WLAs.
- Copper was identified as a possible problem pollutant in the Class II inspections reported in the Phase I report. Measurable dissolved and total recoverable copper concentrations were also found in the Snohomish River and Ebey Slough during the 1993 metals screening survey reported in the Phase I report. Measurable concentrations of dissolved and total recoverable copper have also been found by AMS at station 07A090 located in the Snohomish River near the city of Snohomish. It is recommended that a complete review of existing copper data and an intensive survey of the Snohomish River Estuary be conducted to assess if a copper TMDL and associated WLAs and LAs need to be established.

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Figures

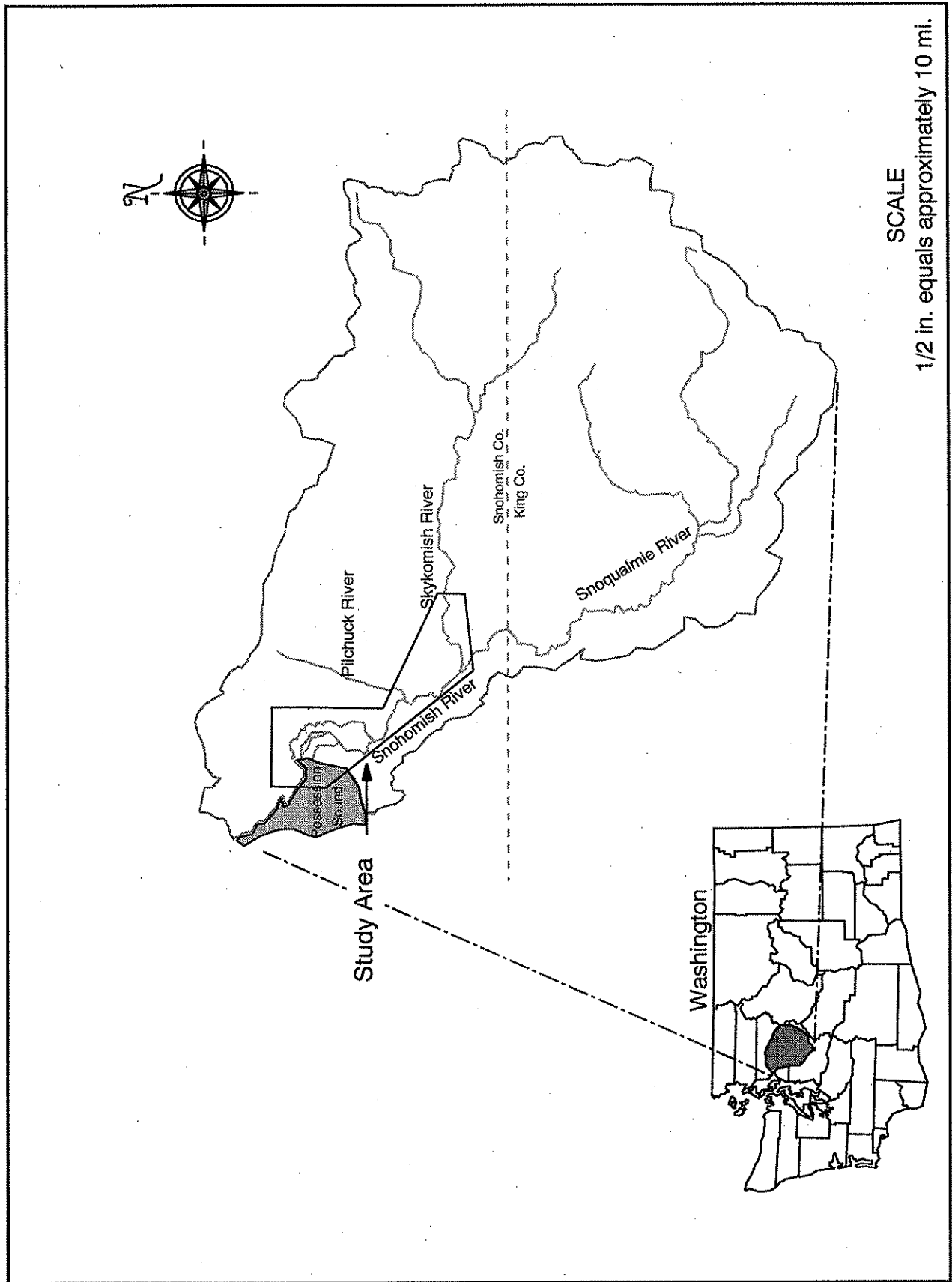


Figure 1. Snohomish River Basin and TMDL study area location map.

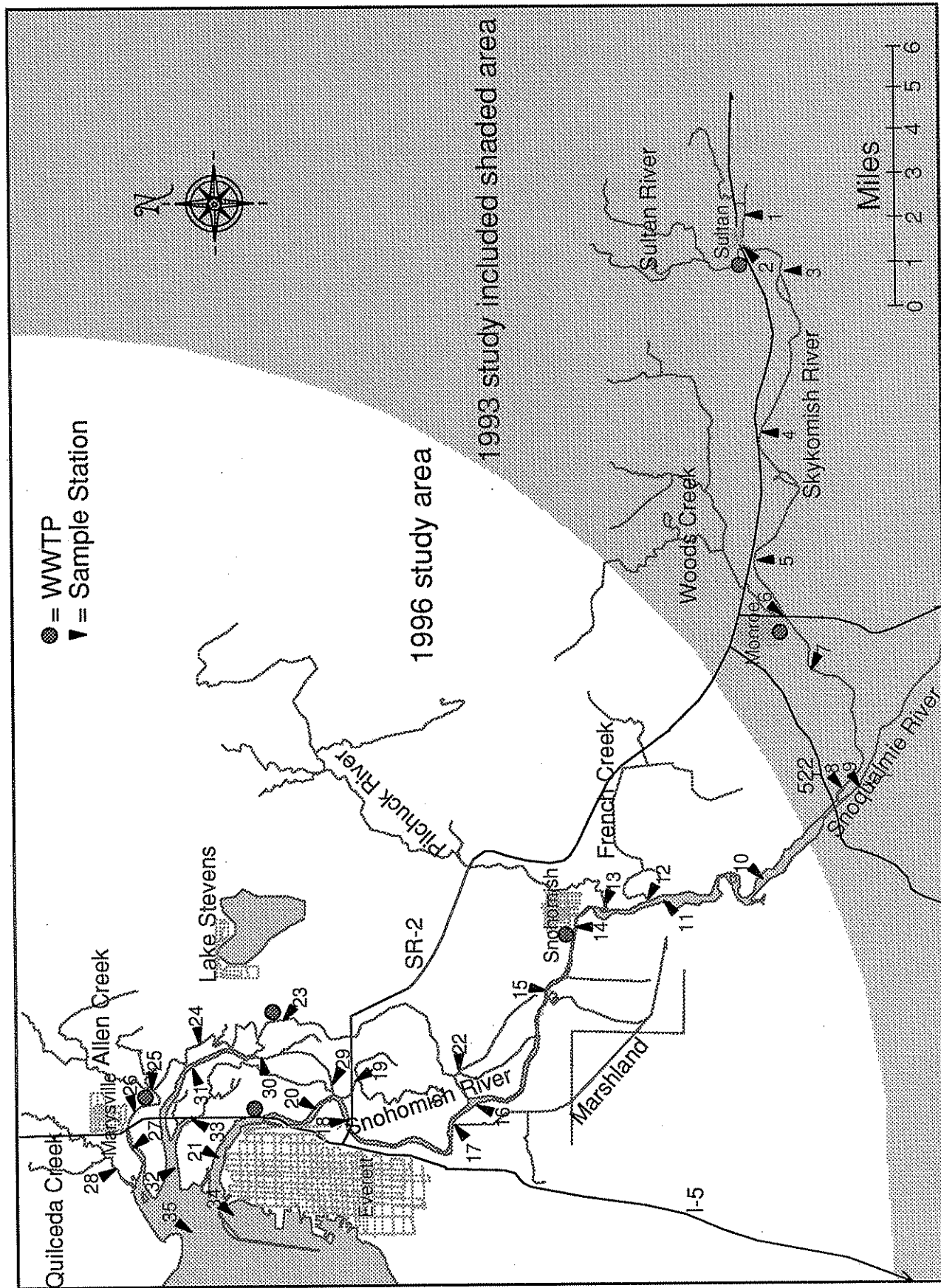


Figure 2. Study area map with sampling stations and wastewater treatment plants (WWTPs) annotated.

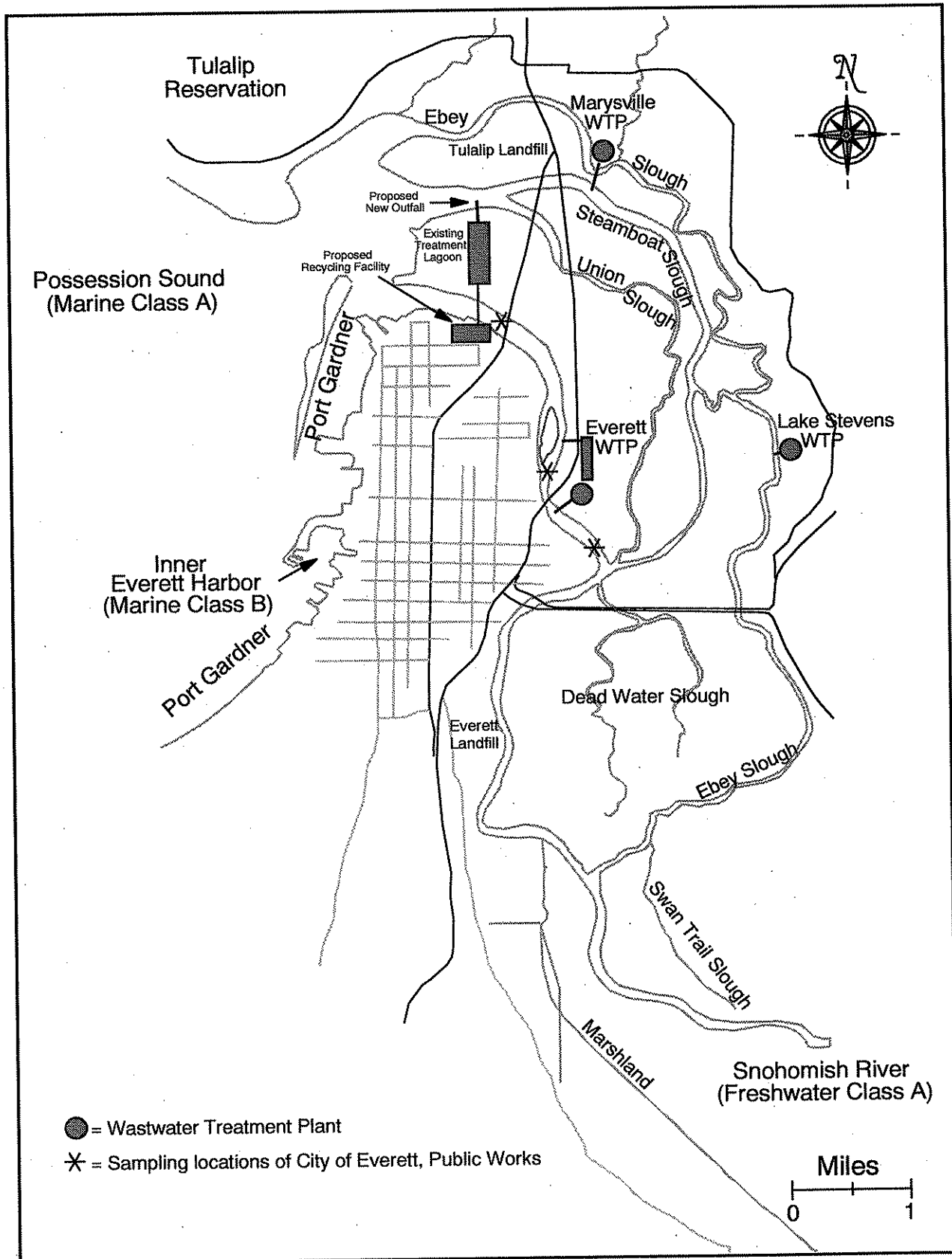


Figure 3. Map of sloughs and near shore estuary within study area.

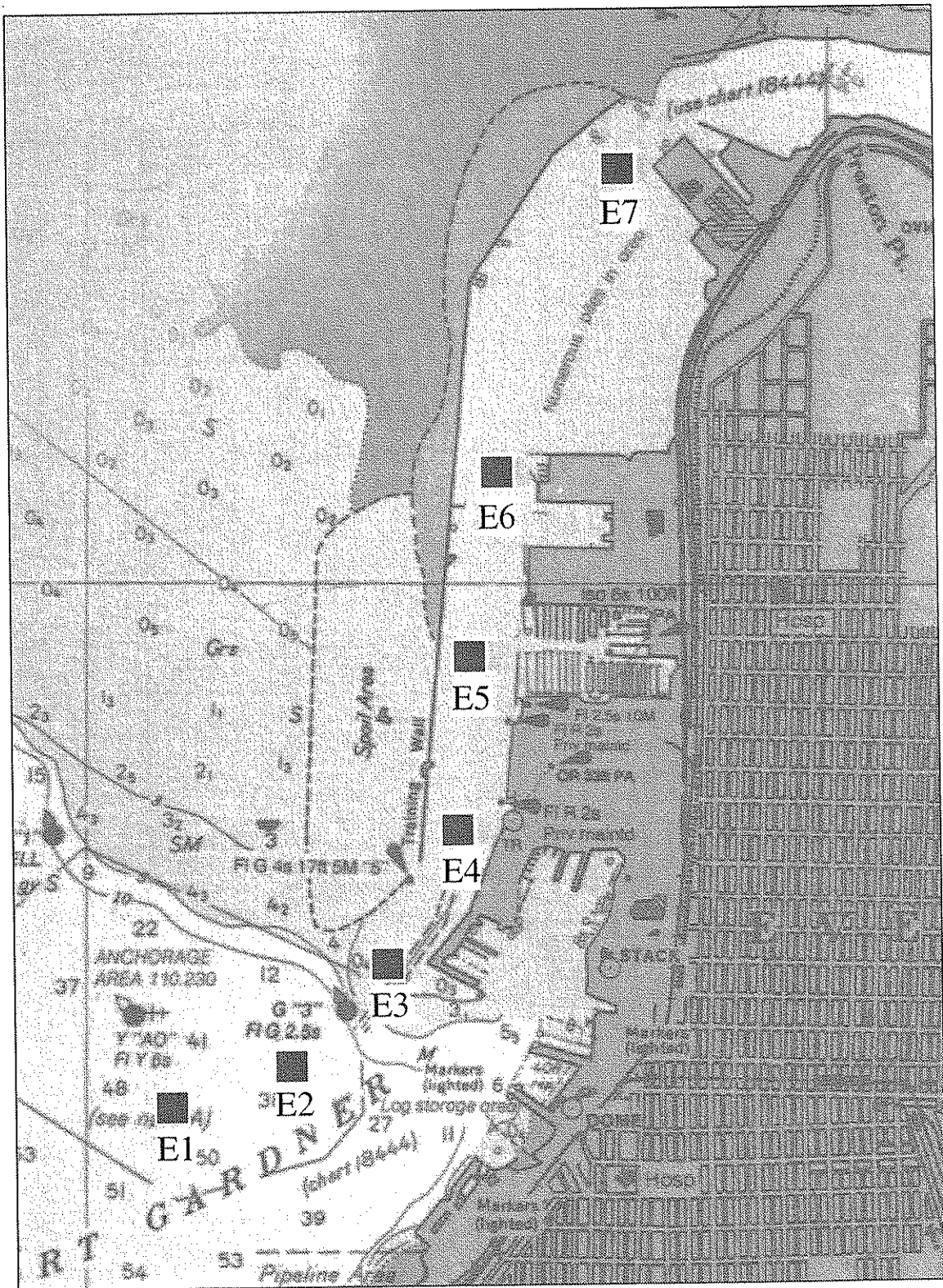


Figure 4. Sampling stations in the lower river and Possession Sound.

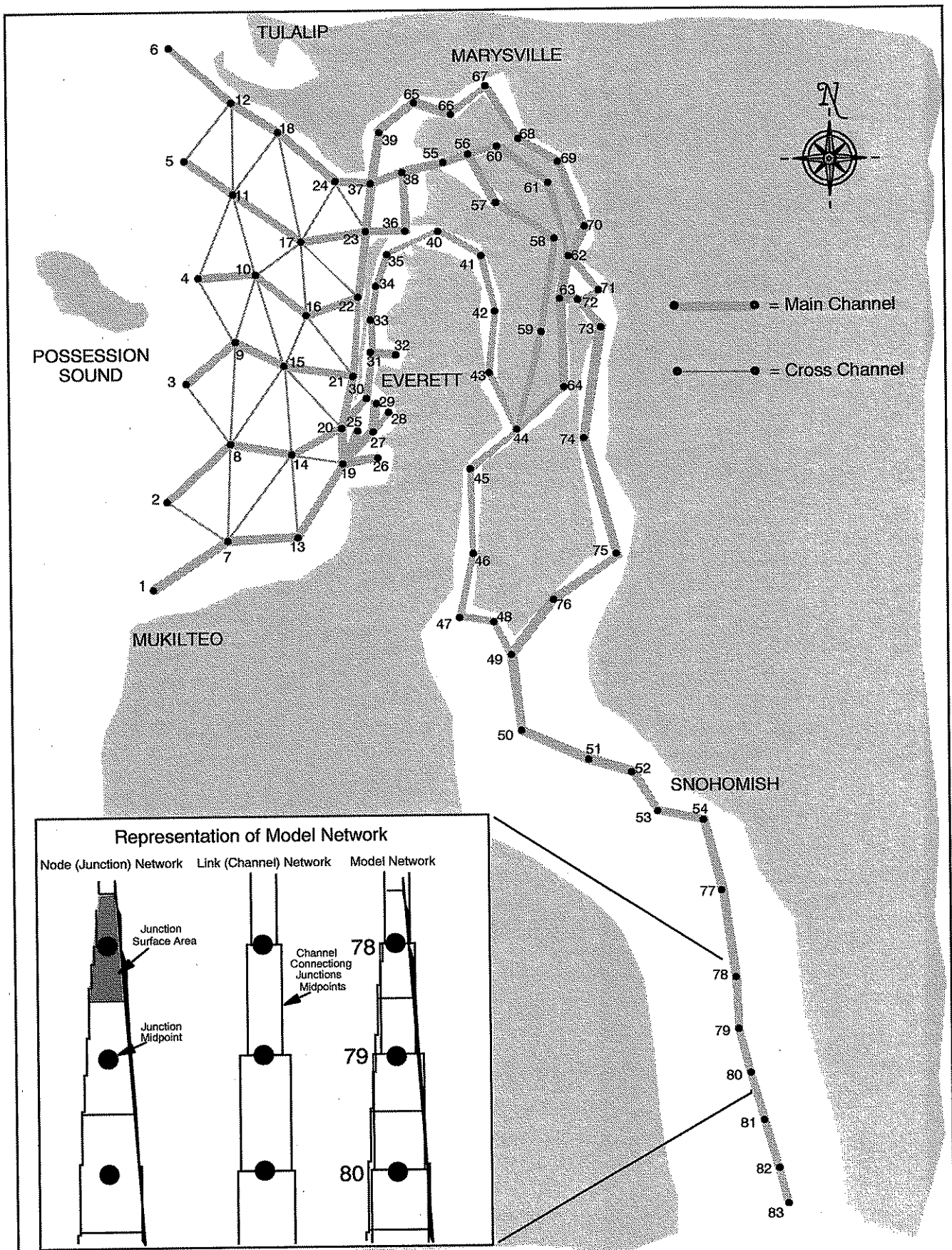


Figure 5. Link-Node network for the Snohomish River Estuary hydrodynamic model.

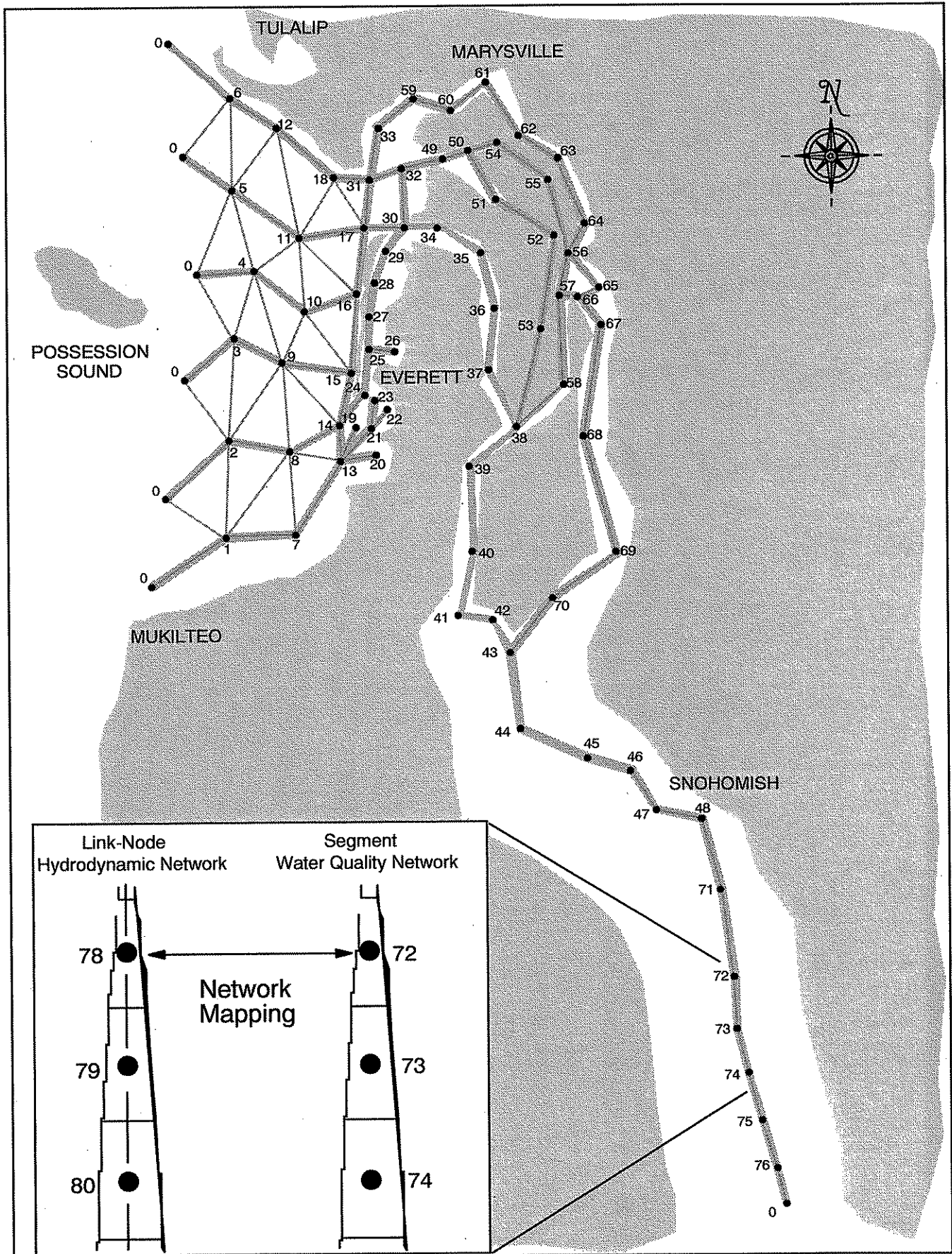


Figure 6. Segment network for the Snohomish River Estuary water quality model.

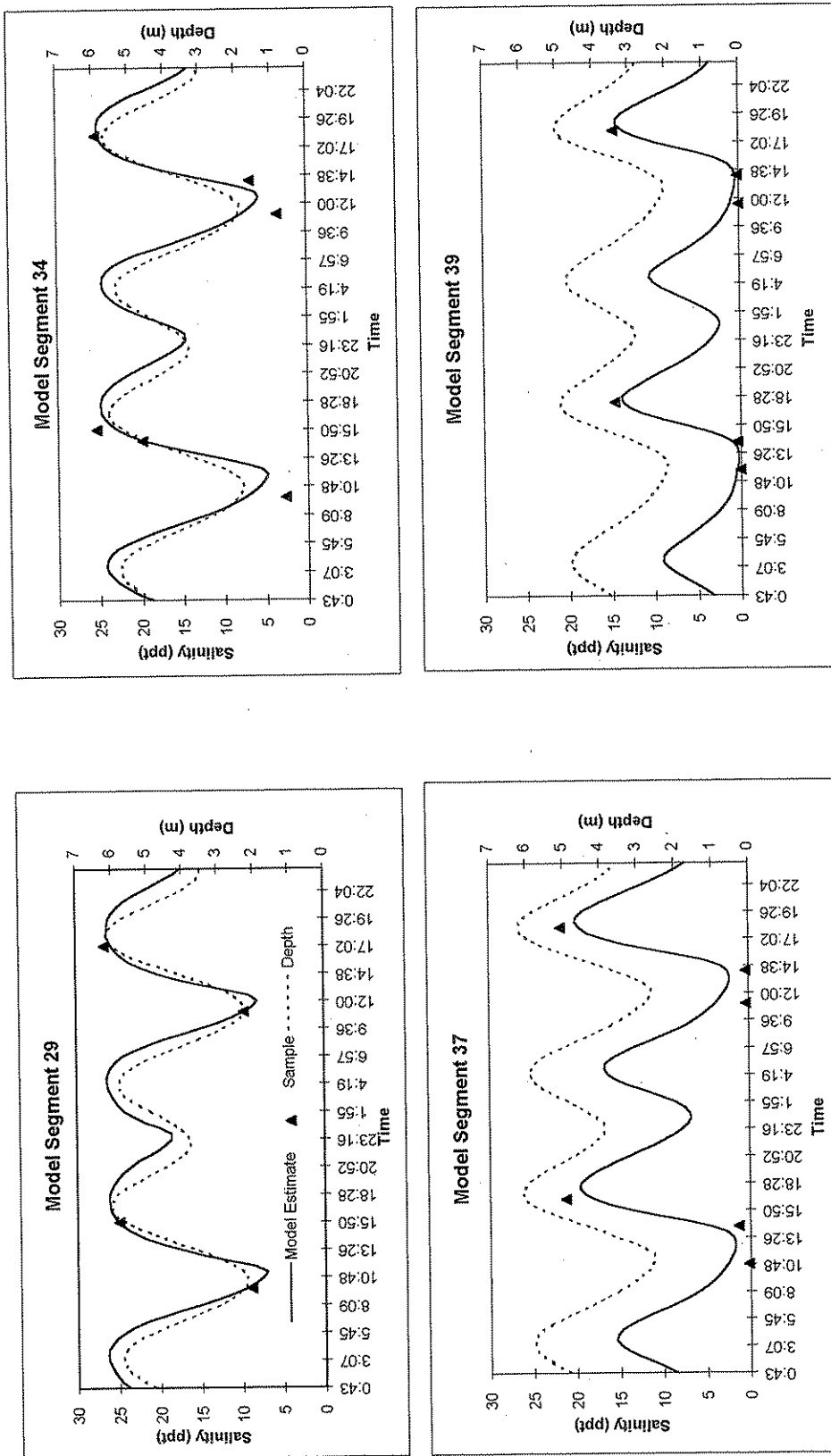


Figure 7. Verification of Eutro5 to predict salinity in the Snohomish River Estuary.

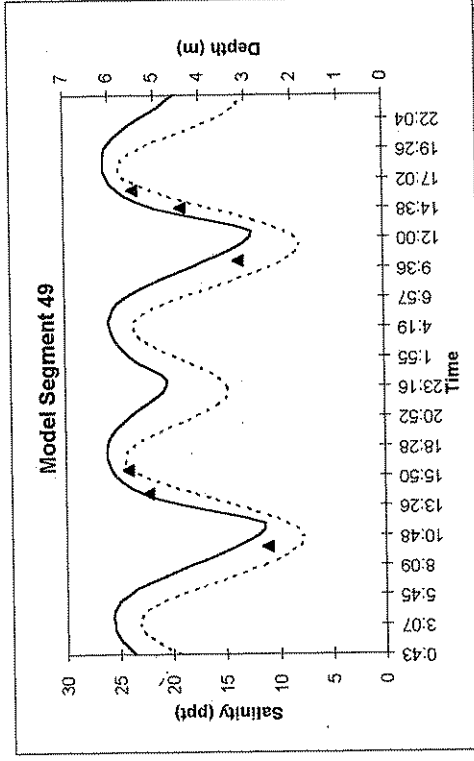
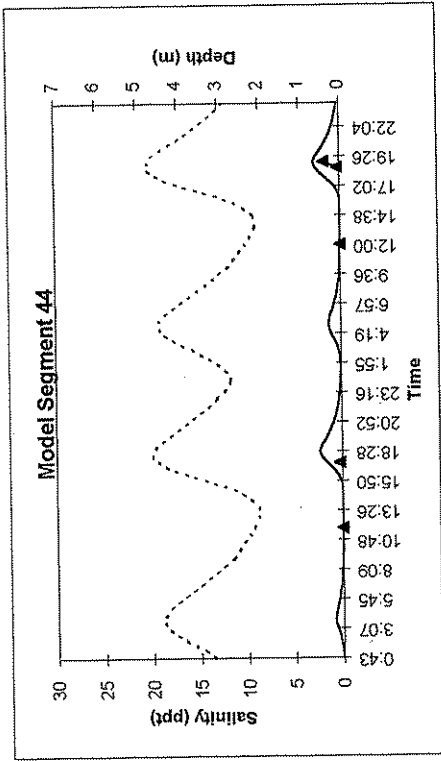
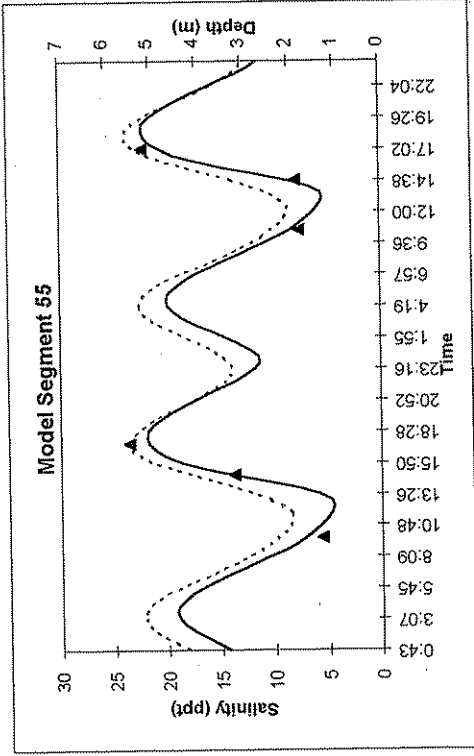
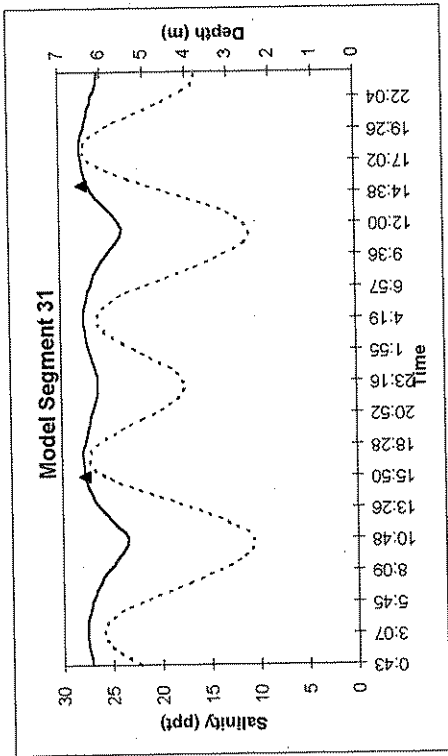


Figure 8. Verification of Eutro5 to predict salinity in the Snohomish River Estuary.

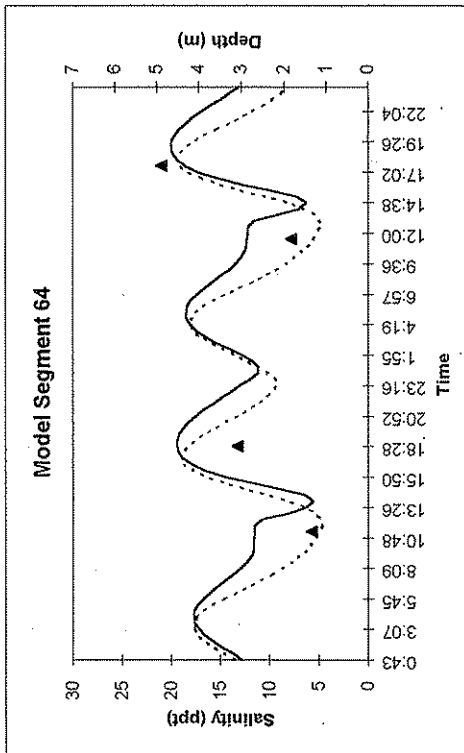
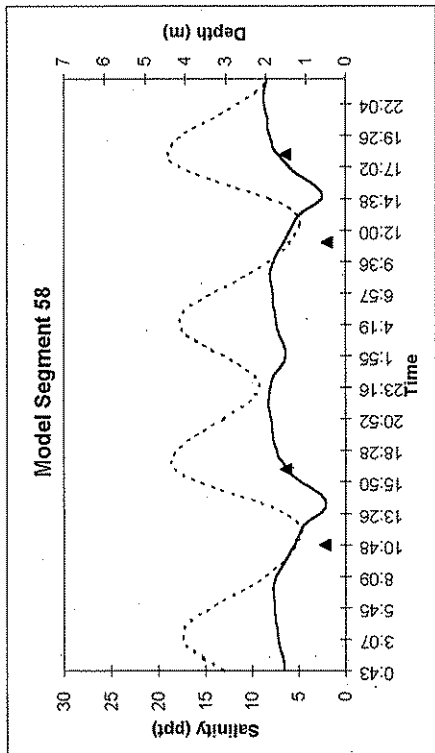
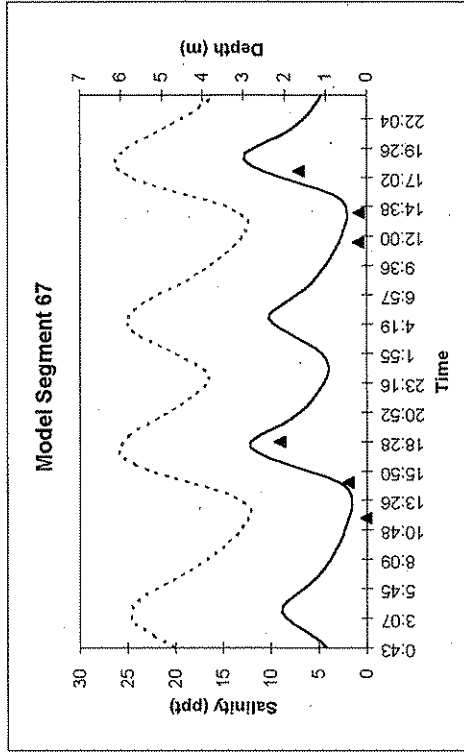
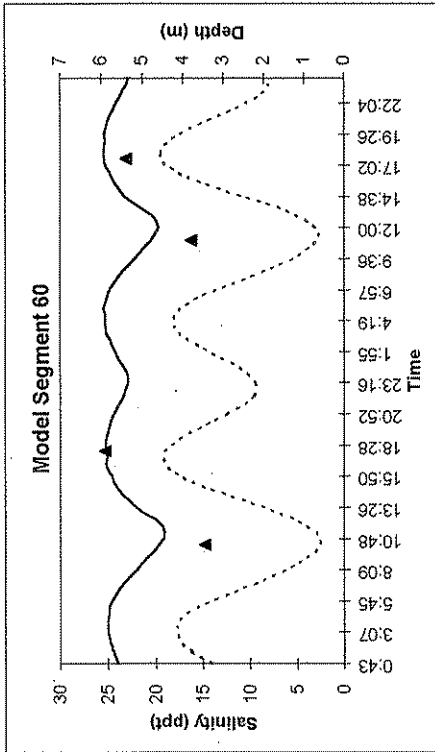


Figure 9. Verification of Eutro5 to predict salinity in the Snohomish River Estuary.

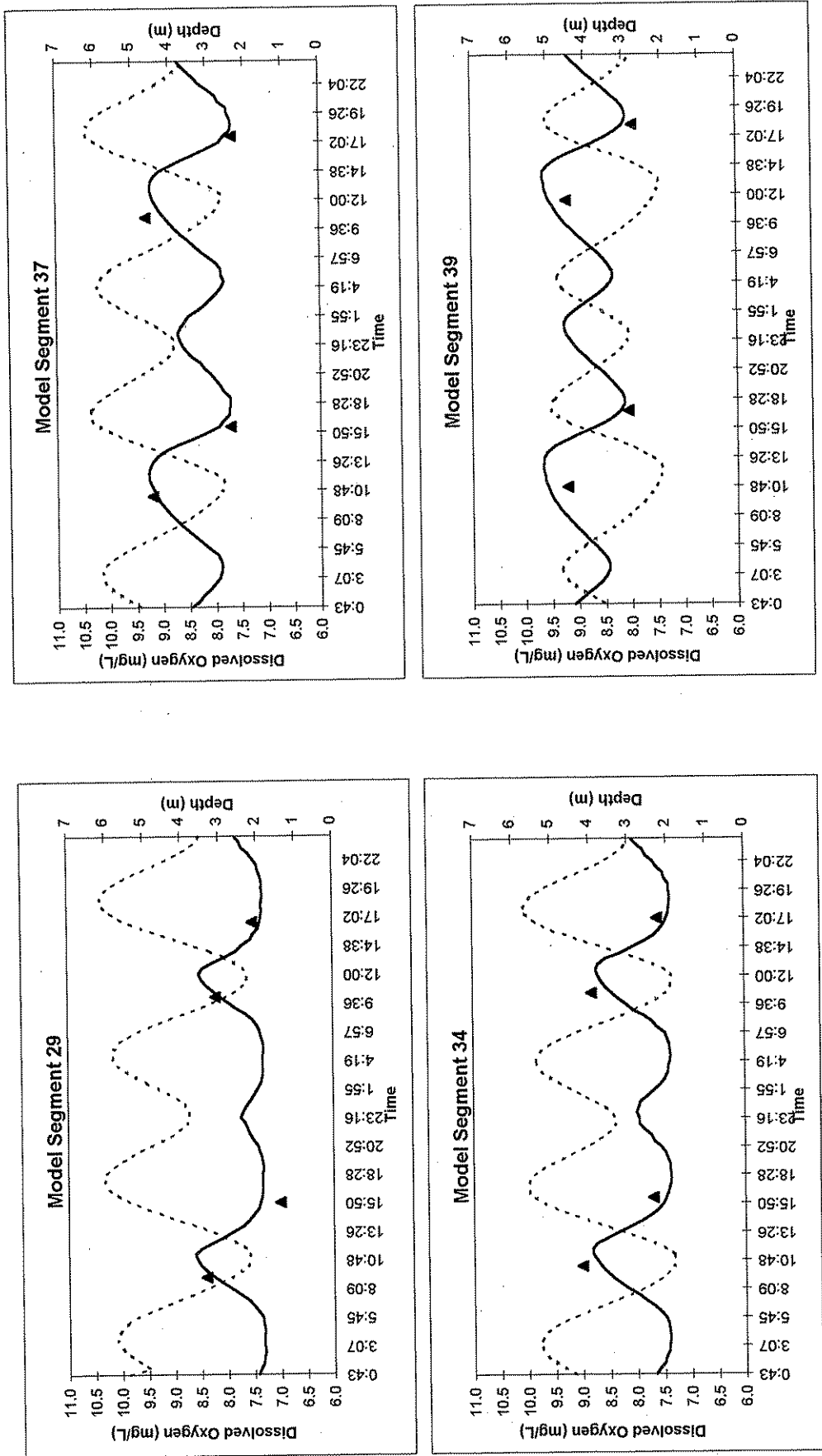


Figure 10. Verification of EUTRO5 to predict dissolved oxygen in the Snohomish River Estuary

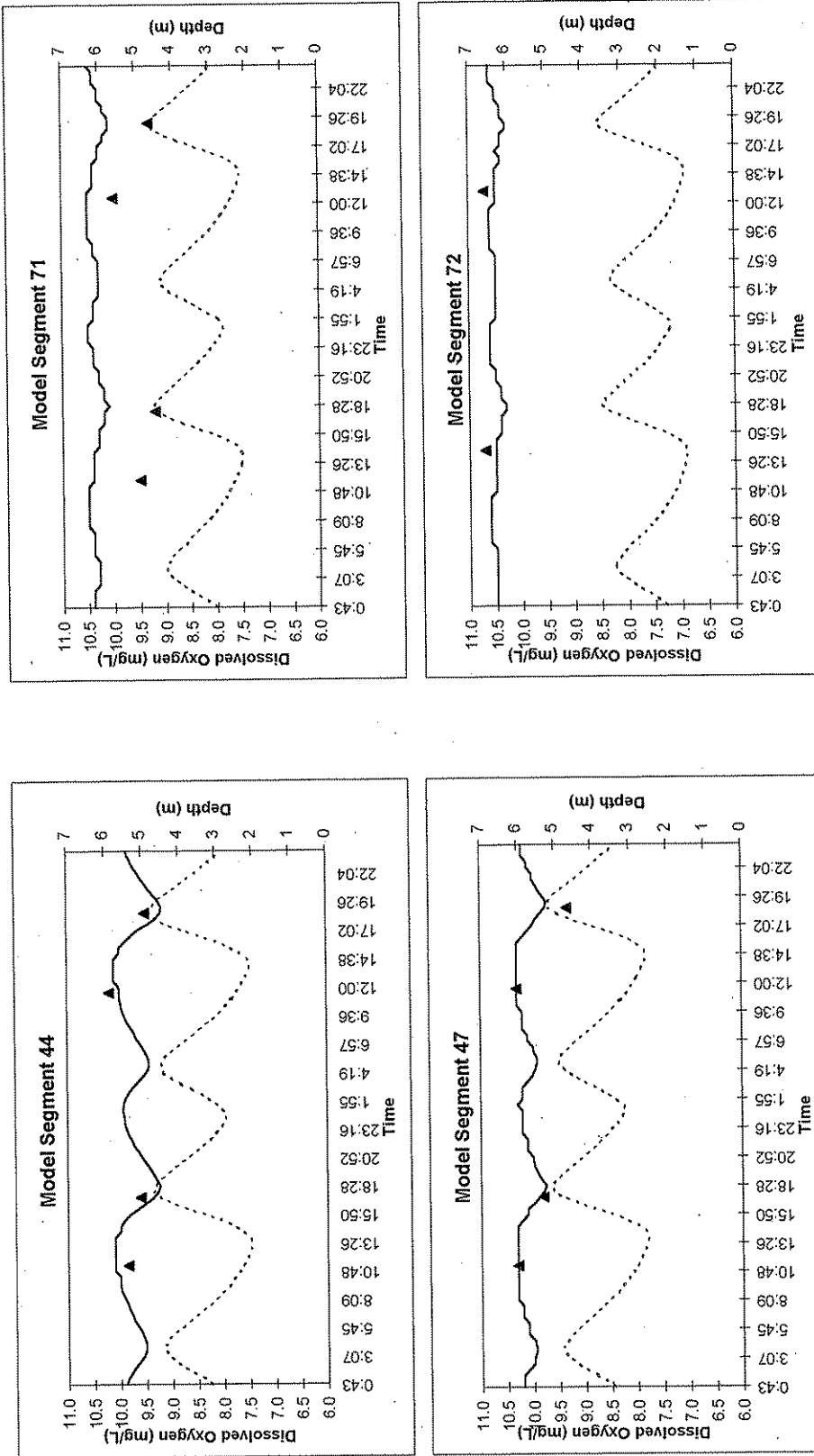


Figure 11. Verification of EUTROS to predict dissolved oxygen in the Snohomish River Estuary

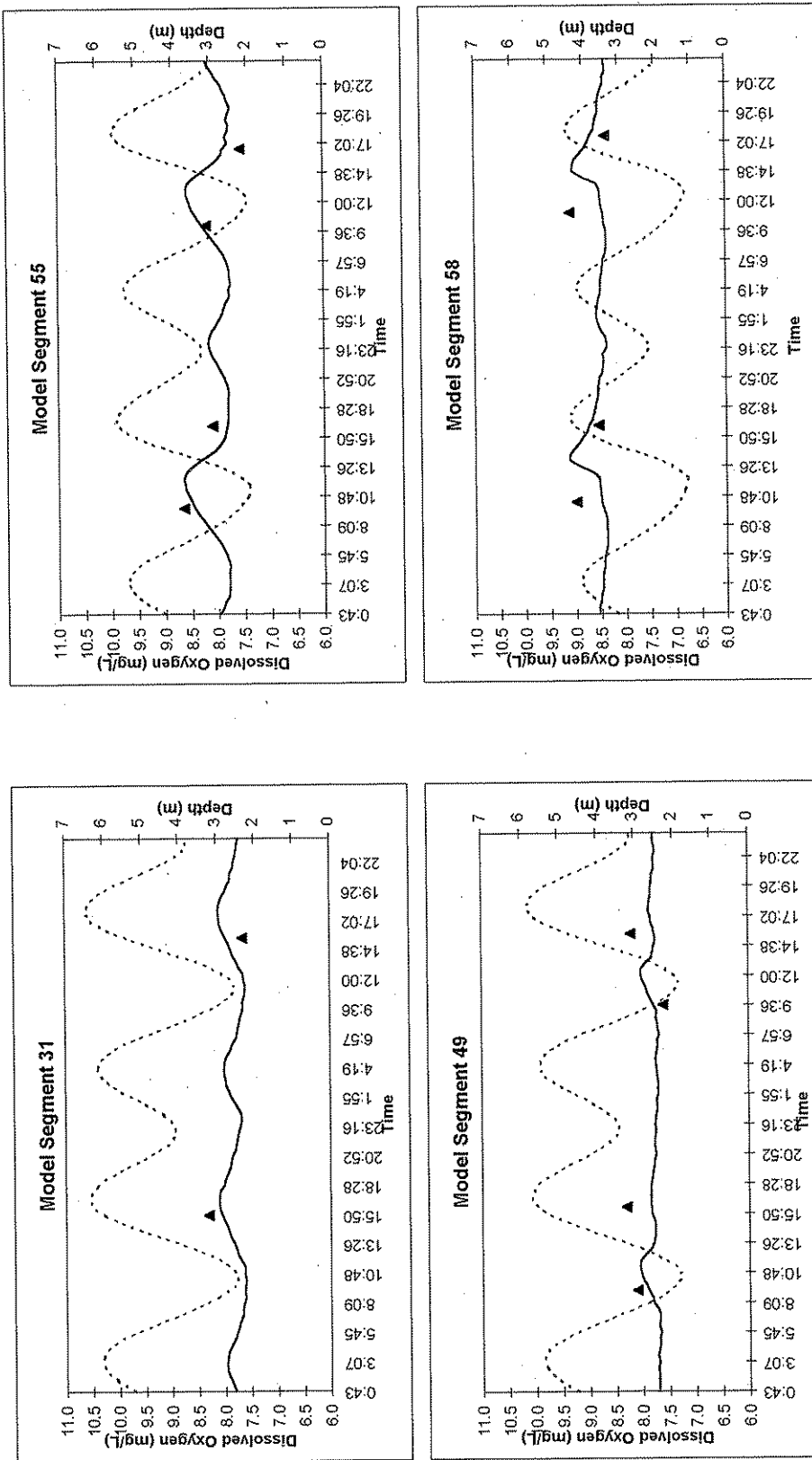


Figure 12. Verification of EUTROS to predict dissolved oxygen in the Snohomish River Estuary

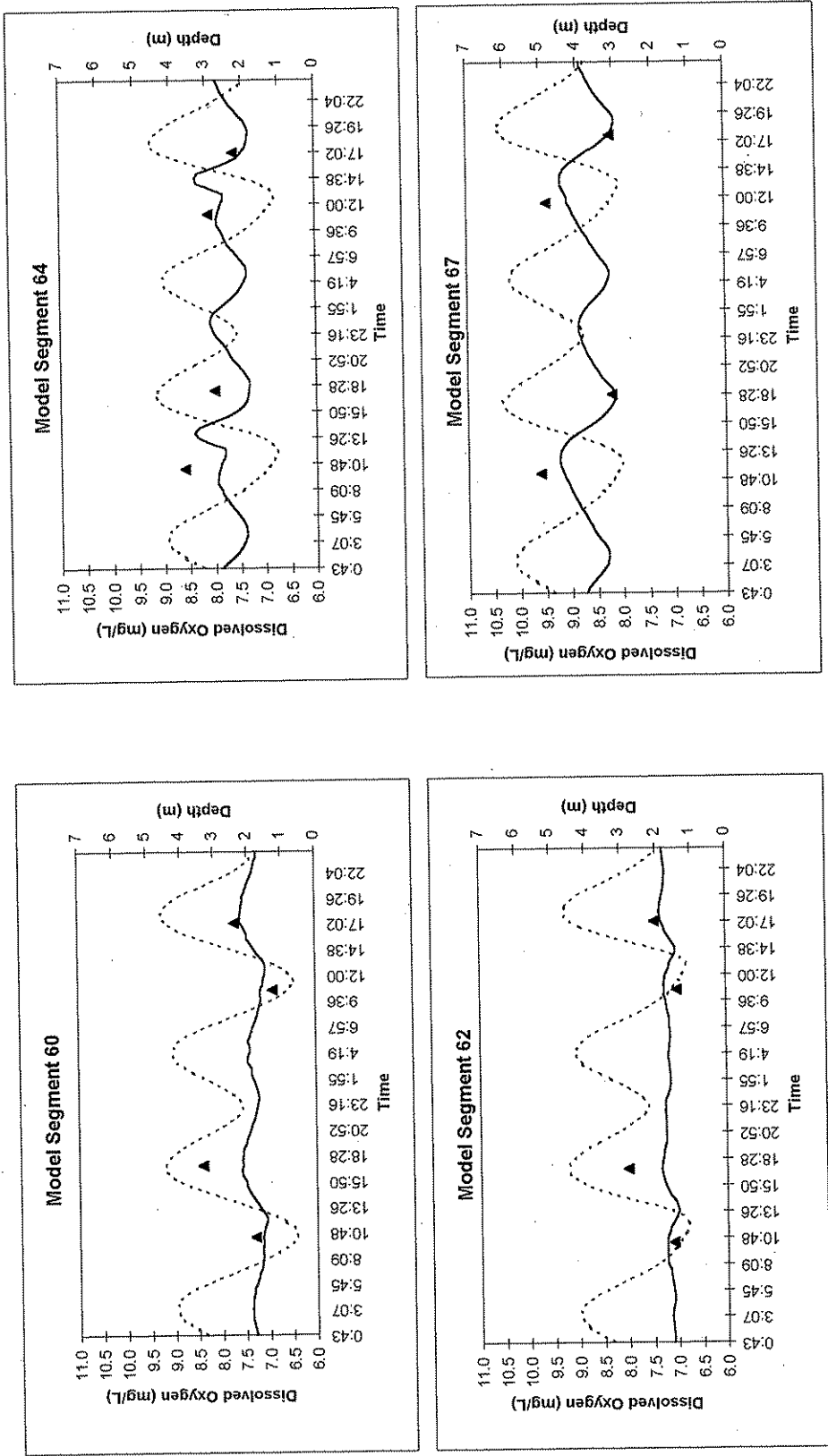


Figure 13. Verification of EUTRO5 to predict dissolved oxygen in the Snohomish River Estuary

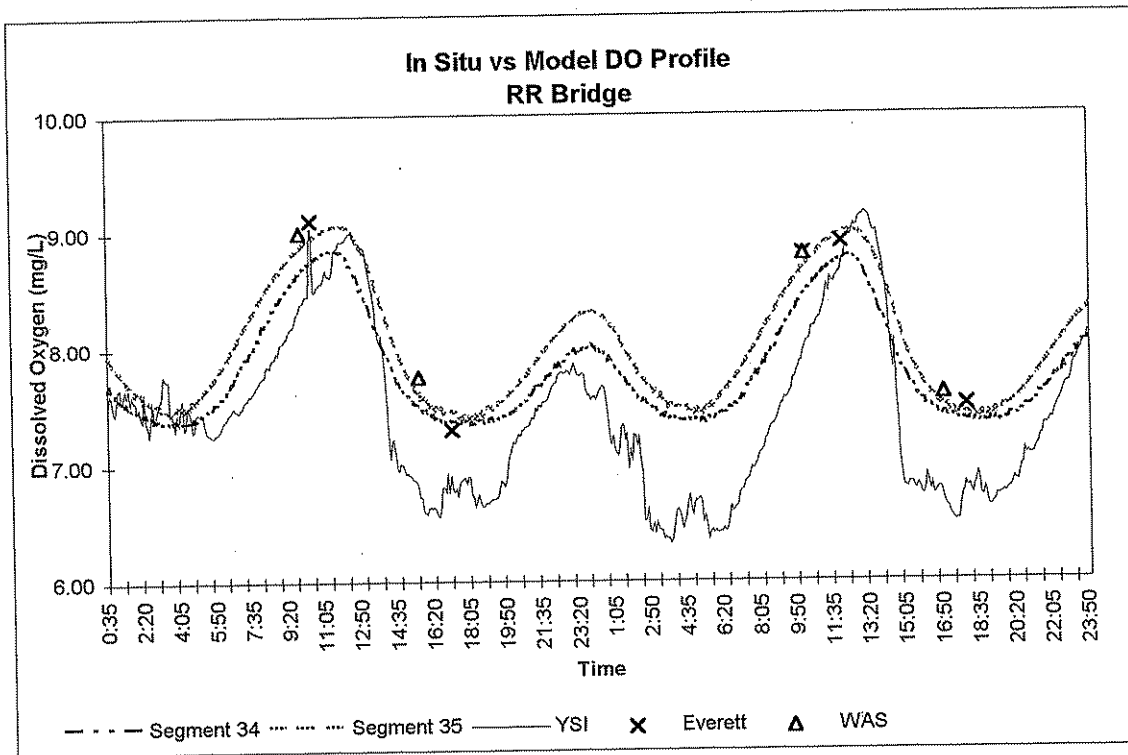
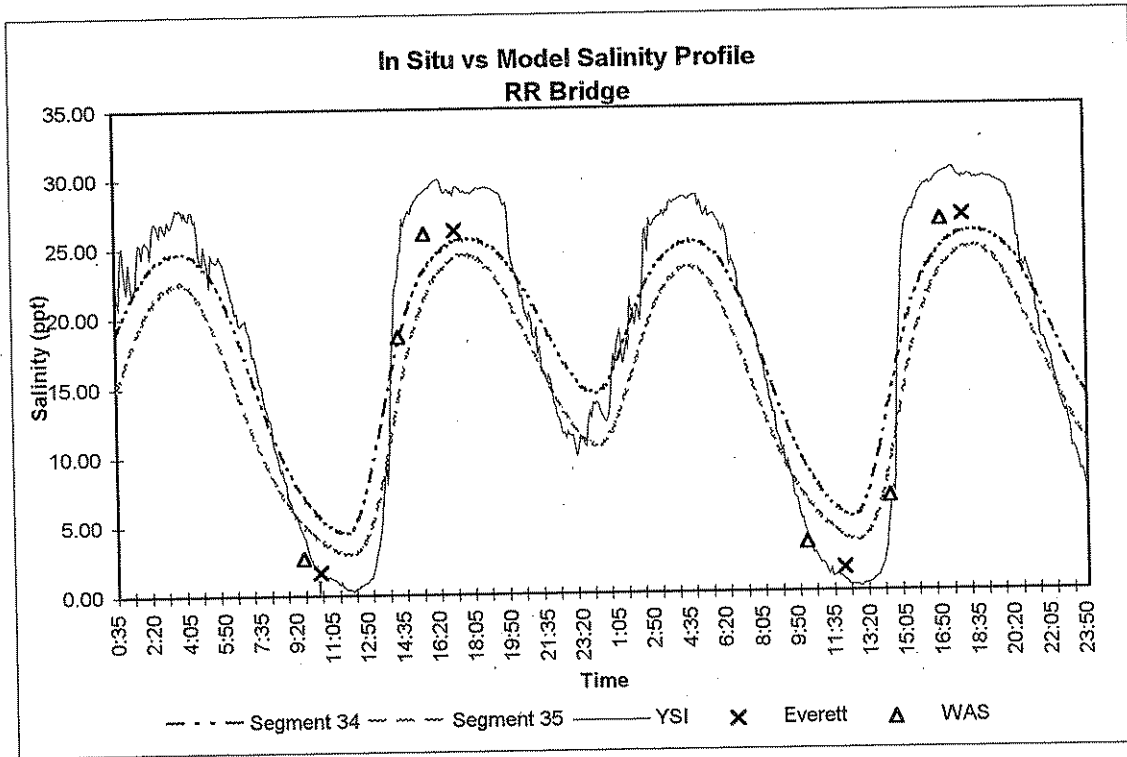


Figure 14. Comparison of EUTRO5 predicted values to continuous *in situ* and grab sample data.

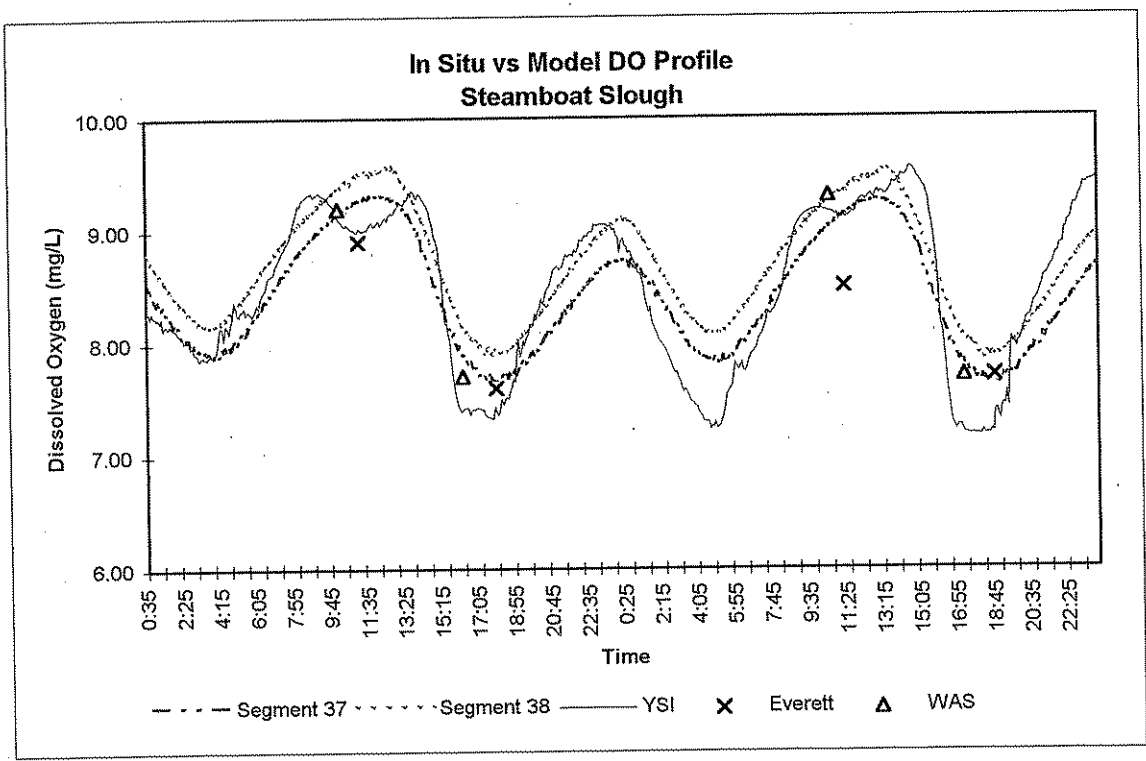
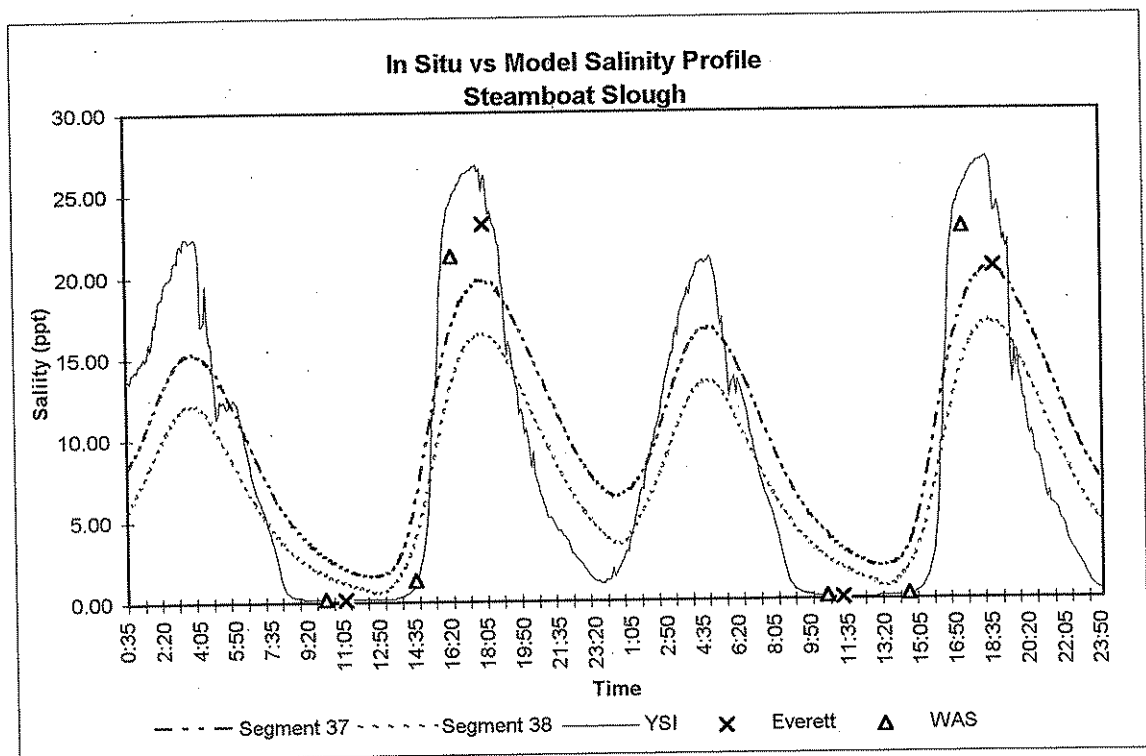


Figure 15. Comparison of EUTRO5 predicted values to continuous *in situ* and grab sample data.

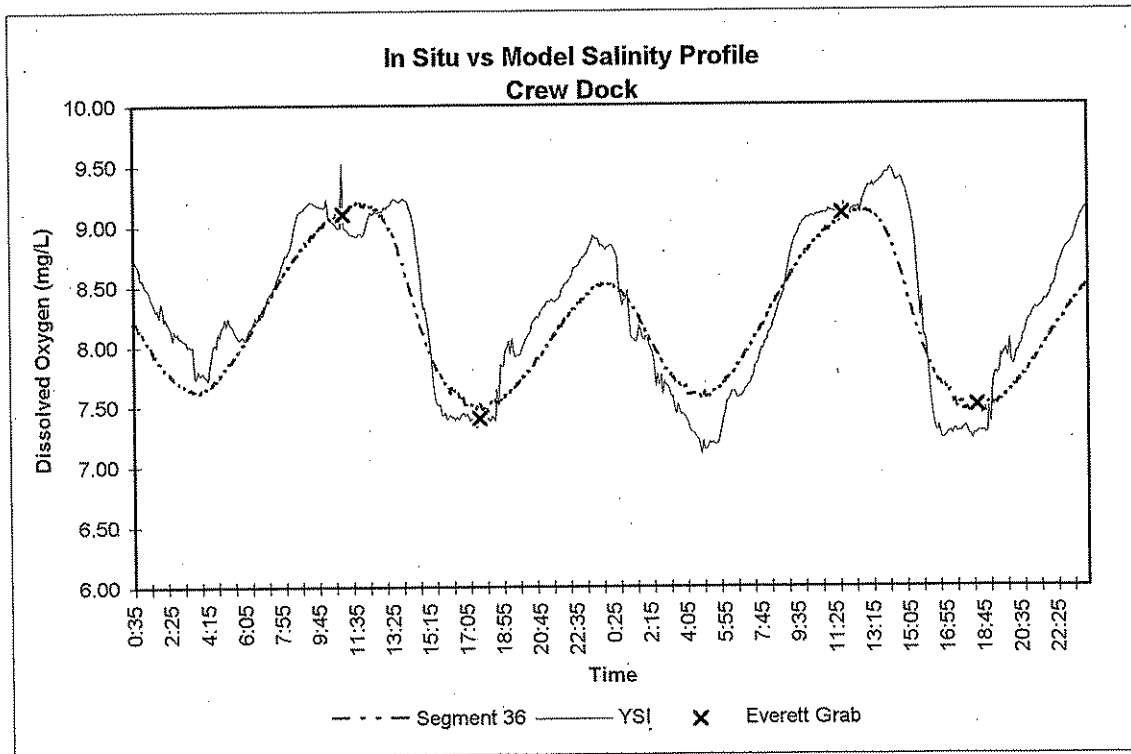
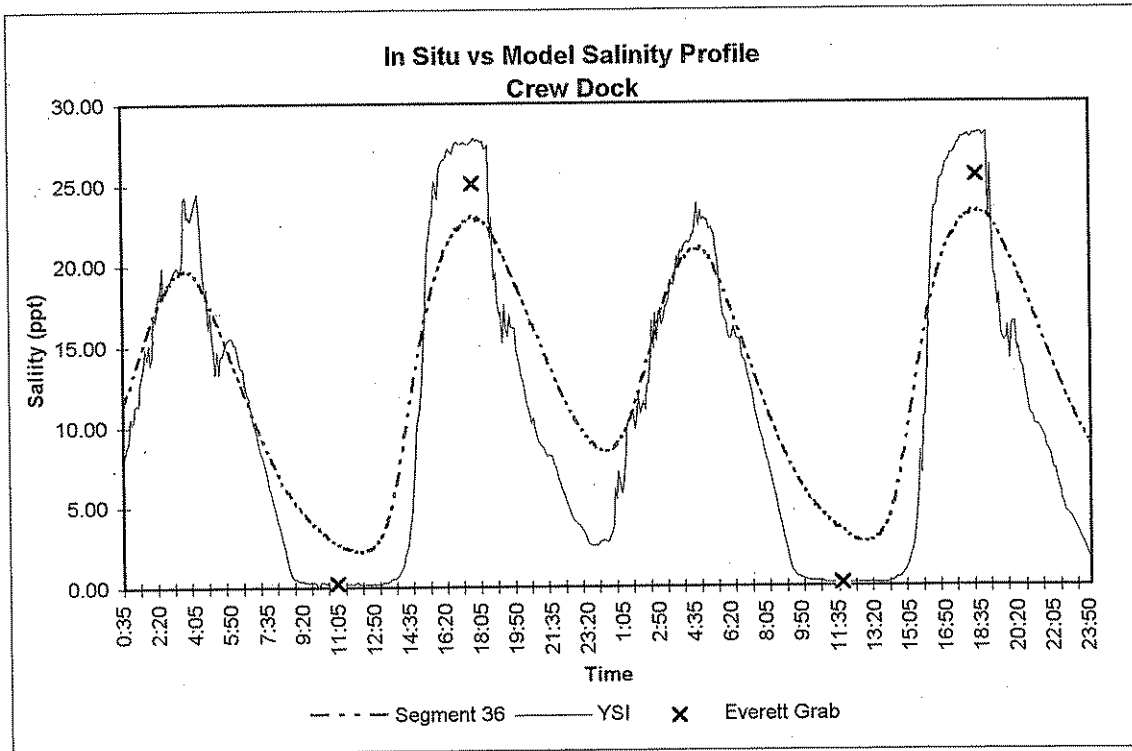


Figure 16. Comparison of EUTRO5 predicted values to continuous *in situ* and grab sample data.

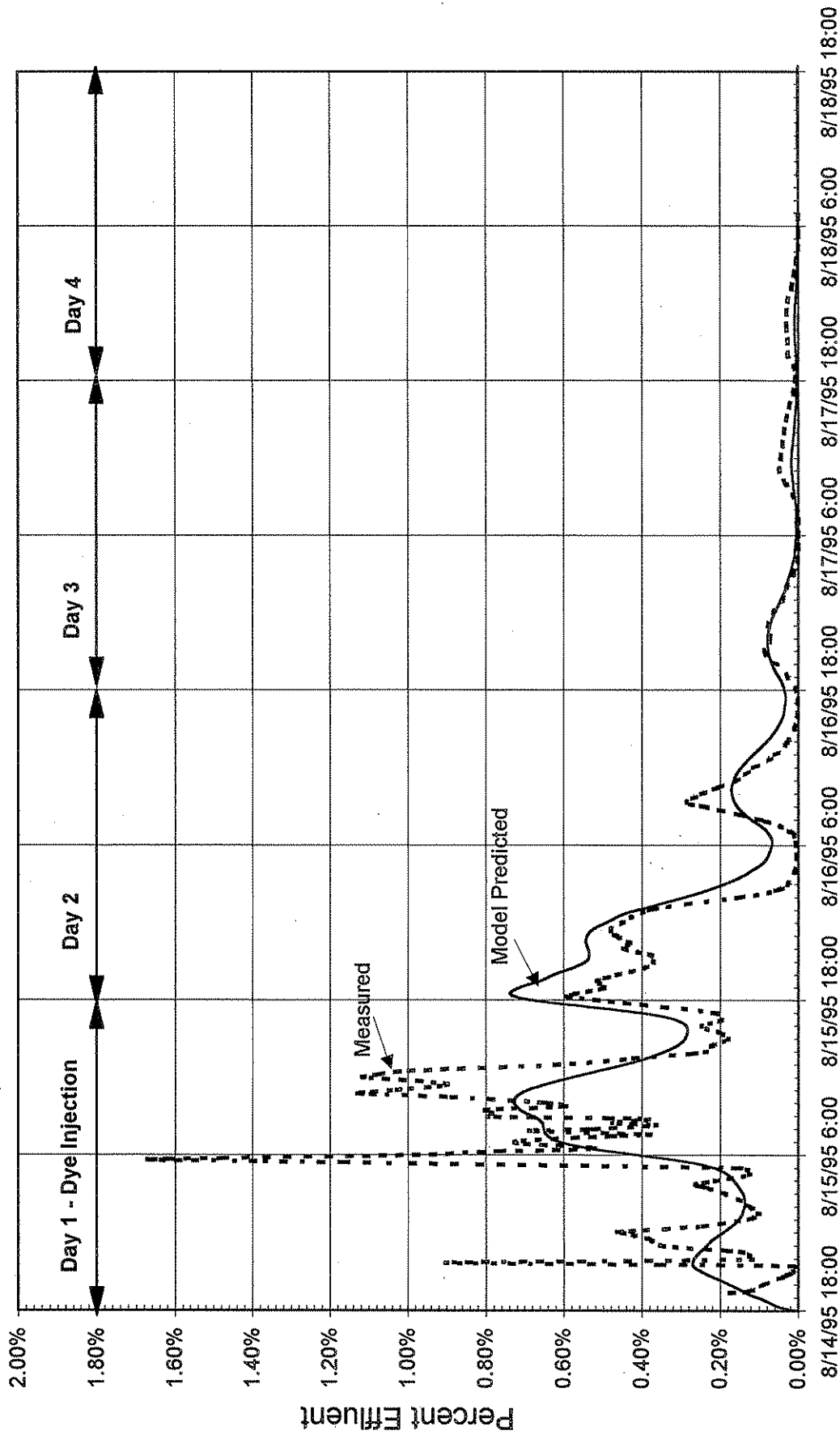


Figure 17. Measured and EUTRO5 model predicted dye concentrations as percent effluent in the Snohomish River near Langus Park.

Dissolved Oxygen Concentrations--Snohomish River

Under Critical Conditions

With and Without Loads

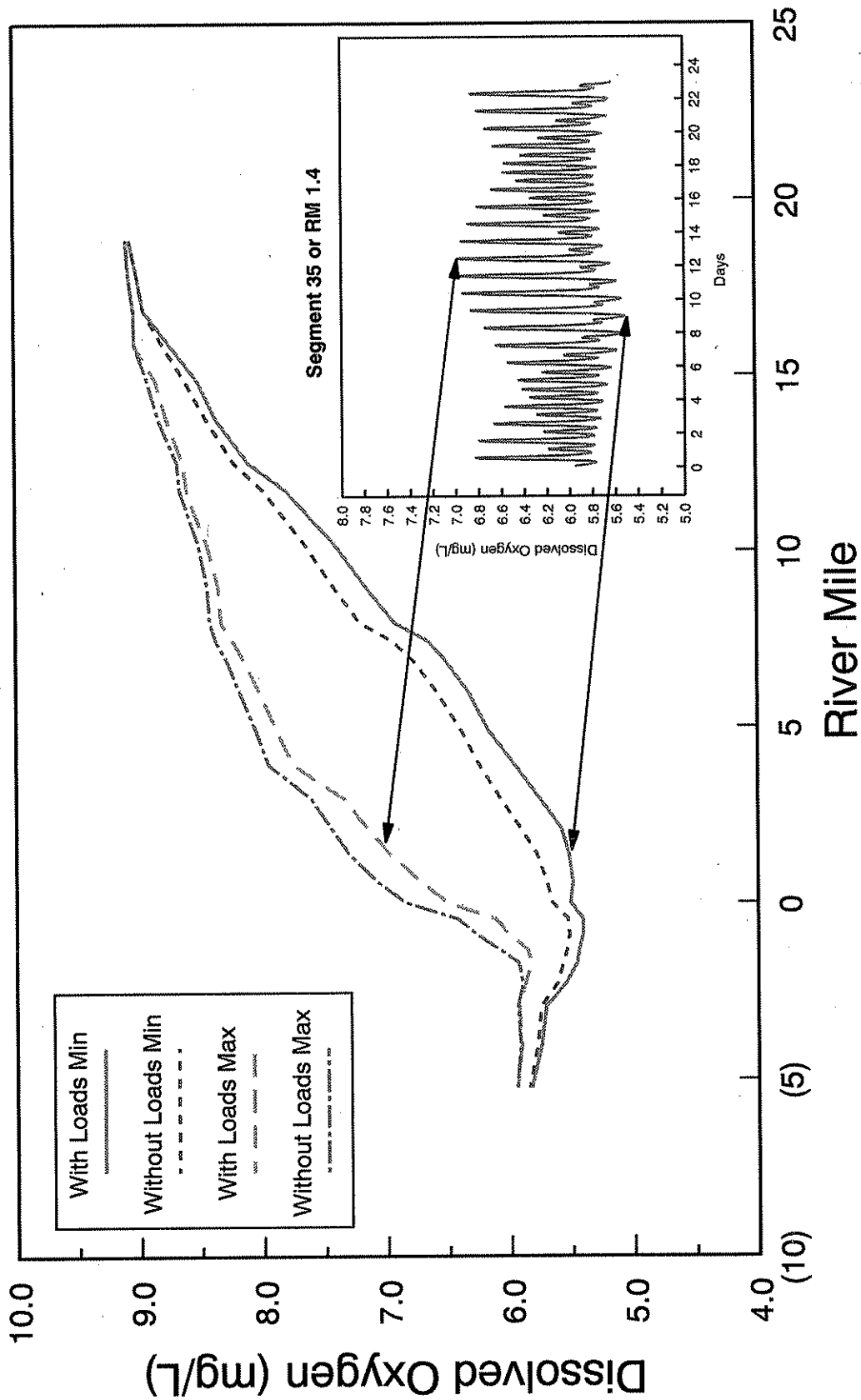


Figure 18. EUTRO5 predicted dissolved oxygen concentrations for the Snohomish River under critical conditions, with and without BOD loads.

Dissolved Oxygen Concentrations--Steamboat S.

Under Critical Conditions With and Without Loads

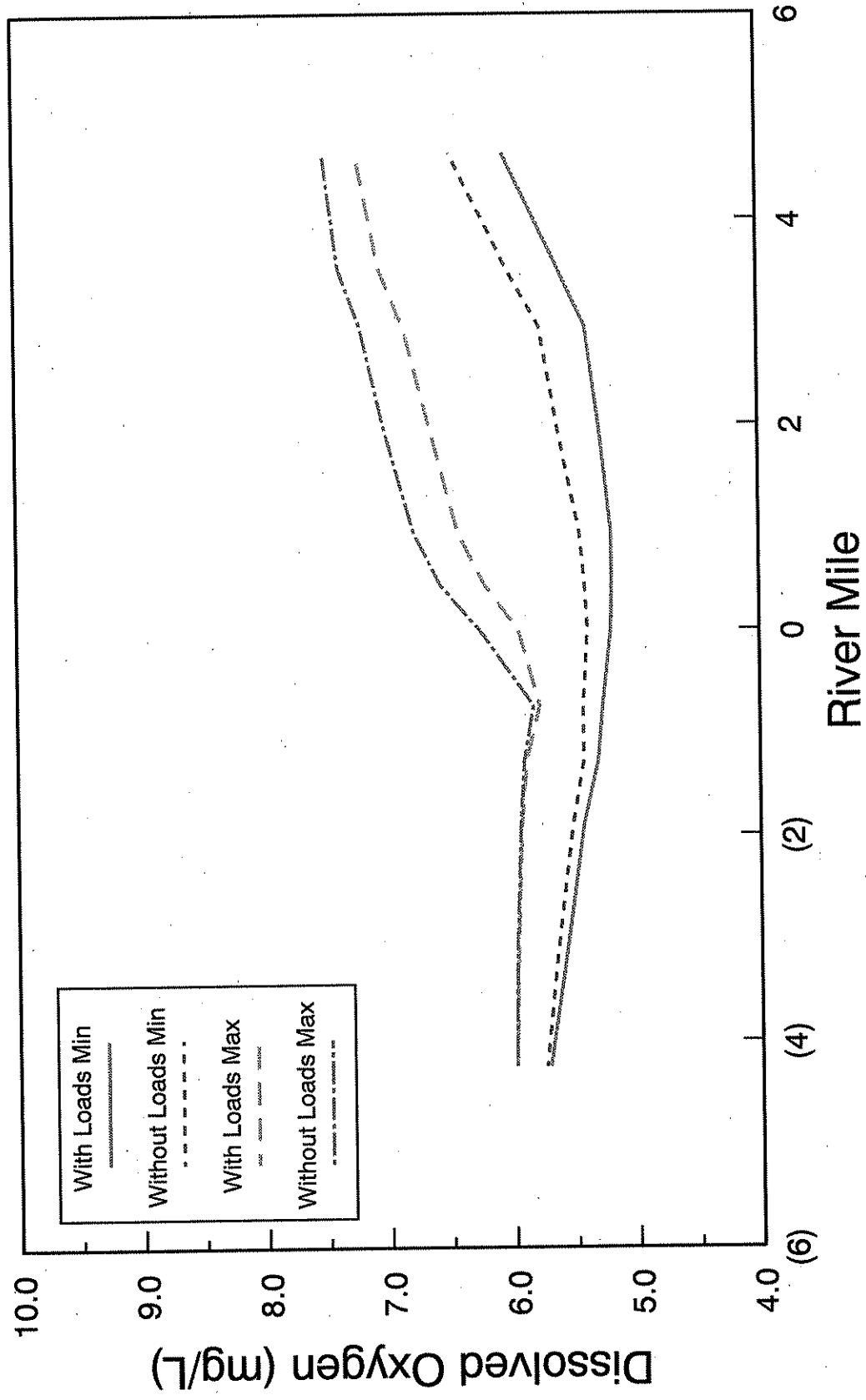


Figure 19. EUTRO5 predicted dissolved oxygen concentrations for Steamboat Slough under critical conditions, with and without BOD loads.

Dissolved Oxygen--Ebey Slough

Under Critical Conditions

With and Without Loading

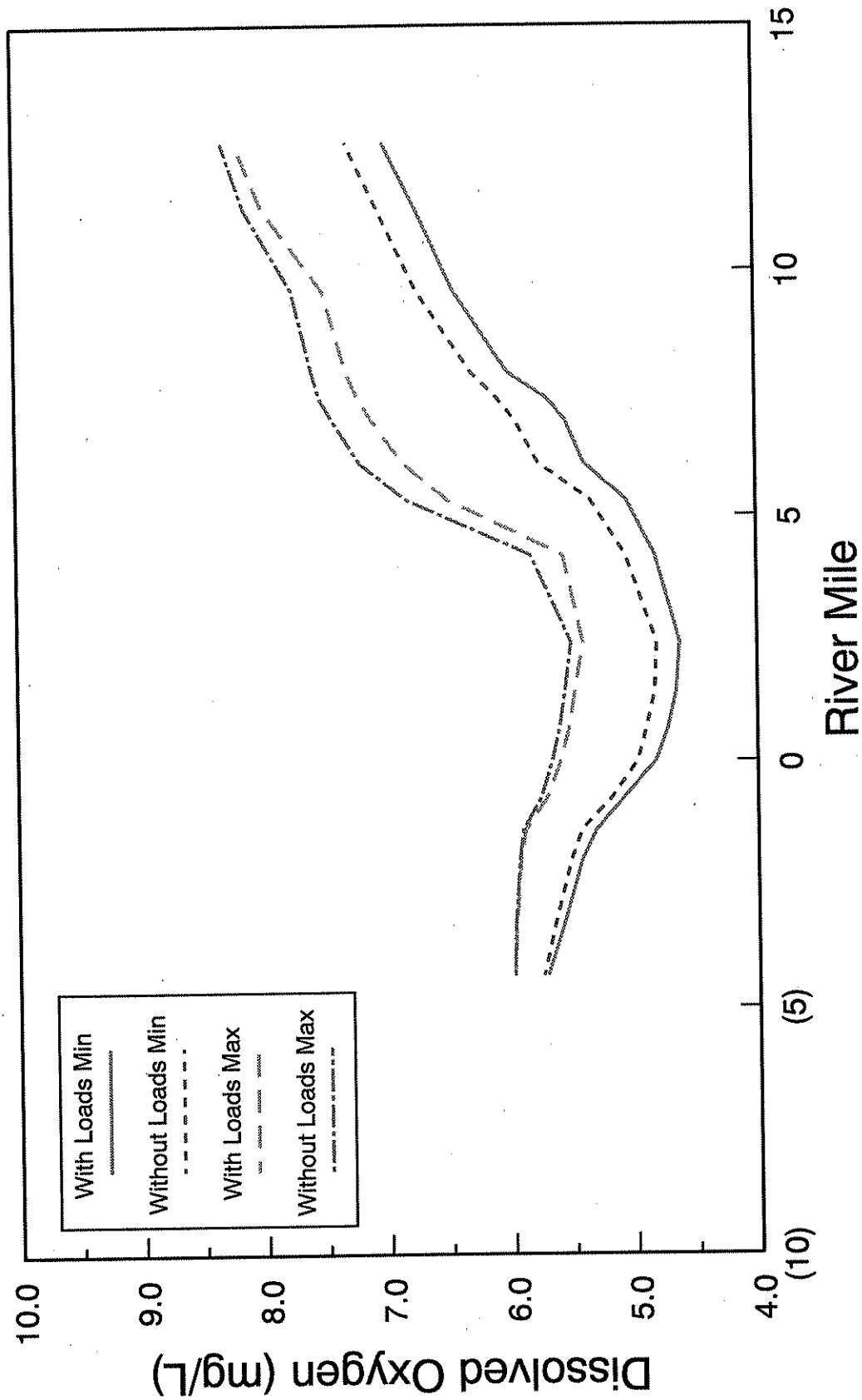


Figure 20. EUTRO5 predicted dissolved oxygen concentrations for Ebey Slough under critical conditions, with and without BOD loads.

Example of Model Intermittent Discharge Period

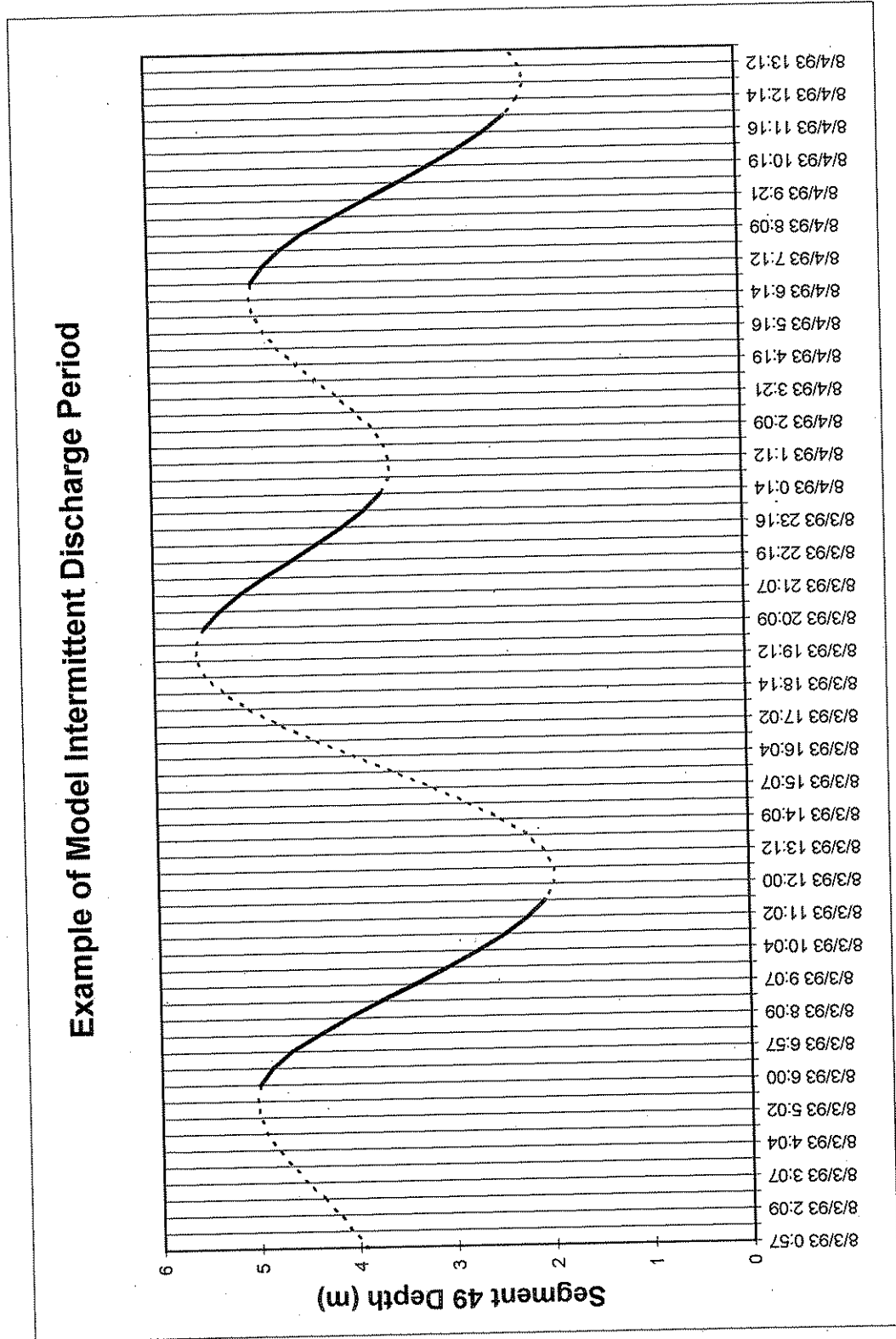


Figure 21. Example of the intermittent discharge periods (solid line on graph) for the Smith Island Treatment Plant.

Tables

Table 1. Summary of field and laboratory measurements of water, target precision or detection limits, and methods.

Parameter (all measurements of water)	Precision Limit (for field measurements) or Reporting Limit (for general chemistry)	Method ¹
Field Measurements		
Velocity	± 0.05 f/s	Current Meter
pH	± 0.1 SU	Field Meter/Electrode
Temperature	± 0.2 °C	Alcohol Thermometer
Dissolved Oxygen	± 0.06 mg/L	Gas Probe/Winkler Titration
Specific Conductivity	± 5% of reading	Field Meter/Conductivity Bridge
General Chemistry		
Salinity ¹		
Ammonia nitrogen	0.01 mg/L	EPA 350.1
Nitrate+nitrite nitrogen	0.01 mg/L	EPA 353.2
Total persulfate nitrogen	0.01 mg/L	SM 4500 NO3-F Modified
Orthophosphate	0.01 mg/L	EPA 365.3
Total phosphorus	0.01 mg/L	EPA 365.3
Chlorophyll α	0.05 g/L	SM 10200H(3), Fluorometer
Ultimate Biochemical Oxygen Demand (BOD ₃₅)	2 mg/L	NCASI

¹ Salinity was analyzed with an Autosol[®] salinometer by the Routine Chemistry Laboratory at the University of Washington.

EPA= Environmental Protection Agency

SM= Standard Methods

NCASI= National Council of the Paper Industry for Air and Stream Improvement, Inc.

Table 2. Field Replicate Precision.

Parameter	Root Mean Square of the Coefficient of Variation (%) of Replicates	
	1993 Data	1996 Data
Conductivity	1.4	1.0
Temperature	0.9	0.8
Dissolved Oxygen	3.7	0.9
pH	1.3	0.7
Total Persulfate Nitrogen	5.4	4.0
Ammonia	14.2	17.4
Nitrite-Nitrate	1.9	6.9
Total Phosphorus	15.5	12.6
Ortho-Phosphorus	3.9	2.7
Turbidity	61.4	
Total Suspended Solids	2.0	
Total Organic Carbon	9.3	
Chlorophyll <i>a</i>	29.7	14.1
Phaeophytin <i>a</i>	44.4	17.6
Total Hardness	4.3	
Fecal Coliform	45.1	

Table 3. Root mean square error (RMSE) between the recalibrated model predicted (P) and observed (O) values for salinity, dissolved oxygen, chlorophyll *a*, ammonia, and phosphorus (for comparison, values from the Phase I report are presented).

Parameter	n (number of pairs)	RMSE ^a	RMSE	Range Observed
		Recalibrated model	from Phase I report	
Salinity (ppt)	38	2.10 ^b	2.18	0.0 - 27.3
Dissolved Oxygen (mg/L)	56	0.32	0.23	6.6 - 10.2
Chlorophyll <i>a</i> (µg/L)	26	1.50 ^c	1.55	1.4 - 7.4
Ammonia (mg/L)	52	0.0185 ^d	0.0167	0.005 ^e - 0.140
Total Phosphorus (mg/L)	52	0.0050	0.0050	0.005 ^e - 0.049

^a $RMSE = [\sqrt{\sum(P_i - O_i)^2/n}]$

^b For salinity values ≤ 3 ppt RMSE is 0.77.

^c RMSE is 0.66 µg/L excluding two sites in Ebey Slough associated with water quality model segments 60 and 67.

^d RMSE is 0.0107 mg/L excluding two sites in Ebey Slough associated with water quality model segments 60 and 62.

^e Minimum value is one-half the detection limit.

Table 4. 1996 confirmation conditions for water quality of upstream and downstream boundaries and pointloads (tributaries and WWTPs).

	Flow (cfs) ^a	Ammonia (mg/L)	Nitrate (mg/L)	Ortho P (mg/L)	Chlorophyll <i>a</i> (µg/L)	Non-algal CBODU (mg/L)	Dissolved Oxygen (mg/L)	Organic N (mg/L)	Organic P (mg/L)	Salinity (ppt)
Upstream Boundary	1840	0.005	0.134	0.005	2.8	1.2	9.5	0.047	0.005	0.0
French Creek	0.0 ^b	0.135	0.637	0.016	30.0	6.0	8.1	0.223	0.054	0.0
Pilchuck River	80	0.005	0.290	0.005	1.6	1.2	10.3	0.079	0.005	0.0
Snohomish WWTP	0.79	0.460	7.58	4.55	0.0	12.0	8.7	2.38	0.005	0.0
Marshland	0.0 ^b	0.117	1.23	0.034	14	5.4	9.2	0.370	0.090	0.0
Deadwater Slough	0.0 ^b	0.237	0.226	0.018	15.7	10.8	9.8	0.519	0.080	3.3
Everett WWTP (Mechanical)	9.04	11.4	10.2	4.05	0.0	17.5	4.8	3.00	0.230	0.0
Everett WWTP (Lagoon)	8.22	17.3	0.050	4.76	0.0	20.0	5.4	3.30	0.030	0.0
Swan Trail Slough	0.0 ^b	3.83	0.069	0.015	7.0	16.2	1.2	0.546	0.215	0.1
Lake Stevens WWTP	1.92	17.3	0.005	1.07	0.0	12.0	6.1	.005	0.250	0.0
Marysville WWTP	2.02	1.90	3.80	2.39	0.0	30.0	1.7	10.8	2.21	0.0
Allen Creek	0.0 ^c	0.010	0.923	0.073	1.4	1.2	1.1	0.133	0.018	0.0
Quilceda Creek	24.0	0.018	1.08	0.056	2.4	1.2	9.5	0.142	0.036	0.6
Tulalip Landfill		8.14 ^d								
Downstream Boundary	NA	0.005	0.186	0.035	4.1	2.5	7.2 ^e	0.113	0.014	29.3

^a Flows in model input as m³/s

^b Tributaries controlled by pumping stations; were not pumping during survey.

^c No flow measured during survey.

^d Value listed as lbs/day based on EPA estimate of 2971 lbs/year (EPA, 1996).

^e EUTRO5 model boundaries 1-5 set at 7.2; boundary 6 set at 8.0 based on sampling station PS35.

Table 5. Summary of EUTRO5 model constants for eutrophication kinetics.

Model Constant	Description	Units	Value Used	Temperature Correction Coefficient
K12C	Nitrification rate	day ⁻¹	0.2	1.085
KNIT	Half-saturation constant for nitrification-oxygen limitation	mg O ₂ /L	2.0	
K20C	Denitrification rate	day ⁻¹	0.09	1.045
KNO3	Half-saturation constant for denitrification oxygen limitation	mg O ₂ /L	0.1	
K1C	Saturated growth rate of phytoplankton	day ⁻¹	1.40	1.066
LGHTS	Light formulation switch: LGHTS = 1 use Di Toro et al. (1971) formulation.		1	
CCHL	Carbon-to-chlorophyll ratio	mg C/mg chl <i>a</i>	75	
IS1	Saturation light intensity for phytoplankton growth.	Ly/day	100	
KMNG1	Nitrogen half-saturation constant for nitrogen for phytoplankton growth, which also affects ammonia preference.	mg N/L	0.03	
KMPG1	Phosphorous half-saturation constant for phytoplankton growth.	mg PO ₄ -P/L	0.002	
NCRB	Nitrogen-to-carbon ratio in phytoplankton	mg N/L	0.25	
PCRB	Phosphorus-to-carbon ratio in phytoplankton	mg P/L	0.025	
K1RC	Endogenous respiration rate of phytoplankton	day ⁻¹	0.092	1.045
K1D	Non-predatory phytoplankton death rate	day ⁻¹	0.02	
KDC	CBOD deoxygenation rate	day ⁻¹	0.152	1.047
OCRB	Oxygen to carbon ratio in phytoplankton	mg O ₂ /mg C	2.67	
K71C	Mineralization rate of dissolved organic nitrogen	day ⁻¹	0.10	1.047
FON	Fraction of dead and respired phytoplankton nitrogen recycled to organic nitrogen		1.0	
K83C	Mineralization rate of dissolved organic phosphorus	day ⁻¹	0.10	1.047
FOP	Fraction of dead and respired phytoplankton phosphorus recycled to organic phosphorus		1.0	

Di Toro et. al. 1971 referenced in the WASP5 users manual.

Table 6. Root mean square error (RMSE) between the confirmation model predicted (P) and observed (O) values for salinity, dissolved oxygen, chlorophyll *a*, ammonia, and phosphorus.

Parameter	n (number of pairs)	RMSE ^a Recalibrated model	Range Observed
Salinity (ppt)	46	3.00 ^b	0.0 - 27.7
Dissolved Oxygen (mg/L)	64	0.39	7.0 - 10.7
Chlorophyll <i>a</i> (µg/L)	42	4.43	1.5 - 26.3
Ammonia (mg/L)	59	0.0163	0.005 ^c - 0.112
Total Phosphorus (mg/L)	59	0.0050	0.005 ^c - 0.294

^a $RMSE = [\sqrt{\sum(P_i - O_i)^2/n}]$

^b For salinity values ≤ 3 ppt RMSE is 2.4.

^c Minimum value is one-half the detection limit; only one Total Phosphorus value exceeded 0.072 mg/L.

Table 7. Critical conditions for water quality of upstream and downstream boundaries and pointloads (tributaries and WWTPs).

	Flow (cfs) ^a	Ammonia (mg/L)	Nitrate (mg/L)	Ortho P (mg/L)	Chlorophyll <i>a</i> (µg/L)	Non-algal CEODU (mg/L)	Dissolved Oxygen (mg/L)	Organic N (mg/L)	Organic P (mg/L)	Salinity (ppt)
Upstream Boundary	1051	0.050	0.340	0.005	1.6	1.2	9.1	0.117	0.020	0.0
French Creek	16.2	0.219	1.08	0.023	1.9	6.0	3.5	0.806	0.059	0.0
Pilchuck River	55.8	0.030	0.400	0.005	1.6	1.2	9.2	0.134	0.020	0.0
Snohomish WWTP	4.34 ^b	21.0	7.58	4.55	0.0	90.0	8.7 ^e	12.8	4.55	0.0
Marshland	16.2	0.231	1.23	0.052	14.1	5.4	4.4	0.519	0.135	0.0
Deadwater Slough	5.51	0.752	0.226	0.018	15.7	10.8	9.8	1.28	0.136	0.0
Everett WWTP (Mechanical)	12.4 ^c	24.0	10.16	4.05	0.0	65.7	4.8 ^e	4.0	4.05	0.0
Everett WWTP (Lagoon)	16.3 ^c	26.0	0.950	4.76	0.0	100.0	5.4 ^e	25.0	4.76	0.0
Swan Trail Slough	1.62	3.83	0.069	0.015	11.7	16.2	1.2	0.470	0.099	0.0
Lake Stevens WWTP	3.72 ^b	17.3	0.170	2.17	0.0	90.0	6.1 ^e	4.0	1.07	0.0
Marysville WWTP	9.46 ^b	21.0	3.94	4.73	0.0	90.0	1.7 ^e	10.0	2.39	0.0
Allen Creek	0.0 ^d	0.025	0.923	0.900	1.3	1.2	1.0	0.282	0.025	0.0
Quilceda Creek	4.5	0.023	1.08	0.064	1.6	1.2	7.2	0.219	0.039	0.0
Downstream Boundary	NA	0.095	0.300	0.060	0.9	2.5	6.0	0.126	0.040	29.5

^a Flows in model input as m³/s.

^b Maximum monthly flow listed in current annual permit.

^c Summer low flow seasonal permit limit.

^d No flow measured during survey, assumed no flow during critical conditions.

^e WWTP effluent measured dissolved oxygen concentration during 1996 survey.

Table 8. Critical 7-day average low flows in cfs.

Location	7Q10 _{Annual}	7Q20 _{Jul-Oct}	7Q20 _{Nov-Jun}
Skykomish @ Goldbar	442	401	563
Sultan @ Sultan	40	29	120
Skykomish @ Sultan	580	530	822
Woods Creek @ Monroe	14	13	20
Skykomish @ Monroe	701	653	993
Snoqualmie @ Mouth	436	398	675
Snohomish @ RM 20.4	1137	1051	1668
Pilchuck @ Snohomish	58	56	73
Allen Creek @ Ebey S.	-	0	-
Quelceda Creek	-	4.5	-

Table 9. Contribution of point and nonpoint sources of BOD to the total dissolved oxygen (DO) deficit at low (L) ebbing and high (H) slack tide at selected water quality model segments.

Model Seg #	DO (mg/L) Deficit ^a		% Contribution to Dissolved Oxygen Deficit ^b															
	DO (mg/L) Deficit ^a		Everett WWTP		Marysville WWTP		Lake Stevens WWTP		Snohomish WWTP		French Creek		Marshland		Deadwater S.		Other ^c	
	L	H	L	H	L	H	L	H	L	H	L	H	L	H	L	H	L	H
34	0.41	0.23	51	71	<5	<5	<5	<5	19	11	9	<5	11	7	<5	<5	<5	<5
36	0.35	0.39	40	77	<5	<5	6	<5	23	10	11	<5	14	5	<5	<5	<5	<5
43	0.11	0.32	<5	37	<5	6	9	<5	33	22	59	9	8	9	<5	<5	<5	7
46	0.08	0.24	<5	27	<5	<5	7	<5	22	28	78	19	<5	14	<5	<5	<5	<5
50	0.38	0.23	21	18	24	45	18	24	18	11	8	<5	5	4	<5	<5	<5	<5
55	0.41	0.36	17	17	17	44	24	22	22	11	7	<5	5	3	<5	<5	<5	<5
56	0.39	0.40	18	17	13	37	21	22	28	14	8	<5	6	4	<5	<5	<5	<5
57	0.37	0.44	16	18	11	32	19	23	30	16	11	5	8	5	<5	<5	<5	<5
58	0.30	0.46	27	48	6	10	10	13	27	18	13	6	10	6	<5	<5	<5	<5
64	0.37	0.31	19	19	19	35	21	26	21	13	8	<5	5	3	<5	<5	<5	<5
66	0.35	0.42	19	20	11	31	19	25	27	16	14	<5	8	4	<5	<5	<5	<5
68	0.28	0.33	18	21	7	18	14	21	32	21	18	9	7	3	<5	<5	<5	7

^a Deficit determined as the difference between DO at critical conditions with point and nonpoint sources of BOD, and critical conditions without point sources of BOD and nonpoint sources of BOD set at estimated background conditions (i.e., ammonia = 0.005 mg/L and CBOD = 1.2 mg/L).

^b Model and rounding error may account for 0.01 mg/L or 5% at a deficit of 0.20 mg/L.

^c Other contributions are Quiceda Creek, Pilehuck River, organic nitrogen from all sources, rounding and model error.

Table 10. Model predicted changes in dissolved oxygen (DO) concentrations at low (L) ebbing and high (H) slack tide after three (I, II, III) different point source BOD loading reduction alternatives. Dissolved oxygen deficits are calculated as the difference between alternative predicted concentration and the predicted concentration without BOD loading.

Model Seg #	Critical Conditions																					
	Without BOD Loads ^a				With BOD Loads ^b				I Ammonia set at 2 mg/L				II Ammonia set at 2 mg/L, BOD ₅ at 33 mg/L				III Ammonia set at 5 mg/L, BOD ₅ at 27 mg/L					
	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H		
34	7.13	5.72	6.72	5.49	-0.41	-0.23	6.88	5.60	-0.25	-0.12	6.93	5.63	-0.20	-0.09	6.93	5.63	-0.20	-0.09	6.93	5.63	-0.20	-0.09
36	7.50	5.97	7.15	5.58	-0.35	-0.39	7.28	5.76	-0.22	-0.21	7.32	5.82	-0.18	-0.15	7.32	5.82	-0.18	-0.15	7.32	5.82	-0.18	-0.15
43	8.44	7.25	8.33	6.93	-0.11	-0.32	8.36	7.05	-0.08	-0.20	8.36	7.09	-0.08	-0.16	8.36	7.08	-0.08	-0.17	8.36	7.08	-0.08	-0.17
46	8.61	7.86	8.53	7.62	-0.08	-0.24	8.54	7.70	-0.07	-0.16	8.54	7.73	-0.07	-0.13	8.54	7.72	-0.07	-0.14	8.54	7.72	-0.07	-0.14
50	6.60	5.42	6.22	5.19	-0.38	-0.23	6.36	5.29	-0.24	-0.13	6.42	5.33	-0.18	-0.09	6.42	5.33	-0.18	-0.09	6.42	5.33	-0.18	-0.09
55	7.05	5.64	6.64	5.28	-0.41	-0.36	6.79	5.43	-0.26	-0.21	6.85	5.48	-0.20	-0.16	6.84	5.48	-0.21	-0.16	6.84	5.48	-0.21	-0.16
56	7.24	5.79	6.85	5.39	-0.39	-0.40	6.99	5.55	-0.25	-0.24	7.04	5.61	-0.20	-0.18	7.04	5.61	-0.20	-0.18	7.04	5.61	-0.20	-0.18
57	7.41	6.05	7.04	5.61	-0.37	-0.44	7.17	5.75	-0.24	-0.27	7.22	5.85	-0.19	-0.20	7.22	5.85	-0.19	-0.20	7.22	5.85	-0.19	-0.20
58	7.52	6.50	7.22	6.04	-0.30	-0.46	7.32	6.25	-0.20	-0.25	7.36	6.31	-0.16	-0.19	7.36	6.31	-0.16	-0.19	7.36	6.31	-0.16	-0.19
64	6.86	5.36	6.49	5.05	-0.37	-0.31	6.63	5.17	-0.23	-0.19	6.68	5.22	-0.18	-0.14	6.68	5.22	-0.18	-0.14	6.68	5.22	-0.18	-0.14
66	7.56	6.11	7.21	5.69	-0.35	-0.42	7.33	5.87	-0.23	-0.24	7.38	5.94	-0.18	-0.17	7.38	5.94	-0.18	-0.17	7.38	5.94	-0.18	-0.17
68	7.77	6.76	7.49	6.43	-0.28	-0.33	7.58	6.55	-0.19	-0.21	7.62	6.60	-0.15	-0.16	7.61	6.60	-0.16	-0.16	7.61	6.60	-0.16	-0.16

^a Critical conditions without point sources of BOD and nonpoint sources of BOD set at estimated background conditions.

^b Critical conditions with point and nonpoint sources of BOD.

Note: Segment 57 is the "critical" segment for the modeled system (i.e., requires the most load reduction to meet the criteria).

Table 11. Model predicted changes in dissolved oxygen (DO) concentrations at low (L) ebbing and high (H) slack tide after point source BOD loading reduction alternatives including the proposed Smith Island Treatment Plant (SITP) discharge. Dissolved oxygen deficits are calculated as the difference between scenario alternative predicted concentration and the predicted concentration without BOD loading.

Model Seg #	Critical Conditions Without BOD Loads ^a		II WWTPs Ammonia set at 2 mg/L, BOD ₅ at 33 mg/L (same as II in Table 10)				IV WWTPs Ammonia set at 2 mg/L, BOD ₅ at 33 mg/L; SITP Ammonia set at 2mg/L, BOD ₅ at 63 mg/L				V WWTPs Ammonia set at 2 mg/L, BOD ₅ at 26 mg/L; SITP Ammonia set at 2mg/L, BOD ₅ at 26 mg/L					
	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	Deficit (mg/L) L	Deficit (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	DO (mg/L) L	DO (mg/L) H	Deficit (mg/L) L	Deficit (mg/L) H
	34	7.13	5.72	6.93	5.63	-0.20	-0.09	6.93	5.63	-0.20	-0.09	6.95	5.64	-0.18	-0.08	
36	7.50	5.97	7.32	5.82	-0.18	-0.15	7.31	5.81	-0.19	-0.16	7.33	5.84	-0.17	-0.13		
43	8.44	7.25	8.36	7.09	-0.08	-0.16	8.36	7.08	-0.08	-0.17	8.37	7.10	-0.07	-0.15		
46	8.61	7.86	8.54	7.73	-0.07	-0.13	8.54	7.72	-0.07	-0.14	8.55	7.74	-0.06	-0.12		
50	6.60	5.42	6.42	5.33	-0.18	-0.09	6.39	5.26	-0.21	-0.16	6.43	5.31	-0.17	-0.11		
55	7.05	5.64	6.85	5.48	-0.20	-0.16	6.82	5.42	-0.23	-0.22	6.86	5.48	-0.19	-0.16		
56	7.24	5.79	7.04	5.61	-0.20	-0.18	7.03	5.55	-0.21	-0.24	7.06	5.61	-0.18	-0.18		
57	7.41	6.05	7.22	5.85	-0.19	-0.20	7.20	5.78	-0.21	-0.26	7.23	5.85	-0.18	-0.20		
58	7.52	6.50	7.36	6.31	-0.16	-0.19	7.35	6.29	-0.17	-0.21	7.37	6.33	-0.15	-0.17		
64	6.86	5.36	6.68	5.22	-0.18	-0.14	6.66	5.17	-0.20	-0.19	6.69	5.22	-0.17	-0.14		
66	7.56	6.11	7.38	5.94	-0.18	-0.17	7.37	5.89	-0.19	-0.22	7.39	5.95	-0.17	-0.16		
68	7.77	6.76	7.62	6.60	-0.15	-0.16	7.61	6.58	-0.16	-0.18	7.63	6.61	-0.14	-0.15		
49 ^b	6.29	5.40	6.13	5.32	-0.16	-0.08	6.10	5.25	-0.19	-0.15	6.14	5.30	-0.15	-0.10		
32 ^b	5.83	5.44	5.81	5.37	-0.02	-0.07	5.80	5.32	-0.03	-0.12	5.81	5.36	-0.02	-0.08		
31 ^b	5.91	5.44	5.90	5.38	-0.01	-0.06	5.90	5.34	-0.01	-0.10	5.90	5.37	-0.01	-0.07		

^a Critical conditions without point sources of BOD and nonpoint sources of BOD set at estimated background conditions.

^b Model segment 49 receives SITP effluent, segments 31 and 32 are downstream (seaward) of the discharge (see Figure 6).

Note: Segment 57 is the "critical" segment for the modeled system (i.e., requires the most load reduction to meet the criteria).

Table 12. Summary of recommended TMDLs and WLAs for water quality-based permit limit development

	Daily Maximum BOD ₅		Daily Maximum Ammonia	
	mg/L	pounds/day	mg/L	pounds/day
1. WLAs without Smith Island Treatment Plant				
Everett Mechanical WWTP	27	1805	5	334
Everett Lagoon WWTP	27	2372	5	439
Marysville WWTP	27	1377	5	255
Lake Steven's Sewer District	27	541	5	100
Snohomish WWTP	27	632	5	117
2. WLAs with Smith Island Treatment Plant				
Everett Mechanical WWTP	26	1738	2	134
Everett Lagoon WWTP	26	2284	2	176
Marysville WWTP	26	1326	2	102
Lake Steven's Sewer District	26	521	2	40
Snohomish WWTP	26	608	2	47
Smith Island Treatment Plant	26	2732	2	210
Background/Nonpoint LA	--	3862 ^a	--	133 ^a
Total Maximum Daily Load #1	--	10589	--	1378
Total Maximum Daily Load #2	--	9209	--	842

^a Includes Snohomish River and tributary loads; does not include seaward loads.

Table 13. Results of sensitivity analysis of EUTRO5 model predictions for dissolved oxygen changes at low (L) ebbing and high (H) slack tide..
% change after a 10% change in model input parameters

Model Steg#	Critical Conditions with Loads Dissolved Oxygen (mg/L)		River Flow		Upstream DO		Downstream DO		Nitrification Rate		Segment Temp.		Deoxygenation Rate		SOD	
	L	H	L	H	L	H	L	H	L	H	L	H	L	H	L	H
	34	6.72	5.49	1.9	0.9	1.9	0.4	1.5	7.1	0.2	0.2	4.8	3.1	0.6	0.9	1.8
36	7.15	5.58	1.7	1.4	2.7	0.9	0.6	5.6	0.3	0.4	4.8	4.3	0.4	0.9	1.5	1.8
43	8.33	6.93	0.7	1.6	4.1	2.2	0	1.4	0	0.3	3.4	4.5	0.2	0.4	0.7	1.4
46	8.53	7.62	0.5	1.4	4.6	3.0	0	0.5	0	0.1	2.8	4.2	0.1	0.4	0.5	1.2
50	6.22	5.19	1.6	0.8	1.3	0.4	2.1	6.4	0.2	0.2	4.8	4.1	0.8	1.0	2.4	2.3
55	6.64	5.28	2.0	1.3	2.0	0.8	1.1	5.3	0.3	0.4	5.1	4.6	0.6	1.0	2.3	2.8
56	6.85	5.39	1.9	1.9	2.2	0.7	0.6	4.1	0.3	0.2	5.1	5.2	0.6	1.1	2.0	3.2
57	7.04	5.61	1.4	2.0	2.6	1.3	0.4	3.0	0.3	0.4	5.1	5.5	0.4	0.9	1.9	2.9
58	7.22	6.04	1.7	2.2	2.6	2.2	0.4	1.5	0.3	0.5	5.0	6.1	0.4	0.8	1.7	2.7
64	6.49	5.05	2.0	1.6	1.7	0.6	1.2	4.0	0.2	0.2	5.1	6.5	0.6	1.0	2.3	4.2
66	7.21	5.69	1.9	2.1	2.6	1.2	0.4	2.8	0.1	0.4	4.9	5.6	0.6	0.9	1.8	3.0
68	7.49	6.43	1.5	1.7	2.9	1.9	0.3	1.2	0.1	0.3	4.7	5.4	0.4	0.6	1.5	2.0

Appendix A

Tides for August 27 and 28

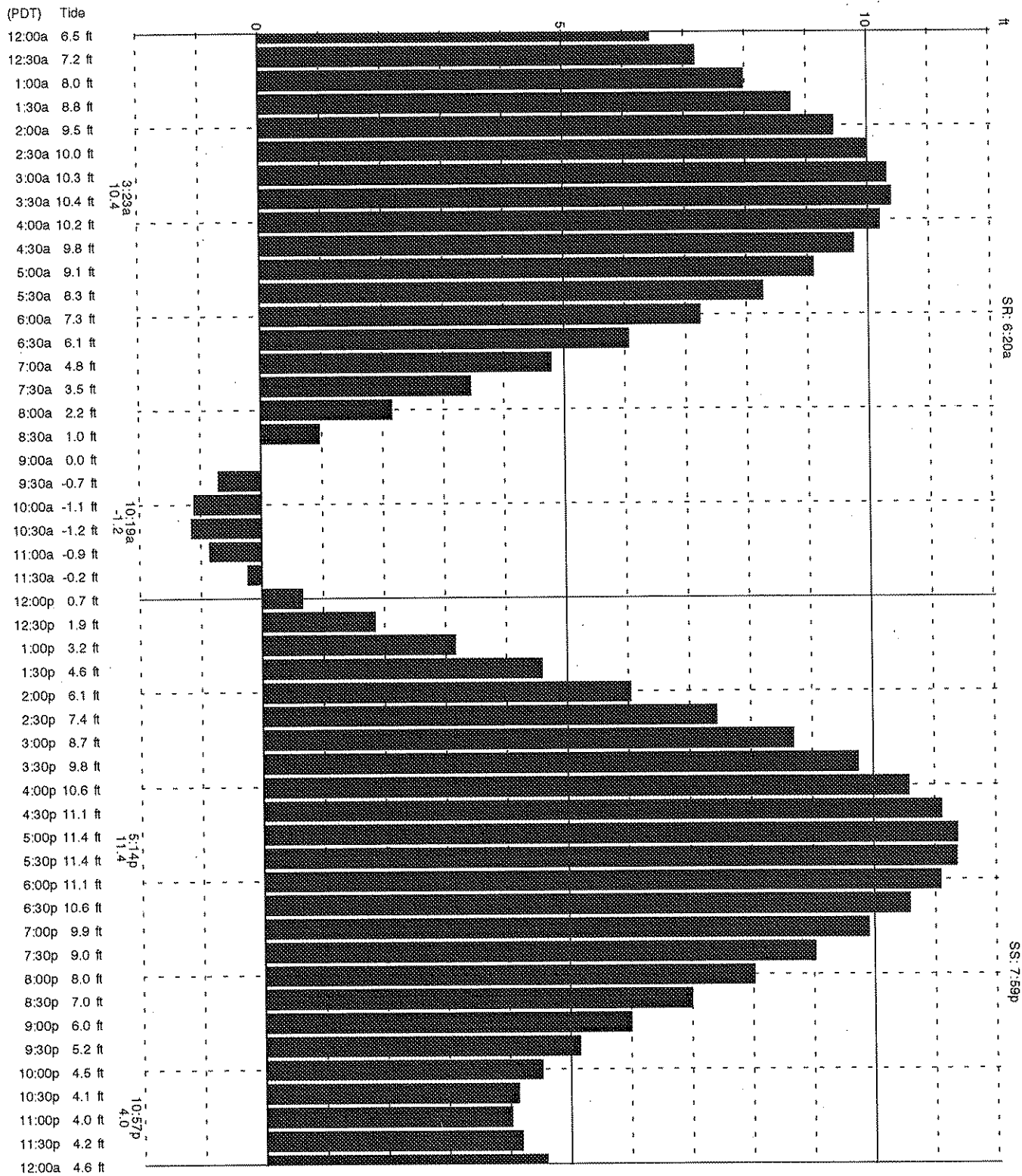
Tides-Tulalip

based on Seattle, Washington (NOAA)
48° 3.90 N 122° 17.30 W

Tuesday, August 27, 1996

Average Tides
Mean Range: 7.5 ft
MHHW: 11.1 ft
Mean Tide: 6.4 ft

Daily Highs & Lows
3:23a 10.4 ft High
10:19a -1.2 ft Low
5:14p 11.4 ft High
10:57p 4.0 ft Low



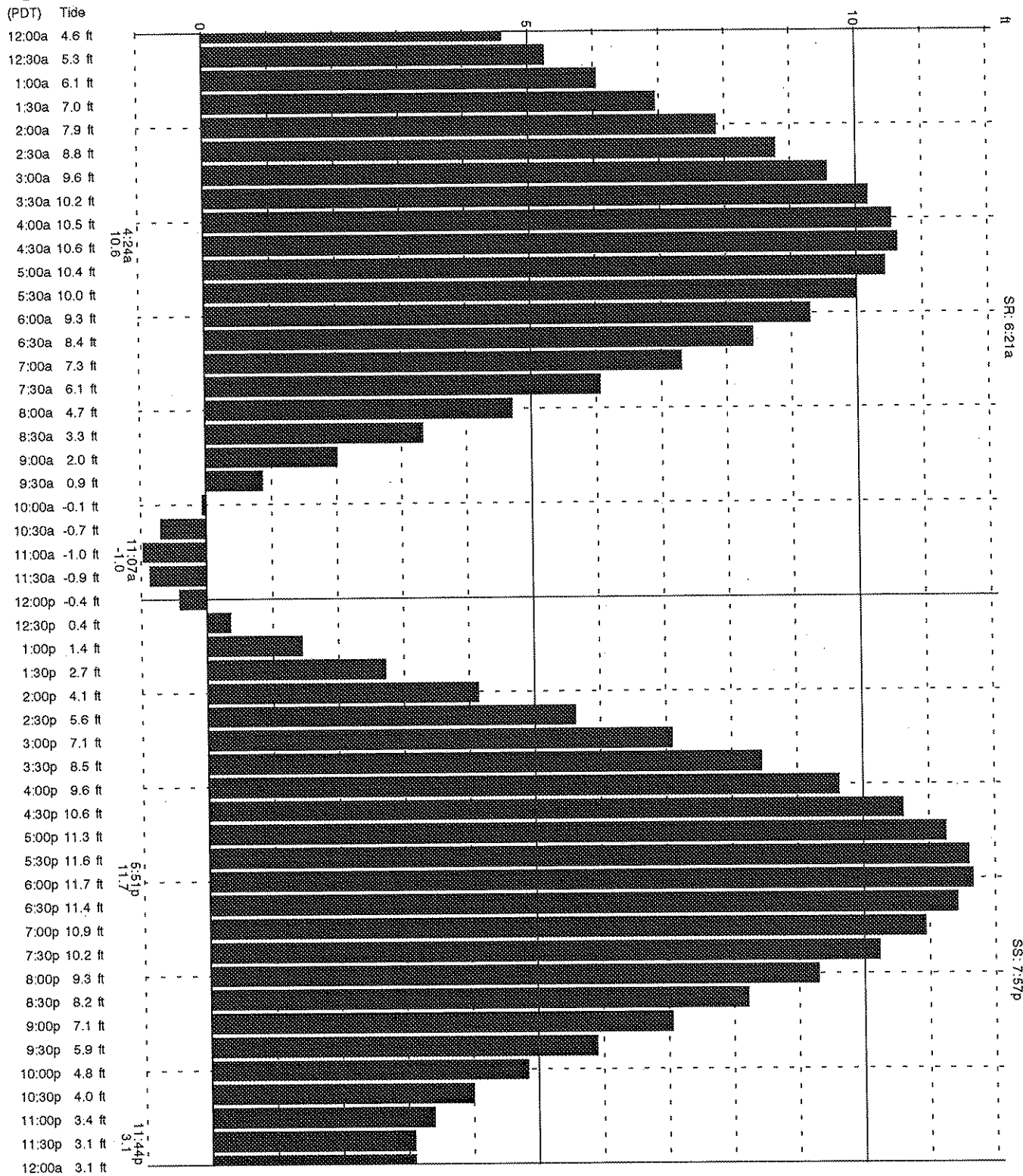
Tides-Tulalip

based on Seattle, Washington (NOAA)
48° 3.90 N 122° 17.30 W

Wednesday, August 28, 1996

Average Tides
Mean Range: 7.5 ft
MHHW: 11.1 ft
Mean Tide: 6.4 ft

Daily Highs & Lows
4:24a 10.6 ft High
11:07a -1.0 ft Low
5:51p 11.7 ft High
11:44p 3.1 ft Low



Appendix B

Sampling Schedules

Appendix B.1. Snohomish River TMDL Sampling Schedule

TEAM 1 8/27/96

Site	Location	Station		Sample Depth	Site Depth	Secchi Disc	pH Cond		Salinity	DO	UBOD	NO3-					Chloro a	
		ID	LAB ID Tide ^a				Temp					NH3	NO2	TPN	Ortho-P	TP		
14	Upriver of Snohomish WTP	SNO14	358600 Low	Surface	X	X	X		X			X	X	X	X	X	X	
15	1 mi. below Snohomish WTP	SNO15	358601 Low	Surface	X	X	X		X			X	X	X	X	X	X	
16	Upriver of Ebey Slough	SNO16	358602 Low	Surface	X	X	X		X			X	X	X	X	X	X	
			REP16	358603 Low	Surface	X	X	X		X			X	X	X	X	X	X
			358604 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	X
18	By Hwy 2 bridge	SNO18	358605 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358606 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358607 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
20	Upriver of Everett WTP	SNO20	358608 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358609 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358610 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
21	1 mi. downriver Everett WTP	SNO21	358611 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358612 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358613 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
30	Steamboat S. upriver Ebey S. junction	STM30	358614 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358615 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358616 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
34	Port Gardner	POG34	358617 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358618 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358619 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X	
			358620 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X		

^a Sampled both high and low tide; surface samples at low tide; surface and bottom samples at high tide

TEAM 1 8/28/96

Site	Location	Station		Sample Depth	Site Depth	Secchi Disc	pH Cond		Salinity	DO	UBOD	NO3-					Chloro a
		ID	LAB ID Tide ^a				Temp					NH3	NO2	TPN	Ortho-P	TP	
14	Upriver of Snohomish WTP	SNO14	358700 Low	Surface	X	X	X		X			X	X	X	X	X	X
15	1 mi. below Snohomish WTP	SNO15	358701 Low	Surface	X	X	X		X			X	X	X	X	X	X
16	Upriver of Ebey Slough	SNO16	358702 Low	Surface	X	X	X		X			X	X	X	X	X	X
			358703 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358704 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X
18	By Hwy 2 bridge	SNO18	358705 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358706 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358707 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X
20	Upriver of Everett WTP	SNO20	358708 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358709 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358710 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X
21	1 mi. downriver Everett WTP	SNO21	358711 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			REP21	358712 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X
			358713 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358714 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X
30	Steamboat S. upriver Ebey S. junction	STM30	358715 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358716 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358717 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X
34	Port Gardner	POG34	358718 Low	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358719 High	Surface	X	X	X	X	X	X	X	X	X	X	X	X	X
			358720 High	Bottom	X	X	X	X	X	X	X	X	X	X	X	X	X

^a Sampled both high and low tide; surface samples at low tide; surface and bottom samples at high tide

Appendix B.1. Snohomish River TMDL Sampling Schedule.

TEAM 2 8/27/96

Site	Location	Station		Sample Depth	Site Depth	Secchi Disc	pH		Salinity	DO	UBOD	NO3-			Ortho-P	TP	Chloro a	
		ID	LAB ID Tide ^a				Cond	Temp				NH3	NO2	TPN				
23	Upriver of Lake Stevens WTP	EBE23	358621	Low	Surface	X	X	X		X		X	X	X	X	X	X	
			358622	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358623	High	Bottom	X		X	X	X		X	X	X	X	X	X	X
24	Ebey S. downriver Lake Stevens WTP	EBE24	358624	Low	Surface	X	X	X		X		X	X	X	X	X	X	
			358625	High	Surface	X	X	X		X		X	X	X	X	X	X	X
			358626	High	Bottom	X		X	X	X		X	X	X	X	X	X	X
26	Downriver Allen Creek	EBE26	358627	Low	Surface	X	X	X		X		X	X	X	X	X	X	
			358628	High	Surface	X	X	X		X		X	X	X	X	X	X	X
			358629	High	Bottom	X		X	X	X		X	X	X	X	X	X	X
27	Mouth of Ebey Slough	EBE27	358630	Low	Surface	X	X	X		X		X	X	X	X	X	X	
			358631	High	Surface	X	X	X		X		X	X	X	X	X	X	X
			358632	High	Bottom	X		X	X	X		X	X	X	X	X	X	X
31	Steamboat S. Upriver Marysville WTP	STM31	358633	Low	Surface	X	X	X		X		X	X	X	X	X	X	
			358634	Low	Surface	X	X	X		X		X	X	X	X	X	X	X
		REP	358635	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
32	Mouth of Steamboat Slough	STM32	358636	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358637	Low	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358638	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
33	Union Slough at I-5	UNS33	358639	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358640	Low	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358641	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
35	Possession Sound	PS35	358642	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358643	High	Surface	X	X	X	X	X	X		X	X	X	X	X	X
			358644	High	Bottom	X		X	X	X	X		X	X	X	X	X	X

* Sampled both high and low tide; surface samples at low tide; surface and bottom samples at high tide.

TEAM 2 8/28/96

Site	Location	Station		Sample Depth	Site Depth	Secchi Disc	pH		Salinity	DO	UBOD	NO3-			Ortho-P	TP	Chloro a	
		ID	LAB ID Tide ^a				Cond	Temp				NH3	NO2	TPN				
23	Upriver of Lake Stevens WTP	EBE23	358721	Low	Surface	X	X	X		X		X	X	X	X	X	X	
			REP23	358722	Low	Surface	X	X	X		X		X	X	X	X	X	X
			358723	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
24	Ebey S. downriver Lake Stevens WTP	EBE24	358724	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358725	Low	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358726	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
26	Downriver Allen Creek	EBE26	358727	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358728	Low	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358729	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
27	Mouth of Ebey Slough	EBE27	358730	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358731	Low	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358732	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
31	Steamboat S. Upriver Marysville WTP	STM31	358733	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358734	Low	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358735	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
32	Mouth of Steamboat Slough	STM32	358736	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358737	Low	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358738	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
33	Union Slough at I-5	UNS33	358739	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358740	Low	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358741	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
35	Possession Sound	PS35	358742	High	Bottom	X		X	X	X		X	X	X	X	X	X	
			358743	High	Surface	X	X	X	X	X		X	X	X	X	X	X	X
			358744	High	Bottom	X	X	X	X	X		X	X	X	X	X	X	X

* Sampled both high and low tide; surface samples at low tide; surface and bottom samples at high tide.

Appendix B.1. Snohomish River TMDL Sampling Schedule.

TEAM 3 8/27/96

Site	Location	Station		Flow	Secchi Disc	pH		DO	UBOD	NH3	NO3-		TPN	Ortho-P	TP	Chloro a
		ID	LAB ID			Cond	Temp				NO2	NO2				
10	Downriver of Confluence	SNO10	358645			X	X	X		X	X	X	X	X	X	X
11	Upriver of French Creek	SNO11	358646			X	X	X	X	X	X	X	X	X	X	X
		REP11	358647			X	X	X	X	X	X	X	X	X	X	X
12	French Creek	FRN12	358648	a		X	X	X	X	X	X	X	X	X	X	X
13	Pilchuck River	PIL13	358649	X		X	X	X	X	X	X	X	X	X	X	X
17	Marshland	MAR1	358650	a		X	X	X	X	X	X	X	X	X	X	X
19	Deadwater Slough	DED19	358651	a		X	X	X	X	X	X	X	X	X	X	X
22	Swan Trail Slough	STS22	358652	a		X	X	X	X	X	X	X	X	X	X	X
25	Allen Creek	ALL25	358653	X		X	X	X	X	X	X	X	X	X	X	X
28	Quilceda Creek	QIL28	358654	X		X	X	X	X	X	X	X	X	X	X	X

(a) note if water was being pumped.

TEAM 3 8/28/96

Site	Location	Station		Flow	Secchi Disc	pH		DO	BOD35	NH3	NO3-		TPN	Ortho-P		TP	Chloro a
		ID	LAB ID			Cond	Temp				NO2	NO2		P	P		
10	Downriver of Confluence	SNO10	358745			X	X	X		X	X	X	X	X	X	X	X
11	Upriver of French Creek	SNO11	358746			X	X	X		X	X	X	X	X	X	X	X
12	French Creek	FRN12	358747	a		X	X	X		X	X	X	X	X	X	X	X
13	Pilchuck River	PIL13	358748	X		X	X	X		X	X	X	X	X	X	X	X
17	Marshland	MAR1	358749	a		X	X	X		X	X	X	X	X	X	X	X
		REP17	358750			X	X	X		X	X	X	X	X	X	X	X
19	Deadwater Slough	DED19	358751	a		X	X	X		X	X	X	X	X	X	X	X
22	Swan Trail Slough	STS22	358752	a		X	X	X		X	X	X	X	X	X	X	X
25	Allen Creek	ALL25	358753	X		X	X	X		X	X	X	X	X	X	X	X
28	Quilceda Creek	QIL28	358754	X		X	X	X		X	X	X	X	X	X	X	X

(a) note if water was being pumped.

Appendix B.2. Snohomish River TMDL WWTP effluent Sampling Schedule.

Parameter	Composite Sample	Grab Sample 1	Grab Sample 2
Solids⁴	X	X	X
Total solids			
Total non-volatile solids			
Total suspended solids			
Total non-volatile suspended solids			
BOD₅	X	X	X
UBOD	X		
NH₃	X	X	X
NO₃-NO₂	X	X	X
TKN	X	X	X
Ortho-P	X	X	X
TP	X	X	X
Temperature	X	X	X
pH	X	X	X
DO		X	X
Chlorine free & combined		X	X

Appendix C

1996 Ecology Survey Data

Appendix C.1. Laboratory data and field measurements for samples collected August 27 and 28, 1996

Date	Station	LAB ID	ID	Time (hour)	Tide	Sample Depth (m)	Cond (umho/cm)	Sal (o/oo)	Calibrated Sal (o/oo)	Measured Sal (o/oo)	Temp (C°)	DO (mg/L)	pH (S.U.)	TPN (mg/L)	NH3 (mg/L)	NO2-NO3 (mg/L)	NO2 (mg/L)	TP (mg/L)	Ortho-P (mg/L)	BOD35 (mg/L)	Chloro a (ug/L)	Pheno a (ug/L)	
27-Aug	SNO18			1437		a	24300	19.9			18.0												
27-Aug	SNO20	1431	a	1800		a	1800	1.3			18.5												
27-Aug	SNO21	1420	a	610		a	610	0.4			18.0												
27-Aug	EBE23	1505	a	3070		a	3070	1.9			19.4												
27-Aug	STM31	1450	a	19700		a	19700	13.7			18.5												
27-Aug	STM32	1437	a	29300		a	29300	22.2			17.0												
28-Aug	SNO18	1438	a	425		a	425	0.2			19.9												
28-Aug	SNO20	1433	a	520		a	520	0.3			20.0												
28-Aug	SNO21	1421	a	9250		a	9250	6.8			20.0												
28-Aug	EBE23	1424	a	1100		a	1100	0.8			19.3												
28-Aug	STM31	1440	a	11500		a	11500	7.8			19.2												
28-Aug	STM32	1450	a	26700		a	26700	19.0			18.9												
27-Aug	358234 E1			1231	LOW	0.5					9.5	0.148	0.01 U	0.01 U	0.01 U	0.01 U	0.01 U	0.047	0.017		12.9	0.9	
27-Aug	358235 E1			1236	LOW	5					6.6	0.335	0.01 U	0.01 U	0.287	0.01 U	0.051	0.058			1.2	0.8	
27-Aug	358236 E1			1241	LOW	10					5.6	0.356	0.01 U	0.01 U	0.301	0.01 U	0.066	0.060			0.8	1.0	
27-Aug	358241 E1			1803	HIGH	0.5					10.3	0.120	0.01 U	0.01 U	0.01 U	0.01 U	0.028	0.013			11.3	1.1	
27-Aug	358242 E1			1808	HIGH	5					10.4	0.125	0.01 U	0.01 U	0.01 U	0.01 U	0.022	0.011			12.6	1.9	
27-Aug	358243 E1			1813	HIGH	10					7.6	0.277	0.01 U	0.01 U	0.196	0.01 U	0.041	0.045			5.4	1.3	
27-Aug	358232 E3			1130	LOW	0.5					8.1				0.07	0.148		0.035			5.2	2.6	
27-Aug	358233 E3			1135	LOW	3					8.0				0.013	0.099		0.035			7.3	1.8	
27-Aug	358239 E3			1803	HIGH	0.5					10.8	0.01 U	0.01 U	0.01 U	0.01 U	0.020	0.01 U	0.020	0.01 U		7.6	1.0	
27-Aug	358240 E3			1813	HIGH	5					9.5	0.083	0.01 U	0.01 U	0.192	0.01 U	0.032	0.032			12.0	1.5	
27-Aug	358230 E5			1045	LOW	0.5					8.3	0.084	0.01 U	0.01 U	0.131	0.01 U	0.040	0.040			4.6	3.5	
27-Aug	358231 E5			1050	LOW	3					9.7	0.01 U	0.01 U	0.025	0.01 U	0.025	0.200	0.200			10.4	1.4	
27-Aug	358237 E5			1630	HIGH	0.5					8.0	0.280	0.01 U	0.01 U	0.158	0.01 U	0.040	0.040			6.8	1.3	
27-Aug	358238 E5			1635	HIGH	5					7.6	0.306	0.01 U	0.01 U	0.141	0.01 U	0.041	0.041			6.3	1.5	
28-Aug	358250 E1			1142	LOW	0.5					6.5	0.305	0.01 U	0.01 U	0.212	0.01 U	0.051	0.048			3.5	1.4	
28-Aug	358249 E1			1147	LOW	5					6.7	0.296	0.01 U	0.01 U	0.240	0.01 U	0.062	0.058			2.4	1.7	
28-Aug	358248 E1			1152	LOW	10					8.5	0.281	0.01 U	0.01 U	0.139	0.01 U	0.043	0.041			8.2	1.2	
28-Aug	358260 E1			1820	HIGH	0.5					8.1	0.352	0.01 U	0.01 U	0.192	0.01 U	0.049	0.048			4.7	1.4	
28-Aug	358259 E1			1825	HIGH	5					7.4				0.01 U	0.228	0.054	0.054			2.1	0.9	
28-Aug	358247 E3			1830	HIGH	10					7.3	0.103	0.01 U	0.01 U	0.192	0.01 U	0.046	0.046			2.9	3.4	
28-Aug	358246 E3			1043	LOW	3					7.7	0.01 U	0.01 U	0.164	0.01 U	0.164	0.045	0.045			5.4	2.0	
28-Aug	358257 E3			1706	HIGH	0.5					8.1	0.017	0.01 U	0.01 U	0.123	0.01 U	0.042	0.042			7.1	0.6	
28-Aug	358256 E3			1711	HIGH	5					7.2	0.106	0.01 U	0.01 U	0.206	0.01 U	0.050	0.050			5.4	1.3	
28-Aug	358245 E5			1009	LOW	0.5					7.9	0.081	0.01 U	0.01 U	0.193	0.01 U	0.049	0.049			3.3	3.8	
28-Aug	358244 E5			1014	LOW	3					6.8	0.081	0.01 U	0.01 U	0.182	0.01 U	0.046	0.046			4.2	6.6	
28-Aug	358255 E5			1634	HIGH	0.5					7.3	0.01 U	0.01 U	0.191	0.01 U	0.191	0.050	0.050			4.3	1.1	
28-Aug	358254 E5			1639	HIGH	5					6.4	0.01 U	0.01 U	0.250	0.01 U	0.250	0.057	0.057			2.5	1.3	

a Represent average of profile data

U The analyte was not detected at or above the reported limit

Appendix C.3. Laboratory data and field measurements for samples collected from WWTP effluents.

General Chemistry Results from City of Everett WWTPs, 8/96.

	Location:	EffEV1-E	EffEV1-1	EffEV1-2	EffEV2-E	EffEV2-1	EffEV2-2
	Type:	Comp	Grab	Grab	Comp	Grab	Grab
	Date:	8/27-28	8/27	8/28	8/27-28	8/27	8/28
	Time:	24 hour	pm	am	24 hour	pm	am
Parameter	Lab Log #: 3581	-45	-46	-47	-50	-51	-52
FIELD SAMPLING							
SOLIDS 4 (mg/L)							
TS		337	362	472	414	392	342
TNVS		221	144 J	240	278	251	233
TSS		7	12	6	9	9	12
TNVSS		1	1	2	3	2	1
BOD5 (mg/L)		10 J	14 J	10 J	12 J	14 J	12 J
NH3-N (mg/L)		17.3	17.1	12.2	11.4	10.7	17.6
NO2+NO3-N (mg/L)		0.179	0.229	10.9	11.3	11.2	0.110
NO2 (mg/L)		0.133	0.163	1.27	1.14	1.10	0.083
Total Kjeldahl N (mg/L)		20.6	21.4	15.0	14.4	13.1	22.8
Phosphate - Ortho (mg/L)		4.76	4.70	4.09	4.05	4.13	4.76
Phosphate - Total (mg/L)		4.79	4.60	3.88	4.28	4.25	2.01
FIELD OBSERVATIONS							
Flow (MGD)		5.30			5.63		
Temperature (°C.)		2.5*	22.0	21.1	3.0*	20.7	20.5
pH (s.u.)		7.59	7.60	7.54	6.93	6.90	6.68
D.O. (mg/L)		--	5.4	4.8	--	3.9	3.65

EffEV1 - Effluent to outfall 015; EffEV2 - Effluent to outfall 015a; -E - Ecology sampler.

J means the analyte was positively identified. The associated numerical result is an estimate.

* - Iced composite sample.

General Chemistry Results from Lake Stevens' S.D. WWTP, 8/96.

	Location:	EffLS-E	EffLS-1	EffLS-2
	Type:	Comp	Grab	Grab
	Date:	8/27-28	8/27	8/28
	Time:	24 hour	pm	am
Parameter	Lab Log #: 3581	-35	-36	-37
FIELD SAMPLING				
SOLIDS 4 (mg/L)				
TS		251	256	249
TNVS		134	133	156
TSS		5	4	4
TNVSS		1 U	1 U	1 U
BOD5 (mg/L)		6 J	4 J	4 J
NH3-N (mg/L)		17.3	13.7	12.7
NO2+NO3-N (mg/L)		0.070	0.180	0.282
NO2 (mg/L)		0.089	0.198	0.346
Total Kjeldahl N (mg/L)		17.0	16.0	16.5
Phosphate - Ortho (mg/L)		1.070	0.677	0.618
Phosphate - Total (mg/L)		1.320	0.894	0.848
FIELD OBSERVATIONS				
Flow (MGD)		1.24		
Temperature (°C.)		5.4	19.0	19.3
pH (s.u.)		7.31	7.38	6.99
D.O. (mg/L)		--	6.7	6.1

EffLS - Effluent; -E - Ecology sampler.

-1 - Grab sample taken on 8/27; -2 - Grab sample taken on 8/28.

U means the analyte was not detected at or above the reported result.

J means the analyte was positively identified. The numerical result is an estimate.

* - Iced composite sample.

Appendix C.3. Laboratory data and field measurements for samples collected from WWTP effl

General Chemistry Results from City of Marysville WWTP, 8/96.

	Location:	EffMA-E	EffMA-1	EffMA-2
	Type:	Comp	Grab	Grab
	Date:	8/27-28	8/27	8/28
	Time:	24 hour	pm	am
Parameter	Lab Log #: 3581	-40	-41	-42
FIELD SAMPLING				
SOLIDS 4 (mg/L)				
TS		419	*	459
TNVS		244	*	184
TSS		70	*	49
TNVSS		8	*	5
BOD5 (mg/L)		15 G	*	15 G
NH3-N (mg/L)		1.91	*	2.72
NO2+NO3-N (mg/L)		5.40	*	4.88
NO2 (mg/L)		1.60	*	1.03
Total Kjeldahl N (mg/L)		12.7	*	11.4
Phosphate - Ortho (mg/L)		2.39	*	2.21
Phosphate - Total (mg/L)		4.60	*	3.83
FIELD OBSERVATIONS				
Flow (MGD)		1.3		
Temperature (°C.)		6.2**	*	21.4
pH (s.u.)		7.36	*	7.11
D.O. (mg/L)		--	*	1.65

EffMA - Effluent; -E - Ecology sampler.

-1 - Grab sample taken on 8/27; -2 - Grab sample taken on 8/28.

G means that the value is greater than the result reported due to insufficient dilution of the samples. Result is a calculated estimate.

* - Samples not collected or observations not made.

** - Iced composite sample.

General Chemistry Results from City of Snohomish WWTP, 8/96.

	Location:	EffSN-E	EffSN-1	EffSN-2
	Type:	Comp	Grab	Grab
	Date:	8/27-28	8/27	8/28
	Time:	24 hour	pm	am
Parameter	Lab Log #: 3581	-30	-31	32
FIELD SAMPLING				
SOLIDS 4 (mg/L)				
TS		263	*	279
TNVS		146	*	159
TSS		6	*	5
TNVSS		1	*	1
BOD5 (mg/L)		6 J	*	3 J
NH3-N (mg/L)		0.455	*	0.490
NO2+NO3-N (mg/L)		7.76	*	8.04
NO2 (mg/L)		0.181	*	0.18
Total Kjeldahl N (mg/L)		2.84	*	2.42
Phosphate - Ortho (mg/L)		4.55	*	4.51
Phosphate - Total (mg/L)		4.27	*	4.44
FIELD OBSERVATIONS				
Flow (MGD)		0.51		
Temperature (°C.)		6.4**	*	20.6
pH (s.u.)		7.69	*	7.51
D.O. (mg/L)		--	*	8.7

EffSN - Effluent; -E - Ecology sampler.

-1 - Grab sample taken on 8/27; -2 - Grab sample taken on 8/28.

J means the analyte was positively identified. The associated numerical result is an estimate.

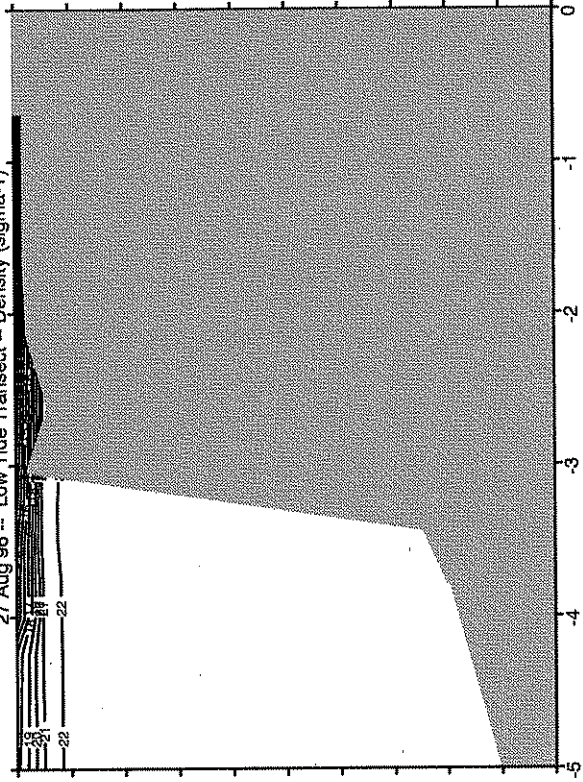
* - Samples not collected or observations not made.

** - Iced composite sample.

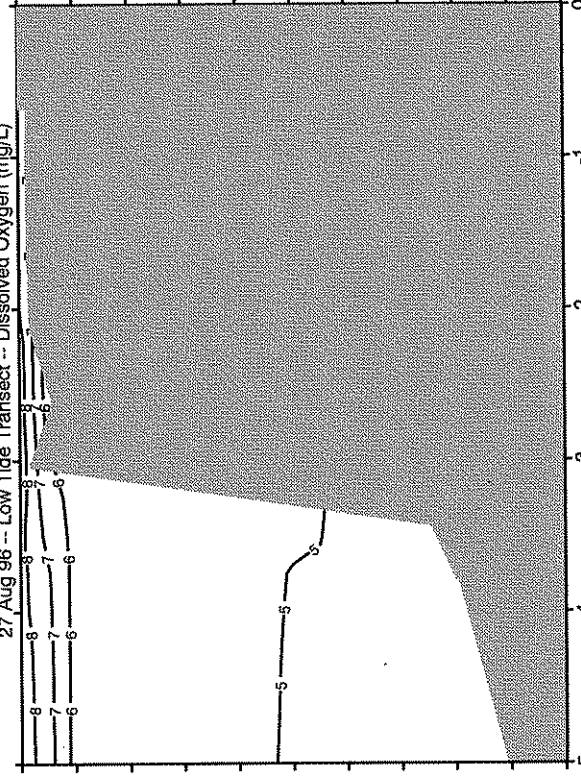
Appendix D

Contour Plots of 1996 Profile Transect Data

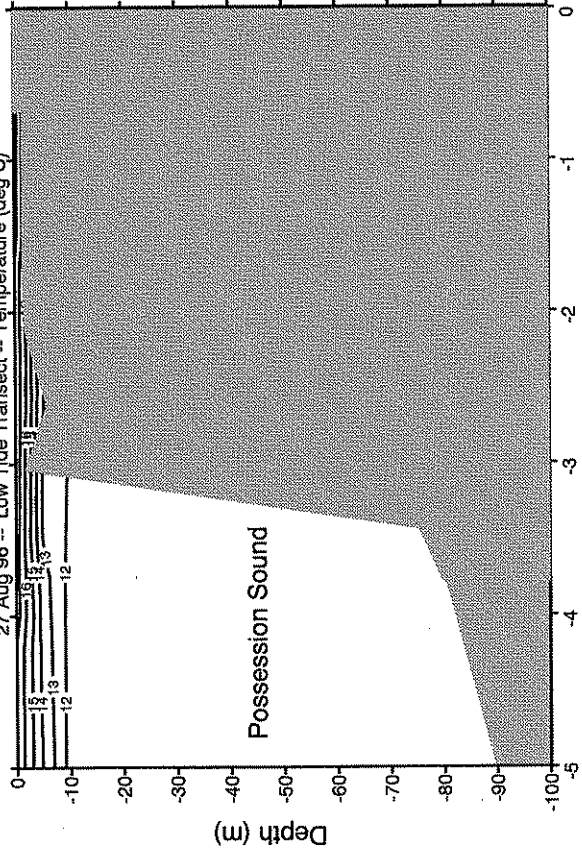
27 Aug 96 -- Low Tide Transect -- Density (sigma-T)



27 Aug 96 -- Low Tide Transect -- Dissolved Oxygen (mg/L)

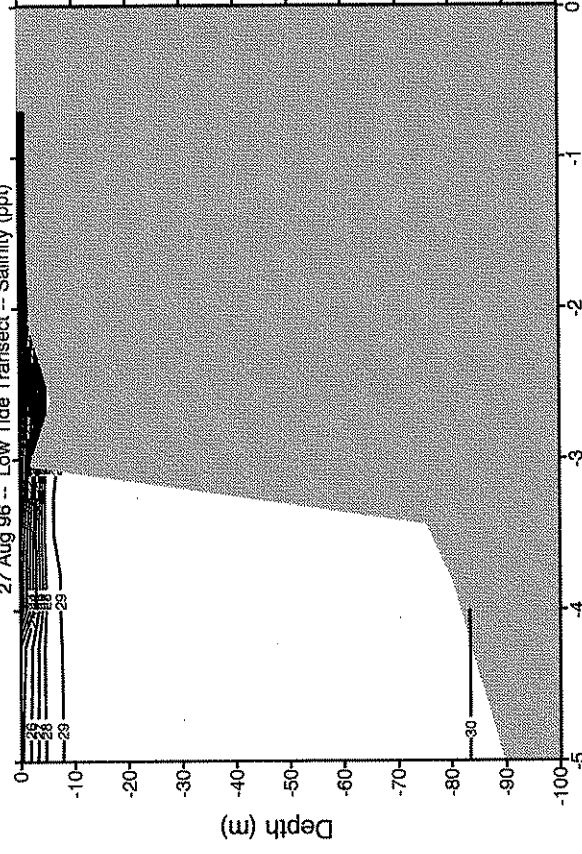


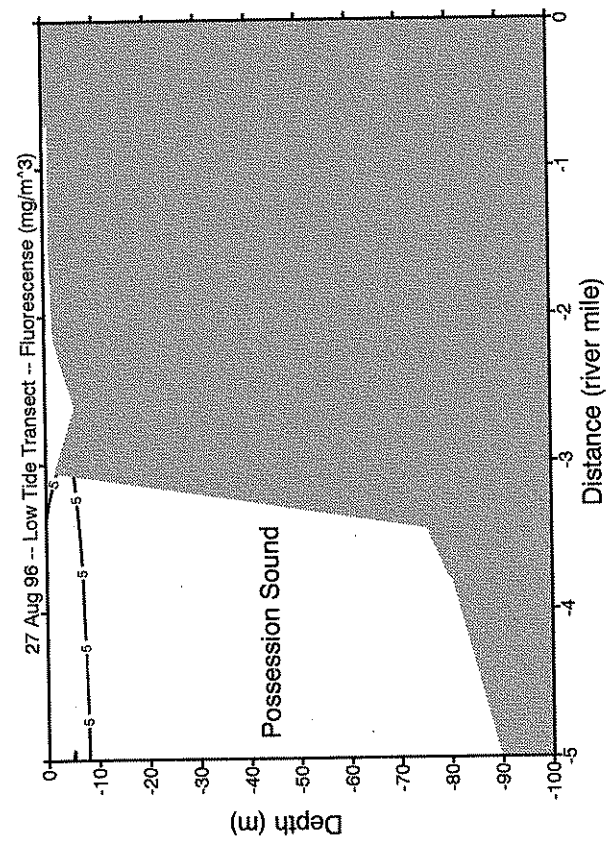
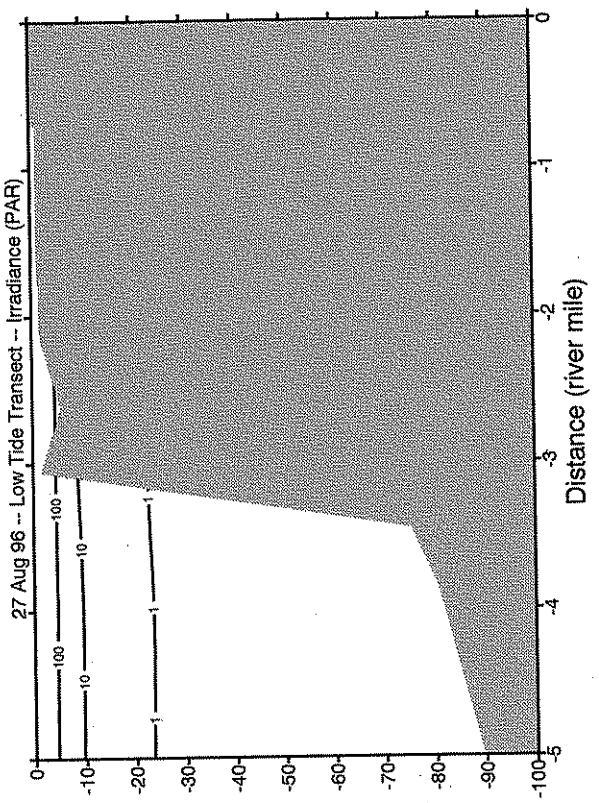
27 Aug 96 -- Low Tide Transect -- Temperature (deg C)

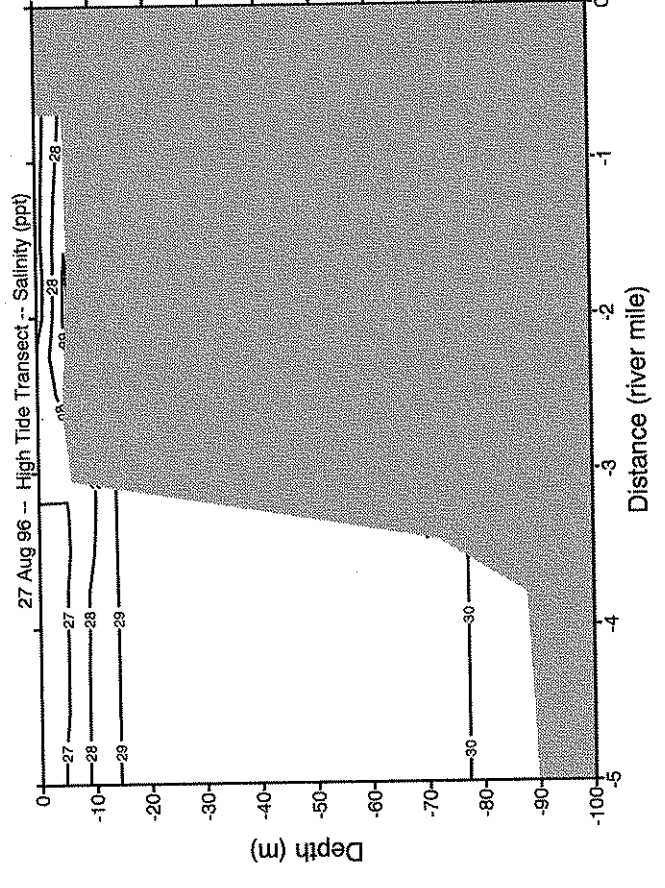
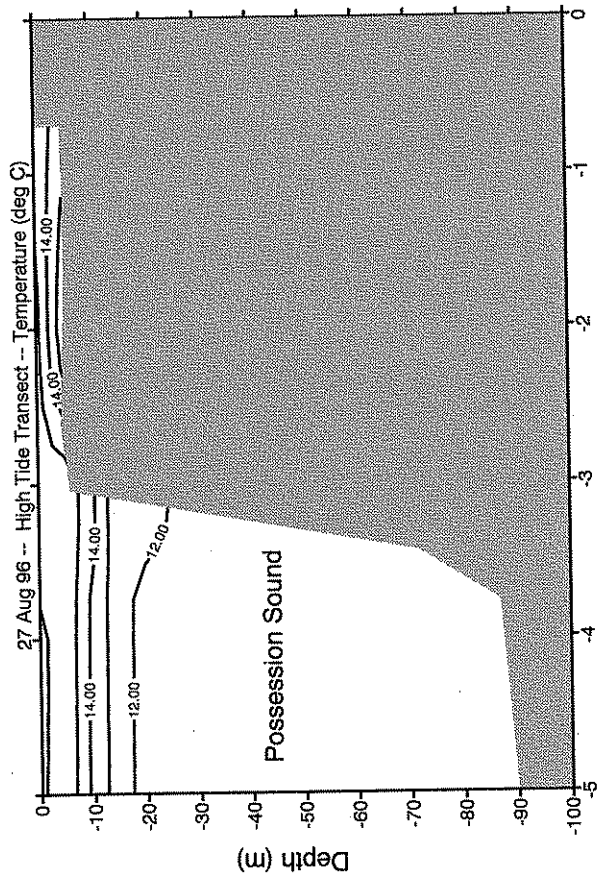
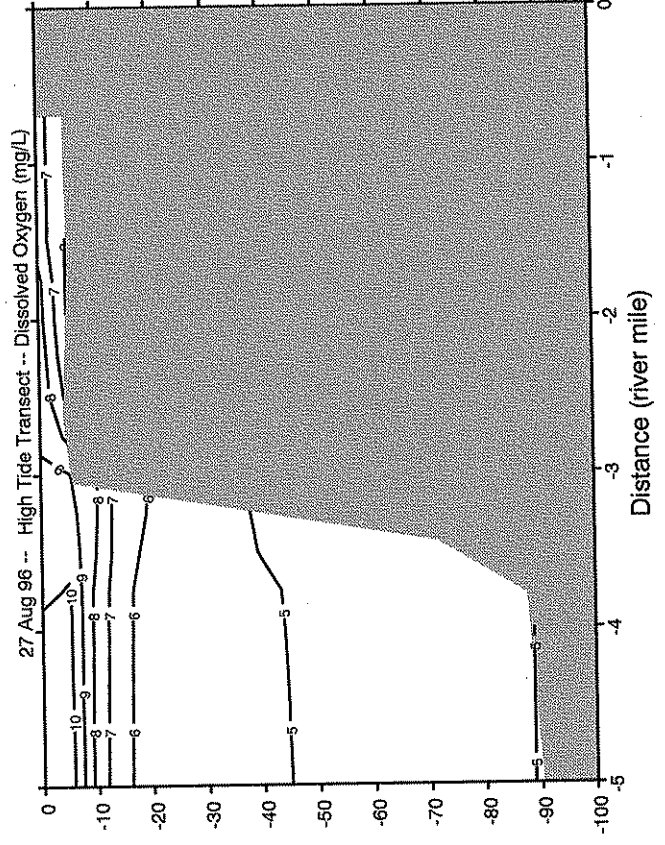
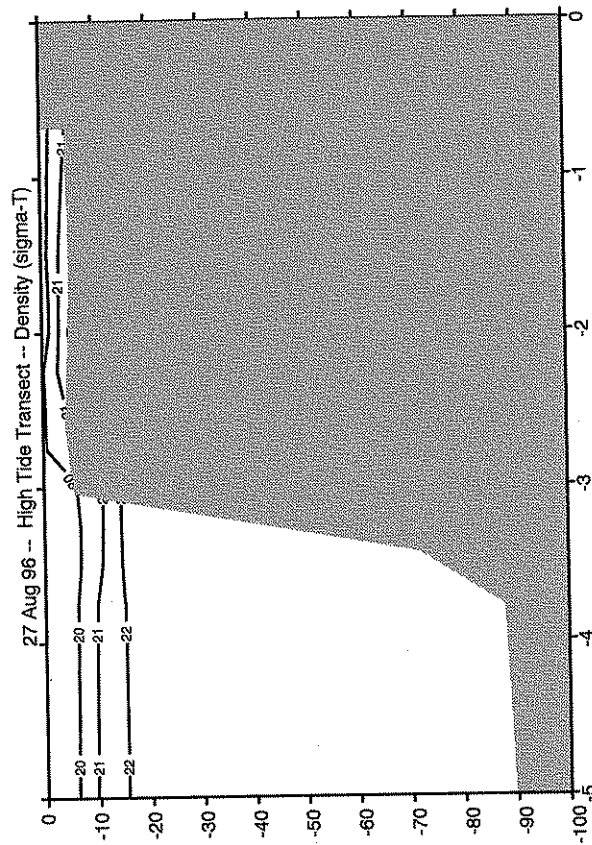


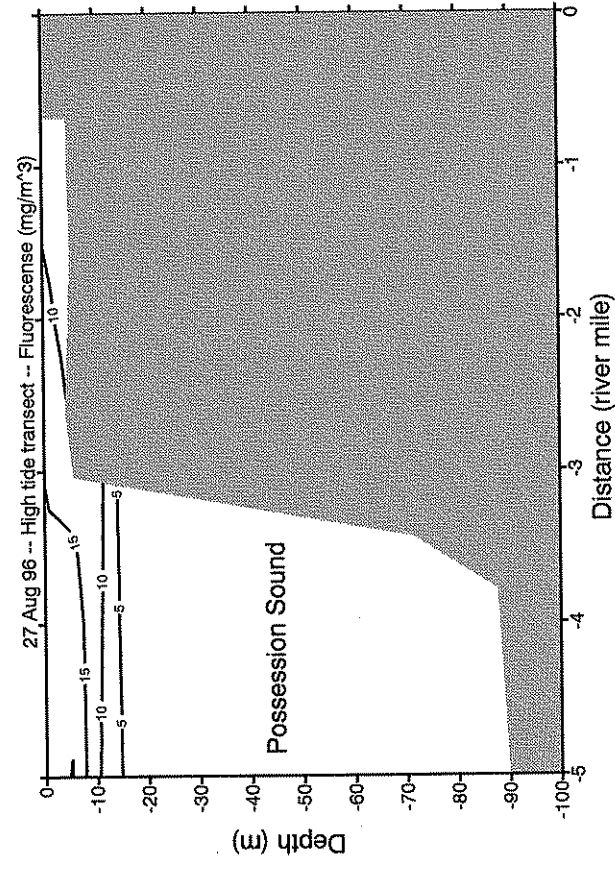
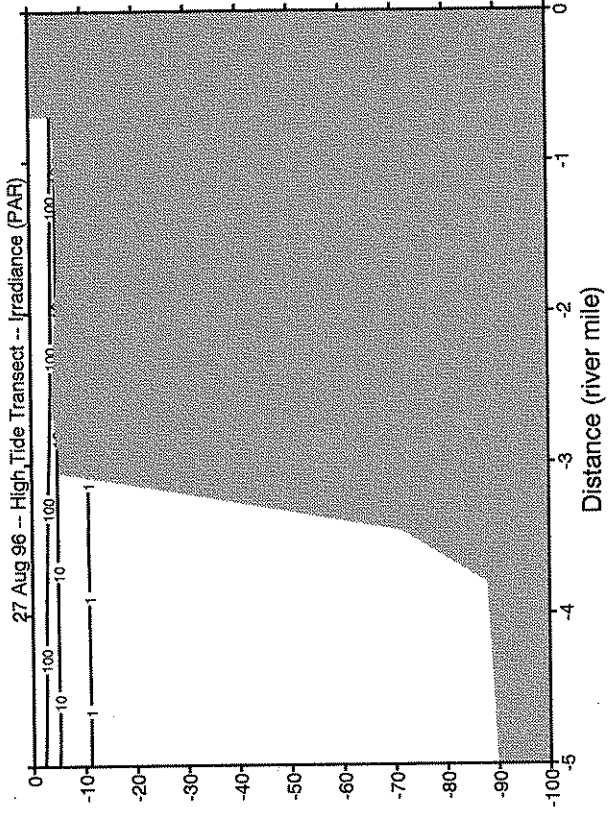
Possession Sound

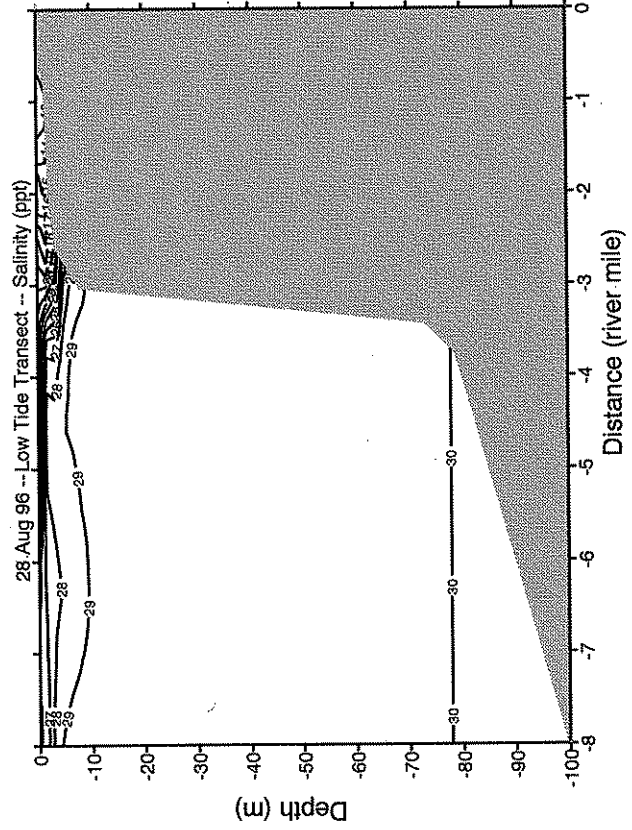
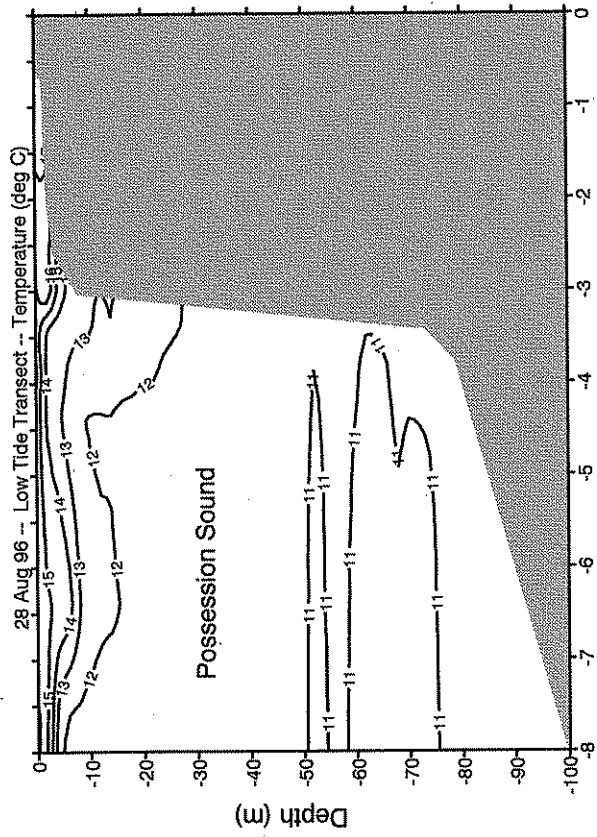
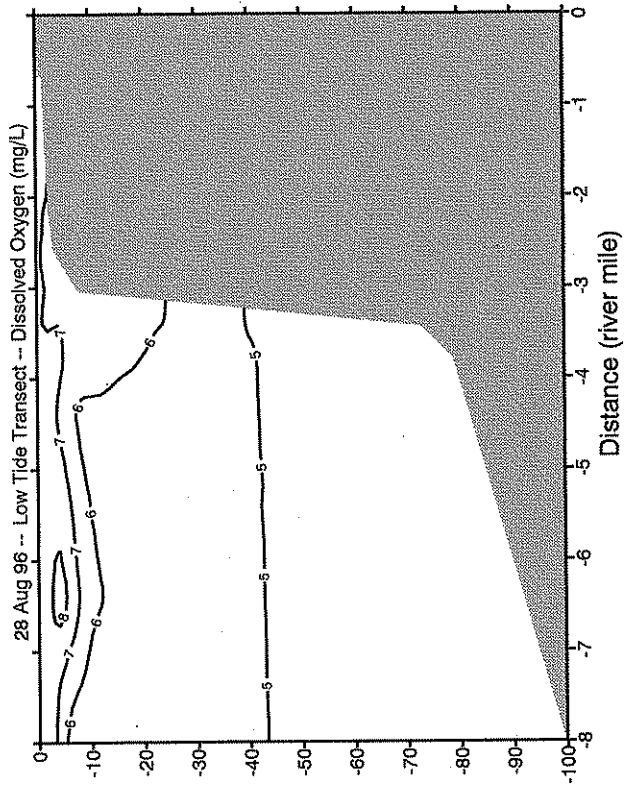
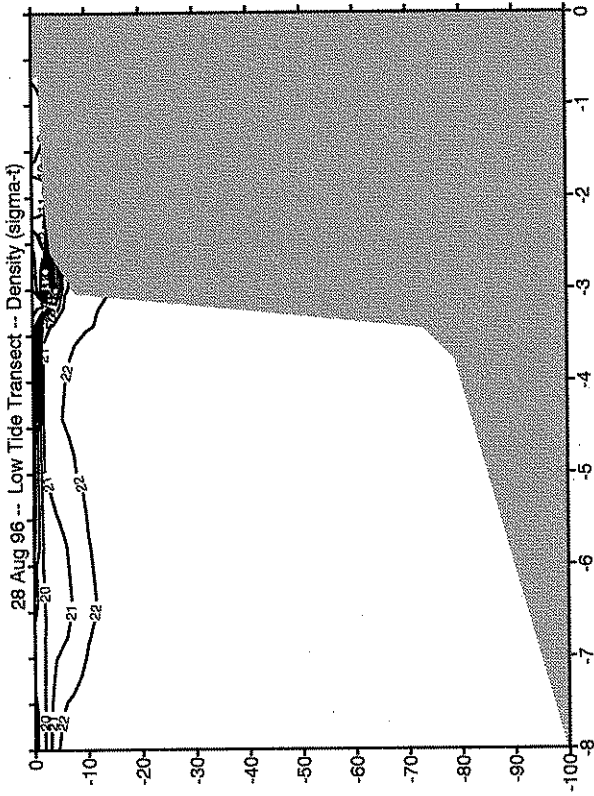
27 Aug 96 -- Low Tide Transect -- Salinity (ppt)

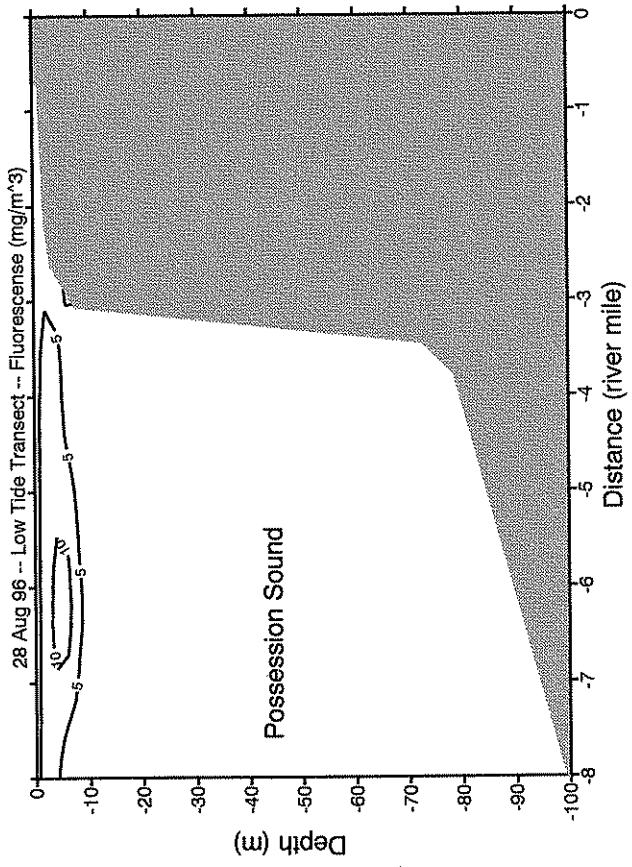
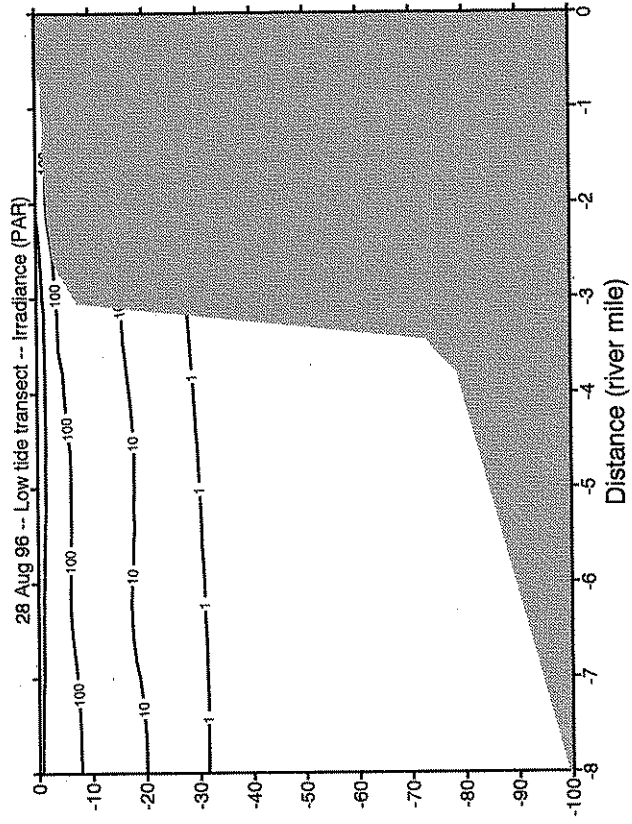


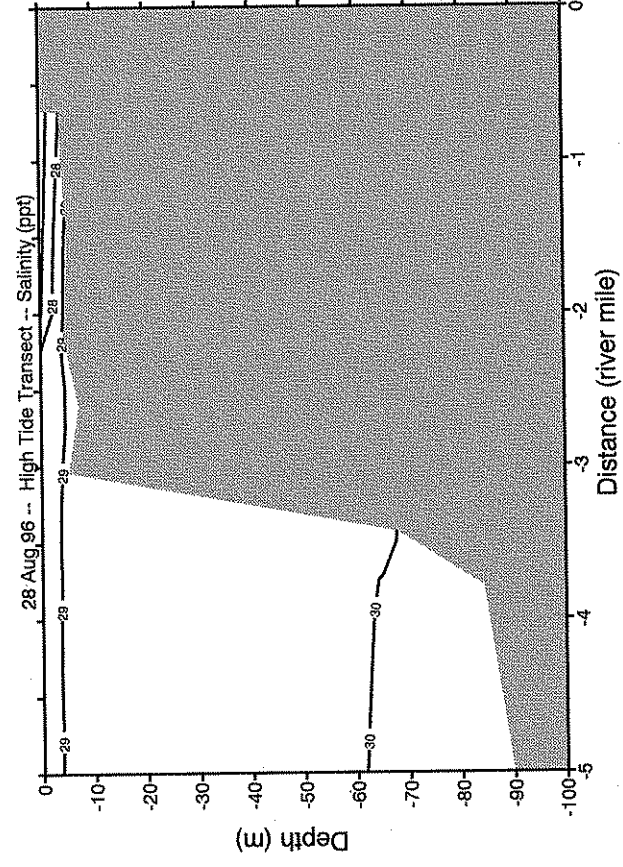
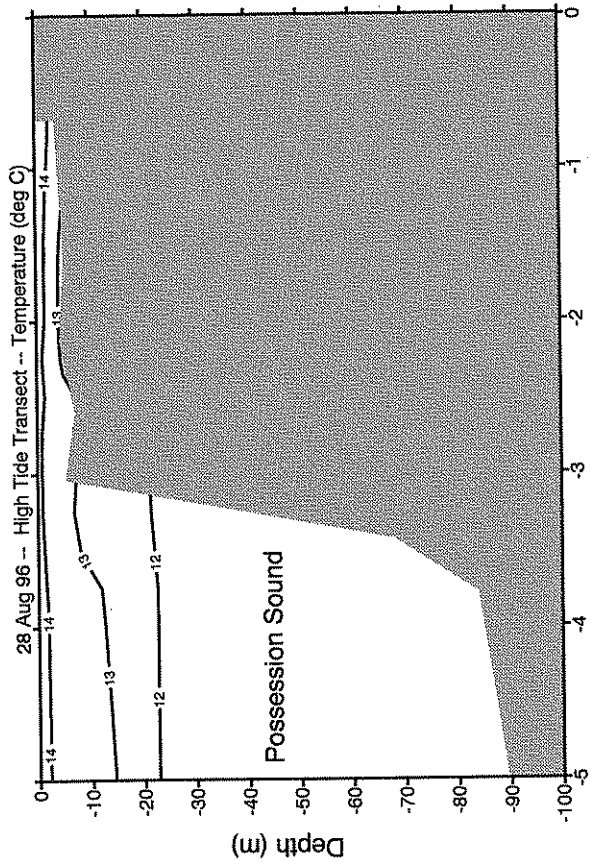
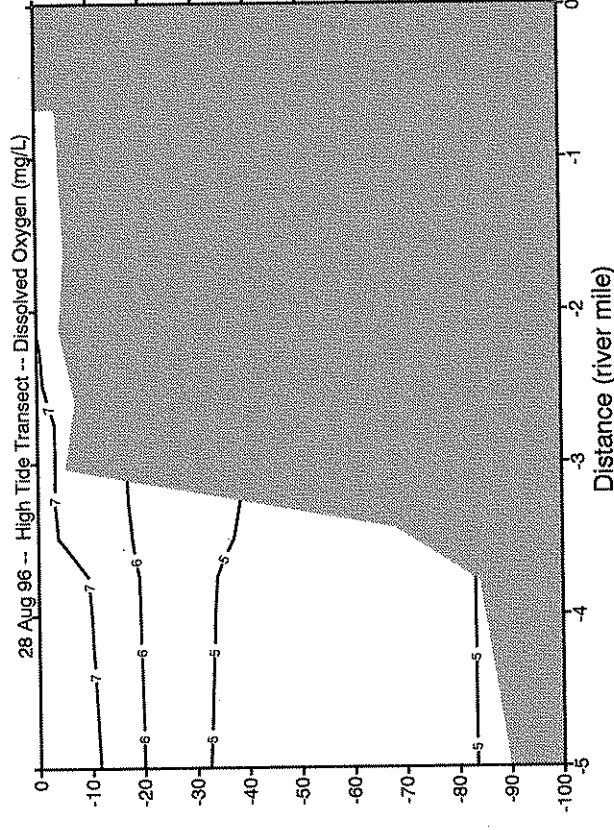
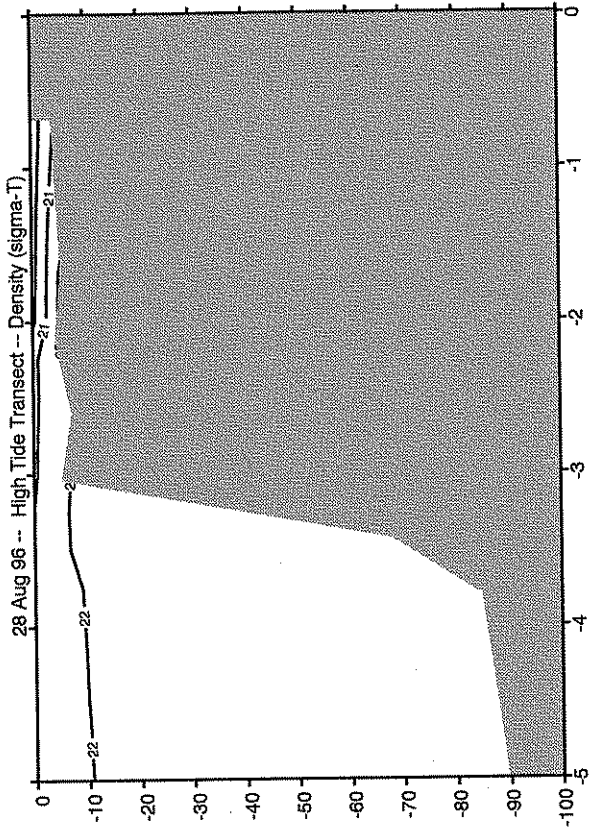


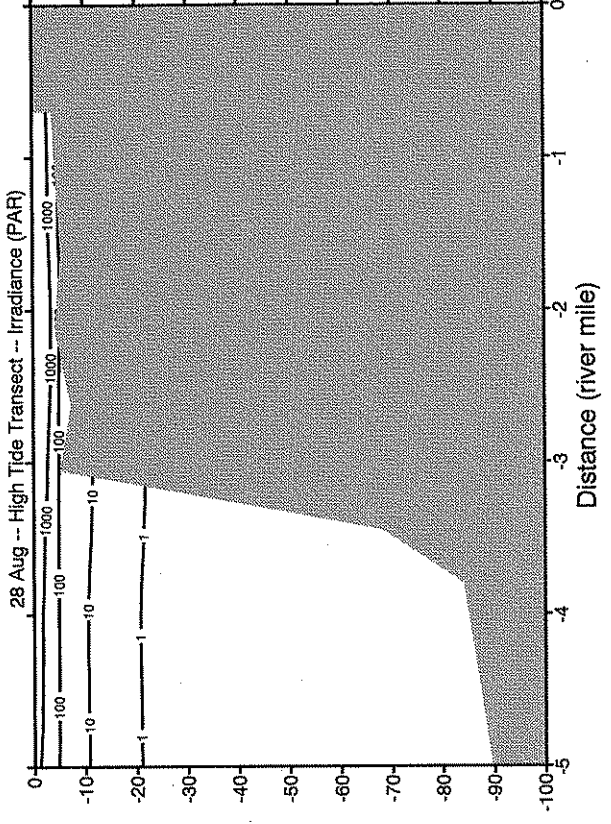
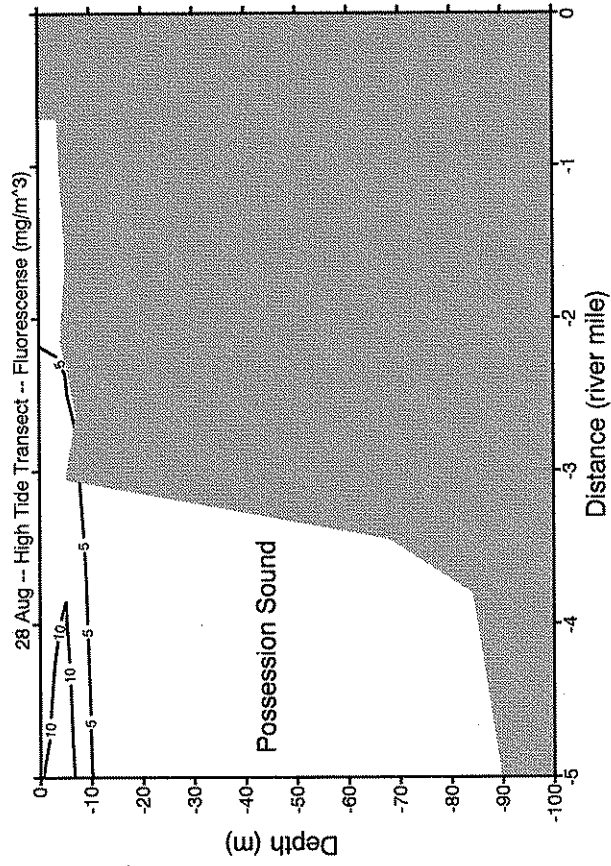












Appendix E

WASP5 Input Files

DYNHYD5 Input File

Critical Conditions

SNOTIDEL.INP, TIME=0 CRITICAL CONDITIONS HYDRODYNAMIC FILE FOR EUTROWASE
 ROBERT CUSIMANO, WA. DEPT. OF ECOLOGY, October 25, 1996

*****A: PROGRAM CONTROL DATA*****

83 115 0000 45.0 5 1 0000 25.0 0000 [Time Step= 45 sec.]

*****B: PRINTOUT CONTROL DATA*****

1.0 1.0 0

1 6 9 11 15 17

*****C: SUMMARY CONTROL DATA*****

1 1 0100 25.0 4 2

*****D: JUNCTION DATA*****

Junction Number	Head (m)	Surface Area (m ²)	Depth MLLW (m)	Channel number entering junction																
1	1.6943	2226443.	-5.0	1	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
2	1.6976	2101893.	-5.0	2	3	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
3	1.7010	1854725.	-5.0	4	5	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
4	1.7045	1731141.	-5.0	6	7	8	0	0	0	0	0	0	0	0	0	0	0	0	0	0
5	1.7080	1669349.	-5.0	9	10	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
6	1.7116	2226443.	-5.0	11	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
7	1.6929	2803443.	-5.0	1	2	12	13	14	0	0	0	0	0	0	0	0	0	0	0	0
8	1.6955	3025846.	-5.0	3	4	14	15	16	17	0	0	0	0	0	0	0	0	0	0	0
9	1.6990	2198943.	-5.0	5	6	16	18	19	0	0	0	0	0	0	0	0	0	0	0	0
10	1.7023	2045937.	-5.0	7	19	20	21	31	32	0	0	0	0	0	0	0	0	0	0	0
11	1.7052	2713145.	-5.0	8	9	21	22	33	34	0	0	0	0	0	0	0	0	0	0	0
12	1.7067	2208976.	-5.0	10	11	22	35	0	0	0	0	0	0	0	0	0	0	0	0	0
13	1.6925	1932227.	-5.0	12	23	25	0	0	0	0	0	0	0	0	0	0	0	0	0	0
14	1.6944	2374524.	-5.0	13	15	23	24	26	27	0	0	0	0	0	0	0	0	0	0	0
15	1.6977	2262487.	-5.0	17	18	20	24	28	29	30	0	0	0	0	0	0	0	0	0	0
16	1.7002	1886242.	-0.9	30	31	41	42	43	0	0	0	0	0	0	0	0	0	0	0	0
17	1.7017	2328539.	-0.9	32	33	37	38	39	40	41	0	0	0	0	0	0	0	0	0	0
18	1.7032	1826042.	-3.9	34	35	36	37	0	0	0	0	0	0	0	0	0	0	0	0	0
19	1.6931	1370368.	-2.2	25	26	44	45	48	49	0	0	0	0	0	0	0	0	0	0	0
20	1.6940	999976.	-5.0	27	28	49	50	54	0	0	0	0	0	0	0	0	0	0	0	0
21	1.6964	1320202.	-2.1	29	43	50	51	0	0	0	0	0	0	0	0	0	0	0	0	0
22	1.6984	1375385.	-1.9	40	42	51	52	0	0	0	0	0	0	0	0	0	0	0	0	0
23	1.6975	1367860.	-0.9	39	52	53	61	91	0	0	0	0	0	0	0	0	0	0	0	0
24	1.6976	1293447.	-3.2	36	38	91	92	0	0	0	0	0	0	0	0	0	0	0	0	0
25	1.6928	128676.	-2.1	48	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
26	1.6930	147822.	-1.9	44	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
27	1.6915	349195.	-1.5	45	46	47	0	0	0	0	0	0	0	0	0	0	0	0	0	0
28	1.6914	115298.	-1.5	46	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
29	1.6914	116719.	-1.5	47	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
30	1.6929	285110.	-5.0	54	55	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
31	1.6916	330761.	-4.6	55	56	57	0	0	0	0	0	0	0	0	0	0	0	0	0	0
32	1.6915	156769.	-2.4	56	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
33	1.6909	161451.	-3.7	57	58	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
34	1.6903	352166.	-3.4	58	59	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
35	1.6902	360108.	-2.4	59	60	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
36	1.6905	537696.	-1.2	60	61	64	108	0	0	0	0	0	0	0	0	0	0	0	0	0
37	1.6934	663111.	-3.2	53	62	63	92	0	0	0	0	0	0	0	0	0	0	0	0	0
38	1.6890	1067031.	-2.6	62	79	108	0	0	0	0	0	0	0	0	0	0	0	0	0	0
39	1.6893	1336282.	-2.3	63	93	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
40	1.6867	336079.	-2.4	64	65	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
41	1.6814	188683.	-2.4	65	66	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
42	1.6782	182003.	-2.2	66	67	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
43	1.6761	186832.	-3.7	67	68	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
44	1.6814	249220.	-0.5	68	69	84	90	0	0	0	0	0	0	0	0	0	0	0	0	0
45	1.7724	182976.	-1.6	69	70	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
46	1.8112	156024.	-1.6	70	71	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
47	1.8498	111288.	-1.6	71	72	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
48	1.8646	91403.	-1.2	72	73	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
49	1.8845	121938.	-0.5	73	74	107	0	0	0	0	0	0	0	0	0	0	0	0	0	0
50	2.0382	174131.	-1.2	74	75	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
51	2.1236	167059.	-1.2	75	76	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
52	2.1573	113393.	-1.4	76	77	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
53	2.2189	97025.	-1.0	77	78	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
54	2.2615	176054.	-2.0	78	109	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
55	1.6849	449458.	-1.8	79	80	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
56	1.6802	203386.	-2.0	80	81	85	0	0	0	0	0	0	0	0	0	0	0	0	0	0
57	1.6753	73188.	-0.5	81	82	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0

58	1.6553	61591.	-0.5	82	83	0	0	0	0	0	0	0
59	1.6490	58953.	-0.5	83	84	0	0	0	0	0	0	0
60	1.6755	209072.	-2.1	85	86	0	0	0	0	0	0	0
61	1.6664	213638.	-1.9	86	87	0	0	0	0	0	0	0
62	1.6589	181067.	-1.7	87	88	99	100	0	0	0	0	0
63	1.6607	114302.	-1.7	88	89	102	0	0	0	0	0	0
64	1.6675	82867.	-0.4	89	90	0	0	0	0	0	0	0
65	1.6849	476240.	-1.2	93	94	0	0	0	0	0	0	0
66	1.6831	280913.	-0.7	94	95	0	0	0	0	0	0	0
67	1.6707	110083.	-0.7	95	96	0	0	0	0	0	0	0
68	1.6613	148623.	-1.1	96	97	0	0	0	0	0	0	0
69	1.6463	128584.	-0.9	97	98	0	0	0	0	0	0	0
70	1.6506	77249.	-0.5	98	99	0	0	0	0	0	0	0
71	1.6593	48980.	-0.5	100	101	0	0	0	0	0	0	0
72	1.6622	57857.	-2.0	101	102	103	0	0	0	0	0	0
73	1.6637	115385.	-2.1	103	104	0	0	0	0	0	0	0
74	1.6781	173080.	-3.1	104	105	0	0	0	0	0	0	0
75	1.7043	150098.	-1.1	105	106	0	0	0	0	0	0	0
76	1.8204	116614.	-1.0	106	107	0	0	0	0	0	0	0
77	2.3168	193875.	-0.7	109	110	0	0	0	0	0	0	0
78	2.3827	203587.	0.5	110	111	0	0	0	0	0	0	0
79	2.8786	204176.	1.5	111	112	0	0	0	0	0	0	0
80	3.3530	192477.	1.7	112	113	0	0	0	0	0	0	0
81	3.9414	180116.	2.4	113	114	0	0	0	0	0	0	0
82	5.1500	187842.	4.3	114	115	0	0	0	0	0	0	0
83	6.2099	92928.	4.5	115	0	0	0	0	0	0	0	0

*****E: CHANNEL DATA*****

Channel Number	Length (m)	Width (m)	Depth (m)	Direction Degrees	Mannings N	Initial Velocity	Connecting Junction
1	2306.	1931.	6.69	64.	0.010	0.03860	1 7
2	2113.	1.	6.70	114.	0.010	0.05980	2 7
3	2177.	1931.	6.70	56.	0.010	0.05280	2 8
4	1985.	1.	6.70	134.	0.010	0.03420	3 8
5	1921.	1931.	6.70	59.	0.010	0.04430	3 9
6	1921.	1.	6.70	140.	0.010	-0.02310	4 9
7	1793.	1931.	6.70	88.	0.010	0.01080	4 10
8	2306.	1.	6.70	29.	0.010	0.02940	4 11
9	1729.	1931.	6.71	115.	0.010	0.02330	5 11
10	2049.	1.	6.71	47.	0.010	0.07410	5 12
11	2306.	1931.	6.71	122.	0.010	0.03530	6 12
12	2177.	1448.	6.69	88.	0.010	0.03050	7 13
13	2817.	1.	6.69	45.	0.010	0.03210	7 14
14	2049.	1.	6.69	3.	0.010	0.01110	7 8
15	1921.	1448.	6.69	97.	0.010	0.04610	8 14
16	2306.	1.	6.70	4.	0.010	-0.02410	8 9
17	2497.	1.	6.70	42.	0.010	-0.01340	8 15
18	1601.	1448.	6.70	111.	0.010	0.04030	9 15
19	1601.	1.	6.70	20.	0.010	0.04450	9 10
20	2177.	1.	6.70	155.	0.010	0.01220	10 15
21	1985.	1.	6.70	338.	0.010	0.05260	10 11
22	2049.	1.	6.71	358.	0.010	0.05370	11 12
23	1921.	1.	6.69	354.	0.010	-0.00316	13 14
24	2177.	1.	6.70	354.	0.010	-0.06020	14 15
25	2306.	1042.	3.69	40.	0.010	0.02700	13 19
26	1665.	1.	3.69	96.	0.010	0.07110	14 19
27	1793.	1042.	6.69	70.	0.010	0.04500	14 20
28	2306.	100.	2.69	128.	0.010	0.10900	15 20
29	2049.	1042.	2.80	96.	0.021	0.03740	15 21
30	1409.	1.	2.60	30.	0.024	0.14200	15 16
31	1857.	1448.	2.60	119.	0.028	0.00063	10 16
32	1665.	1.	2.60	59.	0.032	0.15300	10 17
33	2369.	1448.	2.60	116.	0.032	0.01140	11 17
34	2049.	1.	3.60	44.	0.037	0.13400	11 18
35	1601.	1448.	5.60	113.	0.037	0.02880	12 18
36	2177.	1042.	4.90	122.	0.035	0.02460	18 24
37	2562.	1.	2.60	344.	0.037	0.08770	17 18
38	1793.	1.	2.60	39.	0.034	0.15600	17 24
39	2049.	1042.	2.60	82.	0.037	-0.00512	17 23
40	2049.	1.	2.60	125.	0.037	-0.08040	17 22
41	1665.	1.	2.60	353.	0.032	0.12600	16 17
42	1729.	1042.	2.60	75.	0.032	-0.01460	16 22
43	2049.	1.	2.60	133.	0.028	-0.08840	16 21
44	1089.	402.	3.19	80.	0.010	0.00343	19 26
45	1217.	163.	3.19	53.	0.010	0.03330	19 27
46	768.	392.	3.19	43.	0.017	0.00276	27 28
47	768.	312.	3.19	8.	0.017	0.00353	27 29
48	896.	141.	3.69	32.	0.010	0.00851	19 25

49	768.	758.	3.69	360.	0.010	0.01250	19	20
50	1216.	758.	6.70	14.	0.021	0.06190	20	21
51	1857.	758.	3.80	4.	0.029	0.09840	21	22
52	1537.	758.	2.60	9.	0.037	0.08750	22	23
53	1089.	440.	2.59	8.	0.037	0.12600	23	37
54	1025.	497.	6.70	44.	0.022	0.00256	20	30
55	931.	343.	6.29	7.	0.022	-0.00442	30	31
56	643.	398.	4.19	97.	0.026	0.00474	31	32
57	653.	348.	6.34	360.	0.024	-0.02160	31	33
58	686.	403.	5.24	9.	0.026	-0.02940	33	34
59	730.	475.	4.09	25.	0.026	-0.01860	34	35
60	781.	322.	3.49	46.	0.024	-0.10100	35	36
61	1114.	326.	2.59	90.	0.037	0.06040	23	36
62	951.	719.	2.49	76.	0.042	0.04090	37	38
63	1062.	690.	2.49	11.	0.027	0.07150	37	39
64	905.	322.	3.49	90.	0.027	-0.05680	36	40
65	1251.	142.	4.08	112.	0.035	-0.15700	40	41
66	1145.	169.	3.98	161.	0.035	-0.14900	41	42
67	1383.	133.	4.63	186.	0.035	-0.15900	42	43
68	1526.	121.	3.63	147.	0.038	-0.23800	43	44
69	1571.	102.	2.63	240.	0.037	-0.39600	44	45
70	1781.	103.	3.39	176.	0.036	-0.31800	45	46
71	1425.	93.	3.43	194.	0.034	-0.37000	46	47
72	930.	101.	3.26	93.	0.031	-0.31600	47	48
73	894.	104.	2.52	140.	0.031	-0.32500	48	49
74	1622.	105.	2.61	169.	0.031	-0.56700	49	50
75	1995.	96.	3.28	107.	0.031	-0.46700	50	51
76	1308.	109.	3.44	101.	0.031	-0.37000	51	52
77	1144.	74.	3.39	136.	0.032	-0.52900	52	53
78	1253.	87.	3.74	98.	0.034	-0.39200	53	54
79	1201.	620.	2.79	82.	0.048	0.01630	38	55
80	708.	217.	3.58	75.	0.055	0.00254	55	56
81	1333.	68.	2.93	142.	0.052	0.02200	56	57
82	2005.	27.	2.17	114.	0.052	0.00355	57	58
83	2220.	31.	2.15	191.	0.052	-0.04270	58	59
84	2416.	21.	2.02	18.	0.052	0.11700	59	44
85	828.	195.	3.73	79.	0.055	-0.01970	56	60
86	1629.	158.	3.67	117.	0.055	-0.04520	60	61
87	1588.	107.	3.46	159.	0.046	-0.10000	61	62
88	883.	78.	2.86	195.	0.036	-0.19200	62	63
89	1897.	49.	2.21	175.	0.036	-0.05930	63	64
90	1637.	49.	1.97	55.	0.036	0.08520	64	44
91	1284.	1.	3.59	319.	0.036	0.16100	24	23
92	978.	763.	3.10	91.	0.033	0.02630	24	37
93	1200.	392.	2.49	57.	0.034	0.04960	39	65
94	1039.	464.	2.63	106.	0.034	0.01820	65	66
95	1235.	65.	2.38	58.	0.034	0.03880	66	67
96	1599.	87.	2.57	139.	0.036	-0.00240	67	68
97	1868.	55.	2.65	113.	0.036	-0.06510	68	69
98	1619.	55.	2.35	149.	0.036	-0.12300	69	70
99	1201.	46.	2.75	34.	0.036	0.15000	70	62
100	1425.	46.	2.76	129.	0.036	-0.15600	62	71
101	726.	46.	2.91	248.	0.036	-0.15700	71	72
102	994.	61.	3.01	94.	0.036	-0.15700	63	72
103	819.	70.	3.71	126.	0.036	-0.18800	72	73
104	2710.	64.	4.27	192.	0.036	-0.23200	73	74
105	2683.	64.	3.79	157.	0.036	-0.25700	74	75
106	2180.	59.	2.56	240.	0.036	-0.39200	75	76
107	1773.	59.	2.55	45.	0.036	0.28100	76	49
108	1301.	1.	2.49	353.	0.052	0.06100	36	38
109	2012.	86.	3.14	158.	0.041	-0.33100	54	77
110	1768.	121.	1.95	167.	0.045	-0.29000	77	78
111	1609.	121.	1.63	167.	0.052	-0.47000	78	79
112	1609.	133.	1.52	167.	0.052	-0.43300	79	80
113	1609.	106.	1.60	240.	0.052	-0.50300	80	81
114	1609.	118.	1.20	140.	0.052	-0.59900	81	82
115	1609.	115.	1.28	140.	0.052	-0.57600	82	83

*****F: CONSTANT INFLOW DATA*****

20

Junction Number	Inflow (m ³)	
83	-29.8	[Upstream Boundary Inflow]
82	-0.03	
81	-0.03	
80	-0.03	
79	-0.03	
78	-0.76	[French Creek Inflow]

77 -1.65 [Pilchuk River Inflow]
 73 -0.05
 65 -0.13
 60 -0.19
 54 -0.07
 53 -0.01
 52 -0.01
 51 -0.01
 50 -0.01
 49 -0.01
 48 -0.46
 44 -0.16
 43 -0.35
 42 -0.46

variable inflow data

0

*****G: SEAWARD BOUNDARY DATA*****

3		1 100		20		0		0		0		1.0	
Junction	Time	Head	Junction	Time	Head	Junction	Time	Head	Junction	Time	Head	Junction	Time
	(hr)	(m)		(hr)	(m)		(hr)	(m)		(hr)	(m)		(hr)
1	0000	1.6943	1	538	2.8651	1	1219	-0.1829	1	1930	3.4138		
2	103	1.4326	2	624	2.8042	2	1251	0.0000	2	1953	3.3833		
3	138	1.2497	3	710	2.7432	3	1326	0.2743	3	2016	3.3223		
4	214	1.0668	4	758	2.6518	4	1402	0.5486	4	2041	3.2614		
5	256	0.9144	5	852	2.5603	5	1440	0.8839	5	2108	3.1699		
6	337	0.7620	6	951	2.4689	6	1519	1.2192	6	2138	3.1090		
7	426	0.6096	7	1101	2.4079	7	1605	1.5850	7	2213	3.0175		
8	519	0.4877	8	1227	2.4079	8	1703	1.8898	8	2254	2.9261		
9	617	0.3658	9	1403	2.4994	9	1813	2.1336	9	2339	2.8651		
10	715	0.1829	10	1519	2.7127	10	1932	2.2555	11	32	2.8346		
11	810	0.0000	11	1607	2.8956	11	2047	2.2250	12	130	2.8346		
12	901	-0.1829	12	1649	3.0785	12	2142	2.1031	13	233	2.8956		
13	949	-0.3353	13	1718	3.2614	13	2231	1.8898	14	328	2.9870		
14	1033	-0.3962	14	1746	3.3833	14	2313	1.6154	15	428	3.0785		
15	1115	-0.3962	15	1815	3.4747	15	2355	1.2497	16	524	3.1394		
16	1157	-0.2743	16	1844	3.5662	17	36	0.8839	17	622	3.1699		
17	1239	0.0305	17	1914	3.6271	18	123	0.5486	18	722	3.1394		
18	1325	0.3048	18	1950	3.6271	19	211	0.2438	19	825	3.0481		
19	1410	0.7315	19	2025	3.5662	20	300	0.0611	20	931	2.9261		
20	1500	1.1582	20	2107	3.4747	21	356	-0.0611	21	1050	2.8042		
21	1554	1.5545	21	2152	3.3223	22	500	-0.0914	22	1217	2.7737		
22	1701	1.8898	22	2244	3.1394	23	607	-0.0914	23	1349	2.8346		
23	1827	2.0726	23	2349	2.9871	24	716	-0.0914	24	1502	2.9871		
24	1958	2.0726	25	102	2.8346	25	821	-0.0914	25	1601	3.1091		
25	2114	1.9507	26	214	2.8042	26	917	-0.0915	26	1644	3.2309		
3		2 100		20		0		0		1.0			
1	0000	1.6976	1	537	2.8651	1	1218	-0.1829	1	1929	3.4138		
2	102	1.4326	2	623	2.8042	2	1250	0.0000	2	1952	3.3833		
3	137	1.2497	3	709	2.7432	3	1325	0.2743	3	2015	3.3223		
4	213	1.0668	4	757	2.6518	4	1401	0.5486	4	2040	3.2614		
5	255	0.9144	5	851	2.5603	5	1439	0.8839	5	2107	3.1699		
6	336	0.7620	6	950	2.4689	6	1518	1.2192	6	2137	3.1090		
7	425	0.6096	7	1100	2.4079	7	1604	1.5850	7	2212	3.0175		
8	518	0.4877	8	1226	2.4079	8	1702	1.8898	8	2253	2.9261		
9	616	0.3658	9	1402	2.4994	9	1812	2.1336	9	2338	2.8651		
10	714	0.1829	10	1518	2.7127	10	1931	2.2555	11	31	2.8346		
11	809	0.0000	11	1606	2.8956	11	2046	2.2250	12	129	2.8346		
12	900	-0.1829	12	1648	3.0785	12	2141	2.1031	13	232	2.8956		
13	948	-0.3353	13	1717	3.2614	13	2230	1.8898	14	327	2.9870		
14	1032	-0.3962	14	1745	3.3833	14	2312	1.6154	15	427	3.0785		
15	1114	-0.3962	15	1814	3.4747	15	2354	1.2497	16	523	3.1394		
16	1156	-0.2743	16	1843	3.5662	17	35	0.8839	17	621	3.1699		
17	1238	0.0305	17	1913	3.6271	18	122	0.5486	18	721	3.1394		
18	1324	0.3048	18	1949	3.6271	19	210	0.2438	19	824	3.0481		
19	1409	0.7315	19	2024	3.5662	20	259	0.0611	20	930	2.9261		
20	1459	1.1582	20	2106	3.4747	21	355	-0.0611	21	1049	2.8042		
21	1553	1.5545	21	2151	3.3223	22	459	-0.0914	22	1216	2.7737		
22	1700	1.8898	22	2243	3.1394	23	606	-0.0914	23	1348	2.8346		
23	1826	2.0726	23	2348	2.9871	24	715	-0.0914	24	1501	2.9871		
24	1957	2.0726	25	101	2.8346	25	820	-0.0914	25	1600	3.1091		
25	2113	1.9507	26	213	2.8042	26	916	-0.0915	26	1643	3.2309		
3		3 100		20		0		0		1.0			
1	0000	1.7010	1	536	2.8651	1	1217	-0.1829	1	1928	3.4138		
2	101	1.4326	2	622	2.8042	2	1249	0.0000	2	1951	3.3833		
3	136	1.2497	3	708	2.7432	3	1324	0.2743	3	2014	3.3223		
4	212	1.0668	4	756	2.6518	4	1400	0.5486	4	2039	3.2614		

5	254	0.9144	5	850	2.5603	5	1438	0.8839	5	2106	3.1699
6	335	0.7620	6	949	2.4689	6	1517	1.2192	6	2136	3.1090
7	424	0.6096	7	1059	2.4079	7	1603	1.5850	7	2211	3.0175
8	517	0.4877	8	1225	2.4079	8	1701	1.8898	8	2252	2.9261
9	615	0.3658	9	1401	2.4994	9	1811	2.1336	9	2337	2.8651
10	713	0.1829	10	1517	2.7127	10	1930	2.2555	11	30	2.8346
11	808	0.0000	11	1605	2.8956	11	2045	2.2250	12	128	2.8346
12	859	-0.1829	12	1647	3.0785	12	2140	2.1031	13	231	2.8956
13	947	-0.3353	13	1716	3.2614	13	2229	1.8898	14	326	2.9870
14	1031	-0.3962	14	1744	3.3833	14	2311	1.6154	15	426	3.0785
15	1113	-0.3962	15	1813	3.4747	15	2353	1.2497	16	522	3.1394
16	1155	-0.2743	16	1842	3.5662	17	34	0.8839	17	620	3.1699
17	1237	0.0305	17	1912	3.6271	18	121	0.5486	18	720	3.1394
18	1323	0.3048	18	1948	3.6271	19	209	0.2438	19	823	3.0481
19	1408	0.7315	19	2023	3.5662	20	258	0.0611	20	929	2.9261
20	1458	1.1582	20	2105	3.4747	21	354	-0.0611	21	1048	2.8042
21	1552	1.5545	21	2150	3.3223	22	458	-0.0914	22	1215	2.7737
22	1659	1.8898	22	2242	3.1394	23	605	-0.0914	23	1347	2.8346
23	1825	2.0726	23	2347	2.9871	24	714	-0.0914	24	1500	2.9871
24	1956	2.0726	25	100	2.8346	25	819	-0.0914	25	1559	3.1091
25	2112	1.9507	26	212	2.8042	26	915	-0.0915	26	1642	3.2309
3	4 100	20	0	0	1.0						
1	0000	1.7045	1	535	2.8651	1	1216	-0.1829	1	1927	3.4138
2	100	1.4326	2	621	2.8042	2	1248	0.0000	2	1950	3.3833
3	135	1.2497	3	707	2.7432	3	1323	0.2743	3	2013	3.3223
4	211	1.0668	4	755	2.6518	4	1359	0.5486	4	2038	3.2614
5	253	0.9144	5	849	2.5603	5	1437	0.8839	5	2105	3.1699
6	334	0.7620	6	948	2.4689	6	1516	1.2192	6	2135	3.1090
7	423	0.6096	7	1058	2.4079	7	1602	1.5850	7	2210	3.0175
8	516	0.4877	8	1224	2.4079	8	1700	1.8898	8	2251	2.9261
9	614	0.3658	9	1400	2.4994	9	1810	2.1336	9	2336	2.8651
10	712	0.1829	10	1516	2.7127	10	1929	2.2555	11	29	2.8346
11	807	0.0000	11	1604	2.8956	11	2044	2.2250	12	127	2.8346
12	858	-0.1829	12	1646	3.0785	12	2139	2.1031	13	230	2.8956
13	946	-0.3353	13	1715	3.2614	13	2228	1.8898	14	325	2.9870
14	1030	-0.3962	14	1743	3.3833	14	2310	1.6154	15	425	3.0785
15	1112	-0.3962	15	1812	3.4747	15	2352	1.2497	16	521	3.1394
16	1154	-0.2743	16	1841	3.5662	17	33	0.8839	17	619	3.1699
17	1236	0.0305	17	1911	3.6271	18	120	0.5486	18	719	3.1394
18	1322	0.3048	18	1947	3.6271	19	208	0.2438	19	822	3.0481
19	1407	0.7315	19	2022	3.5662	20	257	0.0611	20	928	2.9261
20	1457	1.1582	20	2104	3.4747	21	353	-0.0611	21	1047	2.8042
21	1551	1.5545	21	2149	3.3223	22	457	-0.0914	22	1214	2.7737
22	1658	1.8898	22	2241	3.1394	23	604	-0.0914	23	1346	2.8346
23	1824	2.0726	23	2346	2.9871	24	713	-0.0914	24	1459	2.9871
24	1955	2.0726	25	59	2.8346	25	818	-0.0914	25	1558	3.1091
25	2111	1.9507	26	211	2.8042	26	914	-0.0915	26	1641	3.2309
3	5 100	20	0	0	1.0						
1	0000	1.7080	1	534	2.8651	1	1215	-0.1829	1	1926	3.4138
2	59	1.4326	2	620	2.8042	2	1247	0.0000	2	1949	3.3833
3	134	1.2497	3	706	2.7432	3	1322	0.2743	3	2012	3.3223
4	210	1.0668	4	754	2.6518	4	1358	0.5486	4	2037	3.2614
5	252	0.9144	5	848	2.5603	5	1436	0.8839	5	2104	3.1699
6	333	0.7620	6	947	2.4689	6	1515	1.2192	6	2134	3.1090
7	422	0.6096	7	1057	2.4079	7	1601	1.5850	7	2209	3.0175
8	515	0.4877	8	1223	2.4079	8	1659	1.8898	8	2250	2.9261
9	613	0.3658	9	1359	2.4994	9	1809	2.1336	9	2335	2.8651
10	711	0.1829	10	1515	2.7127	10	1928	2.2555	11	28	2.8346
11	806	0.0000	11	1603	2.8956	11	2043	2.2250	12	126	2.8346
12	857	-0.1829	12	1645	3.0785	12	2138	2.1031	13	229	2.8956
13	945	-0.3353	13	1714	3.2614	13	2227	1.8898	14	324	2.9870
14	1029	-0.3962	14	1742	3.3833	14	2309	1.6154	15	424	3.0785
15	1111	-0.3962	15	1811	3.4747	15	2351	1.2497	16	520	3.1394
16	1153	-0.2743	16	1840	3.5662	17	32	0.8839	17	618	3.1699
17	1235	0.0305	17	1910	3.6271	18	119	0.5486	18	718	3.1394
18	1321	0.3048	18	1946	3.6271	19	207	0.2438	19	821	3.0481
19	1406	0.7315	19	2021	3.5662	20	256	0.0611	20	927	2.9261
20	1456	1.1582	20	2103	3.4747	21	352	-0.0611	21	1046	2.8042
21	1550	1.5545	21	2148	3.3223	22	456	-0.0914	22	1213	2.7737
22	1657	1.8898	22	2240	3.1394	23	603	-0.0914	23	1345	2.8346
23	1823	2.0726	23	2345	2.9871	24	712	-0.0914	24	1458	2.9871
24	1954	2.0726	25	58	2.8346	25	817	-0.0914	25	1557	3.1091
25	2110	1.9507	26	210	2.8042	26	913	-0.0915	26	1640	3.2309
3	6 100	20	0	0	1.0						
1	0000	1.7116	1	533	2.8651	1	1214	-0.1829	1	1925	3.4138
2	58	1.4326	2	619	2.8042	2	1246	0.0000	2	1948	3.3833
3	133	1.2497	3	705	2.7432	3	1321	0.2743	3	2011	3.3223

4	209	1.0668	4	753	2.6518	4	1357	0.5486	4	2036	3.2614
5	251	0.9144	5	847	2.5603	5	1435	0.8839	5	2103	3.1699
6	332	0.7620	6	946	2.4689	6	1514	1.2192	6	2133	3.1090
7	421	0.6096	7	1056	2.4079	7	1600	1.5850	7	2208	3.0175
8	514	0.4877	8	1222	2.4079	8	1658	1.8898	8	2249	2.9261
9	612	0.3658	9	1358	2.4994	9	1808	2.1336	9	2334	2.8651
10	710	0.1829	10	1514	2.7127	10	1927	2.2555	11	27	2.8346
11	805	0.0000	11	1602	2.8956	11	2042	2.2250	12	125	2.8346
12	856	-0.1829	12	1644	3.0785	12	2137	2.1031	13	228	2.8956
13	944	-0.3353	13	1713	3.2614	13	2226	1.8898	14	323	2.9870
14	1028	-0.3962	14	1741	3.3833	14	2308	1.6154	15	423	3.0785
15	1110	-0.3962	15	1810	3.4747	15	2350	1.2497	16	519	3.1394
16	1152	-0.2743	16	1839	3.5662	17	31	0.8839	17	617	3.1699
17	1234	0.0305	17	1909	3.6271	18	118	0.5486	18	717	3.1394
18	1320	0.3048	18	1945	3.6271	19	206	0.2438	19	820	3.0481
19	1405	0.7315	19	2020	3.5662	20	255	0.0611	20	926	2.9261
20	1455	1.1582	20	2102	3.4747	21	351	-0.0611	21	1045	2.8042
21	1549	1.5545	21	2147	3.3223	22	455	-0.0914	22	1212	2.7737
22	1656	1.8898	22	2239	3.1394	23	602	-0.0914	23	1344	2.8346
23	1822	2.0726	23	2344	2.9871	24	711	-0.0914	24	1457	2.9871
24	1953	2.0726	25	57	2.8346	25	816	-0.0914	25	1556	3.1091
25	2109	1.9507	26	209	2.8042	26	912	-0.0915	26	1639	3.2309

*****H: WIND DATA*****

0

*****I: EVAPORATION OR GEOMETRY DATA*****

0

*****J: JUNCTION GEOMETRY DATA*****

0

*****K: CHANNEL GEOMETRY DATA*****

0

*****L: MAP TO WASP5*****

Junction EUTRO5 Segment

Number	Number
0	83
1	0
2	0
3	0
4	0
5	0
6	0
7	1
8	2
9	3
10	4
11	5
12	6
13	7
14	8
15	9
16	10
17	11
18	12
19	13
20	14
21	15
22	16
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24	18
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69	63
70	64
71	65
72	66
73	67
74	68
75	69
76	70
77	71
78	72
79	73
80	74
81	75
82	76
83	0

EUTRO5 Input File

SNOFIN1.INP: TIME=0 CRITICAL CONDITIONS WATER QUALITY INPUT FILE (TETRA TECH VERSION)
 *****ROBERT CUSIMANO, WA. DEPT. OF ECOLOGY, March 4, 1997*****

NOSE NOSY ICFL MFLA JMAS NEGS INTY ADFA ZDAY HHMM TELG ***A: MODEL OPTIONS***
 80 9 0 0 1 0 0 0.00 1.0 0100 0

14 24 29 30 34 44
 1
 0.021 24.0
 2
 7.0 7.0 0.021 24.0
 0 0 0 0 0 0 0 0
 1 0
 6 1.000 1.000

*****B: EXCHANGES***

42

Exchange Area	Length (m)	Segment Pair	
13903	2306	0 1	<<<Possession Sound>>>
13903	2177	0 2	
13903	1921	0 3	
13903	1793	0 4	
13903	1729	0 5	
13903	2306	0 6	
6	1970	1 2	
10425	1947	1 7	
6	2509	1 8	
6	2096	2 3	
10425	1682	2 8	
6	2194	2 9	
6	1450	3 4	
10425	1432	3 9	
6	1765	4 5	
6	2007	4 9	
10425	1647	4 10	
6	1437	4 11	
6	1869	5 6	
10425	2123	5 11	
6	1810	5 12	
10425	1434	6 12	
6	1715	7 8	
9502	1986	7 13	
6	1828	8 9	
6	1477	8 13	
7502	1548	8 14	
6	1230	9 10	
6	2092	9 14	
7502	1940	9 15	
7502	1971	12 18	
5458	708	13 14	
1015	803	13 19	
1524	1014	13 20	
1181	1056	13 21	
5458	1115	14 15	
5458	1621	15 16	
2830	642	21 22	
2246	651	21 23	
3578	937	14 24	
2470	931	24 25	
2265	643	25 26	

2
 Dispersion Value (m²/sec) Dispersion Value (m²/sec)
 120.00 0.0 120.00 365.

12

6	1507	10	11
6	1812	10	15
7502	1474	10	16
6	2320	11	12
6	1969	11	16
7502	1847	11	17
6	1572	11	18
5458	1391	16	17
1584	961	17	31
6	1284	17	18

	2457	879	31	32	
	2370	1062	31	33	
2					
	40.00	0.0	40.00		365.
10					
	2040	1201	32	49	<<<Lower Steamboat Slough>>>
	779	750	49	50	
	198	1333	50	51	
	743	828	50	54	
	659	1629	54	55	
	1176	1200	33	59	
	1444	1114	17	30	<<<Lower Snohomish River>>>
	1069	765	29	30	
	6	1195	30	32	
	1444	925	30	34	
2					
	5.00	0.0	5.00		365.
4					
	2032	637	25	27	<<<Lower Snohomish River>>>
	1910	670	27	28	
	2061	750	28	29	
	4043	978	18	31	<<<Mouth to Sloughs>>>
2					
	80.00	0.0	80.00		365.
21					
	68	2005	51	52	<<<Union and Lower Ebey Slough>>>
	77	2220	52	53	
	423	1588	55	56	
	262	922	56	57	
	124	1201	64	56	
	214	1425	56	65	
	132	1797	57	58	
	257	831	57	66	
	121	1597	58	38	
	56	2416	53	38	
	231	1773	43	70	
	1067	1039	59	60	<<<Upper Ebey Slough>>>
	142	1235	60	61	
	192	1599	61	62	
	121	2868	62	63	
	121	1819	63	64	
	133	750	65	66	
	335	819	66	67	
	243	2710	67	68	
	275	2683	68	69	
	211	2180	69	70	
2					
	1.00	0.0	1.00		365.
21					
	460	1271	34	35	
	533	1165	35	36	
	626	1403	36	37	
	500	1546	37	38	<<<Middle Snohomish River>>>
	372	1551	38	39	
	400	1761	39	40	
	365	1405	40	41	
	380	910	41	42	
	314	874	42	43	
	326	1622	43	44	<<<Upper Snohomish River>>>
	362	1995	44	45	
	429	1308	45	46	
	288	1144	46	47	
	372	1253	47	48	
	398	2012	48	71	
	350	1768	71	72	
	181	1609	72	73	
	200	1609	73	74	
	159	1609	74	75	
	130	1609	75	76	
	139	1609	0	76	
2					
	1.00	0.0	1.00		365.
0	0	0	0	0	0

2 0 1.0
 1.0000 1.0000

*****C: VOLUMES***

Segment Number	Bottom Segment	Segment Volume (m ³)			Average Depth	
1	77	2.58E+07	0	0	7.2	0
2	77	2.37E+07	0	0	7.2	0
3	77	2.03E+07	0	0	7.2	0
4	77	2.09E+07	0	0	7.2	0
5	77	2.32E+07	0	0	7.2	0
6	77	2.28E+07	0	0	7.2	0
7	77	2.00E+07	0	0	7.2	0
8	77	1.67E+07	0	0	7.2	0
9	77	1.61E+07	0	0	7.2	0
10	77	1.62E+07	0	0	7.2	0
11	77	2.01E+07	0	0	7.2	0
12	77	1.66E+07	0	0	7.2	0
13	77	1.32E+07	0	0	7.2	0
14	77	1.38E+07	0	0	7.2	0
15	77	1.61E+07	0	0	7.2	0
16	77	1.58E+07	0	0	7.2	0
17	77	7.07E+06	0	0	3.4	0
18	77	1.09E+07	0	0	7.2	0
19	77	4.09E+05	0	0	7.2	0
20	77	1.42E+06	0	0	7.2	0
21	77	2.40E+06	0	0	7.2	0
22	77	9.81E+05	0	0	7.2	0
23	77	7.77E+05	0	0	7.2	0
24	77	2.54E+06	0	0	6.8	0
25	77	2.44E+06	0	0	6.8	0
26	77	5.31E+05	0	0	4.6	0
27	77	1.08E+06	0	0	4.9	0
28	77	1.96E+06	0	0	4.6	0
29	77	1.83E+06	0	0	4.6	0
30	78	1.76E+06	0	0	3.4	0
31	79	4.38E+06	0	0	3.4	0
32	79	3.02E+06	0	0	4.0	0
33	79	1.47E+06	0	0	2.4	0
34	78	7.37E+05	0	0	3.2	0
35	78	7.81E+05	0	0	4.6	0
36	78	4.91E+05	0	0	2.7	0
37	78	1.10E+06	0	0	5.9	0
38	78	5.48E+05	0	0	2.2	0
39	78	6.95E+05	0	0	3.8	0
40	78	5.93E+05	0	0	3.8	0
41	78	4.23E+05	0	0	3.8	0
42	78	3.11E+05	0	0	3.4	0
43	80	2.80E+05	0	0	2.3	0
44	80	6.17E+05	0	0	3.4	0
45	80	5.69E+05	0	0	3.4	0
46	80	4.09E+05	0	0	3.6	0
47	80	3.11E+05	0	0	3.2	0
48	80	5.95E+05	0	0	4.2	0
49	79	1.80E+06	0	0	4.0	0
50	79	8.53E+05	0	0	2.2	0
51	79	1.96E+05	0	0	2.7	0
52	79	1.65E+05	0	0	2.7	0
53	79	1.59E+05	0	0	2.7	0
54	79	8.99E+05	0	0	4.3	0
55	79	8.75E+05	0	0	4.1	0
56	79	7.07E+05	0	0	3.9	0
57	79	3.31E+05	0	0	2.9	0
58	79	2.14E+05	0	0	2.6	0
59	79	1.62E+06	0	0	3.4	0
60	79	8.15E+05	0	0	2.9	0
61	79	3.19E+05	0	0	2.9	0
62	79	4.91E+05	0	0	3.3	0
63	79	3.99E+05	0	0	3.1	0
64	79	2.09E+05	0	0	2.7	0
65	79	1.33E+05	0	0	2.7	0
66	79	2.43E+05	0	0	4.2	0
67	79	5.01E+05	0	0	4.3	0
68	79	9.09E+05	0	0	5.3	0
69	79	4.93E+05	0	0	3.3	0
70	79	3.85E+05	0	0	2.7	0
71	80	7.55E+05	0	0	3.9	0
72	80	5.29E+05	0	0	2.6	0
73	80	3.47E+05	0	0	1.7	0

74	80	1	3.47E+05	0	0	1.8	0
75	80	1	3.07E+05	0	0	1.7	0
76	80	1	3.19E+05	0	0	1.7	0
77	0	3	1.00E+08	0	0	0.1	0
78	0	3	1.00E+08	0	0	0.1	0
79	0	3	1.00E+08	0	0	0.1	0
80	0	3	1.00E+08	0	0	0.1	0

3 1FINTIDE1.HYD *****D: FLOWS***
0 0 0 0 0 0 0 0 0
7 + * + * + *****E: BOUNDARIES***
1.0 1.0 (System 1--Ammonia)

Ammonia (mg/L)	Starting Day	Ammonia (mg/L)	Ending Day	[same for other "system variables]
.095	0.	.095	365.0	
.095	0.	.095	365.0	
.095	0.	.095	365.0	
.095	0.	.095	365.0	
.095	0.	.095	365.0	
.095	0.	.095	365.0	
.095	0.	.095	365.0	
.050	0.	.050	365.0	

76 7 + * + * + *****E: BOUNDARIES***
1.0 1.0 (System 2--Nitrate)

.300	0.	.300	365.0	
.300	0.	.300	365.0	
.300	0.	.300	365.0	
.300	0.	.300	365.0	
.300	0.	.300	365.0	
.300	0.	.300	365.0	
.300	0.	.300	365.0	
.160	0.	.160	365.0	

76 7 + * + * + *****E: BOUNDARIES***
1.0 1.0 (System 3--Inorg Phos)

.060	0.	.060	365.0	
.060	0.	.060	365.0	
.060	0.	.060	365.0	
.060	0.	.060	365.0	
.060	0.	.060	365.0	
.060	0.	.060	365.0	
.060	0.	.060	365.0	
.005	0.	.005	365.0	

76 7 + * + * + *****E: BOUNDARIES***
1.0 1.0 (System 4--Chlorophyl-a)

0.9	0.	0.9	365.0	
0.9	0.	0.9	365.0	
0.9	0.	0.9	365.0	
0.9	0.	0.9	365.0	
0.9	0.	0.9	365.0	
0.9	0.	0.9	365.0	
0.9	0.	0.9	365.0	
1.6	0.	1.6	365.0	

76 7 + * + * + *****E: BOUNDARIES***
1.0 1.0 (System 5--CBOD)

1	2				
	2.5	0.	2.5	365.0	
2	2				
	2.5	0.	2.5	365.0	
3	2				
	2.5	0.	2.5	365.0	
4	2				
	2.5	0.	2.5	365.0	
5	2				
	2.5	0.	2.5	365.0	
6	2				
	2.5	0.	2.5	365.0	
76	2				
	1.2	0.	1.2	365.0	
	7	+	*	+	*****E: BOUNDARIES***
	1.0	1.0			(System 6--DO)
1	2				
	6.0	0.	6.0	365.0	
2	2				
	6.0	0.	6.0	365.0	
3	2				
	6.0	0.	6.0	365.0	
4	2				
	6.0	0.	6.0	365.0	
5	2				
	6.0	0.	6.0	365.0	
6	2				
	6.0	0.	6.0	365.0	
76	2				
	9.1	0.	9.1	365.0	
	7	+	*	+	*****E: BOUNDARIES***
	1.0	1.0			(System 7--Org Nitrogen)
1	2				
	.126	0.	.126	365.0	
2	2				
	.126	0.	.126	365.0	
3	2				
	.126	0.	.126	365.0	
4	2				
	.126	0.	.126	365.0	
5	2				
	.126	0.	.126	365.0	
6	2				
	.126	0.	.126	365.0	
76	2				
	.117	0.	.117	365.0	
	7	+	*	+	*****E: BOUNDARIES***
	1.0	1.0			(System 8--Org Phos)
1	2				
	.040	0.	.040	365.0	
2	2				
	.040	0.	.040	365.0	
3	2				
	.040	0.	.040	365.0	
4	2				
	.040	0.	.040	365.0	
5	2				
	.040	0.	.040	365.0	
6	2				
	.040	0.	.040	365.0	
76	2				
	.020	0.	.020	365.0	
	7	+	*	+	*****E: BOUNDARIES***
	1.0	1.0			(System 9--Salinity)
1	2				
	29.5	0.	29.5	365.0	
2	2				
	29.5	0.	29.5	365.0	
3	2				
	29.5	0.	29.5	365.0	
4	2				
	29.5	0.	29.5	365.0	
5	2				
	29.5	0.	29.5	365.0	
6	2				
	29.5	0.	29.5	365.0	
76	2				

Ammonia (lbs/day)	Starting day	Ammonia (lbs/day)	Ending day	
0.0	0.	0.0	365.0	
13	+ *	+		*****F: WASTE LOADS (System 1 - Ammonia)
1.00	0.4535			[lbs/day to kilograms/day conversion factor]
72				(French Creek)
19.2	0.0	19.2	365.0	
71				(Pilchuck River)
9.02	0.0	9.02	365.0	
48				(Snohomsih WTP)
491.2	0.0	491.2	365.0	
42				(Marshland)
20.2	0.0	20.2	365.0	
38				(Deadwater Slough)
22.2	0.0	22.2	365.0	
60				(Tulalip Landfill--Ebey)
4.07	0.0	4.07	365.0	
50				(Tulalip Landfill--Steamboat)
4.07	0.0	4.07	365.0	
70				(Swan Trail Slough)
33.4	0.0	33.4	365.0	
67				(Lake Stevens WTP)
346.9	0.0	346.9	365.0	
54				(Marysville WTP)
1070.8	0.0	1070.8	365.0	
37				(Everett Mech WTP)
1737.7	0.0	1737.7	365.0	
36				(Everett Lagoon WTP)
2108.6	0.0	2108.6	365.0	
59				(Quilceda Creek)
0.56	0.0	0.56	365.0	
11	+ *	+		*****F: WASTE LOADS (System 2 - Nitrate)
1.00	0.4535			
72				(French Creek)
94.6	0.0	94.6	365.0	
71				(Pilchuck River)
120.3	0.0	120.3	365.0	
48				(Snohomsih WTP)
177.3	0.0	177.3	365.0	
42				(Marshland)
107.4	0.0	107.4	365.0	
38				(Deadwater Slough)
6.2	0.0	6.2	365.0	
70				(Swan Trail Slough)
0.06	0.0	0.06	365.0	
67				(Lake Stevens WTP)
3.4	0.0	3.4	365.0	
54				(Marysville WTP)
200.9	0.0	200.9	365.0	
37				(Everett Mech WTP)
679.1	0.0	679.1	365.0	
36				(Everett Lagoon WTP)
83.5	0.0	83.5	365.0	
59				(Quilceda Creek)
26.2	0.0	26.2	365.0	
11	+ *	+		*****F: WASTE LOADS (System 3 - Inorg Phos)
1.00	0.4535			
72				(French Creek)
2.0	0.0	2.0	365.0	
71				(Pilchuck River)
1.5	0.0	1.5	365.0	
48				(Snohomsih WTP)
106.4	0.0	106.4	365.0	
42				(Marshland)
4.6	0.0	4.6	365.0	
38				(Deadwater Slough)
0.53	0.0	0.53	365.0	
70				(Swan Trail Slough)
0.13	0.0	0.13	365.0	
67				(Lake Stevens WTP)
43.5	0.0	43.5	365.0	
54				(Marysville WTP)
241.2	0.0	241.2	365.0	
37				(Everett Mech WTP)
270.7	0.0	270.7	365.0	
36				(Everett Lagoon WTP)
418.2	0.0	418.2	365.0	

59	2				(Quilceda Creek)
	1.6	0.0	1.6	365.0	
	11	+ *	+		*****F: WASTE LOADS (System 4 - Chlorophyll)
	1.00	0.4535			
72	2				(French Creek)
	2.6	0.0	2.6	365.0	
71	2				(Pilchuck River)
	0.57	0.0	0.57	365.0	
48	2				(Snohomsih WTP)
	0.0	0.0	0.0	365.0	
42	2				(Marshland)
	1.3	0.0	1.3	365.0	
38	2				(Deadwater Slough)
	3.1	0.0	3.1	365.0	
70	2				(Swan Trail Slough)
	0.1	0.0	0.1	365.0	
67	2				(Lake Stevens WTP)
	0.0	0.0	0.0	365.0	
54	2				(Marysville WTP)
	0.0	0.0	0.0	365.0	
37	2				(Everett Mech WTP)
	0.0	0.0	0.0	365.0	
36	2				(Everett Lagoon WTP)
	0.0	0.0	0.0	365.0	
59	2				(Quilceda Creek)
	0.06	0.0	0.06	365.0	
	11	+ *	+		*****F: WASTE LOADS (System 5 - CBOD)
	1.00	0.4535			
72	2				(French Creek)
	523.9	0.0	523.9	365.0	
71	2				(Pilchuck River)
	360.9	0.0	360.9	365.0	
48	2				(Snohomsih WTP)
	2105.3	0.0	2105.3	365.0	
42	2				(Marshland)
	471.5	0.0	471.5	365.0	
38	2				(Deadwater Slough)
	320.7	0.0	320.7	365.0	
70	2				(Swan Trail Slough)
	141.5	0.0	141.5	365.0	
67	2				(Lake Stevens WTP)
	1804.6	0.0	1804.6	365.0	
54	2				(Marysville WTP)
	4589.0	0.0	4589.0	365.0	
37	2				(Everett Mech WTP)
	4391.1	0.0	4391.1	365.0	
36	2				(Everett Lagoon WTP)
	8785.7	0.0	8785.7	365.0	
59	2				(Quilceda Creek)
	65.6	0.0	65.6	365.0	
	11	+ *	+		*****F: WASTE LOADS (System 6 - DO)
	1.00	0.4535			
72	2				(French Creek)
	305.6	0.0	305.6	365.0	
71	2				(Pilchuck River)
	2767.0	0.0	2767.0	365.0	
48	2				(Snohomsih WTP)
	203.5	0.0	203.5	365.0	
42	2				(Marshland)
	384.2	0.0	384.2	365.0	
38	2				(Deadwater Slough)
	291.0	0.0	291.0	365.0	
70	2				(Swan Trail Slough)
	10.5	0.0	10.5	365.0	
67	2				(Lake Stevens WTP)
	122.3	0.0	122.3	365.0	
54	2				(Marysville WTP)
	86.7	0.0	86.7	365.0	
37	2				(Everett Mech WTP)
	320.8	0.0	320.8	365.0	
36	2				(Everett Lagoon WTP)
	474.4	0.0	474.4	365.0	
59	2				(Quilceda Creek)
	174.6	0.0	174.6	365.0	
	11	+ *	+		*****F: WASTE LOADS (System 7 - Org N)
	1.00	0.4535			
72	2				(French Creek)

70.4	0.0	70.4	365.0	
71 2				(Pilchuck River)
40.3	0.0	40.3	365.0	
48 2				(Snohomsih WTP)
299.4	0.0	299.4	365.0	
42 2				(Marshland)
45.4	0.0	45.4	365.0	
38 2				(Deadwater Slough)
38.0	0.0	38.0	365.0	
70 2				(Swan Trail Slough)
4.1	0.0	4.1	365.0	
67 2				(Lake Stevens WTP)
80.2	0.0	80.2	365.0	
54 2				(Marysville WTP)
550.7	0.0	550.7	365.0	
37 2				(Everett Mech WTP)
267.2	0.0	267.2	365.0	
36 2				(Everett Lagoon WTP)
2196.4	0.0	2196.4	365.0	
59 2				(Quilceda Creek)
5.3	0.0	5.3	365.0	
11	+ *	+		*****F: WASTE LOADS (System 8 - Org P)
1.00	0.4535			
72 2				(French Creek)
5.2	0.0	5.2	365.0	
71 2				(Pilchuck River)
6.0	0.0	6.0	365.0	
48 2				(Snohomsih WTP)
106.4	0.0	106.4	365.0	
42 2				(Marshland)
11.8	0.0	11.8	365.0	
38 2				(Deadwater Slough)
4.0	0.0	4.0	365.0	
70 2				(Swan Trail Slough)
0.9	0.0	0.9	365.0	
67 2				(Lake Stevens WTP)
21.5	0.0	21.5	365.0	
54 2				(Marysville WTP)
121.7	0.0	121.7	365.0	
37 2				(Everett Mech WTP)
270.7	0.0	270.7	365.0	
36 2				(Everett Lagoon WTP)
418.2	0.0	418.2	365.0	
59 2				(Quilceda Creek)
0.9	0.0	0.9	365.0	
11	+ *	+		*****F: WASTE LOADS (System 9 - Salinity)
1.00	0.4535			
72 2				(French Creek)
0.0	0.0	0.0	365.0	
71 2				(Pilchuck River)
0.0	0.0	0.0	365.0	
48 2				(Snohomsih WTP)
0.0	0.0	0.0	365.0	
42 2				(Marshland)
0.0	0.0	0.0	365.0	
38 2				(Deadwater Slough)
0.0	0.0	0.0	365.0	
70 2				(Swan Trail Slough)
0.0	0.0	0.0	365.0	
67 2				(Lake Stevens WTP)
0.0	0.0	0.0	365.0	
54 2				(Marysville WTP)
0.0	0.0	0.0	365.0	
37 2				(Everett Mech WTP)
0.0	0.0	0.0	365.0	
36 2				(Everett Lagoon WTP)
0.0	0.0	0.0	365.0	
59 2				(Quilceda Creek)
0.0	0.0	0.0	365.0	
0	+ *	+ *	+	(System F.2 - NPS loads)

5 + * + * + *****G: PARAMETERS***
 SAL 2 1.0 TMP SG 3 1.0 KESG 5 1.0 SOD1D 9 1.0
 SODAT 12 1.0

Segment
Number

Segment Number	Average Salinity	Temperature	Light Extinction Coefficient	Sediment Oxygen Demand
1	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
2	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
3	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
4	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
5	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
6	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
7	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
8	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
9	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
10	25.8 TMP SG	15.9 KESG	0.85 SOD1D	1.5
11	25.8 TMP SG	15.9 KESG	0.85 SOD1D	1.5
12	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
13	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
14	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
15	25.8 TMP SG	15.9 KESG	0.85 SOD1D	0.0
16	25.8 TMP SG	15.9 KESG	0.85 SOD1D	1.5
17	25.8 TMP SG	15.9 KESG	0.85 SOD1D	1.5
18	25.8 TMP SG	19.1 KESG	0.85 SOD1D	0.0
19	25.8 TMP SG	19.1 KESG	0.85 SOD1D	1.5
20	24.3 TMP SG	19.1 KESG	0.85 SOD1D	1.5
21	24.3 TMP SG	19.1 KESG	0.85 SOD1D	1.5
22	24.3 TMP SG	19.1 KESG	0.85 SOD1D	1.5
23	24.3 TMP SG	19.1 KESG	0.85 SOD1D	1.5

SAL	2	24.3	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	1.5
SODAT	12	1.08									
24											
SAL	2	24.3	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
25											
SAL	2	24.3	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
26											
SAL	2	24.3	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
27											
SAL	2	23.4	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
28											
SAL	2	23.4	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
29											
SAL	2	23.2	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
30											
SAL	2	23.2	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
31											
SAL	2	23.0	TMPSG	3	19.2	KESG	5	0.85	SOD1D	9	1.5
SODAT	12	1.08									
32											
SAL	2	21.1	TMPSG	3	19.2	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
33											
SAL	2	21.1	TMPSG	3	19.2	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
34											
SAL	2	12.0	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
35											
SAL	2	10.2	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
36											
SAL	2	8.2	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
37											
SAL	2	6.5	TMPSG	3	19.1	KESG	5	0.85	SOD1D	9	2.8
SODAT	12	1.08									
38											
SAL	2	0.0	TMPSG	3	19.4	KESG	5	0.85	SOD1D	9	2.4
SODAT	12	1.08									
39											
SAL	2	0.0	TMPSG	3	19.4	KESG	5	0.85	SOD1D	9	2.4
SODAT	12	1.08									
40											
SAL	2	0.0	TMPSG	3	19.4	KESG	5	0.85	SOD1D	9	2.4
SODAT	12	1.08									
41											
SAL	2	0.0	TMPSG	3	19.4	KESG	5	0.85	SOD1D	9	2.4
SODAT	12	1.08									
42											
SAL	2	0.0	TMPSG	3	19.4	KESG	5	0.85	SOD1D	9	2.4
SODAT	12	1.08									
43											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	0.85	SOD1D	9	2.4
SODAT	12	1.08									
44											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	0.85	SOD1D	9	2.4
SODAT	12	1.08									
45											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	2.4
SODAT	12	1.08									
46											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	2.4
SODAT	12	1.08									
47											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	2.4
SODAT	12	1.08									
48											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	2.4
SODAT	12	1.08									

49											
SAL	2	14.0	TMPSG	3	19.3	KESG	5	0.69	SOD1D	9	2.8
SODAT	12	1.08									
50											
SAL	2	12.0	TMPSG	3	19.3	KESG	5	0.69	SOD1D	9	2.8
SODAT	12	1.08									
51											
SAL	2	10.0	TMPSG	3	19.3	KESG	5	0.69	SOD1D	9	3.5
SODAT	12	1.08									
52											
SAL	2	8.0	TMPSG	3	19.3	KESG	5	0.69	SOD1D	9	3.5
SODAT	12	1.08									
53											
SAL	2	6.0	TMPSG	3	19.3	KESG	5	0.69	SOD1D	9	3.5
SODAT	12	1.08									
54											
SAL	2	9.7	TMPSG	3	19.3	KESG	5	0.69	SOD1D	9	2.8
SODAT	12	1.08									
55											
SAL	2	9.0	TMPSG	3	19.3	KESG	5	0.69	SOD1D	9	2.8
SODAT	12	1.08									
56											
SAL	2	2.0	TMPSG	3	19.3	KESG	5	0.69	SOD1D	9	2.8
SODAT	12	1.08									
57											
SAL	2	1.0	TMPSG	3	19.8	KESG	5	0.69	SOD1D	9	2.8
SODAT	12	1.08									
58											
SAL	2	0.5	TMPSG	3	19.8	KESG	5	0.69	SOD1D	9	2.8
SODAT	12	1.08									
59											
SAL	2	16.0	TMPSG	3	19.3	KESG	5	1.20	SOD1D	9	2.8
SODAT	12	1.08									
60											
SAL	2	13.5	TMPSG	3	19.3	KESG	5	1.20	SOD1D	9	2.8
SODAT	12	1.08									
61											
SAL	2	11.8	TMPSG	3	19.3	KESG	5	1.20	SOD1D	9	4.5
SODAT	12	1.08									
62											
SAL	2	6.0	TMPSG	3	19.3	KESG	5	1.20	SOD1D	9	4.5
SODAT	12	1.08									
63											
SAL	2	5.5	TMPSG	3	19.3	KESG	5	1.20	SOD1D	9	2.8
SODAT	12	1.08									
64											
SAL	2	1.0	TMPSG	3	19.3	KESG	5	1.20	SOD1D	9	2.8
SODAT	12	1.08									
65											
SAL	2	1.0	TMPSG	3	19.8	KESG	5	1.10	SOD1D	9	2.8
SODAT	12	1.08									
66											
SAL	2	1.0	TMPSG	3	19.8	KESG	5	1.10	SOD1D	9	2.8
SODAT	12	1.08									
67											
SAL	2	1.0	TMPSG	3	19.8	KESG	5	1.10	SOD1D	9	2.4
SODAT	12	1.08									
68											
SAL	2	0.3	TMPSG	3	19.8	KESG	5	1.10	SOD1D	9	2.4
SODAT	12	1.08									
69											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.10	SOD1D	9	2.4
SODAT	12	1.08									
70											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.10	SOD1D	9	2.4
SODAT	12	1.08									
71											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	1.5
SODAT	12	1.08									
72											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	0.5
SODAT	12	1.08									
73											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	0.5
SODAT	12	1.08									
74											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	0.5

SODAT	12	1.08									
75											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	0.5
SODAT	12	1.08									
76											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	1.51	SOD1D	9	0.5
SODAT	12	1.08									
77											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	10.0	SOD1D	9	0.1
SODAT	12	1.08									
78											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	10.0	SOD1D	9	0.1
SODAT	12	1.08									
79											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	10.0	SOD1D	9	0.1
SODAT	12	1.08									
80											
SAL	2	0.0	TMPSG	3	19.8	KESG	5	10.0	SOD1D	9	0.1
SODAT	12	1.08									
+	*	+	*	+	*	+					
GLOBAL		0									
AMMONIA-N		1									
nitrif		3									
K12C		11	0.20		K12T		12	1.085			
KNIT		13	2.0								
NITRATE-N		1									
denitrif		3									
K20C		21	0.09		K20T		22	1.045			
KNO3		23	0.1								
O-PO4		0									
PHYTO-C		4									
growth		2									
K1C		41	1.40		K1T		42	1.066			
light		5									
LGHTS		43	1.0		PHIMX		44	720.0			
XKC		45	0.0170		CCHL		46	75.0			
IS1		47	100.								
nutrients		4									
KMNG1		48	0.030		KMPG1		49	0.002			
NCRB		58	0.250		PCRB		57	0.025			
death		3									
K1RC		50	0.092		K1RT		51	1.045			
K1D		52	0.10								
CBOD		1									
deoxygen		2									
KDC		71	0.152		KDT		72	1.047			
DO		1									
ratio		1									
OCRB		81	2.67								
ORGANIC-N		1									
mineralize		3									
K71C		91	0.10		K71T		92	1.047			
FON		95	1.0								
ORGANIC-P		1									
mineralize		3									
K83C		100	0.10		K83T		101	1.047			
FOP		104	1.0								
SALINITY		0									
8		*	+	*	+						
ITOT	2	5									
392.											
F	2	6	0.0		392.		365.0				
0.58											
WIND	2	7	0.0		0.58		365.0				
0.0											
TFNH4	2	13	0.0		0.0		365.0				
0.0											
TFPO4	2	14	0.0		0.0		365.0				
0.0											
ZOO	2	19	0.0		0.0		365.0				
0.0											
AIRTM	2	21	0.0		0.0		365.0				
22.0											
XICEV	2	22	0.0		22.0		365.0				
1.0											
0.0					1.0		365.0				

*****H: CONSTANTS***

I: TIME FUNCTIONS*

SYSTEM-1 (AMMONIA)

Segment Ammonia		Segment Ammonia	
Number	(mg/L)	Number	(mg/L)
1	0.942E-01	1.0	2 0.940E-01
4	0.949E-01	1.0	5 0.942E-01
7	0.941E-01	1.0	8 0.949E-01
10	0.942E-01	1.0	11 0.942E-01
13	0.942E-01	1.0	14 0.941E-01
16	0.949E-01	1.0	17 0.949E-01
19	0.949E-01	1.0	20 0.940E-01
22	0.942E-01	1.0	23 0.942E-01
25	0.944E-01	1.0	26 0.946E-01
28	0.101E+00	1.0	29 0.111E+00
31	0.629E-01	1.0	32 0.760E-01
34	0.199E+00	1.0	35 0.266E+00
37	0.209E+00	1.0	38 0.182E+00
40	0.761E-01	1.0	41 0.594E-01
43	0.521E-01	1.0	44 0.389E-01
46	0.446E-01	1.0	47 0.415E-01
49	0.951E-01	1.0	50 0.114E+00
52	0.969E-01	1.0	53 0.118E+00
55	0.115E+00	1.0	56 0.146E+00
58	0.153E+00	1.0	59 0.505E-01
61	0.439E-01	1.0	62 0.473E-01
64	0.118E+00	1.0	65 0.150E+00
67	0.145E+00	1.0	68 0.841E-01
70	0.676E-01	1.0	71 0.450E-01
73	0.452E-01	1.0	74 0.450E-01
76	0.450E-01	1.0	77 0.459E-01
79	0.459E-01	1.0	80 0.459E-01

3 0.0 50.0 ***J: INITIAL CONDS***
 [same for other "initial conditions"]

Number	(mg/L)	Number	(mg/L)
1.0	3 0.947E-01	1.0	6 0.944E-01
1.0	6 0.944E-01	1.0	9 0.947E-01
1.0	9 0.947E-01	1.0	12 0.940E-01
1.0	12 0.940E-01	1.0	15 0.944E-01
1.0	15 0.944E-01	1.0	18 0.944E-01
1.0	18 0.944E-01	1.0	21 0.942E-01
1.0	21 0.942E-01	1.0	24 0.945E-01
1.0	24 0.945E-01	1.0	27 0.944E-01
1.0	27 0.944E-01	1.0	30 0.155E+00
1.0	30 0.155E+00	1.0	33 0.540E-01
1.0	33 0.540E-01	1.0	36 0.248E+00
1.0	36 0.248E+00	1.0	39 0.139E+00
1.0	39 0.139E+00	1.0	42 0.597E-01
1.0	42 0.597E-01	1.0	45 0.448E-01
1.0	45 0.448E-01	1.0	48 0.366E-01
1.0	48 0.366E-01	1.0	51 0.867E-01
1.0	51 0.867E-01	1.0	54 0.138E+00
1.0	54 0.138E+00	1.0	57 0.148E+00
1.0	57 0.148E+00	1.0	60 0.487E-01
1.0	60 0.487E-01	1.0	63 0.670E-01
1.0	63 0.670E-01	1.0	66 0.146E+00
1.0	66 0.146E+00	1.0	69 0.758E-01
1.0	69 0.758E-01	1.0	72 0.734E-01
1.0	72 0.734E-01	1.0	75 0.469E-01
1.0	75 0.469E-01	1.0	78 0.419E-01
1.0	78 0.419E-01	1.0	

SYSTEM-2 (NITRATE)

1	0.256E+00	1.0	2 0.255E+00
4	0.256E+00	1.0	5 0.255E+00
7	0.256E+00	1.0	8 0.256E+00
10	0.257E+00	1.0	11 0.256E+00
13	0.256E+00	1.0	14 0.256E+00
16	0.257E+00	1.0	17 0.256E+00
19	0.257E+00	1.0	20 0.257E+00
22	0.257E+00	1.0	23 0.257E+00
25	0.257E+00	1.0	26 0.256E+00
28	0.257E+00	1.0	29 0.256E+00
31	0.252E+00	1.0	32 0.249E+00
34	0.259E+00	1.0	35 0.253E+00
37	0.230E+00	1.0	38 0.203E+00
40	0.165E+00	1.0	41 0.162E+00
43	0.160E+00	1.0	44 0.157E+00
46	0.157E+00	1.0	47 0.159E+00
49	0.246E+00	1.0	50 0.242E+00
52	0.221E+00	1.0	53 0.216E+00
55	0.224E+00	1.0	56 0.216E+00
58	0.209E+00	1.0	59 0.240E+00
61	0.225E+00	1.0	62 0.221E+00
64	0.218E+00	1.0	65 0.206E+00
67	0.195E+00	1.0	68 0.177E+00
70	0.165E+00	1.0	71 0.158E+00
73	0.159E+00	1.0	74 0.159E+00
76	0.160E+00	1.0	77 0.160E+00
79	0.160E+00	1.0	80 0.160E+00

3 0.0 50.0 ***J: INITIAL CONDS***

1.0	3 0.255E+00	1.0	6 0.254E+00
1.0	6 0.254E+00	1.0	9 0.256E+00
1.0	9 0.256E+00	1.0	12 0.254E+00
1.0	12 0.254E+00	1.0	15 0.256E+00
1.0	15 0.256E+00	1.0	18 0.252E+00
1.0	18 0.252E+00	1.0	21 0.257E+00
1.0	21 0.257E+00	1.0	24 0.257E+00
1.0	24 0.257E+00	1.0	27 0.257E+00
1.0	27 0.257E+00	1.0	30 0.257E+00
1.0	30 0.257E+00	1.0	33 0.246E+00
1.0	33 0.246E+00	1.0	36 0.243E+00
1.0	36 0.243E+00	1.0	39 0.183E+00
1.0	39 0.183E+00	1.0	42 0.163E+00
1.0	42 0.163E+00	1.0	45 0.158E+00
1.0	45 0.158E+00	1.0	48 0.158E+00
1.0	48 0.158E+00	1.0	51 0.232E+00
1.0	51 0.232E+00	1.0	54 0.240E+00
1.0	54 0.240E+00	1.0	57 0.207E+00
1.0	57 0.207E+00	1.0	60 0.237E+00
1.0	60 0.237E+00	1.0	63 0.213E+00
1.0	63 0.213E+00	1.0	66 0.202E+00
1.0	66 0.202E+00	1.0	69 0.168E+00
1.0	69 0.168E+00	1.0	72 0.155E+00
1.0	72 0.155E+00	1.0	75 0.159E+00
1.0	75 0.159E+00	1.0	78 0.160E+00
1.0	78 0.160E+00	1.0	

SYSTEM-3 (INORGANIC PHOSPHORUS)

1	0.593E-01	1.0	2 0.592E-01
4	0.592E-01	1.0	5 0.592E-01
7	0.596E-01	1.0	8 0.595E-01
10	0.596E-01	1.0	11 0.599E-01
13	0.599E-01	1.0	14 0.599E-01
16	0.599E-01	1.0	17 0.593E-01
19	0.599E-01	1.0	20 0.599E-01
22	0.511E-01	1.0	23 0.511E-01
25	0.526E-01	1.0	26 0.524E-01
28	0.554E-01	1.0	29 0.566E-01
31	0.505E-01	1.0	32 0.514E-01
34	0.682E-01	1.0	35 0.733E-01
37	0.546E-01	1.0	38 0.401E-01
40	0.126E-01	1.0	41 0.957E-02
43	0.796E-02	1.0	44 0.531E-02
46	0.644E-02	1.0	47 0.640E-02
49	0.525E-01	1.0	50 0.526E-01

3 0.0 50.0 ***J: INITIAL CONDS***

1.0	3 0.590E-01	1.0	6 0.599E-01
1.0	6 0.599E-01	1.0	9 0.594E-01
1.0	9 0.594E-01	1.0	12 0.591E-01
1.0	12 0.591E-01	1.0	15 0.597E-01
1.0	15 0.597E-01	1.0	18 0.592E-01
1.0	18 0.592E-01	1.0	21 0.590E-01
1.0	21 0.590E-01	1.0	24 0.596E-01
1.0	24 0.596E-01	1.0	27 0.538E-01
1.0	27 0.538E-01	1.0	30 0.625E-01
1.0	30 0.625E-01	1.0	33 0.490E-01
1.0	33 0.490E-01	1.0	36 0.669E-01
1.0	36 0.669E-01	1.0	39 0.269E-01
1.0	39 0.269E-01	1.0	42 0.987E-02
1.0	42 0.987E-02	1.0	45 0.659E-02
1.0	45 0.659E-02	1.0	48 0.612E-02
1.0	48 0.612E-02	1.0	51 0.448E-01
1.0	51 0.448E-01	1.0	

52	0.400E-01	1.0	53	0.394E-01	1.0	54	0.538E-01	1.0
55	0.425E-01	1.0	56	0.398E-01	1.0	57	0.354E-01	1.0
58	0.392E-01	1.0	59	0.477E-01	1.0	60	0.471E-01	1.0
61	0.447E-01	1.0	62	0.435E-01	1.0	63	0.404E-01	1.0
64	0.408E-01	1.0	65	0.343E-01	1.0	66	0.323E-01	1.0
67	0.295E-01	1.0	68	0.168E-01	1.0	69	0.134E-01	1.0
70	0.116E-01	1.0	71	0.404E-02	1.0	72	0.502E-02	1.0
73	0.490E-02	1.0	74	0.497E-02	1.0	75	0.489E-02	1.0
76	0.506E-02	1.0	77	0.500E-02	1.0	78	0.500E-02	1.0
79	0.500E-02	1.0	80	0.500E-02	1.0			
SYSTEM 4 (CHLOROPHYL a)				3	0.0	50.0	***J: INITIAL CONDS***	
1	0.90	1.0	2	0.90	1.0	3	0.90	1.0
4	0.90	1.0	5	0.90	1.0	6	0.90	1.0
7	0.96	1.0	8	0.80	1.0	9	0.83	1.0
10	0.87	1.0	11	0.85	1.0	12	0.85	1.0
13	0.81	1.0	14	0.88	1.0	15	0.83	1.0
16	0.84	1.0	17	0.88	1.0	18	0.82	1.0
19	0.82	1.0	20	0.81	1.0	21	0.82	1.0
22	0.83	1.0	23	0.83	1.0	24	0.84	1.0
25	0.81	1.0	26	0.81	1.0	27	0.84	1.0
28	0.86	1.0	29	0.85	1.0	30	0.89	1.0
31	0.85	1.0	32	0.81	1.0	33	0.81	1.0
34	0.88	1.0	35	0.85	1.0	36	0.85	1.0
37	0.86	1.0	38	0.80	1.0	39	0.83	1.0
40	0.89	1.0	41	0.87	1.0	42	0.83	1.0
43	0.90	1.0	44	1.00	1.0	45	1.00	1.0
46	1.00	1.0	47	1.00	1.0	48	1.00	1.0
49	0.87	1.0	50	0.84	1.0	51	0.87	1.0
52	0.85	1.0	53	0.81	1.0	54	0.87	1.0
55	0.98	1.0	56	0.81	1.0	57	0.82	1.0
58	0.80	1.0	59	0.84	1.0	60	0.84	1.0
61	0.87	1.0	62	0.86	1.0	63	0.89	1.0
64	0.89	1.0	65	0.85	1.0	66	0.89	1.0
67	0.88	1.0	68	0.81	1.0	69	0.83	1.0
70	1.47	1.0	71	1.20	1.0	72	1.50	1.0
73	1.40	1.0	74	1.40	1.0	75	1.50	1.0
76	1.50	1.0	77	1.50	1.0	78	1.60	1.0
79	1.60	1.0	80	1.60	1.0			
SYSTEM-5 (CBOD)				3	0.0	50.0	***J: INITIAL CONDS***	
1	0.242E+01	1.0	2	0.243E+01	1.0	3	0.246E+01	1.0
4	0.248E+01	1.0	5	0.244E+01	1.0	6	0.249E+01	1.0
7	0.245E+01	1.0	8	0.249E+01	1.0	9	0.240E+01	1.0
10	0.246E+01	1.0	11	0.241E+01	1.0	12	0.240E+01	1.0
13	0.248E+01	1.0	14	0.248E+01	1.0	15	0.237E+01	1.0
16	0.245E+01	1.0	17	0.241E+01	1.0	18	0.220E+01	1.0
19	0.245E+01	1.0	20	0.247E+01	1.0	21	0.246E+01	1.0
22	0.246E+01	1.0	23	0.225E+01	1.0	24	0.232E+01	1.0
25	0.242E+01	1.0	26	0.222E+01	1.0	27	0.242E+01	1.0
28	0.243E+01	1.0	29	0.221E+01	1.0	30	0.243E+01	1.0
31	0.244E+01	1.0	32	0.239E+01	1.0	33	0.240E+01	1.0
34	0.125E+01	1.0	35	0.140E+01	1.0	36	0.235E+01	1.0
37	0.121E+01	1.0	38	0.113E+01	1.0	39	0.201E+01	1.0
40	0.139E+01	1.0	41	0.133E+01	1.0	42	0.149E+01	1.0
43	0.146E+01	1.0	44	0.126E+01	1.0	45	0.185E+01	1.0
46	0.113E+01	1.0	47	0.127E+01	1.0	48	0.118E+01	1.0
49	0.236E+01	1.0	50	0.221E+01	1.0	51	0.238E+01	1.0
52	0.242E+01	1.0	53	0.224E+01	1.0	54	0.207E+01	1.0
55	0.247E+01	1.0	56	0.224E+01	1.0	57	0.203E+01	1.0
58	0.246E+01	1.0	59	0.248E+01	1.0	60	0.238E+01	1.0
61	0.243E+01	1.0	62	0.222E+01	1.0	63	0.267E+01	1.0
64	0.242E+01	1.0	65	0.223E+01	1.0	66	0.221E+01	1.0
67	0.200E+01	1.0	68	0.246E+01	1.0	69	0.222E+01	1.0
70	0.236E+01	1.0	71	0.182E+01	1.0	72	0.222E+01	1.0
73	0.169E+01	1.0	74	0.179E+01	1.0	75	0.226E+01	1.0
76	0.120E+01	1.0	77	0.128E+01	1.0	78	0.228E+01	1.0
79	0.118E+01	1.0	80	0.128E+01	1.0			
SYSTEM-6 (DO)				3	0.0	50.0	***J: INITIAL CONDS***	
1	0.600E+01	1.0	2	0.600E+01	1.0	3	0.600E+01	1.0
4	0.600E+01	1.0	5	0.600E+01	1.0	6	0.600E+01	1.0
7	0.595E+01	1.0	8	0.591E+01	1.0	9	0.592E+01	1.0
10	0.596E+01	1.0	11	0.597E+01	1.0	12	0.593E+01	1.0
13	0.591E+01	1.0	14	0.591E+01	1.0	15	0.593E+01	1.0
16	0.594E+01	1.0	17	0.593E+01	1.0	18	0.598E+01	1.0
19	0.598E+01	1.0	20	0.598E+01	1.0	21	0.597E+01	1.0
22	0.586E+01	1.0	23	0.596E+01	1.0	24	0.598E+01	1.0
25	0.581E+01	1.0	26	0.571E+01	1.0	27	0.581E+01	1.0
28	0.583E+01	1.0	29	0.571E+01	1.0	30	0.581E+01	1.0

31	0.582E+01	1.0	32	0.591E+01	1.0	33	0.582E+01	1.0
34	0.586E+01	1.0	35	0.595E+01	1.0	36	0.623E+01	1.0
37	0.584E+01	1.0	38	0.594E+01	1.0	39	0.683E+01	1.0
40	0.685E+01	1.0	41	0.742E+01	1.0	42	0.742E+01	1.0
43	0.714E+01	1.0	44	0.754E+01	1.0	45	0.771E+01	1.0
46	0.781E+01	1.0	47	0.837E+01	1.0	48	0.857E+01	1.0
49	0.689E+01	1.0	50	0.608E+01	1.0	51	0.615E+01	1.0
52	0.682E+01	1.0	53	0.629E+01	1.0	54	0.613E+01	1.0
55	0.686E+01	1.0	56	0.654E+01	1.0	57	0.674E+01	1.0
58	0.685E+01	1.0	59	0.640E+01	1.0	60	0.625E+01	1.0
61	0.587E+01	1.0	62	0.584E+01	1.0	63	0.593E+01	1.0
64	0.684E+01	1.0	65	0.682E+01	1.0	66	0.688E+01	1.0
67	0.682E+01	1.0	68	0.738E+01	1.0	69	0.750E+01	1.0
70	0.685E+01	1.0	71	0.826E+01	1.0	72	0.863E+01	1.0
73	0.913E+01	1.0	74	0.910E+01	1.0	75	0.910E+01	1.0
76	0.910E+01	1.0	77	0.910E+01	1.0	78	0.910E+01	1.0
79	0.910E+01	1.0	80	0.910E+01	1.0			
SYSTEM-7 (ORGANIC NITROGEN)			3	0.0		50.0	***J: INITIAL CONDS***	
1	0.126E+00	1.0	2	0.126E+00	1.0	3	0.125E+00	1.0
4	0.126E+00	1.0	5	0.126E+00	1.0	6	0.125E+00	1.0
7	0.126E+00	1.0	8	0.125E+00	1.0	9	0.125E+00	1.0
10	0.126E+00	1.0	11	0.128E+00	1.0	12	0.126E+00	1.0
13	0.126E+00	1.0	14	0.127E+00	1.0	15	0.126E+00	1.0
16	0.127E+00	1.0	17	0.131E+00	1.0	18	0.128E+00	1.0
19	0.126E+00	1.0	20	0.126E+00	1.0	21	0.125E+00	1.0
22	0.126E+00	1.0	23	0.126E+00	1.0	24	0.131E+00	1.0
25	0.138E+00	1.0	26	0.138E+00	1.0	27	0.146E+00	1.0
28	0.156E+00	1.0	29	0.163E+00	1.0	30	0.193E+00	1.0
31	0.130E+00	1.0	32	0.138E+00	1.0	33	0.130E+00	1.0
34	0.220E+00	1.0	35	0.244E+00	1.0	36	0.241E+00	1.0
37	0.181E+00	1.0	38	0.201E+00	1.0	39	0.126E+00	1.0
40	0.106E+00	1.0	41	0.844E-01	1.0	42	0.910E-01	1.0
43	0.834E-01	1.0	44	0.758E-01	1.0	45	0.796E-01	1.0
46	0.793E-01	1.0	47	0.786E-01	1.0	48	0.766E-01	1.0
49	0.144E+00	1.0	50	0.150E+00	1.0	51	0.146E+00	1.0
52	0.153E+00	1.0	53	0.164E+00	1.0	54	0.156E+00	1.0
55	0.150E+00	1.0	56	0.149E+00	1.0	57	0.143E+00	1.0
58	0.160E+00	1.0	59	0.133E+00	1.0	60	0.134E+00	1.0
61	0.139E+00	1.0	62	0.141E+00	1.0	63	0.147E+00	1.0
64	0.149E+00	1.0	65	0.141E+00	1.0	66	0.137E+00	1.0
67	0.133E+00	1.0	68	0.110E+00	1.0	69	0.999E-01	1.0
70	0.943E-01	1.0	71	0.602E-01	1.0	72	0.671E-01	1.0
73	0.664E-01	1.0	74	0.666E-01	1.0	75	0.660E-01	1.0
76	0.673E-01	1.0	77	0.620E-01	1.0	78	0.620E-01	1.0
79	0.620E-01	1.0	80	0.620E-01	1.0			
SYSTEM-8 (ORGANIC PHOSPHORUS)			3	0.0		50.0	***J: INITIAL CONDS***	
1	0.398E-01	1.0	2	0.398E-01	1.0	3	0.398E-01	1.0
4	0.399E-01	1.0	5	0.398E-01	1.0	6	0.397E-01	1.0
7	0.391E-01	1.0	8	0.393E-01	1.0	9	0.396E-01	1.0
10	0.395E-01	1.0	11	0.394E-01	1.0	12	0.393E-01	1.0
13	0.390E-01	1.0	14	0.391E-01	1.0	15	0.392E-01	1.0
16	0.394E-01	1.0	17	0.392E-01	1.0	18	0.390E-01	1.0
19	0.390E-01	1.0	20	0.390E-01	1.0	21	0.387E-01	1.0
22	0.387E-01	1.0	23	0.387E-01	1.0	24	0.389E-01	1.0
25	0.387E-01	1.0	26	0.386E-01	1.0	27	0.386E-01	1.0
28	0.386E-01	1.0	29	0.386E-01	1.0	30	0.389E-01	1.0
31	0.389E-01	1.0	32	0.285E-01	1.0	33	0.386E-01	1.0
34	0.291E-01	1.0	35	0.286E-01	1.0	36	0.274E-01	1.0
37	0.153E-01	1.0	38	0.116E-01	1.0	39	0.269E-02	1.0
40	0.124E-01	1.0	41	0.292E-01	1.0	42	0.285E-02	1.0
43	0.154E-01	1.0	44	0.212E-01	1.0	45	0.133E-02	1.0
46	0.128E-01	1.0	47	0.222E-01	1.0	48	0.104E-02	1.0
49	0.180E-01	1.0	50	0.174E-01	1.0	51	0.165E-01	1.0
52	0.149E-01	1.0	53	0.138E-01	1.0	54	0.170E-01	1.0
55	0.154E-01	1.0	56	0.139E-01	1.0	57	0.129E-01	1.0
58	0.129E-01	1.0	59	0.183E-01	1.0	60	0.183E-01	1.0
61	0.180E-01	1.0	62	0.177E-01	1.0	63	0.166E-01	1.0
64	0.150E-01	1.0	65	0.126E-01	1.0	66	0.121E-01	1.0
67	0.114E-01	1.0	68	0.198E-01	1.0	69	0.180E-01	1.0
70	0.221E-01	1.0	71	0.156E-01	1.0	72	0.199E-01	1.0
73	0.299E-01	1.0	74	0.198E-01	1.0	75	0.192E-01	1.0
76	0.203E-01	1.0	77	0.200E-01	1.0	78	0.200E-01	1.0
79	0.200E-01	1.0	80	0.200E-01	1.0			
SYSTEM-9 (SALINITY)			3	0.0		50.0	***J: INITIAL CONDS***	
1	0.298E+02	1.0	2	0.298E+02	1.0	3	0.298E+02	1.0
4	0.298E+02	1.0	5	0.298E+02	1.0	6	0.298E+02	1.0
7	0.296E+02	1.0	8	0.297E+02	1.0	9	0.292E+02	1.0

10	0.291E+02	1.0	11	0.295E+02	1.0	12	0.293E+02	1.0
13	0.294E+02	1.0	14	0.293E+02	1.0	15	0.297E+02	1.0
16	0.298E+02	1.0	17	0.291E+02	1.0	18	0.291E+02	1.0
19	0.295E+02	1.0	20	0.265E+02	1.0	21	0.283E+02	1.0
22	0.294E+02	1.0	23	0.264E+02	1.0	24	0.286E+02	1.0
25	0.292E+02	1.0	26	0.271E+02	1.0	27	0.273E+02	1.0
28	0.294E+02	1.0	29	0.277E+02	1.0	30	0.263E+02	1.0
31	0.247E+02	1.0	32	0.215E+02	1.0	33	0.238E+02	1.0
34	0.161E+02	1.0	35	0.705E+01	1.0	36	0.669E+01	1.0
37	0.316E+01	1.0	38	0.100E-24	1.0	39	0.366E+00	1.0
40	0.100E-24	1.0	41	0.673E-01	1.0	42	0.188E-01	1.0
43	0.100E-24	1.0	44	0.409E-05	1.0	45	0.640E-05	1.0
46	0.426E-05	1.0	47	0.162E-05	1.0	48	0.169E-05	1.0
49	0.168E+02	1.0	50	0.128E+02	1.0	51	0.114E+02	1.0
52	0.637E+01	1.0	53	0.484E+01	1.0	54	0.992E+01	1.0
55	0.555E+01	1.0	56	0.272E+01	1.0	57	0.197E+01	1.0
58	0.481E+01	1.0	59	0.215E+02	1.0	60	0.206E+02	1.0
61	0.167E+02	1.0	62	0.147E+02	1.0	63	0.860E+01	1.0
64	0.481E+01	1.0	65	0.150E+01	1.0	66	0.132E+01	1.0
67	0.920E+00	1.0	68	0.100E-24	1.0	69	0.123E+00	1.0
70	0.132E-01	1.0	71	0.100E-24	1.0	72	0.100E-24	1.0
73	0.100E-24	1.0	74	0.100E-24	1.0	75	0.100E-24	1.0
76	0.100E-24	1.0	77	0.100E-24	1.0	78	0.100E-24	1.0
79	0.100E-24	1.0	80	0.100E-24	1.0			